

Emissions Control Catalysis

Edited by Ioannis V. Yentekakis and Philippe Vernoux Printed Edition of the Special Issue Published in *Catalysts*



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Special Issue Editors

Ioannis V. Yentekakis Philippe Vernoux

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About the Special Issue Editors

Ioannis V. Yentekakis is currently a Full Professor of Physical Chemistry and Catalysis, School of Environmental Engineering, Technical University of Crete (TUC), Greece. He received a Chemical Engineering Diploma (1983) and a Ph.D. (1988) from the University of Patras (UP). His academic career is associated with Princeton University (Post-Doc), ICE-HT/FORTH (Senior Researcher), the Department of Chemical Engineering UP (Assistant Professor), and Cambridge University, U.K. (Visiting Professor). His current interest concerns development of novel nano-structured catalysts and eco-friendly processes with applications in added-value chemicals production, environmental protection, renewable energy generation, fuel cells, hydrogen energy, CO2 hydrogenation, natural gas (NG), and biogas valorization. He has authorized more than 110 journal publications (>4500 citations, h-index = 38), 3 international patents, 10 books, and has Guest Edited 5 Special Issues. He has served as department chairman, member of the University Council and the Senate, and chairman or organizer of many scientific conferences. He has coordinated more than 30 funded research projects.

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Editorial Emissions Control Catalysis

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1. Overview

Important advances have been achieved over the past years in agriculture, industrial technology, energy, and health, which have contributed to human well-being. However, some of these improvements in our lives were accompanied by a great threat to the environment and public health, with photochemical smog, stratospheric ozone depletion, acid rain, global warming, and finally climate change being the most well-known major issues, as a result of a variety of pollutants emitted through these human activities. The indications are that we are already at a tipping point that might lead to non-linear, sudden environmental change on a global scale.

Aiming to ensure that "we live well, within the planet's ecological limits" the United Nation's Sustainable Development Goals [1] and the EU's Environmental Action Plan [2] include calls to action and priority objectives to protect life on land and water and tackling climate change. In a full harmony, scientists around the world are developing tools and techniques to understand, monitor, protect, and improve the environment, and to ensure that our achievements must not adversely affect our environment and at the same time we must mitigate any damage that has already occurred. In turn, new tools, technology, and advanced materials are continuously added to our quiver that enable us to effectively control emissions, either of mobile (e.g., cars) or stationary (industrial) sources, and to improve the quality of outdoor and indoor air, with catalysis to play a major role on these efforts, i.e., as a leading technology for improving quality of life, health, and environment.

Emissions Control Catalysis in the area of Environmental Catalysis [3,4] is continuously growing up, providing novel multifunctional, nano-structured materials with active metal species at sizes of the level of nanoparticles, nanoclusters or even single atoms [5–8]. These materials are also often promoted by several ways (i.e., by surface- or support-induced promotion [9–13] and by electrochemical promotion [14–16], among others) in order to be very active and selective for the abatement of a variety of pollutants and greenhouse gases. Regarding the latter, representative cases concern the abatement of CO, NO_x, N₂O, NH₃, CH₄, higher hydrocarbons, volatile organic compounds (VOCs), particulate matter (PM), and specific pollutants emitted by industry (e.g., SO_x, H₂S, dioxins, and aromatic hydrocarbons) or landfill and wastewater treatment plants (biogas).

The concept of Cyclic Economy is currently of growing attention in emission control catalysis strategies for the production of useful chemicals and fuels from the controlled pollutants. The CO or CO_2 hydrogenation for the production of CH_4 or liquid fuels, biogas or methane reforming for the production of syngas, and/or H_2 , oxidative coupling of methane for the production of ethylene are a few examples of such efforts.

2. Special Issue Contributions and Highlights

The present "Emissions Control Catalysis" Special Issue was successful in collecting 21 high-quality contributions by several research groups around the world, which covers recent research progress in the field of the catalytic control of air pollutants emitted by mobile or stationary sources, not limited only to the abatement of pollutants but also including research dedicated to cyclic economy strategies. These papers among others are ranging from the synthesis and physicochemical-textural-structural characterization of the materials, activity-selectivity-durability evaluation under environmentally important reactions, fundamental understanding of structure-activity relationships or other metal-metal and metal-support interactions on the multifunctional materials involved. In particular, three comprehensive review articles covering several major topics and directions in emissions control catalysis subject were published in this special issue. The review of the guest editors (Yentekakis I., and Vernoux P.) and co-workers [17] concerns mainly the of CO, CHs, and NO_x emissions abatement from stoichiometric, lean burn, and diesel engines exhausts, addressing the literature that concerns the electropositive promotion by alkalis or alkaline earths of platinum group metals (PGMs) that have found to be extremely effective for the related three-way and lean-burn reactions and catalytic chemistry. The review of Smirniotis P. and co-workers [18] concerns the selective catalytic reduction of NO_x with NH₃ (NH₃-SCR of NO_x) focusing to low temperature applications that is a highly desirable perspective, and finally the review of Bogaerts A. and co-workers [19] covering a hot recent trend in emissions control implicating cyclic economy strategies, that is the conversion and utilization of CO₂ for the production of value-added chemicals. On the other hand, a major part of contributions (9/21) concerns original research on nitrogen oxides reduction processes [20–28], reflecting the fact that this topic still remains hot among the targets of environmental catalysis. Five out of 21 studies concern CO and hydrocarbons oxidation processes [29–33] while the remained 4/21 concern CO₂ capture/recycling processes under the view of cyclic economy [34–37].

Review Contributions

Alkali metals as surface-promoters of platinum group metals (PGMs) have shown remarkable promotional effects on the vast majority of reactions related to the emissions control catalysis either they are applied by traditional routes (e.g., impregnation) or electrochemically, via the so called, concept of the electrochemical promotion of catalysis (EPOC) or non-faradaic electrochemical modification of catalytic activity (NEMCA effect) [14,15]. In the comprehensive review of Yentekakis et al. [17] published in this Special Issue, more than 120 papers on the subject were collected and discussed comparatively after classification in groups, on the basis of the specific reaction promoted, e.g., (CO, HCs, or H₂)-SCR of NO_x, and CO or hydrocarbons oxidation. The authors also present and analyze, by means of indirect (kinetics) and direct (spectroscopic) evidences, a mechanistic model for the mode of action of electropositive promoters, which consistently interprets all the observed promoting phenomena. Concluding, they claim that this very pronounced (in some cases extraordinarily) alkali-induced promotion in emissions control catalysis prompts to the development of novel catalyst formulations for a more efficient and cost-effective control of the emissions of automotives and stationary combustion processes.

Since the emission standards for NO_x are becoming more stringent to keep our atmosphere clean and severe pollution regulations being imposed around the world, the challenge of a cost-effective, i.e., low-temperature, selective catalytic reduction (LT-SCR) of NO_x by NH_3 with a combination of high NO_x activity and N_2 selectivity in a wide operation temperature window and good resistance to SO_2/H_2O have attracted paramount attention. Supported and mixed transition metal oxides have been widely investigated for LT-SCR technology. However, these catalytic materials have some drawbacks, especially in terms of catalyst poisoning by H_2O or/and SO_2 . Hence, the development of catalysts for the LT-SCR process is still under active investigation throughout seeking better performance. Extensive research efforts have been made to develop new advanced materials for this technology. The comprehensive review of Smirniotis and co-workers [18] collected and comparatively analyses more than 140 publications in the topic, covering the description of the influence of operating conditions, materials, and promoters on the LT-SCR performance, as well as active sites, reaction intermediates, and mechanistic implications evidenced by using isotopic labeling and in situ FT-IR studies.

Global climate change as a result of the greenhouse effects of the increasing emissions of the so-called greenhouse gases, with CO_2 to be the most representative one, is of major concern currently. CO_2 capture, conversion, and utilization is one of the possible solutions to reduce CO_2 concentration in the atmosphere. Among other methods, this can be accomplished by direct catalytic hydrogenation of CO_2 , producing value-added products. In their comprehensive review, Bogaerts A. and co-workers [19], focused mainly in the last 5-years progress on the topic summarized the literature research (~125 papers) and the current priorities on CO_2 hydrogenation to CO, CH_4 , CH_3OH , DME, olefins, and higher hydrocarbons value-added chemicals by heterogeneous catalysis and plasma catalysis. Although, its energy efficiency is still deterring for commercialization, plasma catalysis has currently attracted sufficient attention, due to its simple operating conditions (ambient temperature and atmospheric pressure) and unique advantages in activating inert molecules; its potential advantages and current limitations on CO_2 hydrogenation process have been accounted in the review.

3. Original Contributions

3.1. NO_x Abatement Related Results

A major part of the contributions (9/21) involves research on the selective catalytic reduction (SCR) of NO and/or its direct catalytic decomposition. In particular:

Shen D. and co-workers [20] investigated the deNO_x activity on a series of bimetallic Cu–Mn molecular sieve catalysts (Cu–Mn/SAPO-34) with different loadings of Cu and Mn components during the selective catalytic reduction (SCR) of NO with NH₃ at low temperatures (ca. 120–330 °C), including the effects of H₂O vapor and/or SO₂. Among the catalysts tested, the performance of 2 wt% Cu-6 wt% Mn/SAPO-34 one found to be superior, achieving 72% NO conversion at 120 °C and even better (90%) at 180–330 °C. The reversible negative effect of H₂O on NO conversion was attributed to the competitive adsorption of H₂O and NH₃ on Lewis acid sites; this poisoning was diminished upon increasing the reaction temperature to 300 °C. A permanent poisoning effect of SO₂ on deNO_x activity found is strongly dependent on the reaction temperature, becomes more pronounced at lower ones and is further enhanced by H₂O co-feed; this is assigned to the formation of (NH₄)₂SO₄, which results in the plug of active sites and a decrease of surface area.

Gao Y. and co-workers [21] synthesized Mn-Co/TiO₂ and Mn-Fe/TiO₂ nanocatalysts by a hydrothermal method which were characterized by a variety of methods including Brunner–Emmet–Teller (BET)/Barrett–Joyner–Halenda (BJH) analysis of N₂ adsorption/desorption isotherms at –196 °C, transmission electron microscope (TEM), X-ray diffraction (XRD), H₂-temperature-programmed reduction (TPR), NH₃-temperature-programmed desorption (TPD), and X-ray photoelectron spectroscopy (XPS) that enabled the authors for a comprehensive comparison of the catalysts nanostructure characteristics and their de-NOx catalytic performance, gained insight into the structure-activity relationships. The Mn–Co/TiO₂ catalyst offered superior structure characteristics than Mn–Fe/TiO₂: higher surface area and active components distribution, diminished crystallinity, reduced nanoparticle size and also higher Mn⁴⁺/Mnⁿ⁺ ratio, confirming its better oxidation ability and larger amount of Lewis and Brønsted acid sites on the Mn–Co/TiO₂ surface. As a result, Mn–Co/TiO₂ nanocatalyst displayed superior SCR of NO with NH₃ on both activity and selectivity in the temperature range of 75–250 °C. Kinetics data revealed that both Eley–Rideal (E–R) and Langmuir–Hinshelwood (L–H) mechanisms were implicated in NH₃-SCR process over Mn–Fe/TiO₂ and Mn–Co/TiO₂ catalysts.

Han J. and co-workers [22] investigated the enhanced deNO_x performance and stability of sulfated sintered ore catalysts (SSOC) during the selective catalytic reduction of NO_x with NH₃. The maximum deNOx efficiency found was about 92% at 300 °C, NH₃/NO = 1 and 5000 h⁻¹ gas hourly space velocity (GHSV). A systematic characterization of the materials by means of X-ray

fluorescence spectrometry (XRF), Brunauer–Emmett–Teller (BET) analyzer, X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), and diffuse reflectance infrared spectroscopy (DRIFTS) was conducted to provide an in depth understanding of the NH₃-SCR reaction mechanism and to explain the denitration performance and stability of SSOC. The existence of more Brønsted acid sites at the surface of SSOC found to be responsible for the improved adsorption capacity of NH₃ and NO over the SSOC surface that accomplished the formation of amide species (–NH₂), NH₄⁺ species, NO₂ molecules in a gaseous or weakly adsorbed state, and nitrates. The reaction between –NH₂, NH₄⁺, and NO (E–R mechanism) and the reaction of the coordinated ammonia with the adsorbed NO₂ (L–H mechanism) were attributed to NO_x reduction.

Song C. and co-workers [23] studied the promotional effect of Ce and/or Zr incorporation by ion exchange on Cu/ZMS-5 catalysts for the NH₃-SCR of NO. The cerium and zirconium addition promotes the activity of catalysts; the cerium-rich catalysts exhibiting superior SCR activities compared to the zirconium-rich ones. The improved low temperature activity of the CuCe_xZr_{1-x}O_y/ZSM-5 catalysts in comparison to the unpromoted Cu/ZSM-5 (the former achieving >95% NO conversion at 175–468 °C, the later at 209–405 °C) was attributed to an increase of the reactive lattice oxygen content and reducibility of the catalysts via the Ce³⁺/Ce⁴⁺ redox couple and its interaction with cooper species. Moreover, the presence of zirconium in the catalysts promotes surface copper enrichment, prevents copper crystallization and causes suppression of N₂O formation, increasing N₂-selectivity of the system.

He H. and co-workers [24] synthesized $CeZr_{0.5}Ti_aO_x$ (with a = 0, 1, 2, 5, and 10) catalysts by a stepwise precipitation approach, which were studied on the NH₃-SCR of NO_x. In all contents, Ti addition was beneficial to the deNO_x catalytic performance. Superior behavior was obtained by the $CeZr_{0.5}Ti_2O_x$ particular catalyst composition. Controlling pH and precipitation time during the two steps involved in the preparation method enabled the authors to achieve a catalyst with enhanced acidity (favorable for NH₃ adsorption in NH₃-SCR processes) and high dispersion of CeO₂ onto the surface of ZrO_2 -TiO₂ synthesized first. The as-prepared $CeZr_{0.5}Ti_2O_x$ catalyst was characterized by superior redox properties, enhanced adsorption and activation of NO_x and NH₃ and enhanced surface adsorbed oxygen such as $O_2^{2^-}$ and O^- belonging to defect-oxide or a hydroxyl-like group; all these factors positively affecting its SCR of NO_x with NH₃ performance.

Olson L. and co-workers [25] investigated the poisoning effect of a phosphorous containing atmosphere on the NO_x storage capacity of a Pt/Ba/Al₂O₃ structured (i.e., washcoated on a ceramic monolith) catalyst. A significant loss of the NO_x storage capacity was caused by phosphorous exposure characterized by a progressively decreasing axial distribution of phosphorous concentration from the inlet to the outlet of the monolith. The values of the specific surface area and pore volume of phosphorous-poisoned monolithic catalysts followed an inverse order: lower at the inlet, higher at the outlet of the monolith. Additional features of the axial phosphorous accumulation detected were: a higher surface accumulation at the inlet of the monolith mostly appeared in the form of P_4O_{10} , the presence of more metaphosphate (PO_3^-) in the middle section of the monolith, and a less surface accumulation of phosphorous at the outlet of the monolith due to its extended diffusion into the washcoat. In respect to the poisoning effect of phosphorous on the SCR of NO_x it was revealed that the formations of N₂ and N₂O were decreased in favor of NH₃ production; the reaction is more influenced by the phosphorous poisoning than the ammonia formation from the stored nitrates.

The direct NO decomposition activity on PdO or PtO supported on Co_3O_4 spinel was studied by Reddy et al. [26] in an attempt to discover means of enhancing the activity of Co_3O_4 spinel, one of the most active single-element oxide catalysts for NO decomposition at high temperatures (typically > 650 °C). In fact, the authors demonstrated a four-fold higher promotion on the NO decomposition activity of PdO- rather than of PtO-modified Co_3O_4 at 650 °C, accompanied by superior selectivity towards N₂ as well. Structural and surface analysis measurements using a variety of methods (e.g., XRD, XPS, H₂-TPR, and in situ FT-IR) showed an enhanced reducibility of PdO/ Co_3O_4 with an increased thermal stability of surface adsorbed NO_x species, both considered to contribute on the promotion observed. In opposite, PtO enters into the Co_3O_4 structure, without notable influences on the redox and NO adsorption properties of Co_3O_4 , resulting in marginal promotion compared to PdO. The PdO promotion followed volcano behavior with an optimal PdO loading of 3 wt%.

Zhang et al. [27] investigated the NO_x storage capacity of a series of Pd/BEA catalysts with various Pd loadings for cold-start applications. In situ FTIR measurements using CO and NH₃ enable the authors to identify two isolated Pd²⁺ species, Z⁻-Pd²⁺-Z⁻ and Z⁻-Pd(OH)⁺, on exchanged sites of zeolites, as the main active sites for NO trapping. Among these active sites a superior NO_x storage capacity of Z⁻-Pd²⁺-Z⁻ was demonstrated, which is caused by the different resistance to H₂O. Atom utilization of Pd can be improved by using lower Pd loading, with an optimum at 0.5 wt%, since this leads to a sharp decline of Z⁻-Pd(OH)⁺ attributed to the 'exchange preference' for Z⁻-Pd²⁺-Z⁻ in BEA.

Finally, Ingel et al. [28] in a different approach for controlling N₂O emissions in ammonia oxidation process at high temperature, proposed the design of experiments and response surface methodology to study this process. The reactor's load, the temperature of reaction and the number of catalytic gauzes were selected as independent variables, whereas ammonia oxidation efficiency and N2O concentration in nitrous gases were assumed as dependent variables (response). Statistically significant mathematical models were developed from the achieved results, which describe the effect of independent variables on the analyzed responses. The ammonia oxidation efficiency value depends on the reactor's load and the number of catalytic gauzes but not on the temperature in the studied range (870–910 $^{\circ}$ C). The concentration of N₂O in nitrous gases depends on all three parameters. The developed models were used for the multi-criteria optimization with the application of desirability function. Sets of parameters were achieved for which optimization assumptions were met: maximization of ammonia oxidation efficiency and minimization of the N₂O amount being formed in the reaction. As authors claim, the presented methodology can be used to minimize the primary N_2O emission at high ammonia oxidation efficiency. It can be applied for optimization of operating parameters of ammonia oxidation reactor with two types of catalysts: catalytic gauzes and catalyst for high temperature of N₂O decomposition. As a result, it is possible to obtain the set of independent variables ensuring low N₂O emission and to meet the binding environmental regulations.

3.2. CO, CH₄, and Other Hydrocarbons Oxidation Reactions

Avgouropoulos and co-workers [29] synthesized a series of atomically dispersed copper-ceria nanocatalysts via appropriate tuning of a novel hydrothermal method and investigated their activity on CO oxidation, which was found to be strongly dependent on the nanostructured morphology, oxygen vacancy concentration, and nature of atomically dispersed Cu²⁺ clusters. A number of techniques including electron paramagnetic resonance (EPR) spectroscopy, X-ray diffraction (XRD), N₂ adsorption, scanning electron microscopy (SEM), Raman spectroscopy, and ultraviolet-visible diffuse reflectance spectroscopy (UV-Vis DRS) were employed in the characterization of the synthesized materials. The aim was to find the key factors that govern the physicochemical properties of the synthesized materials during preparation and then to provide convincing structure-activity correlations. Elevated temperatures and low concentrations of NaOH (≤ 0.1 M) during preparation have led to more active ceria-based catalysts in CO oxidation. This was explained with the obtained morphology, the nature of oxygen vacancies and dispersed copper species, and to a lesser extent, with the specific surface area of the materials and the concentration of defects.

Briois and co-workers [30] in an attempt to provide means of replacing expensive platinum group metals with cost-effective perovskite type materials for catalytic oxidation reactions, prepared thin catalytic coatings of Sr and Ag doped lanthanum perovskites, $La_{1-x-y}Sr_xAg_yCoO_{3-\alpha}$ (x = 0.13-0.28, and y = 0.14-0.48), on alumina substrates by using the cathodic co-sputtering magnetron method in reactive condition. Such thin porous catalytic film arrangements can optimize the surface/bulk ratio by combining a large gas exposure surface area with an extremely low loading, thus saving raw materials. The sputtering method was optimized to generate crystallized and thin perovskite

films. The authors found that high Ag contents has a strong impact on the morphology of the coatings, favoring the growth of covering films with a porous wire-like morphology that showed a good catalytic activity for CO oxidation. Interestingly, the optimal composition ($La_{0.40}Sr_{0.1}Ag_{0.48}Co_{0.93}O_3$) displayed similar catalytic performance than this of a Pt film, and was also efficient for CO and NO abatement in a simulated Diesel exhaust gas mixture, demonstrating the promising catalytic properties of such nanostructured thin sputtered perovskite films.

Methane, a substantially more potent greenhouse gas than CO_2 , is the main compound of natural gas, which is lately used with a continuous increased rate in various industrial processes and as an alternative fuel for heavy-duty transportation not excluding light-duty tracks. As a consequence, control of CH_4 emissions via catalytic deep oxidation has attracted considerable renewed attention, given that methane appears the lowest reactivity among alkanes. Under this view, Iliopoulou and co-workers [31] synthesized a series of novel Co–Ce mixed oxide catalysts in an effort to enhance synergistic effects that could improve their redox and oxygen storage properties and, thus, their activity in methane deep oxidation. The effect of the synthesis method (hydrothermal or precipitation) and Co loading (0, 2, 5, and 15 wt%) on the catalytic efficiency and stability was investigated. Hydrothermally synthesized Co_3O_4/CeO_2 catalysts appeared superior performance due to their improved physicochemical properties (smaller crystallite size, larger surface area, and enhanced reducibility). In respect to Co loading, the optimum performance was observed over a 15 wt% Co/CeO₂ catalyst, which also presented sufficient tolerance to water presence.

Based on the fact that Pd is one of the most active catalysts for complete methane oxidation Baranova and coworkers [32] used the concept of electrochemical promotion of catalysis (EPOC) to further promote the reaction over palladium nano-structured catalysts deposited on yttria-stabilized zirconia (YSZ) solid electrolyte. Anodic polarization (O^{2-} supply to the catalyst) resulted to a rate enhancement up to ~3 at 450 °C with an apparent Faradaic efficiency as high as 3000 (for a current application as low as 1 µA). Electrochemical promotion on this catalytic system showed persistent behavior the catalyst remained under promotion for a long period of time after interruption of the external bias induced EPOC. Increasing polarization time resulted in a longer-lasting persistent promotion (p-EPOC); more time was required for the reaction rate to reach its initial un-promoted value. The phenomenon was attributed to the continuing promotion by the stored oxygen in palladium oxide formed during the anodic polarization.

Taking into account that sulfur poisoning is one of the most important factors deteriorating the efficiency of diesel exhaust after-treatment systems and that bare TiO₂ appears high sulfur resistivity, Zhang et al. [33] prepared TiO₂–CeO₂ composites by co-precipitation and studied their sulfur resistance and catalytic activity in the oxidation of diesel soluble organic fraction (SOF). They found that TiO₂-modification of CeO₂ significantly improves the catalytic SOF purification efficiency of CeO₂ besides the fact that this ceria doping does not downgrade the excellent sulfur resistance of bare TiO₂; the prepared TiO₂–CeO₂ characterization by X-ray diffraction (XRD) and Raman spectroscopy indicate that cerium ions can enter into the TiO₂ lattice, without forming complete CeO₂ crystals. Moreover, as confirmed by XPS and H₂-TPR, the synthesized TiO₂–CeO₂ composites appeared enhanced oxygen storage capacities (OSC) that considered responsible for their better SOF oxidation activity.

3.3. CO₂ Capture/Recycling: Combining Emissions Control with Added-Value Chemical Production (Cyclic Economy)

De Lucas-Consuegra and co-workers [34] developed a low-temperature (below 90 °C) proton exchange membrane (Sterion) electrochemical cell for the electrocatalytic conversion of gaseous CO_2 to liquid fuels. This novel system achieved gas-phase electrocatalytic reduction of CO_2 over a Cu-based cathode by using water electrolysis-derived protons generated in-situ on an IrO₂ anode. Three Cu-activated carbon cathodes with varying Cu loading (10, 20, and 50 wt% Cu–AC), and thus particle size, were tested. Products distribution was a function of the Cu loading and particle size

of the Cu–AC cathode; methyl formate, acetaldehyde, and methanol were being the main reaction products, respectively, over 50, 20, and 10 wt% Cu–AC. The membrane electrode assembly (MEA) containing the cathode with the largest Cu loading and particle size (50 wt% Cu–AC, 40 nm) showed the highest CO₂ electrocatalytic activity per mole of Cu (and the lowest energy consumption values for the conversion of CO₂, 119 kW·h·mol⁻¹), which was attributed to the lower Cu–CO bonding strength over large Cu particles.

Wang H., Lu J.-X. and co-workers [35] fabricated an electrocatalytic cell consisting of a 0.5 M KHCO₃ aqueous solution as electrolyte saturated with CO₂ by bubbling, a CuO/TiO₂-Nafion as working electrode, and Pt as counter and reference electrodes in order to study the electroreduction of CO₂ to added-value multi-carbon oxygenate products (ethanol, acetone, and n-propanol). The non-noble metal electrocatalyst CuO/TiO₂ was in situ reduced to Cu/TiO₂, which efficiently catalyzed CO₂ reduction, offering a maximum overall faradaic efficiency of 47.4% at a potential of -0.85 V vs. reversible hydrogen electrode (RHE). The catalytic activity for CO₂ electroreduction was strongly dependent on the CuO contents of the catalysts as-prepared, resulting in different electroactive surface areas. The significantly improved CO₂ reduction activity of CuO/TiO₂ was attributed to the high CO₂ adsorption ability of TiO₂ component of the working electrode.

Since hydrogen is currently considered as an efficient and environmentally benign energy carrier, among others, considerable attention is played by scientists worldwide for its sufficient production from hydrocarbon feedstocks (natural gas (NG), liquefied petroleum gas (LPG), etc.), biogas and bio-alcohols. To this end, Tang D. and co-workers [36] applied chemical looping reforming (CLR) as a prospective alternative for hydrogen production via ethanol steam reforming, which is characterized by energy efficiency and inherent CO₂ capture. Taking into account that oxygen carriers (OCs) with sufficient oxygen mobility and sintering resistance still remain the main challenges for the development of high-performance materials in the CLR process, the authors explore the performance of Ni/CeO₂ nanorod (NR) synthesized by a hydrothermal method as an OC in the CLR of ethanol. Using a variety of characterization techniques, they showed that the as-prepared Ni/CeO₂-NR possesses the desired properties for CLR, i.e., high Ni dispersion, abundant oxygen vacancies, and strong metal-support interaction, all factors improving catalytic activity. Testing the material in CLR process successfully offered a H₂ selectivity of 80% in 10-cycle stability test. The authors concluded that the small particle size and abundant oxygen vacancies contributed to improve water gas shift reaction, the high oxygen mobility of CeO₂–NR effectively eliminated surface coke on the Ni particle, and the covered interfacial Ni atoms closely anchored on the underlying surface oxygen vacancies on the (111) facets of CeO₂–NR enhance the anti-sintering capability.

Ioannidou et al. [37] contributed with a detailed and comparative catalytic-kinetic study of the performance of modified X-Ni/GDC electrodes (where X = Au, Mo, and Fe), in the form of half-electrolyte supported cells, in the reverse water gas shift reaction (RWGS). The importance of the RWGS reaction (H₂ + CO₂ \rightarrow H₂O + CO) is well known, since it takes place in most of the hydrocarbon processing reactions (e.g., hydrocarbons reforming processes) as well as in co-electrolysis of H₂O and CO₂ (CO₂ utilization) in solid oxide electrolysis cells (SOECs) yielding synthesis gas (CO + H₂), as considered in the present study. Solid oxide electrolysis is a contemporary process for CO₂ capture/recycling, which is proven as an attractive method to provide CO₂ neutral synthetic hydrocarbon fuels. The X-Ni/GDC catalysts were tested at open circuit conditions in order to elucidate their catalytic activity towards the production of CO; one of the products of the H₂O/CO₂ co-electrolysis reaction. The CO production rate increases by increasing the operating temperature and the partial pressure of H₂ in the reaction mixture. Fe and Mo modification enhances CO production, and 2 wt% Fe-Ni/GDC and 3 wt% Mo-Ni/GDC electrodes were superior compared to the other samples, in the whole studied temperature range (800–900 °C) reaching thermodynamic equilibrium. No carbon formation was detected.

Conflicts of Interest: The authors declare no conflict of interest.

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Review



Electropositive Promotion by Alkalis or Alkaline Earths of Pt-Group Metals in Emissions Control Catalysis: A Status Report

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Abstract: Recent studies have shown that the catalytic performance (activity and/or selectivity) of Pt-group metal (PGM) catalysts for the CO and hydrocarbons oxidation as well as for the (CO, HCs or H₂)-SCR of NO_x or N₂O can be remarkably affected through surface-induced promotion by successful application of electropositive promoters, such as alkalis or alkaline earths. Two promotion methodologies were implemented for these studies: the Electrochemical Promotion of Catalysis (EPOC) and the Conventional Catalysts Promotion (CCP). Both methodologies were in general found to achieve similar results. Turnover rate enhancements by up to two orders of magnitude were typically achievable for the reduction of NO_x by hydrocarbons or CO, in the presence or absence of oxygen. Subsequent improvements (ca. 30-60 additional percentage units) in selectivity towards N2 were also observed. Electropositively promoted PGMs were also found to be significantly more active for CO and hydrocarbons oxidations, either when these reactions occur simultaneously with deNOx reactions or not. The aforementioned direct (via surface) promotion was also found to act synergistically with support-mediated promotion (structural promotion); the latter is typically implemented in TWCs through the complex (Ce–La–Zr)-modified γ -Al₂O₃ washcoats used. These attractive findings prompt to the development of novel catalyst formulations for a more efficient and cost-effective control of the emissions of automotives and stationary combustion processes. In this report the literature findings in the relevant area are summarized, classified and discussed. The mechanism and the mode of action of the electropositive promoters are consistently interpreted with all the observed promoting phenomena, by means of indirect (kinetics) and direct (spectroscopic) evidences.

Keywords: platinum; palladium; Rhodium; iridium; NO; N₂O; propene; CO; methane; alkali; alkaline earth; platinum group metals; deNO_x chemistry; lean burn conditions; TWC; catalyst promotion; EPOC

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References

1. Introduction

The development of new catalyst formulations for a more efficient and cost-effective control of unburned hydrocarbons, CO and NO_x pollutants emitted by mobile and stationary combustion processes is currently an urgent need [1–3]. Commercial three-way catalytic converters (TWCs) have been highly successful in controlling NO_x, CO and hydrocarbon emissions from conventional stoichiometric gasoline engines. However, some problems, that emerged during their four decades of implementation, concerning the economy of their production process, their composition, their operational/life time behaviour and their recycling still need to be resolved. The most significant issues of TWC technology are the following:

- Commercial TWCs use formulations based variously on two or three noble metals (Pt, Pd and Rh) where Rh is essential for an efficient control of NO_x emissions (Rh is highly effective and therefore the key component in TWCs for the NO dissociation/reduction; Pt and Pd, although very active for CO and hydrocarbons oxidations, are almost ineffective for NO_x reduction). However, Rh is rare, therefore its successful reduction or even replacement by another De-NO_x efficient catalyst in TWC formulations is highly desirable.
- 2. A small but significant portion of NO is still converted to the undesirable by-product N₂O in commercial TWCs; Nitrous oxide emission control, due to N₂O harmful impact on stratospheric ozone depletion and its outstanding global warming potential (the latter is about 310 and 21 times greater than that of CO₂ and CH₄, respectively [4,5]), is a current challenge in environmental catalysis technology.
- 3. TWCs recycling for the noble metals recovery is nowadays marginally profitable, mainly due to the costly final separation of the recovered noble metals; simpler TWCs formulations (i.e., consisting of only one noble metal) are expected to substantially reduce both TWCs' production and recycling costs.
- 4. Commercial TWCs are no longer efficient in controlling NO_x emissions from the advanced lean-burn and diesel engines that operate at net-oxidising conditions.

In general, catalytic systems with suitable efficiency under oxygen rich conditions are a great challenge in emissions control technology. Besides lean-burn and diesel engines' effluent gases, such conditions are also typical in several stationary processes of significant environmental footprint, such as in soil fuels combustion processes for heat production in industry, in solid wastes combustion processes and certain chemical processes such as ammonia oxidation processes (e.g., nitric acid plants), adipic acid and other specific industries.

Prompted by these issues, numerous efforts are currently reported in open literature, by several research groups, on discovering means for enhancing the catalytic performance of Pt and Pd or even Rh, for reactions related to CO, NO_x and hydrocarbons abatement under stoichiometric and/or oxygen rich conditions. A notable part of these studies is focussed on the reduction or replacement of the Rh usage due to its scarcity and consequently higher cost compared to Pt and Pd.

It has been recently shown that the catalytic performance of precious metals can be strongly promoted by alkalis and alkaline earths for reactions related to emissions control catalysis. Two promotion methods were implemented towards this aim: (i) the Electrochemical Promotion of Catalysis (EPOC) concept or NEMCA effect (Non-Faradaic Electrochemical Modification of Catalytic Activity), proposed three decades ago by Vayenas and co-workers (e.g., [6–13]) and (ii) the Conventional Catalysts' Promotion (CCP), applicable by means of highly dispersed (supported) catalysts, as typically used in industrial applications (e.g., [14,15]). On these bases, rational/more efficient catalyst formulations and catalytic systems have been designed with significantly widened operational and/or implementation windows.

The present review summarises and comparatively analyses these literature results, paying attention on both fundamental and practical issues emerged. Mechanistic considerations (involving the mode of action of electropositive promoters in emissions control catalytic chemistry) are described in confrontation with recent spectroscopic and surface analysis studies evidencing their rightness.

2. Promotion Methodologies and Catalysts Formulations/Designs

Materials formulations and designs, experimental setup and other specific characteristics and/or rules regarding the two promotion methods, EPOC and CCP, are briefly described below.

2.1. Electrochemical Catalysts Formulations. The Electrochemical Promotion of Catalysis (EPOC) Concept

It has been shown by Vayenas and co-workers (e.g., [6,7]) that the chemisorptive and catalytic properties of metal catalysts, in the form of thin polycrystalline porous metal films, interfaced with solid electrolytes (X^z-ionic conductors: X is the type of ion transferred, z is its positive or negative charge) are subjected to in situ controlled electrochemical promotion via external bias applications, that is, currents or voltages imposed between the catalyst metal film and a separate counter electrode deposited on the same solid electrolyte specimen, which constitute a solid state electrochemical galvanic cell schematically shown in Figure 1.

In this way, applying small electric currents (typically some tens of μ A units) or potentials (typically -2 to +2 V) between the working (catalyst film) and the counter electrodes via a Galvanostat/Potentiostat (Figure 1), ions are supplied from (or to) the solid electrolyte to (or from) the catalyst-electrode surface. Ions flux direction is determined by the direction of the applied current in combination with their charge, while their flux rate by the Faraday's law, $r_z = I/zF$ (I is the applied current in A; z is the ion's charge; F is the Faraday's constant = 96,485 Cb/mol). As we shall see in detail in Section 3.3, there is compelling evidence that these ions (together with their compensating charge in the metal thus forming surface dipoles) spill over (migrate) onto the gas-exposed catalyst surface, establishing an effective electrochemical double layer on the catalytically active surface altering its electron availability (work function). This strongly affects the binding strength of chemisorbed reactants and intermediates resulting to a significant modification on catalytic reaction rate (and or selectivity), typically greatly larger than the rate of ions flux that created these modifications

(i.e., non-Faradaic behaviour), the so-called NEMCA (Non-Faradaic Electrochemical Modification of Catalytic Activity) phenomenon or EPOC (Electrochemical romotion of Catalysis).



Figure 1. Schematic representation of a solid state electrochemical cell, through which the EPOC concept can be applied. X: type of ion transferred from the solid electrolyte to the catalyst surface and z: ion charge (e.g., O^{2-} , Na^+ , K^+ , H^+).

The EPOC was then applied to a large number of catalytic reactions [6,7,13], including reactions related to emissions control catalysis (e.g., [16–22]). Numerous EPOC studies regarding electrochemically imposed electropositive promotion by alkalis of Pt-group metals' catalysed emissions control reactions have also been reported, which are subject of the present review.

Alkali-conducting solid electrolytes such as Na⁺ or K⁺ substituted β - and β'' -aluminas, (Na or K)_{1+X}Al₁₁O_{17+X/2} and (Na or K)_{1+X}M_XAl_{11-X}O₁₇, respectively [6], are used for the construction of electrochemical galvanic cells with the following configuration (Figure 1):

PGM-film electrode / alkali-conducting solid electrolyte / inert counter electrode (e.g., Au) (1)

These cells are exposed to the reacting gas mixture by means of a continuous flow well-mixed reactor, as that shown in Figure 2, described as "single-pellet" (reactor) configuration and behaved similarly to a classical Continuous-Stirred-Tank-Reactor (CSTR) [12]. Controlled amounts of promoting species (alkali cations) can therefore be in situ supplied through the solid electrolyte onto the catalyst surface, causing significant alterations on the catalytic properties of the latter.

For the sake of simplicity, inert (in respect to the studied reaction) counter and reference electrodes (e.g., Au) are typically applied in EPOC studies conducted in reactor configurations similar to this shown in Figure 2; avoiding their participation in the catalytic reaction vectors (activity, selectivity) simplifies the data analysis, interpretation and understanding. It is also worth noting that besides their role as an in situ and reversible promoting species sources, solid electrolytes involved in the configuration of EPOC galvanic cells do not actually play any additional role on the promotional phenomena obtained.



Figure 2. The single pellet reactor configuration for electrochemical promotion studies [12].

- 2.1.1. Operation Modes of the Electrochemical Promotion of Catalysis (EPOC)
- (i) Galvanostatic transient operation:

One can use the galvanostatic transient mode of EPOC operation, described in detail in ref. [16], in order to be able to calculate the coverage, θ , of the promoting species (electrochemically supplied) on the catalyst surface as a function of time. Briefly, transient galvanostatic operation concerns the application of a constant current, *I*, between the catalyst and counter electrodes via the galvanostat. In cells of the type described by the formula (1), imposition of a constant negative current corresponds to a constant flux of alkali cations to the catalyst surface at a rate equal *I*/*F* (in moles of alkali/s).

The obtained alkali coverage, θ_{Alk} , can therefore be calculated from Faraday's law [6,16],

$$\theta_{\text{Alk}} (\%) = 100 \cdot (-It/FN_0) \tag{2}$$

where, *t* is the time of current (*I*) application, *F* is the Faraday's constant and N_o is the number of catalyst active sites independently measured via surface titration. Recording the metal-solid electrolyte polarization transient during the galvanostatic experiment that is, the time variation of catalyst potential by means of a reference electrode, V_{WR} (Figure 1), this permits establishment of the relationship $\theta_{Alk}(V_{WR})$ between alkali coverage and catalyst potential V_{WR} (Figure 3).

This relationship is in general dependent on gaseous composition but this dependence is relatively weak because the catalyst work function $e\Phi$ varies according to [6,7,11]:

$$e\Delta V_{WR} = \Delta(e\Phi) \tag{3}$$

and is determined primarily by the coverage of alkali, due to the large dipole moment of alkalis on the Pt-group metal surfaces.

Galvanostatic operation concerns a continuous supply and accumulation of alkali on the catalyst surface (Equation (2)). Current interruption at a desired alkali coverage can then provide an optimally alkali-promoted catalyst surface. Alkali species are not typically consumed during the catalytic reaction, which allows working under stable reaction conditions. Hence, the promotional effect induced by the alkali coverage remains practically constant, even after the interruption of the applied current.



Figure 3. Dependence of alkali coverage on catalyst potential V_{WR} under certain reaction conditions (Reproduced with permission from Ref. [19]. Copyright 1997, Elsevier).

It is worth noting that galvanostatic EPOC operation provides a rapid method for assessing the response of the reaction rate to promoter coverage (Figure 4); it consequently offers a very quick and easy way to find, at any set of conditions, the optimal promoter loading. This information could then be used for the direct design of optimally promoted conventional catalysts formulations for practical applications (Figure 4) [21].

(ii) Potentiostatic operation (a steady-state mode of operation):

The potentiostatic mode of EPOC operation corresponds to catalyst operation under the imposition of a constant catalyst polarization (V_{WR}) via the potentiostat. This is a steady-state mode of operation: since the desired catalyst potential V_{WR} is imposed via the potentiostat, a very rapid transfer of the appropriate amount of alkali cations to the catalyst surface takes place, which supplies the required amount of alkali in order to develop and stabilize the externally imposed polarization value V_{WR} ; then the current is sharply vanished, asymptotically approaching the zero value.

Since the $\theta_{Alk}(V_{WR})$ relationship between alkali coverage and catalyst potential can be obtained through a galvanostatic transient conducted at the desired reaction conditions (as described in (*i*); e.g., Figure 3), the potentiostatic mode of operation is more suitable and commonly used in EPOC studies. It allows to work under steady-state conditions at several constant catalyst potentials, that is, at several constant coverages of the promoter. The steady-state reaction rate response on promoted catalyst surfaces, upon varying of the other reaction parameters (e.g., temperature, reactants' composition, etc.) can then be studied (e.g., [6,7]).



Figure 4. Comparison of electrochemically promoted (EPOC) and conventionally promoted (CCP) Pt with respect to N₂-selectivity under the NO + C_3H_6 reaction at T = 375 °C and [NO] = 1.3%, $[C_3H_6] = 0.6\%$ feed. Similar well-correlated dependences of the variation of N₂-selectivity on nominal sodium coverage (θ_{Na} , upper abscissa) were obtained by the two methods of promotion (EPOC and CCP) applied. EPOC was performed in a Pt/(Na) β "Al₂O₃/Au galvanic cell (Reprinted with permission from Ref. [22]. Copyright 2000, Elsevier).

2.1.2. Certain Characteristics of EPOC Concept:

(i) Reversibility of the electrochemically imposed promotional phenomena:

By reversing the direction of the applied current in the galvanic cell represented in Figure 1 or even by setting the potential value that corresponds to an alkali-free catalyst surface (Figure 3), the catalyst restores its initial un-promoted intrinsic properties. In other words, the electrochemically imposed promotional phenomena are totally reversible; EPOC provides an in situ, controlled and reversible way of catalyst promotion [6,7].

(ii) EPOC can be used as a fast probe for the optimal promoter loading determination:

Inter alia, EPOC can be successfully used as an effective research tool for evaluating the influence of a candidate promoter to a catalytic system and at the same time for determining the optimal promoter loading (coverage) at any reaction conditions of our interest. After the understanding that electrochemical and conventional promotions in catalysis, EPOC and CCP, are subjected to the same physicochemical rules (this has been proved both experimentally [21,22] and theoretically [23,24]), the findings and knowledge obtained by the application of the electrochemical promotion can be successfully applied for the design of effective conventional catalyst formulations for practical applications [21,22,25]. This strategy enabled us to design novel catalyst formulations extremely active and selective in emission control catalysis, for example, optimally alkali-promoted noble metal catalysts [21].

2.2. Conventional Catalyst Formulations. The Conventional Catalysts Promotion (CCP) Method

The wet impregnation method was used in most cases for the production of conventional type (i.e., highly dispersed on large surface area carriers), alkali-promoted noble metal catalysts. Typical carriers used were γ -Al₂O₃, SnO₂, yttria stabilized zirconia (YSZ) and rare earth oxides (REOs: La₂O₃, CeO₂) modified γ -Al₂O₃. The preparation of the conventional promoted catalysts was performed either by two subsequent impregnation steps (where the noble metal was first deposited, followed by that of the promoter) or by simultaneous, one-step impregnation, in a solution containing both the noble metal and the promoter precursors, typically nitrate but also chloride salts). When using only nitrate precursors, pre-treatment procedures typically involve a high temperature (ca. 500-600 °C) treatment step in air for the decomposition of the nitrates (active phase and promoter) and a subsequent treatment of the resulted catalysts under reaction conditions (i.e., under the studied reaction) for several hours or days in order to ensure stable operation. On the other hand, when a chloride precursor is used, pre-treatment in H₂ flow at temperatures typically 400–450 $^{\circ}$ C for several hours (>2 h) is applied as a first step for the precursor decomposition and the removal of the residual chlorine (e.g., [14,15]); the inhibitory behaviour of the latter on the reactions related to emissions control catalysis is well known. Then, for performance stabilization purposes, operation at certain reaction conditions is applied, as well [14,15].

2.2.1. Estimation of the Promoter Coverage

The coverage of the promoter species on the active catalyst surface is a significant factor not only for ranking but also for better understanding its promoting effects. Its knowledge is also useful for comparison purposes between electrochemically and conventionally promoted catalysts. For the former case (EPOC) we have already shown (Equation (2)) that the coverage of the alkali promoter can be directly estimated via the Faraday law applied to the data of a galvanostatic transient. In the latter case (CCP), a "nominal" percentage surface coverage of the alkali promoter (θ_{Alk}) could be calculated from the alkali wt% content of the catalyst (e.g., [14,21]) by the assumption that all the promoter is present at the surface and distributed uniformly over the entire available area (noble metal + support), without any incorporation into the bulk and that an one-to-one correlation between alkali adatoms and active metal sites exists. Hence, the following equation can be used:

$$\theta_{Alk}$$
 (%) = (wt% of Alkali on the catalyst)·($N_{AV}/M_{Alk}\cdot A \cdot d$) (4)

where N_{AV} is the Avogadro number, M_{Alk} is the molecular weight of alkali used, A is the BET catalyst surface area (m²/g) and d is the surface density (atoms/m²) of the active metal (PGM) (e.g., $d_{Pt} = 1.53 \times 10^{19}, d_{Pd} = 1.27 \times 10^{19}, d_{Rh} = 1.33 \times 10^{19}$ for {111} crystallites). Considering d in Equation (4) as the surface density of the alkali adatom used (i.e., $8.84 \times 10^{19}, 3.314 \times 10^{19}, 1.8 \times 10^{19}, 1.115 \times 10^{19}$ or 1.45×10^{19} for Li, Na, K, Cs or Rb, respectively; values calculated on the basis of the alkali ionic radius), one can probably be led to a better estimation of the nominal alkali coverage [22,25]. In any case and since the issues about the alkali distribution (preference between active metal or support surface, formation of 2D and/or 3D aggregates) are not clear yet, it is apparent that Equation (4) leads to a quite rough estimation of alkali coverage, which however, in the absence of more rigorous estimations or experimental measurements, can be of some interest and applicable as a first approximation upon comparing the EPOC and CCP methods of promotion on a catalytic system (e.g., [22,25]).

The fixed bed, single pass, quartz (or stainless steel) tube reactor configuration is typically used for testing the conventional type catalysts (in the form of powder, small particle clusters or even structured honeycomb monolithic configurations) at the desired reactions. On-line gas chromatography, on-line mass spectroscopy and continuous NO_x chemiluminescence's analysis were used for the analysis of the influent reactor streams in most studies.

2.2.2. EPOC and CCP Comparison Issues

In order to have a solid basis for comparison between electrochemically promoted or conventional promoted catalysts, the rate enhancement ratio (ρ), is introduced in both cases, defined as:

$$\rho = r_x / r_o \quad (x = EPOC \text{ or } CCP) \tag{5}$$

r_o is the unpromoted rate and r_x is the promoted by alkali rate, where the alkali species are either supplied electrochemically (x = EPOC) or conventionally (x = CCP). ρ values as high as 420 (42,000% rate increase) have been reported for some important emissions control catalytic systems (e.g., NO + C₃H₆/Pt) by means of alkali promotion. It must be emphasized here that in literature studies (either EPOC or CCP) involving ρ values estimation, particular attention was devoted in order the catalytic activity data to be acquired in the intrinsic kinetic regime, not influenced by mass transfer limitations (i.e., under low conversions, no limiting reactant, differential reactor operation, small catalyst particles, etc.)

In the case of EPOC an additional parameter, the Faradaic efficiency (Λ) is also introduced [6], which can describe the magnitude of promotion; however, this has no meaning in CCP studies:

$$\Lambda = (\mathbf{r}_{\text{EPOC}} - \mathbf{r}_{\text{o}}) / (I/zF) = \Delta \mathbf{r} / (I/zF)$$
(6)

where *F* is the Faradaic constant and *I* is the current. The term I/zF (*z* is the promoting ion charge; *z* = +1 in the case of alkali promotion) corresponds to the rate of promoting ions supplied electrochemically to the catalyst surface according to the Faraday's law.

Another useful parameter to quantify of the magnitude of the EPOC is the promotion index *Pi* [16,26]:

$$Pi = (r_{\rm EPOC} - r_{\rm o}) / \Delta \theta_{\rm i} \tag{7}$$

where $\Delta \theta_i$ is the coverage of the promoting species such as Na⁺.

3. Results and Discussion

3.1. Promotion of Simple, "Model" Reactions

Literature studies concerning "model" (i.e., simple) reactions, catalysed by Pt-group metals under electropositive promotion by alkalis or alkaline earths, conducted either by EPOC or CCP, are listed in Table 1 and discussed below.

3.1.1. CO Oxidation

Several publications concern the promotion of the PGMs-catalysed CO oxidation by alkalis. Significant promotional effects have been achieved either on thin metal film electrocatalysts interfaced with an alkali-contacting solid electrolyte (EPOC), playing the role of in situ source of alkali cations or on conventional type highly dispersed catalysts dosed with several amounts of alkali promoter (CCP). However, the promoting phenomena were found to be strongly depended on the CO/O_2 gas phase composition and on the alkali loading of the catalyst.

In particular, Yentekakis et al. [16] using the continuous flow, well-mixed, single-pellet reactor described in Figure 2 have shown that the rate of CO oxidation can be markedly enhanced via EPOC by up to 600% (at T = 350 °C) under CO-rich conditions (i.e., at conditions where CO oxidation kinetics obey negative order in CO, positive order in O₂) over a Pt film electrocatalyst interfaced with a Na⁺-conducting β ″Al₂O₃ solid electrolyte (Table 1). Those rate enhancements were achieved at sodium coverages on Pt surface of about 2–4%, while higher sodium coverages were found to depress the catalyst activity, thus leading to a "volcano" type behaviour upon increasing the coverage of the promoting species (Figure 5, curve a). However, at O₂-rich conditions (i.e., when the rate is of positive order in CO) only poisoning phenomena occurred upon increasing Na coverage on the

catalyst surface (Figure 5 curve b). According to the rules of catalyst promotion, Na promoters modify the chemisorptive properties of CO and O_2 reactants on Pt active sites. At CO-rich conditions and for Na coverages up to ~4%, the promotional effect was attributed to a Na-induced enhancement in oxygen chemisorption as a result of a strengthening of the Pt–O bond. The rate poisoning behaviour at higher Na coverages was attributed to active sites blocking phenomena due to the formation of a CO–Na–Pt surface complex.



Figure 5. Effect of Na coverage θ_{Na} on the rate of CO oxidation at oxygen lean (a: T = 350 °C, $[O_2] = 6\%$, [CO] = 5.3%; filled symbols) and at oxygen rich ((b): T = 350 °C, $[O_2] = 6\%$, [CO] = 2.8%; open symbols) conditions. The volcano type behaviour upon increased Na coverage is obvious in case (**a**). Data was acquired in a Pt/(Na) β'' Al₂O₃/Au galvanic cell (Reprinted with permission from Ref. [16]; Copyright 1994, Elsevier).

Another EPOC study was reported by De Lucas-Consuegra et al. [27] on the effect of electrochemically imposed potassium (K) on the Pt-catalysed CO oxidation at 200–350 °C. Rate enhancement ratios of about $\rho = 11 (1100\%)$ at T = 278 °C and light-off temperature decreases by ~40 °C were achieved at equimolar reactant composition, [CO] = [O₂] = 5000 ppm (Figure 6 and Table 1).



Figure 6. Effect of catalyst overpotential (VWR) on the maximum rate enhancement ratio (ρ_{max}) and on T₅₀ of CO oxidation on Pt. Conditions: [CO] = [O₂] = 0.5% He balance, F_t = 108 cm³/min. Data was acquired in a Pt/(K) β "Al₂O₃/Au galvanic cell. (Reprinted with permission from Ref. [27]; Copyright 2008, Elsevier).

The maximum promotional effect was found under the application of -2 V; the appropriate parameters necessary for the estimation of the K coverage corresponding to this catalyst polarization are not available. The reversible promotional phenomena were attributed to the formation and decomposition of potassium compounds on the Pt surface (probably potassium carbonates and/or oxides) that enhanced the adsorption of oxygen at the expense of CO on a surface predominately covered by CO under unpromoted conditions in the reaction temperature range studied.

Regarding CCP studies, CO oxidation activity enhancements were also reported by Mirkelamoglu and Karakas [28,29] over conventional type 0.1wt% Na-promoted 1wt% Pd/SnO₂ catalyst tested at two different [CO]/[O] ratios, 1.25 and 0.5. The authors found that in the case of CO-rich gas composition ([CO]/[O] = 1.25), Pd can be up to 2.6-fold times (260%) more active on CO oxidation at 150 °C (Table 1), while non or moderate promotional effects were observed at O₂-rich conditions ([CO]/[O] = 0.5). A surface segregation of Pd atoms increasing the reactive sites for CO oxidation, together with the formation of super-oxide species observed over Na-PdO/SnO₂ were considered to be the origin of the observed promotional effects on the Na-dosed catalyst [29].

The effect of alkali oxide additives (Na₂O, K₂O) on the alumina support of Pt/Al_2O_3 catalysts on CO oxidation was investigated by Lee and Chen [30]. They concluded that the addition of alkalis on Al₂O₃ influences the basicity of the catalyst, which in turn results to higher oxidation activities. Higher promotional effect was achieved by potassium, which resulted in a decrease of CO light-off temperature by 90 °C under CO/O₂ stoichiometric conditions. The study, however, was performed under simulated two-stroke motorcycle emissions (involving C_3H_6 , H_2 , CO_2 and H_2O in the feed, besides CO and O_2) at the stoichiometric point and oxygen-deficient environments. Therefore, this work will be further analysed in the proper Section 3.2.1.

The influence of alkaline earths on CO oxidation activity was also investigated over Pd catalysts supported on Ba-modified γ -Al₂O₃ or CeO₂-ZrO₂ carriers by Tanikawa and Egawa [31]. They found that the activity of Pd/ γ -Al₂O₃ catalyst can be significantly improved by increasing the Ba promoter loading; an amount of 10–15wt% Ba decreases the light-off temperature by ~45 °C compared to the un-modified catalyst (Table 1). However, an adverse effect upon Ba addition was observed for Pd/CeO₂-ZrO₂ catalysts; an amount of 10 wt% Ba leads to an increase of light-off temperature by ~100 °C. The pronounced effect of Ba over Pd/ γ -Al₂O₃ catalysts was ascribed to the weakening of CO adsorption strength, whereas its inhibiting role over Pd/CZ catalysts was attributed to the suppression of Pd interaction with ceria-zirconia support.

In addition, the pronounced effect of electropositive promoters on CO preferential oxidation (PROX) under H₂-rich conditions has been recently demonstrated [32–41]. Minemura et al. [32,33] showed that the addition of alkalis (Li, Na, K, Rb and Cs) in a $2wt\% Pt/\gamma-Al_2O_3$ catalyst results in very significant promotional effects on the selective oxidation of CO under H2-rich environments (Table 1). The promotional effects followed "volcano" type behaviour upon increasing alkali loading, thus providing an optimal loading for each alkali; a 10–15 alkali/Pt loading ratio was found to optimize the promotional effects of Na and K, while the optimal alkali/Pt ratios were 3 and 5 for Cs and Rb, respectively. For a constant alkali/Pt loading equal to 3, the order on the promoting effect on CO conversion (and TOFs) was found to be Cs > Rb > K > Na > Li ~alkali-free Pt, while the order of the selectivity to CO_2 was K > Rb, Na > Cs > Li. At higher alkali loadings (alkali/Pt = 10), the order of the turnover frequency rates of CO oxidation were K > Na > Rb > Cs~Li > alkali-free Pt, with a similar order for the selectivity towards CO2. The changes in the order of the activity of the catalysts were due to over-promotion, that is, for some alkalis (e.g., Rb and Cs) their optimal loading was exceeded under the conditions used. In these studies, turnover rate enhancement ratio values up to $\rho = 10$ for the optimal K promotion at T = 100 °C [32] and even better (up to ρ = 20) at 80 °C [33] were achieved. Further studies by the research group, focused on the optimal promoter (K) when applying different supports (Al₂O₃, SiO₂, ZrO₂, Nb₂O₅ and TiO₂) for the dispersion of Pt, have shown that the most remarkable promoting effects are obtained on Pt/γ -Al₂O₃ [34] (Table 1). Corroborative characterization studies by means of a variety of techniques, e.g., Transmission Electron Microscopy (TEM), Extended X-ray Absorption Fine Structure (EXAFS), X-ray Absorption Near Edge Structure (XANES), Fourier-Tranform Infrared Spectroscopy (FTIR), enable the authors to consistently interpret the promotion features [35,36]. FTIR spectroscopic evidences for a K-induced weakening in the strength of CO adsorption on Pt and drastic changes on the adsorption site of CO (bridge and three-fold hollow CO species) found on the Pt (10wt% K)/ γ -Al₂O₃ catalyst were considered responsible for the observed promotional effects attributed to the change in the electronic state of K-modified platinum [35,36]. OH co-adsorbed species (originating form H₂ and O₂ interaction under PROX conditions) that promote the CO oxidation were found by in situ FTIR on the highly active K-modified Pt/γ -Al₂O₃ catalyst [36]. The inhibiting effects on PROX reaction under over-promotion conditions (beyond the optimal promoter loading) were attributed to aggregation of Pt metal particles caused by the larger amount of alkali metal used [36].

CO preferential oxidation was also found to be very pronounced over Pt clusters dispersed on alkali (Na, Rb or Cs)-modified SiO₂ by Pedrero et al. [37]. Optimum promotional effects were obtained with Cs at a surface concentration of 1.6 atoms/nm²; CO oxidation turnover rates (TOFs) more than one order of magnitude higher than those obtained on the un-promoted Pt/SiO₂ catalysts were recorded, which were accompanied by CO oxidation selectivity >90% (Table 1). In accordance to the authors, the beneficial effect of alkalis can be interpreted based both on the electronic interactions between

alkali atoms and metals and on the inhibition of spill over-mediated H_2 oxidation pathways induced by alkalis.

The effect of Na promoter on CO PROX reaction was also investigated over bimetallic $PtCo/Al_2O_3$ catalysts by Kwak et al. [38]. They found moderate promotion of PROX activity on monometallic Pt catalyst while substantial rate promotion on the bimetallic PtCo counterpart catalyst up to a sodium loading of 2.0wt%; the CO-PROX selectivity was not influenced in both cases (Table 1). They concluded that the formation of Na-O-Al bonds by incorporation of Na ions into the alumina lattice, that happens at low sodium content, suppresses the formation of the surface spinel cobalt species and promotes the formation of bimetallic Pt-Co species; at higher Na content (ca. 3wt%) bulk sodium particles on the catalyst surface interact with the active species inhibiting PROX activity.

On the other hand, a notable increase on both CO conversion and CO₂ selectivity under PROX conditions was attained with Mg-promoter over Pt/Al_2O_3 catalysts by Cho et al. [39] (Table 1). The maximum values of CO conversion and CO₂ selectivity for Pt-Mg/Al₂O₃ catalysts were 93.1 and 62%, respectively, at 170 °C, compared to 70.2 and 46.8% at 200 °C over the un-promoted Pt/Al₂O₃ catalysts. The effect of CO₂ and H₂O in the feed was also studied in this work under PROX conditions, as well as the CO oxidation reaction in the absence of excess H₂ under Mg-promotion. The superior performance of Mg-modified catalysts was attributed to an increase in Pt electron density, as well as to an increase of hydroxyl groups on catalyst surface.

Finally, de Lucas-Consuegra and co-workers [40] demonstrated the pronounced effect of electropositive modifiers (K⁺ ions) on the preferential oxidation of CO via EPOC by the use of a Pt/K⁺-conducting- β Al₂O₃/Au galvanic cell. At a 195 °C and a feed composition of CO/O₂/H₂ = 0.4%/0.2%/16%, the authors found up to ρ = 1.3 rate enhancements, achievable at a potassium coverage of $\theta_{\rm K}$ ~2–4%. A concomitant increase in CO₂ selectivity of ~10% was recorded as well (Table 1). The effect of the alkali on the chemisorptive bonds of CO, H and O on Pt surface was considered to be the origin of the observed promotional phenomena.

For more information, a short review has been recently published regarding the promotional effect of alkali promoters (CCP) on PGMs (Pt, Ru and Ir)-catalysed CO PROX reaction [41].

A comparative overview of the literature findings included in Section 3.1.1 and in Table 1 leads to the following general remarks: Among PGMs only Pt and Pd were investigated so far for CO oxidation under electropositive promotion (by alkalis or alkaline earths). No studies exist that were performed at similar conditions using both EPOC and CCP methods, preventing us from making a direct one-to-one comparison of the outputs of the two methods. The general view is that alkalines efficiently promote the Pt and Pd catalysed CO oxidation only at conditions where the reaction rate is negative order in CO, positive order in O₂ (the so-called CO-rich conditions at which CO coverage predominates on the catalyst surface; not necessary at $CO/O_2 > 1$), while non, moderate or only poisoning was found under O₂-rich conditions. Even in the former case, the promotion follows volcano type behaviour upon increasing promoter loading, reversed to poisoning for high promoter loadings; optimal promoter loadings were typically low. The maximum rate enhancement ratios were reported for Pt/CO+O₂ catalytic system under EPOC: $\rho = 6$ with Na at 350 °C [16] and $\rho = 11$ with K at 278 °C [27].

Volcano-type promotion was the main feature of the electropositive promotion of PGMs under the preferential CO oxidation (PROX), as well. Moreover, for similar reaction and promoter loading conditions larger alkalis (Cs, Rb, K) offered superior promotion. For this reaction two studies exist that concern usage of the same promoter (K) and active metal (Pt), one performed using CCP the other EPOC. Although ρ values obtained by CCP were significantly larger ($\rho = 10$ [32–36]) than that obtained by EPOC ($\rho = 1.3$ [40]), the applied conditions were quite different, preventing us from making a creative comparison.

The alkaline-induced pronounced enhancement on oxygen adsorption compared to that of CO was a key argument in most of the aforementioned studies upon explaining their findings, whereas the formation of large surface alkali complexes was considered responsible for the rate poisoning at high alkali loadings.

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Ref.	[16]	[27]	[29]	[31]
Promotion highlights and optimal achievements	 ✓ Volcano behaviour of promotion upon increasing θ_{Na}, under CO-rich conditions. ✓ Promotion occurs at θ_{Na} = 2-4%, under CO-rich conditions; moderate inhibition at low θ_{Na} = 2-4%, under O₂-rich conditions; strong inhibition for θ_{Na} > 4% in both cases. ~ ρ = 6 at θ_{Na} = 4%, Γ = 350 °C, CO-rich conditions ([CO] = 5.3%, [O₂] = 6%) 	 V Significant K-induced promotion, providing a system proper for LT CO oxidation. ρ = 11 at T = 278 °C and ΔT₅₀ = -40 °C; [CO] = [O₂] = 500 ppm 	 V Significant promotion at CO-rich conditions; more effective on pre-reduced catalysts Achievements on pre-oxidized catalysts: <i>ρ</i> = 1.5, ΔT₅₀ = -12 °C (at [CO]/[O] = 1.25); <i>ρ</i> = 0.66, ΔT₅₀ = +4 °C (at [CO]/[O] = 0.5) Achievements on pre-reduced catalysts: <i>ρ</i> = 2.6, ΔT₅₀ = -25 °C (at [CO]/[O] = 0.5) 1.3, ΔT₅₀ = -8 °C (at [CO]/[O] = 0.5) 	 ✓ Promotion on Pd(Ba)/γ-Al₂O₃; poisoning on Pd(Ba)/CZ ▶ Promotion: ΔT₅₀ = -45 °C, with 10-15wt% Ba loading on Pd(Ba)/γ-Al₂O₃ ▶ Poisoning: ΔT₅₀ = +100 °C, with 10wt% Ba loading on Pd(Ba)/CZ
Reaction conditions	T = 300-450 ° C [CO] = 0-4% [O2] = 0-6%	T = 200-350 ° C [CO] = [O ₂] = 500 ppm	T = 150 °C [CO]/[O] = 1.25 [CO]/[O] = 0.5 ([CO] = 2400 ppm) Na-loading: 0.1 wt%	$T = 50-250 \ ^{\circ}C$ [CO] = 2.2%, [O ₂] = 1.1% Ba-loadings: 0, 1, 5.5, 10, 15 wt%
Promoter	Na	K	Na	Ba
Catalyst, (promotion method applied)	Pt-film over (Na)β"Al ₂ O ₃ solid electrolyte (EPOC)	Pt-film over (K)β" Al ₂ O ₃ (EPOC)	1 wt% PdO/SnO2 (CCP)	1 wt% Pd/y-Al ₂ O ₃ and 1 wt% Pd/CZ (CCP)
Reactants	CO, O ₂	CO, O ₂	CO, O ₂	C0, O2

Table 1. Electropositive promotion of PGMs-catalysed CO oxidation (CO+O₂) and Preferential CO oxidation (PROX: CO+O₂+excess H₂).

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Table 1. Cont.

Ref.	[32] [33] [34] [36]	[37]	[38]
Promotion highlights and optimal achievements	 Volcano-type promotion upon increasing alkali loading. K was the superior promoter. For A/Pt < 3 the promoting effect of alkalis followed the order: Cs > Rb > K > Na > Li For A/Pt > 5 the promoting effect of alkalis followed the order: K > Na > Rb > Cs > Li An A/Pt=10-15, 3 and 5 optimizes Na or K, Cs and Rb promotion, respectively. \$\rho\$ = 10 with K, with a loading of K/Pt = 10 molar ratio 	 ✓ Volcano-type promotion upon increasing alkali loading. Gs was the superior promoter. ✓ Promotion is maximized at a Cs loading = 1.6 Cs/nm² of catalyst surface. ➤ ρ = 10 and ΔS_{CO} = +30% (from 60%→90%) with an 1.6 Cs/nm² alkali surface density. 	 ✓ Moderate promotion of Na on Pt/Al₂O₃ catalyst. ✓ Significant volcano-type promotion (maximized at 2 wt% Na) on PtCo/Al₂O₃ catalyst. ▲ ΔX_{CO} = 75% (from 25% →100%) on 2 wt% Na promoted PtCo/Al₂O₃.
Reaction conditions	T = 100–160 ° C [CO] = [O ₂] = 0.2% Alkali loadings: A/Pt molar ratio = 0–20	$T = 110 \circ C$ $[CO] = [O_2] = 1\%, [H_2] = 70\%$ Cs surface density = 0-6 atoms/nm ²	T = 25–300 °C [CO] = 0.1%, [O2] = 0.1%, [H2] = 1% Na loading: 0.5–3 wt% Na
Promoter	Li, Na, K, Rb, Cs	Na, Rb, Cs	Na
Catalyst, (promotion method applied)	2 wt% Pt/(y-Al ₂ O ₃ , SiO ₂ , ZrO ₂ , Nb ₂ O ₅ or TiO ₂) (CCP)	1.6 wt% Pt/SiO2 (CCP)	1 wt% Pt/γ-Al ₂ O ₃ 1 wt% Pt-1.8wt%Co/γ-Al ₂ O ₃ (CCP)
Reactants	CO, O ₂ , H ₂ excess	CO, O ₂ , H ₂ excess	CO, O ₂ , H ₂ excess
Ref. [39] [40] **Promotion highlights and optimal achievements** reaction (max $X_{CO} = 93\%$ and max $S_{CO} = 62\%$ Both CO oxidation and PROX reactions were $\Delta T_{50} = -20\ ^\circ\text{C}$ for CO + O₂ reaction at moist For PROX reaction both CO conversion and at 170 $^{\circ}\text{C})$ on 3wt% Mg-promoted catalyst. $\rho = 1.3$ and ~10% increase in CO oxidation The inhibiting effect of H₂ on PROX was selectivity were enhanced in 150-230 °C. K effect allows activating the Pt at lower $\Delta X_{CO} > 35\%$ and $\Delta S_{CO} > 15\%$ for PROX Both PROX activity and selectivity were temperature on CO oxidation selectivity significantly enhanced by K addition. significantly promoted by 3wt% Mg. attenuated. The inhibiting effect of conditions $(2\% H_2O \text{ on stream})$. selectivity, within $\theta_{\rm K} = 2-4\%$. temperatures for PROX. was attenuated. > A A > A > [H₂] = 65% [CO₂] = 20%, [H₂O] = 2% $[CO] = 1\%, [O_2] = 0.75\%,$ $[CO] = 0.4\%, [O_2] = 0.2\%,$ Mg loading = 3 wt% Mg**Reaction conditions** $T = 100-250 \circ C$ $[H_2] = 16\%$ $T = 195 \ ^{\circ}C$ $\theta_{\rm K}=0{-}4\%$ Promoter В Ч Catalyst, (promotion method applied) Pt film on $(K)\beta Al_2O_3$ solid electrolyte $2wt\%Pt/\gamma\text{-}Al_2O_3$ (CCP) (EPOC) CO, O₂, H₂ CO, O₂, H₂ Reactants excess excess

Table 1. Cont.

3.1.2. Light Hydrocarbons Oxidation

(i) Alkenes oxidation

The oxidation of ethylene over polycrystalline Pt films has been extensively studied under electropositive promotion (EPOC) by Na via a Na⁺-conducting $\beta''Al_2O_3$ solid electrolyte, by Vayenas and co-workers [18,42] (Table 2). Both promoting and poisoning effects induced by alkali have been observed: low Na coverages (typically <8%) were found to cause increases in the reaction rate (up to $\rho = 2$), while larger Na coverages led to a significant decrease of rate, which eventually falls well below the initial value corresponding to Na-free Pt surface (Figure 7). The data was interpreted in terms of (i) Na-enhanced oxygen chemisorption and (ii) poisoning of the surface by accumulation of Na compounds.



Figure 7. Rate as a function of Na coverage for both potentiostatic and galvanostatic experiments for $[O_2] = 8.0\%$ [C_2H_4] = 4.2% (balance He) mixture at 352 °C. Data was acquired in a Pt/(Na) β ″Al₂O₃/Au galvanic cell (Reprinted with permission from Ref. [18]; Copyright 1996, Elsevier).

According to the authors, low Na coverage gives rise to an enhanced chemisorption of O_2 at the expense of ethylene, resulting in an increased rate. At high sodium coverages both (i) and (ii) factors operate synergistically to poison the system: the increased strength of the Pt–O bond and the coverage of the catalytic surface by compounds of Na strongly suppress the rate. Kinetic results obtained over a Pt(111)–Na model catalyst, dosed with several amounts of Na by vapour deposition (CCP), were found to mirror the behaviour of the Pt/ β "Al₂O₃ catalyst promoted by the EPOC concept [18]. This

strongly suggested that Na pumped from the solid electrolyte via electrical polarization is indeed the key species for the promoting or poisoning phenomena. In addition, these results demonstrated that electrochemical promotion of catalysis (EPOC) and surface promotion by conventional means (CCP) obey the same physiochemical rules in the case of alkaline promoting cations. This is well documented nowadays but it was unclear at those early steps of electrochemical promotion studies.

The Pt-catalysed propene oxidation has been also extensively studied by Filkin et al. [20] under in situ electrochemical promotion (EPOC) by sodium, using polycrystalline Pt films deposited on Na⁺-conducting β'' Al₂O₃. Depending on propene and oxygen reactants partial pressures, fully reversible promotion and poisoning effects were observed as a function of sodium coverage as follows. As $[O_2]/[C_3H_6]$ decreases, the effects of Na promotion rapidly increase: for $[O_2]/[C_3H_6] = 5.7$, where the activity of Na-free Pt is highest, the maximum enhancement ratio $\rho_{max} = r_{max}$ (on Na-promoted Pt)/r_o(on Na-free Pt) is equal to 1.1, which then increases to $\rho_{max} = 2.3$ at $[O_2]/[C_3H_6] = 2.0$, where Na-free Pt activity is lower. These maxima enhancement ratios where obtained at moderate Na coverages (not specified) while higher Na coverages cause poisoning of the propene oxidation rate. The larger poisoning was obtained in the case of $[O_2]/[C_3H_6] = 5.7$ (Table 2). A Na-modified chemisorption of the reactants considered (Na enhances oxygen chemisorption and inhibits propene chemisorption) on a Langmuir-Hinselwood type reaction accounted all the experimental findings. Using X-ray Photoelectron Spectroscopy (XPS), Auger Electron Spectroscopy (AES) and postreaction Na K edge XANES (X-ray Absorption Near Edge Structure) the authors have demonstrated that under reaction conditions the promoter phase consists of small amounts of 3D sodium carbonate crystallites, while thick layers of this sodium surface compound were responsible for rate inhibition in the poisoning regime. Notably, these promoter and poisoning phases, although stable at reaction temperatures investigated, were readily destroyed via electrochemical Na pumping away from the catalyst by the application of electrical polarizations of opposite sign [20].

Other studies involving the propene oxidation over Pt films using the EPOC concept as well, were reported by Vernoux and co-workers [43,44]. In these studies Pt was interfaced with NASICON (Na₃Zr₂Si₂PO₁₂), an alternative Na⁺-conductor, acting therefore as a Na source for its electrochemical pumping to the Pt catalyst surface. The authors used two distinguished propene/oxygen gas phase compositions, near stoichiometric (C₃H₆/O₂ = 0.04%/0.2%) and excess oxygen (C₃H₆/O₂ = 0.04%/8.3%) conditions, to investigate the effect of Na on the catalytic system under consideration at T ~ 300 °C (Table 2). They found that under oxygen excess Na coverages larger than 3% cause a slight poisoning of the propene oxidation rate ($\rho \sim 0.8$). However, under near-stoichiometric conditions, sodium had strong beneficial effect ($\rho = 3.5$) on the Pt-catalysed propene oxidation rate, which is maximized at a $\theta_{Na} \sim 3.6\%$; the rate was then slightly decreased when higher (up to 6%) sodium coverages were supplied. Cyclic voltammetry studies indicated that sodium promoter exists in the form of Na₂CO₃ and NaHCO₃ surface phases during reaction [43], in agreement with earlier studies by Filkin et al. [20].

In this line, de Lucas-Consuegra et al. [45,46] have more recently studied the propene oxidation under EPOC of a Pt film interfaced with a K⁺-conducting β Al₂O₃ solid electrolyte, that is, using potassium as promoter species. They found that the reaction can be strongly promoted at the temperature interval of ~200–300 °C, under both near-stoichiometric and oxygen-rich conditions. Rate enhancement ratios of the order of ρ = 7 were achieved for a catalyst polarization of about -2 V (Table 2); parameters' values necessary for the estimation of the corresponding alkali coverage are not available in the paper. Cyclic voltammetry, FTIR and SEM-EDX studies enable the authors to identify the formation of stable potassium oxide and superoxide phases upon electrical polarization. These species were responsible for the promotional phenomena. In addition, due to the stability of the potassium surface compounds formed, a permanent promotional effect was observed.

The effect of electropositive modifiers on hydrocarbons oxidation over conventional highly dispersed catalysts (CCP) was also investigated by Yentekakis and co-workers [47]. In particular, they studied the propene oxidation over Na-promoted $Pt(Na)/\gamma$ -Al₂O₃ catalysts under oxygen excess

(1000 ppm propene + 5% O₂). Significant activity enhancements were observed for a Na-loading of ~1.5–2.6 wt % (or 0.68–1.13 × 10⁻³ mmol effective Na/g_{cat}) [47]. Optimal promotion was achieved for 2.6 wt% Na loading, corresponding to a nominal sodium coverage of θ_{Na} ~23% (Equation (4) by using $d_{\text{Na}} = 3.314 \times 10^{19}$ atoms/m²), causing a significant shift of the propene light-off temperature to about 60 °C lower values (Figure 8, Table 1). Rate enhancement ratios as high as ρ ~10 can be calculated from these results. It is worth noting that the aforementioned promotional effects were recorded under oxygen excess conditions, typical of lean-burn combustion engines. The Na-induced enhancement of oxygen gas phase conditions, was considered to be responsible for the observed promotional effect.



Figure 8. The effect of temperature on the conversion of C_3H_6 Na-dosed Pt(Na)/ γ -Al₂O₃ catalysts under $C_3H_6 + O_2$ reaction at oxygen rich conditions: 1000 ppm $C_3H_6 + 5\%$ O₂; catalyst weight 70 mg; total flow rate was varying as $F_t = 210$, 126, 62.5 and 52 cm³/min for C1, C2, C3 and C4 catalysts, respectively, in order to keep the effective contact time of C_3H_6 constant at 4 s. (The effective contact time is defined as: surface Pt atoms/reactant molecules/s) (Reprinted with permission from Ref. [47]; Copyright 2005, Elsevier).

(ii) Alkanes oxidation

An EPOC study regarding the propane oxidation on Pt-based catalysts under Na-promotion supplied via a (Na) β'' Al₂O₃ solid electrolyte has been reported by Kotsionopoulos and Bebelis [48]. The activity was dramatically inhibited by Na-addition even at low Na coverages, over a wide range of propane/oxygen ratios (Table 2). Under stoichiometric conditions ([O₂]/[C₃H₈] = 1%/0.2%) and T = 320 °C, a Na coverage as small as θ_{Na} ~2.5% leads to a significant decrease of the reaction rate of 625 times. The poisoning effect was more pronounced at higher temperatures (e.g., 360, 400 and 440 °C). This poisoning effect of electropositive promoters on Pt for alkanes oxidation can be attributed to the weak adsorption of alkanes compared to oxygen. A latter study published by the same authors dealt with the electrochemical characterization of the aforementioned catalytic system [49]. In this work they observed, in good agreement with some of the studies mentioned above, that the sodium promoter is on the metal surface mainly forming carbonate, bicarbonate and oxide phases.

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Ref.	[42]	[18]	[20]	
Promotion highlights and optimal achievements	 ✓ Only rate inhibition was found at such extreme fuel lean conditions. ➤ A Na coverage of θ_{Na} > ~1.5% suffices to cause a 70% rate inhibition (poisoning). 	 Na induces both promotion and poisoning. Moderate θ_{Na}, typically<8%, causes promotion, larger θ_{Na} causes rate inhibition. Results obtained on a Pt(111)-Na model catalyst mirror the behaviour of the EPOC system. No promotional and/ or poisoning effects occur at high temperatures (T > ~460 °C). p-2 with θ_{Na} = 8% at 352 °C, [O₂] = 16.9%, p-2 with θ_{Na} = 8% at 352 °C, [C₂H₄] = 4.2%, [O₂] = 8% 	 Volcano type promotion optimized at moderate Na coverages. Promotion effects increase with decreasing [O₂]/[C₂H₄] (promotion is superior under fuel-rich conditions). High Na coverages deactivate catalyst due to excessive accumulation of 3D Na₂CO₃ species; these species can be readily removed electrochemically, accomplishing catalyst re-activation. <i>P</i> = 1.1 at [O₂]/[C₂H₄] = 5.7, T = 340 °C; ρ = 2.3 at [O₂]/[C₂H₄] = 2.0, T = 340 °C; ρ = 2.3 at [O₂]/[C₂H₄] = 2.0, T = 340 °C. 	
Reaction conditions	$\begin{array}{l} T=291\ ^{\circ}C;\\ [C_{2}H_{4}]=0.021\%,\ [O_{2}]=\\ 5\%;\\ \theta_{Na}=0-3\% \end{array}$	$T = 300-470 \circ C;$ $[C_2H_4] = 4.2\%, [O_2] = 8$ and 16.2%; $\theta_{Na} = 0-35\%$	T = 340 °C [C ₃ H ₆] = 0.62%, [O ₂] = 0-7%	
Promoter Alkenes ox	Na	Na	Na N	
Catalyst and (Promotion method applied)	Pt-film over (Na)β"Al2O3 (EPOC)	Pt film over (Na)β″Al ₂ O ₃ solid electrolyte (EPOC)	Pt film over (Na) $eta''Al_2O_3$ solid electrolyte (EPOC)	
Reactants	C ₂ H ₄ , O ₂	C ₂ H ₄ , O ₂	C ₃ H ₆ , O ₂	

Table 2. Electropositive promotion of PGMs-catalysed alkenes and alkanes oxidation.

Ref.	[43] [44]	[45] [46]	[47]
Promotion highlights and optimal achievements	 ✓ Strong promotion by Na under near-stoichiometric C₃H₃/O₂ conditions; no promotion or moderate poisoning under oxygen excess (lean-burn) conditions. Under reaction conditions promoter phase exists as Na₂CO₃ and NaHCO₃. > p = 3.5 with θ_{Na} = 3.6% at [C₃H₆]/[O₂] = 0.04/02. T = 300 °C. p = 0.8 with θ_{Na} > ~2% at [C₃H₆]/[O₂] = 0.04/8.3, T = 300 °C. 	 ✓ Strong promotion by K (better than that of Na) under both near-stoichiometric and oxygen-rich conditions. ✓ Formation of stable and more effective promoter phases (potassium oxides and superoxides). > p = 7, propene conversion increased form 10% to 70%, at 200 °C and [C₃H₆]/[O₂] = 2000ppm/7% (O₂-rich) > p = 3 at 220 °C and [C₃H₆]/[O₂] = 2000ppm/1% (near-stoichiometric). 	 ✓ Volcano type promotion under excess oxygen conditions (optimal Na-loading at ~2.6wt% Na). ✓ C₃H₆ oxidation is inhibited for higher Na loadings; poisoning becomes even worse in case of NO co-feed. ➤ ΔT₅₀ = -60 °C at [C₃H₆]/[O₂] = 0.1%/5% (O₂-rich conditions) with 2.6wt% Na (θ_{Na}~23%); <i>p</i>~10 at T = 120-200 °C.
Reaction conditions	$T = 300 \circ C, \theta_{Na} = 06\%$ - C ₃ H ₆ /O ₂ = 0.04%/0.2% (stoich.) - C ₃ H ₆ /O ₂ = 0.04%/8.3% (O ₂ -rich)	T=190-310 °C [C3H6] = 2000 ppm, [O2] = 1-7%	$T = 100-500 \circ C$ $[C_3H_6] = 1000 \text{ ppm}, [O_2]$ $= 5\%$ Na-loadings: 0, 1.6, 2.6, 4.2 wt%
Promoter	a N	К	Na
Catalyst and (Promotion method applied)	Pt film over NaSICon (Na ₃ Zr ₂ Si ₂ PO ₁₂) solid electrolyte (EPOC)	Pt film over (K)βAl ₂ O ₃ solid electrolyte (EPOC)	0.5 wt% Pt/y-Al2O3 (CCP)
Reactants	C ₃ H ₆ , O ₂	C ₃ H ₆ , O ₂	C ₃ H ₆ , O ₂

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Ref.	[48]	[20]			
Promotion highlights and optimal achievements	 Na addition to the catalyst causes only poisoning of the rate. Low θ_{Na} = 1-3% are enough to cause up to 3 orders of magnitude rate poisoning (ρ~0.002). Under reaction conditions Na forms carbonate, bicarbonate and oxide phases. Strong poisoning: ρ~0.002 for θ_{Na} >~2.5% at T = 440 °C. 	 Ø Both beneficial and detrimental effects on CH4 oxidation were found depending on the promoter, catalyst composition and pre-treatment. Promotion was achieved with Ca-doping and after cooling in inert atmosphere. No promotion (even inhibition) was observed by Na and K addition at the specific loadings used. Highest activities for 2.5wt%pd/γ-Al₂O₃ and 5wt%pd/γ-Al₂O₃ were and 0.76 wt% Ca addition, respectively, that is, at a weight ratio Ca:Pd equal to 0.15 in both cases. 			
Reaction conditions cidation (alkane + O ₂)	T = $320-440 \circ C$ [C ₃ H ₈] = 0.2% , [O ₂] = 1% $\theta_{Na} = 0-3\%$	$T = 250-600 \circ C;$ $[CH_4] = 500 \text{ ppm}, [O_2] = 8\%, [H_2O] = 5\%, [CO_2] = 8\%, [H_2O] = 5\%, [CO_2] = 5\%, 1.5 \text{ wt}\%, Ba = 5 \text{ wt}\%, K = 1.48 \text{ wt}\%, Ba = 5 \text{ wt}\%, K = 1.48 \text{ wt}\%, Na = 0.87 \text{ wt}\%$			
Promoter Alkanes ox	Na	Na, K, Ba, Ca			
Catalyst and (Promotion method applied)	Pt film over (Na)β"Al ₂ O ₃ solid electrolyte (EPOC)	5 (or 2.5) wt% Pd/y-Al2O3 (CCP)			
Reactants	C ₃ H ₈ , O ₂	CH4, O2			

Table 2. Cont.

Olsson and co-workers [50] studied the effect of the addition of alkalis (Na, K) and alkaline earths (Ba, Ca) on Pd/Al₂O₃ catalyst for methane combustion (CCP). Both beneficial and detrimental effects on methane oxidation were observed. High loadings of alkaline modifiers caused reaction inhibition. No promotion was induced by Na and K modifiers, while Ba- and Ca- doping caused reaction inhibition. However, when using a specific catalyst pre-treatment, which leads to the removal of surface hydroxyl species, promotional effects of Ca were recorded at Ca loadings where they previously observed reaction inhibition. The Ca/Pd ratio was found to be critical for the aforementioned promotion. A Ca/Pd ratio of 0.15 was favourable. The modification of the redox properties of palladium and the formation of PdO were considered the origin of this promotion, which overcomes the alkaline-induced inhibition of methane adsorption (activation) on increased electronic density catalyst particles surfaces which occurs in parallel.

Comparatively overviewing the literature findings of Section 3.1.2 summarized in Table 2, the following general features and conclusions can emerge.

Alkali addition on PGMs-based catalyst is detrimental for alkanes oxidation reactions that was attributed to the low adsorption propensity of alkanes on PGMs surfaces which becomes worse due to the alkali induced promotion of oxygen adsorption and the concomitant O-poisoning of the surfaces.

On the opposite, alkenes oxidation on PGMs is substantially promoted by electropositive promoters, attributed to the opposite trend of their adsorption compared to alkanes: The alkali-induced enhancement on oxygen adsorption counterbalances its coverage on a PGM surface, predominantly covered by the alkene molecules on the unpromoted surface, resulting to an enhancement on reaction probability (promotion).

The alkenes oxidation followed volcano type promotion upon increasing promoter loading, resulting from the competitive adsorption of the reactants (Langmuir-Hinshelwood type reaction) and the formation of 3D alkali carbonates and their excessive accumulation on the surface at high promoter loadings.

The main promotion characteristics were qualitatively similar between EPOC and CCP studies, although the intensity of the promotion was larger in the latter case: maximum rate enhancement ratios (ρ) obtained were 3.5 and 10 by using EPOC and CCP, respectively. An ultimate contact between nanodispersed active phase particles and promoter species, which is favoured in CCP, is most probably responsible for this superiority.

3.1.3. NO Reduction by CO

The first study of the effect of electropositive promoters on the Pt-catalysed NO reduction by CO was reported by Palermo et al. [17]. The EPOC was applied by means of a Pt/Na⁺-conducting β'' Al₂O₃/Au cell. The promotional effect of Na on the Pt/NO + CO system was strongly dependent on the reaction temperature, gas composition and Na coverage. For instance, at high CO/NO ratios the effect of Na was marginal and according to the authors this was attenuated by CO island formation and limited availability of chemisorbed NO. At low CO/NO ratios the effect of Na was also relatively small, because, as authors conclude, the CO + O reaction is limited by the low CO coverage. However, at the intermediate CO/NO regime, where the above constrains do not apply, Na strongly accelerates the reaction: rate gains of up to $\rho = 13$ for N₂ production compared to the unpromoted Na-free Pt rate were achieved (Table 3). Significant increases to N₂ selectivity of the system (up to 30%) were also obtained (Figure 9). Significant progressive changes in the apparent activation energy of the NO + CO reaction upon varying the promoter coverage, similar to that obtained with the Pt/O₂ + CO catalytic system [16], were also recorded in this study. The observed promotional phenomena, on both rate and selectivity, were attributed to the enhanced NO versus CO chemisorption and the extended dissociation of the chemisorbed NO molecules, both induced by Na-promoter.

Ref.	[17]	[51]	[23]
Promotion highlights and maximum achievements	The promotional effects of Na are maximized at intermediate CO/NO values, while at high and low CO/NO values promotion was attenuated due to the formation of CO islands and the low CO coverage, respectively. Significant Na-induced changes on the reaction apparent activation energy recorded. $\rho_{N_2} \sim 13$, $\rho_{CO_2} \sim 4.8$, $\Delta S_{N_2} \sim 45\%$ at T = 348 °C for $\theta_{Na} = 3-20\%$	Sodium coverage (θ_{Na}), catalyst work function ($\Delta \phi$) and catalyst potential (V_{WR}) are directly correlated; θ_{Na} and $\Delta \phi$ varying linearly with V_{WR} . Low sodium coverages ca. 1–2% are able to create significant promotion of the Rh-catalysed NO+CO (but less than that obtained on Pt catalysts). $\rho_{N_2} \sim 3.1, \rho_{CO_2} \sim 1.4, \rho_{N_2O} = 0.3, \Delta S_{N_2} \sim +50\%$ $(24\% - 80\%)$ at T = 307°C, for $\theta_{Na} - 1-2\%$	All K, Rb, Cs alkalis exert a pronounced promotion; however, Rb promotion is superior to K and Cs. In comparison with Rb, larger molar loadings of K were needed to produce similar reductions in light-off temperature. In both cases promotional effects increased monotonically with increasing alkali loading. $P_{N_2} \sim 110$, $P_{CO_2} \sim 45$, $P_{N_2O} = 7$, $\Delta S_{N_2} = +58\%$ $24\% \rightarrow 82\%$) at T = 350 °C, [CO] = 0.5\%, [NO] = 1\%; $P_{N_2} \sim 45$, $P_{CO_2} \sim 45$, $P_{N_2O} = 7$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 110$, $P_{CO_2} \sim 25$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45$, $P_{CO_2} \sim 15$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45$, $P_{CO_2} \sim 15$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45$, $P_{CO_2} \sim 15$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45$, $P_{CO_2} \sim 15$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45$, $P_{CO_2} \sim 15$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45$, $P_{CO_2} \sim 15$, $P_{N_2O} = 4$, $\Delta S_{N_2} = +51\%$ $P_{N_2} \sim 45\%$ for $N_2 = -150$ °C with R loading >9.7 wt%; $\Delta T_{50} = -115$ °C with K loading >9.7 wt%; $\Delta T_{50} = -115$ °C with Cs loading > 15 wt%, at [NO] = [CO] = 1000 ppm.
	> > A	> A	S S A A A
Reaction conditions	$T = 300-400 \circ C$ $[CO] = 0-1.5\%, [NO] = 0-1.5\%$ $[NO]/[CO] = 0.75\%, 0.75\%, 0.75\%, 0.75\%, 0.8\%, 0.75\%$	T = 307 °C [CO] = [NO] = 1%	T = 150–500 °C [CO] = 0–3%, [NO] = 0–3% Promoter loadings used: wt% RB: 1.9, 9.7, 15.5 wt% K: 2.7, 4.4, 8.8
Promoter	Na	Na	K, Rb, Cs
Catalyst and (Promotion method applied)	Pt film over (Na) $\beta''Al_2O_3$ solid electrolyte (EPOC)	Rh film over (Na)β″Al2O3 solid electrolyte (EPOC)	0.5 wt% Pt/y-Al2O3 (CCP)
Reactants	NO, CO	NO, CO	NO, CO

Table 3. Electropositive promotion of PGMs-catalysed NO reduction by CO (NO + CO).

Ref.	[54]	[22]
Promotion highlights and maximum achievements	Volcano type promotion was obtained with an optimal Na-loading equal to 0.034wt% Na (i.e., Pd/alkali molar ratio = 3/1), taking into account both activity and N ₂ - selectivity. $\rho_{N_2} \sim 2$, $\rho_{CO_2} \sim 15$, $\rho_{N_2} O = 1$, $\Delta S_{N_2} = +20\%$ (from ~50% to ~70%) at T = 332 °C. [CO] = [NO] = 1% for 0.034 wt% Na.	Ba-induced promotional effects on Pd/ γ -Al ₂ O ₃ ; progressively increased activity with increasing Ba-loading. Ba-induced poisoning effects on Pd/CZ; progressively decreased activity with increasing Ba-loading. $\Delta T_{50(CO)} = -40 \circ C$ and $\Delta T_{50(NO)} = -33 \circ C$ with 15wt% Ba on Pd/ γ -Al ₂ O ₃ (promotion) $\Delta T_{50(CO)} = +97 \circ C$ and $\Delta T_{50(NO)} = +99 \circ C$ with 10wt% Ba on Pd/CZ (strong poisoning)
	> A	> > A A
Reaction conditions	$T = 352 \circ C$ [CO] = 0-4%, [NO] = 0-4% Na loadings: 0.017, 0.034, 0.068 and 0.102 wt% (or 6/1, 3.1, 1.5/1 and 1/1 Pd/Alkali)	T = 50-300 °C [CO] = 2.2%, [NO] = 0.2% [O ₂] = 1% Ba-loadings: 0, 1, 5 and 15 wt% Ba
Promoter	Na	В
Catalyst and (Promotion method applied)	0.5wt%Pd/YSZ (CCP)	1 wt% Pd/(γ-Al ₂ O ₃ or CZ) (CCP)
Reactants	NO, CO	NO, CO, O ₂

Table 3. Cont.



Figure 9. Effect of catalyst potential on the CO₂, N₂ and N₂O formation rates and the selectivity towards N₂ during the EPOC by Na of the Pt-catalysed NO + CO reaction. Conditions: [NO] = [CO] = 0.75%, T = 348 °C. Data was acquired in a Pt/(Na) β ″Al₂O₃/Au galvanic cell. (Reprinted with permission from Ref. [17]; Copyright 2005, Elsevier).

Another EPOC (by Na) study using a Rh/(Na) β'' Al₂O₃/Au galvanic cell was performed by Williams et al. [51,52]. They showed that the NO reduction by CO can also be promoted over Rh catalyst. Rate enhancement ratios of $\rho_{N_2} = 3.1$, $\rho_{CO_2} = 1.4$ and $\rho_{N_2O} = 0.3$ were observed at T = 307 °C and stoichiometric CO/NO gas reaction conditions ([NO] = [CO] = 1%) for a sodium coverage as low as 2% (Table 3), together with a large increase in N₂-selectivity: from 24% on the electrochemically cleaned Na-free Rh to 80% on Rh electrochemically dosed with $\theta_{Na} = 2\%$. These promotional effects were attributed to the Na-induced NO dissociation, which was considered as the rate limiting step for the NO reduction by CO process. Using XPS and work function ($\Delta \Phi$) measurements the authors also demonstrated that the Na coverage on the catalyst surface can be controlled through the externally imposed catalyst potential; a linear relationship of catalyst work function and θ_{Na} with catalyst potential was experimentally demonstrated in this study [51,52].

Konsolakis et al. [53] studied the NO + CO reaction on (K, Rb and Cs)-promoted, conventional (CCP), highly dispersed Pt/ γ -Al₂O₃ catalysts, over a wide range of temperature (ca. 150–500 °C), partial pressures of reactants and promoter loadings (Table 3). All these alkalis were found to be very active promoters for the reaction, however, optimal promotion was achieved by rubidium: rate increases by factors as high as $\rho_{N_2} = 110$ and $\rho_{CO_2} = 45$ for the production of N₂ and CO₂, respectively, relative to alkali-free Pt were obtained, accompanied by substantial increases in N₂-selectivity by up to ~60% (Figure 10) [53]. The effects of K and Cs-promotion mirror those of Rb-promotion but they were less pronounced. Light-off measurements in the temperature range of 150–500 °C and under stoichiometric conditions (1000 ppm NO + 1000 ppm CO) showed that the NO conversion performance of alkali-free Pt is very poor (never exceeding 60%), whereas the 0.5wt%Pt(Rb)/ γ -Al₂O₃ catalyst

promoted with 9.7wt% Rb (optimally promoted catalyst), exhibited 100% NO and CO conversions with 100% N₂-selectivity at ~350 °C. In addition, 115 °C decrease on the NO light-off temperature (T₅₀) was obtained for K and Cs, while a larger decrease in T₅₀ by ~150 °C was obtained by Rb-promotion (Figure 11).



Figure 10. (a) Effect of Rb or K content of Pt/γ -Al₂O₃ catalyst on the turnover (TOF) formation rates of N₂, CO₂ and N₂O and on N₂-selectivity at T = 352 °C, [NO] = 1%, [CO] = 0.5%. (b) Corresponding enhancement ratio ρ = TOF/TOF⁰ values. Solid lines and symbols refer to Rb-promotion, dashed lines and open symbols refer to K-promotion. (Reprinted with permission from Ref. [53]; Copyright 2001, Elsevier).



Alkali (K, Rb or Cs) content, mmol/g_{cat}

Figure 11. Effect of K-, Rb- or Cs-promotion on the NO light-off temperature (T_{50}) for Pt/NO + CO system. Conditions: [NO] = 1000 ppm, [CO] = 1000 ppm, total gas flow rate $F_t = 80 \text{ cm}^3\text{STP/min}$; catalyst weight 8(±0.2) mg. (w/F_t = 6×10⁻³ g s/cm³). (Reprinted with permission from Ref. [53]; Copyright 2001, Elsevier).

The observed promotional effects were ascribed to alkali-induced changes in the chemisorption bond strengths of CO, NO and NO dissociation products. The superior promotion of Rb compared to that of K was attributed to its larger size and thus to the greater electric field experienced by co-adsorbed species located at an adjacent site. On the other hand, the unexpected lower efficiency of Cs promotion compared to that of Rb was attributed to a change in the adsorption site of NO due to the very large Cs ion [53].

More recently Konsolakis and Yentekakis [54] reported on the promotional effect of Na on the Pd-catalysed NO + CO reaction (CCP). A series of Na-promoted Pd catalysts were prepared on yttria-stabilized zirconia (YSZ; 10mol% Y₂O₃/ZrO₂) carrier (i.e., Pd(Na)/YSZ), with different loadings of alkali-promoter. Both promoting and poisoning effects, regarding the Na-loading, were found in a wide range of CO/NO ratios. For example at equimolar CO/NO composition ([CO] = [NO] = 1%) and T = 352 °C, the optimal promoter loading was found to be between 0.02–0.03wt% of Na (corresponding to a nominal sodium coverage of $\theta_{Na} \sim 3\%$; Equation (4) with $d_{Na} = 3.314 \times 10^{19}$ atoms/m²) which leads to N₂ and CO₂ formation rate enhancements by up to ~200%, accompanied by a significant improvement in N₂-selectivity of about 20% in comparison to the un-promoted Pd/YSZ catalyst (Table 3). The promotional effects were interpreted in a similar manner to the aforementioned studies [17,51–53].

Tanikawa and Egawa [55] studied the effect of Ba-promoter on NO + CO reaction in the presence of O₂ over Pd dispersed on γ -Al₂O₃ and ceria-zirconia supports (CCP). They found that the activity of Pd/ γ -Al₂O₃ was gradually enhanced by increasing the Ba loading up to 15wt%; the light-off temperatures (T₅₀) for CO oxidation and NO reduction were decreased by 40 and 33 °C, respectively, over the optimally modified Pd(15%Ba)/ γ -Al₂O₃. However, Ba addition on Pd/CZ catalyst decreased its activity for both NO reduction and CO oxidation reactions (Table 3).

Finally, Lepage et al. [56] studied the effect of electropositive promoters (monovalent H⁺, Na⁺, Rb⁺ and Cs⁺ or divalent Mg²⁺, Ca²⁺, Sr²⁺ and Ba²⁺ cations) over Rh nanoparticles supported on a series of Y zeolites for the NO reduction by CO by means of infrared spectroscopy of NO adsorption to monitor the electronic modifications of Rh nanoparticles. The observed decrease in NO reduction ignition temperature (T₁₀) with increasing ionic radius-to-charge ratio was related to the degree of electron transfer from Rh particles to the adsorbed NO molecules. The shift of the IR band of linearly adsorbed NO to lower wavenumbers with increasing the radius/charge ratio of promoter elements substantiated this argument.

A comparative overview of the literature involved in this section and in Table 3 gives prominence to the following general features: (i) Promotion either induced by EPOC or by CCP is substantial and has qualitatively similar characteristics, although CCP leads to much larger rate enhancements; ρ values in the region of 45–110 and around 13 were obtained by using CCP and EPOC, respectively. The ultimate contact and thus the higher interaction of the well dispersed active phase particles with the promoter species may be the cause of the mentioned differences on promotion intensity. (ii) Heavier alkalis were typically more effective than lighter alkalis, with Rb being the best. (iii) Among Pt, Pd and Rh precious metals, alkali promotion was significantly more effective on Pt than on Rh and Pd. (iv) Alkali-induced enhancement of the dissociative chemisorption of NO was considered as a key factor for the interpretation of the aforementioned pronounced promotion.

3.1.4. NO Reduction by Hydrocarbons or H₂

(i) NO reduction by alkenes

NO reduction by hydrocarbons under in situ electrochemical promotion (EPOC) by alkalis was firstly demonstrated by Harkness and Lambert [57] and Yentekakis et al. [58]. Using ethylene as reductant, Harkness and Lambert [57] showed that the rate of the NO + ethylene reaction over a Pt film interfaced with Na⁺-conducting β "Al₂O₃, linearly increased with θ_{Na} al low and intermediate sodium coverages (typically $\theta_{Na} < 0.3$), achieving a plateau for higher values. Rate enhancement values ρ up to 15 were obtained (Table 4).

On the other hand Yentekakis et al. [58] showed very pronounced NO reduction rate enhancements (using propene as the reductant) for the Pt-catalysed NO+propene reaction upon Na promotion over similar Pt/(Na) β "Al₂O₃ electrocatalyst. This effect was studied at 375 °C in a wide range of P_{NO} (0–7 kPa) and $P_{C_3H_6}$ (0–0.4 kPa) partial pressures. The promotional effect of Na was found to be significant in the whole range of conditions studied. The coverage of the promoting species was determined via galvanostatic transients. These transients showed that the promotional effects maximized at a Na coverage of θ_{Na} ~0.4, for which value the maximum enhancements in CO₂ and N₂ rates were ~7 times larger than the rate of the unpromoted Pt, while the N2-selectivity increased from 60% on the clean Pt surface to 80% on the optimal Na-promoted Pt surface (Table 4). Higher sodium coverages ($\theta_{Na} > 1$) caused a gradual decrease of the promoting phenomena. This suggests that nominal sodium coverages greater than one exhibit activity much higher than clean Pt because a substantial amount of Na is present as 3D crystallites, as demonstrated by XPS. A Na-induced strengthening of the NO chemisorpting bond relative to propene was indicated by the kinetic data. The accompanied weakening of the N-O bond and the concomitant extended NO dissociation were considered for the interpretation of the observed promotional effects. XPS data confirmed that under reaction conditions the promoter phase consists of a mixture of NaNO₂ and NaNO₃ 3D crystallites. At high Na loadings these surface promoter compounds inhibit reaction due to active sites blocking phenomena.

EPOC was also used by Williams et al. [59] for the study of NO reduction by propene over Rh film catalyst interfaced with (Na) β "Al₂O₃. They found that Na pumping onto Rh surface at a coverage

as low as ~ 2% causes a very large increase in the N₂-selectivity from 45% on the Na-free Rh surface to 82% on the Na-promoted surface together with beneficial effects on the reaction rate: rate enhancement ratio values of ρ = 2.4, 1.7 and 0.4 were recorded for the N₂, CO₂ and N₂O productions, respectively (Table 4). In a subsequent publication [60] the authors examined the impact of O₂ co-feed. It was found that upon increasing [O₂], the promotional effects of Na were progressively attenuated and finally reversed to poisoning for [O₂] = 2%. Similar considerations as in Reference [57] and [58] were used for the interpretation of the promotional effects, as well.

Using CCP, the group of Yentekakis [14,54] showed that the NO + propene reaction over Pd dispersed on 8mol% Y₂O₃ stabilized ZrO₂ support (YSZ) can be substantially promoted by dosing the catalyst with alkalis. In Reference [14], the rate increased by an order of magnitude accompanied by selectivity increases from ~75% over the Na-free Pd/YSZ catalyst to >95% over the optimally Na-promoted Pd/YSZ catalyst in the temperature interval of 250–450 °C. The promotional effects were found to be optimized at a sodium content of 0.06wt% (corresponding to nominal sodium coverage of ~7%; Equation (4)), while over-promotion attenuated the rate (volcano type behaviour). These promotional effects caused a decrease in the NO light-off temperature (T_{50}) of about 110 °C between the optimally promoted and unpromoted Pd catalyst (Table 4). A strengthening of the chemisorption bond of NO relative to propene that facilitates weakening of the N-O bond of the adsorbed NO molecules and thus their dissociative chemisorption was considered as the origin of the promotional effects. In ref. [54] the effects of Li, Na, K and Cs alkalis on Pd-catalysed NO + C_3H_6 reaction were comparatively studied: Four catalysts dosed with the same alkali/Pd = 1/1.5 atom/atom loading were prepared and their behaviour was compared in a wide range of [NO] at constant $[C_3H_6] = 0.8\%$ and T=380 °C. All these catalyst were found significantly better in terms of activity and selectivity in comparison to the unpromoted Pd/YSZ, in the whole range of conditions investigated; Na provided the most significant promotional effects (Table 4).

Systematically expanding their studies on conventional promoted catalysts (CCP), Yentekakis and co-workers showed that Pt nanoparticles dispersed on γ -Al₂O₃ are subjected to extraordinarily effective promotion by alkalis (Li, Na, K, Rb and Cs) [15,61] and alkaline earths (Ba) [62,63] in both activity and N₂-selectivity during the NO+propene reaction in the temperature interval of TWCs interest (ca. 200–550 °C). In References [15,61] rate increases by two orders of magnitude were achieved, while the selectivity towards N₂ was improved from ~20% over the alkali-free unpromoted Pt catalyst, to >95% over the optimally alkali-promoted catalysts. Best promotional effects were achieved by Rb-promotion, for which rate increases as high as 420-, 280- and 25-fold were obtained for the formation rates of N₂, CO₂ and N₂O, respectively, in comparison to the performance of the unpromoted Pt/ γ -Al₂O₃ catalyst (Table 4). Volcano type behaviour was found to be followed upon increases of promoter loadings for all cases. Thus, promotion was found to be maximized for a Li content of 1.25wt%, a K content of 7.1wt%, a Rb content of 9.7 wt% a Cs content of 15wt% (Figure 12) [61] and a Na content of 4.18wt% [15].

This very spectacular promotion in light-off diagrams (see for example Figure 13) showed that the marginal (~10%) NO conversion efficiency of un-promoted Pt/ γ -Al₂O₃ can be readily proved by alkalis to values up to 100% in the temperature window of TWCs interest, offering at the same time selectivities towards N₂ >95% [15,61]. However, besides temperature, NO_x removal efficiency of automotive catalytic converters is greatly affected by the presence of oxygen in the exhausts gases, in particular under excess O₂ conditions. Therefore, the NO + C₃H₆ + excess O₂ reaction is of higher practical significance than the NO + C₃H₆ one. Hence, it was decided to analyse this reaction and the effect of O₂ in a distinguishable section of the article (Section 3.2).



Figure 12. The effect of Li (a), K (b), Rb (c) and Cs (d) content of 0.5wt%Pt/ γ -Al₂O₃ catalyst on the turnover (TOF) formation rates of N₂, CO₂ and N₂O and on N₂-selectivity. Conditions: T = 375 °C, [NO] = 1.3%, [C₃H₆] = 0.3% (Reprinted with permission from Ref. [61]; Copyright 2001, Elsevier).

In the case of Ba-promotion the optimal Ba-loading was found to be 15.2wt% Ba (Table 4) [62]. Using a 1000 ppm NO + 1000 ppm C_3H_6 feed composition (these concentrations are close to those encountered in practical applications and commonly used in the literature as referred conditions) the authors showed that the very poor NO light-off behaviour of Ba-free Pt/ γ -Al₂O₃ catalysts, which gave a lower than 10% NO conversion in the whole 200–550 °C temperature interval, can be dramatically improved (in a similar manner to alkali-promotion [15,61]), achieving 100% NO conversion at relatively low temperatures accompanied by N₂-selectivities near to 100%. Further studying the Ba-promoted Pt/γ -Al₂O₃ system, the authors focussed on the effect of residual chloride, which concerns catalysts prepared from chlorine-containing precursors [63]; the detrimental effect of residual chloride on deNOx catalysis is well known. Very strikingly, it was found that when Ba is used to promote catalysts prepared from chloride-containing precursors, all the original chloride is retained, yet the catalysts exhibit very strong promotion and in every aspect their behaviours are identical to that of the chloride-free Ba-promoted catalyst [63]. This significant from the practical point of view attenuation of the inhibiting effect of chloride was rationalized in terms of formation of a stable 2D BaCl₂ phase on the catalysts surface, where the promoting effect of Ba overwhelms the poisoning effect of Cl and a large net promotion is resulted.



Figure 13. The conversion of NO (**a**) and the corresponding selectivity towards N₂ (**b**) for Na-promoted 0.5wt%Pt/ γ -Al₂O₃ catalysts, as a function of temperature at constant reactor inlet conditions: [NO] = 1000 ppm, [C₃H₆] = 1000 ppm; total flow rate F_t = 80 cm³STP/min; catalyst weight 8(±0.2) mg. (Reprinted with permission from Ref. [15]; Copyright 1999, Elsevier).

Macleod et al. [64] studied the NO + C_3H_6 reaction over dispersed Rh/ γ -Al₂O₃ catalysts (CCP). Rh is one of the noble metals used in TWCs and its use is related to the effective NO reduction in the three-way catalytic chemistry. They found that NO reduction chemistry on Rh can also be improved by alkalis. For instance, at 375 °C, a 7.3wt% Na content in the catalyst was found to be optimal promoter loading, leading to a ~3-fold increase in Rh activity (Table 4). Significant increases in N₂-selectivity, up to 90% were also obtained, when the unpromoted Rh/ γ -Al₂O₃ appeared to have only 53%. In addition, Na was found to suppress de formation of CO and HCN. Higher Na loadings were poisoning the catalyst bellow to its initial Na-free performance. However, it is worth noticing the substantially lower degree of promotional effect of alkalis on the NO reduction chemistry of Rh compared to that observed for Pt and Pd.

(ii) NO reduction by alkanes

Yentekakis and co-workers [65,66] studied the reduction of NO by methane over Na-dosed Pd catalyst (CCP) dispersed on yttria stabilized zirconia (YSZ) at temperatures between 350–500 °C and a wide range of reactants concentrations ([NO] = 0-1.5%, [CH₄] = 0-25%). It was observed that the reaction clearly exhibits Langmuir-Hinshelwood type kinetics over Pd with characteristic rate maxima reflecting competitive adsorption of NO and methane; NO adsorption however was found to be much more pronounced than that of CH₄ within the temperature range investigated. Sodium was found to cause strong poisoning of the reaction at any loading due to a Na-induced oxygen poisoning of the catalyst (derived by NO decomposition) (Table 4). The very different response of alkali promotion for the NO reduction by alkanes (strong poisoning effects) and by alkenes (strong promotional effects) is attributed to the different trends of the adsorption of the two kinds of hydrocarbons (alkanes and alkenes) on the Pt-group metal surfaces [67,68].

It is worth noting the similarities on the effect of electropositive promotion of Pt-group metals (PGM) between the alkanes or alkenes oxidation by dioxygen (Section 3.1.2) and by NO (Section 3.1.4): In both cases when the oxidation process involves alkenes, electropositive promoters induce strong promotion, whilst when the reductant agent is an alkane, electropositive promoters lead to poisoning. This is understandable in view of the different trend of alkenes and alkanes adsorption on the PGM surfaces (strong and weak, respectively [67,68]) and the influence of the promotion on it [15,20,47,61,65,66] and the fact that in both cases the oxidant species are mainly based on adsorbed atomic oxygen derived by the dissociative chemisorption of molecular oxygen or NO [47].

(iii) NO reduction by H₂ or CO+H₂

Besides hydrocarbons, hydrogen is also one of the gases present in automotive exhaust stream. It could also be externally supplied, by the use of a hydrocarbon reformer or another source of safe in situ hydrogen generator, for the reduction of NO_x from stationary power sources and chemical plants. Notably, the NH₃-Selective Catalytic Reduction process (NH₃-SCR) which is currently on of the most performing technologies for deNO_x processes in stationary power applications and chemical plants, experiences drawbacks related to the storage and slip of ammonia, although it may be pointed out that urea (as an NH₃ carrier) is now commercially available for use on diesel truck and some lighter vehicles. These issues can explain the growing interest and efforts for the use of H₂ as a reducing agent of NO_x in the last years (e.g., [69–72]). Most of these studies have used Pt-group metals as catalysts.

The first EPOC work in this field was performed by Marina et al. [19], who studied the effect of Na on the Pt-catalysed NO reduction by H₂ with (Na) β "Al₂O₃ as a source of Na promoter species. Rate enhancement ratio values up to $\rho_{N_2} = 30$ were achieved for the formation of N₂, accompanied by significant increases in N₂-selectivity. At 375 °C, Na coverage of only 6% increased the N₂-selectivity from ~36% on the clean Na-free Pt surface to ~75% on the Na-promoted Pt surface (Table 4). The promotional effects of Na were applicable in a wide range of H₂/NO gas phase compositions (ca. 0.1–3.0) and temperature (ca. 300–450 °C).

Burch and Coleman [73] studied the effect of sodium for the selective catalytic reduction of NO by hydrogen in the presence of excess oxygen (lean-burn conditions) over Na₂O-modified conventional, highly dispersed Pt/Al₂O₃ and Pt/SiO₂ catalysts (CCP), at temperatures representative of automotive "cold-start" conditions (<200 °C). They found that small sodium loadings significantly increased the NO conversion rate while larger sodium loadings led to severe poisoning. Under oxygen excess and low temperatures (<~140 °C) the selectivity to N₂ was practically unaffected by the addition of sodium, while an adverse effect was found at higher temperatures (Table 4).

Machida et al. [74] investigated the effect of alkalis (Na, K, Cs) and alkaline earths (Mg, Ca, Ba) on Pt/ZSM5 catalysts (CCP) for the selective reduction of NO by H₂ under O₂ excess (0.08% NO, 0.08–0.56% H₂, 10% O₂). The optimum promotional effects were obtained with Na promoter at loadings of 10–15wt%, which exhibited NO_x conversion higher than 90% and N₂ selectivity of ~50%. Based on in situ Diffuse Reflectance Infrared Fourier Transform Spectroscopy (DRIFTS) studies,

it was suggested that Na promotes the adsorption of NO as NO₂-type species, which would be an intermediate towards N_2 formation.

The H₂-SCR process over Pt for automotive emissions control, however, exhibits an important disadvantage: CO, which also coexists in exhaust stream, is strongly adsorbed on Pt surface causing inhibition of NO_x reduction. Pioneer CCP work by Lambert and co-workers [75–78] showed that Pd overcomes this poisoning effect of CO leading to good NO_x conversions at mixed H₂ + CO feeds. In addition, they found that Pd dispersed on 10wt% TiO₂-modified Al₂O₃ offers even more improved deNO_x efficiency in comparison to the Pd/Al₂O₃ catalyst. The effective temperature window of these new catalyst formulations was however relatively narrow (ca. 150–200 °C). In a similar manner, Yentekakis and co-workers [79] studied the lean NO_x reduction by H₂ + CO over potassium-modified Pd/Al₂O₃ catalyst for H₂-SCR process leading to significant enhancements of both NO_x conversion activity and N₂-selectivity of Pd. A potassium loading of ~0.25 wt% was found to optimize the catalyst performance in both activity and selectivity, leading to a catalyst formulation that offered up to ~90% NO_x conversion with a N₂-selectivity of the order of >85% for the temperature range of 180–300 °C (Table 4).

(iv) Direct NO decomposition

Limited studies can be found on this topic, probably due to the necessity of high temperatures for the direct NO decomposition. Wang et al. [80] studied the direct decomposition of NO into nitrogen and oxygen over Na-modified Pd/Al₂O₃ catalysts (CCP) in a flow reactor. They found that the NO reduction activity of Pd was significantly promoted by pre-coating the Al₂O₃ support with NaOH before the impregnation with Pd. A loading of 13wt% NaOH on a 4.8wt% Pd/Al₂O₃ catalyst enhanced the nitrogen yield from 15% up to 60% at T = 627 °C (Table 4). The observed NO decomposition rate enhancement was attributed to the strengthening of the NO adsorption bond on Pd sites.

Overviewing the literature included in Section 3.1.4 and Table 4 the following general remarks can emerge:

- Alkanes oxidation, either using dioxygen (Section 3.1.2) or NO (this section) behaves similarly
 under electropositive promotion by alkalis: addition of alkali on PGM surfaces causes only
 poisoning. Interpretations given are similar in both cases, based on the low adsorption propensity
 of alkanes on PGM surfaces and on the O-poisoning of the surfaces resulted by the alkali-induced
 enhancement of NO dissociative chemisorption.
- On the opposite, electropositive promotion of the NO + alkenes reactions is very pronounced; apparently it is the greater observed, for emissions control catalysis model reactions reviewed in the present work, offering ρ values as high as $\rho_{N_2} = 420$ and $\rho_{CO_2} = 280$.
- Some studies performing one-by-one comparison of EPOC and CCP by means of Na promoted Pt/NO + C₃H₆ system have demonstrated the very close similarities on the promotion output characteristics of the two methods and have convincingly argued for the similar origin of promotion in both cases.
- Once again (see Section 3.1.3) the intensity of CCP was superior to that of EPOC, indicating that
 an intimate contact between active phase and promoter particles succeeded by nanostructured
 catalysts seems to play a significant role on the promotion intensity.
- Heavier alkalis were again (see Section 3.1.3) more effective than the lighter ones; Rb was the best promoter.
- Among Pt, Pd and Rh, alkali promotion was more effective on Pt, lower on Pd and even lower on Rh. As we shall see (Section 3.3) this is related to the initial, unpromoted propensity of each metal on the dissociative adsorption of NO.
- Promotion by alkalis was also found to be substantial on the NO reduction by H₂ as well as on the direct NO decomposition reaction, although only few studies appeared in the open literature so far.
- Catalytic activity promotion was always accompanied by substantial increases on the system selectivity towards N₂.

Ref.	[57]	[58]	[60]	
Promotion highlights and maximum achievements	 Significant promotion in a wide range of gas compositions (C₂H₄/NO = 1-6); in all compositions promotion was monotonically increased by increasing θ_{Na} up to 40%. Significant changes in apparent activation energy by varying θ_{Na}. <i>p</i>_{CO3} = 15 for θ_{Ma} = ~30% and [C₂H₄]=1.5%, [NO] = 0.65% at T = 402 °C 	 Significant volcano type promotion observed with rate increases by an order of magnitude; optimal promotion achieved at θ_{Na} ~30%. Significant changes in apparent activation energy upon varying θ_{Na} were found. Under reaction conditions, the promoter phase consists of a mixture of NaNO₂/NaNO₃, which at high Na - loadings cause poisoning. P_{N2} ~ ρ_{CO2} = 7, ρ_{N2}O = 3, ΔS_{N2} = +20%, (60% → 80%), for θ_{Na} = ~30%, [C3H₆] = 0.6%, [NO] = 1.3%, T = 375 °C 	 Na promotes the Rh-catalysed NO+propene reaction, especially its selectivity towards N₂. The impact of [NO]/[C₃H₆] ratio was studied: promotion was superior for [NO]/[C₃H₆] = 1%/1%. The impact of O₂ co-feed was also tested: upon increasing [O₂] the promotional effects of Na were progressively decreased, reversing to poisoning at [O₂] = 2%. Under promoted conditions the promoter phase consists of arbonate, nitrate or both. P_{N2} = 2.4, <i>P</i>_{CO2} = 1.7, <i>P</i>_{N2} o. 0.4, ΔS_{N2} = +37% (45% -+82%), for θ_{Na} = ~20%, [C₃H₆] = [NO] = 1%, T = 350 °C. 	
s Alkene)		<u> </u>	<u> </u>	
Reaction condition tion by alkenes (NO +	$T = 275-450 ^{\circ}C$ [NO] = 0.5-1%, [C2H] 1-3% $\theta_{Na} = 0-40\%$	$T = 330-400 \circ C$ $[NO] = 0-6\%, [C_3H_6 \\ 0-0.4\% \\ \theta_{Na} = 0-100\%$	T = 350 °C [NO] = 1-9% [C ₃ H ₆ 1% [O ₂] = 0-2%	
Promoter NO reduc	Na	Na	Na	
Catalyst and (Promotion method applied)	Pt film over (Na) β'' Al ₂ O ₃ solid electrolyte (EPOC)	Pt film over (Na) β'' Al ₂ O ₃ solid electrolyte (EPOC)	Rh film over (Na) β'' Al ₂ O ₃ solid electrolyte (EPOC)	
Reactants	NO, C ₂ H ₄ NO, C ₃ H ₆		NO, C ₃ H ₆	

Table 4. Electropositive promotion of PGMs-catalysed NO reduction by hydrocarbons (alkenes or alkanes), H_2 or $H_2 + CO$.

Table 4. Cont.

Ref.	[14] [54]	[15] [61] [21] [22]	
Promotion highlights and maximum achievements	 Rate increases by an order of magnitude were achieved. Promotion followed volcano type behaviour (optimal promotion at ~20% θ_{Na}). In respect to Li, Na, K, Cs alkalis studied Na was superior promoter. Significant variation of the apparent activation energy with Na loading: the promoted induced isokinetic effect was demonstrated. P_{N1} ~ P_{CO1} ~ 7, P_{N2} ~ 15, Δ_{N2} = +20% (75% ~95%), for θ_{Aa} = ~20% (NO] = 1%, (75% ~95%), for θ_{Aa} = ~20%, [NO] = 1%, [C₃H₆] = 0.8%, T = 380 °C. ΔT₅₀ = -110 °C at [C₃H₆] = [NO] = 0.8% with the optimally promoted catalyst, θ_{Na} = 20% (0.068 wt% Na). 	 Extraordinarily effective promotion, by up to two orders of magnitude, with all alkalis (the highest found so far in emissions control catalysis: 420-fold rate increases). Volcano type behaviour; the optimal alkali loading depends on its entity. Among Li, Na, K, Rb, Cs alkalis studied superior promotion was offered by Rb. PN₃ = 420, PCO₃ = 280, PN₃O = 25, ΔSN₃ > +75% (20% → 95%), with 9.7wt% Rb at [NO] = 1.3%, [C₃H₆] = 0.3%, T = 375 °C Optimal alkali loadings: Li: 1.25 wt%, Na: 4.16 wt%, K: 7.1 wt%, Rb, Cs: 15 wt%. 	
\mid	×	" L 80 4.	
Reaction conditions	T = $250-450 ^{\circ}C$ [NO] = $0-8\%$, [C ₃ H ₆] = $0-8\%$ Na loadings: $0-0.102 ^{o}wl$ (nominal $\theta_{Na} = 0-30\%$)	T = 200-500 °C [NO] = 0-7%, [C ₃ H ₆] = 0.3%, 0.6% Alkali loadings: Li: 0-4 wt%; Na: 0-10.4 wt%; K: 0-8. Rb: 0-15.5 wt%; Cs: 0-2 wt%.	
Promoter	Na, Li, K, Cs	Na, Li, K, Rb, Cs	
Catalyst and (Promotion method applied)	0.5 wt% Pd/YSZ (CCP)	0.5 wt% Pt/γ-Al ₂ O ₃ (CCP)	
Reactants	NO, C ₃ H ₆	NO, C ₃ H ₆	

Ref.	[63]		[65] [66]	
Promotion highlights and maximum achievements	Significant, volcano type promotion with rate increase by up to two orders of magnitude; optimal promoter loading at 9.7wt% Ba. Significant changes in apparent activation energy upon varying Ba loading. $\rho_{N_2} = 165, \rho_{CO_2} = 110, \rho_{N_2O} = 9, \Delta S_{N_2} = +49\%$ $(45\% \rightarrow 94\%)$ with 9.7wt% Ba at [NO] = 1.3\%, [C ₃ H ₆] = 0.3\%, T = 450 °C. $\Delta T_{50} = -124$ °C for NO conversion	 Significant, volcano type promotion -but not such impressive as in the case of Pt [15,61–63]; optimal promoter loading: 7.3wt% Na. The formation of C7 and HCN by-products was also suppressed by Na-promotion. PN₃ = 3.3, PCo. = 2.7, PN₃ o = 2.2, ASN₃ = +29% (43% - 72%) with 7.3wt% Na at [NO] = [C₃H₆] = 0.3%, T = 375 °C 		 Pd by itself is an effective catalyst for NO reduction by CH₄. Na induced only poisoning in the whole region of [NO] and [CH₄] concentrations studied; the weak adsorption of CH₄ compared to that of O₂ and the O-poisoning of Pd surface due to the strengthening of the Pd-O bond induced by Na, readily explains this behaviour. Only Na-induced poisoning on the system was obtained.
-		> > A	(ane)	> A
Reaction conditions	T = 200-500 °C [NO] = 0-5%, [C ₃ H ₆] = 0-2% Ba loadings: 0, 3.5, 9.7, 15.2, 22.3wt%	T = 200-550 °C [NO] = 0-1%, [C ₃ H ₆] = 0.1% Na loadings: 0, 1.8, 3.0, 7.3, 11.3 wt%	tion by alkanes (NO + All	$T = 350-450 \ ^{\circ}C$ $[NO] = 0-1.5\%, [CH_4] = 0-20\%,$ $Na loadings: 0, 0.017,$ $0.034, 0.068, 0.102 \ wt\%$
Promoter	ä	Na	NO reduc	Na
Catalyst and (Promotion method applied)	0.5 wt% Pt/γ-Al ₂ O ₃ (CCP)	0.5 wt% Rh/y-Al2O3 (CCP)		0.5 wt% Pd/YSZ (CCP)
Reactants	NO, C ₃ H ₆	NO, C ₃ H ₆		NO, CH4

Table 4. Cont.

Ref.	[19]	[23]	[74]
Promotion highlights and maximum achievements tence or presence of O ₂	Significant, volcano type, promotion effects of Na to NO+H ₂ /Pt catalytic system; optimal promoter coverage $\theta_{Na} \sim 6\%$ Significant variations in the apparent activation energy of the reaction. $\rho_{N_2} = 30$, $\rho_{N_2O} = 6$, $\Delta S_{N_2} = +40\%$ ($36 \rightarrow 75\%$) with $\theta_{Na} \sim 6\%$ at [NO] = 0.63\% [H ₂] = 0.3%, T = 375 °C	The study conducted at lean conditions and at temperatures representative of automotive "cold-start" conditions. Volcano type promotion, optimized at a Na ₂ O loading = 0.27 wt%. Addition of MoO ₃ as co-promoter enhances Pt performance for all the loadings tested (2.5, \rightarrow 90%) with 0.27 wt% NO ₃) $\Delta X_{NO} = 20\% (75\% \rightarrow 90\%)$ with 0.27 wt% Na ₂ O/Al ₂ O ₃ offered X _{NO} = 95% at 100 °C and a ΔT_{50} = -50 °C.	Optimum promotion achieved by Na. Optimal promoter loading: 10–15 wt% Na. NO _x conversion > 90% with a selectivity towards N ₂ ~50%.
he abs	> > A	> > A A	> A
Reaction conditions $D + H_2$, $NO + H_2 + CO$) in t	T = $300-450 \circ C$ [NO] = $0-1.6\%$, [H ₂] = $0-1.\%$ $\theta_{Na} = 0-12\%$	$T = 50-200 \ ^{\circ}C$ $[NO] = 0.1\%, [H_2] = 0.4\%,$ $[O_2] = 6\%$ $Na_2O-laodings: 0, 0.27, 1,$ $S and 10 \ wf%$	T = $40-110 \circ C$ [NO] = 0.08% , [H ₂] = $0.08-0.56\%$, [O ₂] = $10\% O_2$ Na loadings: $0, 3, 5, 10, 15$ and $20 \text{ we}\%$
Promoter and/or CO (N(Na	Na (Mo)	Na, K, Cs, Mg, Ca, Ba
Catalyst and (Promotion method applied) NO reduction by H ₂ :	Pt film on $(Na)\beta''Al_2O_3$ solid electrolyte (EPOC)	1 wt% Pt/y-Al ₂ O ₃ 1 wt% Pt/SiO ₂ (CCP)	1 wt% Pt/ZSM-5 (CCP)
Reactants	NO, H ₂	NO,H2, O2	NO, H ₂ , O ₂

Table 4. Cont.

Ref.	[62]		[80]
Promotion highlights and maximum achievements	 Synergy between support-mediated (modification of Al₂O₃ with TiO₂) and surface-induced (K-addition) promotions providing very pronounced <i>de</i>NOx performance. K-promotion over the doubly-promoted catalyst appeared volcano type behaviour with optimal K-loading=0.25wt%. 90% NOx conversion with >85% N₂-selectivity in a wide range of temperatures (180-300 °C) were avide range avide range of temperatures (180-300 °C) were avide range of		 ✓ Pre-coadding of γ-Al₂O₃ with NaOH causes significant promotion of the direct NO decomposition activity of Pd/γ-Al₂O₃ catalyst. ✓ Promotion was progressively increased with increasing promoter loading. ➤ ΔX_{NO} = 45% (15% → 60%) at T = 625 °C over the 13 wt% NaOH -promoted 4.8 wt% Pd/γ-Al₂O₃ catalyst.
Reaction conditions	T = $50-400 \circ C$ [NO] = 0.1% , [CO] = 25% , [H ₂] = 0.75% , [C2] = $6\% O_2$ loadings: 0, 0.1, 0.25, 0.5, 1.0 and 3.0 wt% K	NO decomposition	T = 550-900 °C [NO] = 4% in He a loadings (as NaOH): 0, 1.3, 4.3, 6.5 and 13 wt% NaOH
Promoter	У У	Direct	Za Za
Catalyst and (Promotion method applied)	0.5 wt% Pd/γ-Al2O3 0.5 wt% Pd/γ-Al2O3-TiO2 (CCP)		(0.6–10 wt%)Pd /y-Al2O3 (CCP)
Reactants	NO, H ₂ , CO, O2		0 N

Table 4. Cont.

3.1.5. N₂O Decomposition and/or Reduction

It is well known that nitrous oxide (N₂O) is absent in automotive engines exhausts gases, although this can be a by-product (in low concentrations) of the NO reduction reactions taking place in the after-treatment devices of cars (catalytic converters), primarily formed during the "cold start" and "intermediate temperature" periods. However, N₂O is a powerful greenhouse gas (global warming potential ~310) and also the dominant stratospheric ozone-depleting gas nowadays. These account for the currently devoted high attention on the N₂O emissions control. Besides mobile sources, other anthropogenic sources of N₂O emissions involve stationary fuels combustion processes, agricultural activities and industrial plants (e.g., adipic and nitric acid plants) where it is produced in fairly high concentrations; the latter rationalizes why a major portion of deN₂O studies concerns gas mixtures with relatively high N₂O concentrations (ca. 0.1–1 vol.%; Table 5), not actually mirror cars emissions.

(i) Direct N₂O decomposition

No studies of electrochemical catalyst promotion (EPOC) were found in literature regarding the direct N_2O decomposition. In the next paragraphs we summarize some examples of the CCP studies for this reaction.

The pronounced effect of electropositive promoters on the direct N₂O decomposition performance (de-N₂O) of noble metals supported catalysts was demonstrated by several groups [81–85] (Table 5). In particular, Haber et al. [81] investigated the effect of Li, Na, K and Cs on N₂O decomposition over Rh catalysts supported on alkali-modified Al₂O₃ carrier. They found that the deposition of alkali metals on Al₂O₃ support results in a notable increase of Rh dispersion, which improved de-N₂O efficiency. In almost all cases the optimum promotional effects were obtained at ~0.078 mol% alkali (Table 5).

Parres-Esclaped et al. [82] studied the N₂O decomposition over Rh catalysts supported either on bare γ -Al₂O₃ or on alkaline earth modified alumina (5 wt% Mg- and Sr-doped γ -Al₂O₃). Enhanced deN₂O performance was obtained with Rh/(Sr)Al₂O₃ catalysts, which achieved total N₂O conversion at 300 and 350 °C, in the absence and presence of O₂, respectively, compared to 375 and 450 °C over un-promoted catalysts (Table 5). However, Mg addition was not beneficial to deN₂O Rh/Al₂O₃ activity. The latter was attributed to the formation of magnesium aluminate, while the former was attributed to the enhanced Rh dispersion (determined by TEM) on account of the Sr-modifier.

Combined structural and surface promotion was applied by Konsolakis et al. [83] to facilitate the deN₂O performance of Pt/γ -Al₂O₃ catalyst. To this end, rare earth oxides (CeO₂, La₂O₃) were used as structural promoters of the γ -Al₂O₃ support and potassium (K) as surface promoter of Pt, providing a doubly-promoted Pt(K)/Al₂O₃-(CeO₂-La₂O₃) catalyst composite. The study was conducted on macro-structured, cordierite honeycomb monolith washcoated catalytic arrangements, suitable for practical implementations. The effect of K-promoter was studied in the range of 0-2wt% K loadings, while the impact of O₂, CO and H₂O in the feed stream was also investigated (Table 5). It was found that incorporation of 20wt% CeO₂-La₂O₃ in the γ -Al₂O₃ support (ACZ) results in dramatic enhancement of the deN₂O catalyst activity which is further improved by potassium addition. The promotional effects of K were found to monotonically increase upon increasing its loading on the structurally promoted Pt/ACZ catalyst in the range of 0-2 wt% studied. As a result, the final doubly-promoted catalyst with 2wt% K loading achieved 100% N₂O conversion at a temperature of only 440 °C; notably the bare Pt/γ -Al₂O₃ did not exceed ~35% N₂O conversion even at 600 °C (Figure 14a). In respect to the impact of CO, O₂, H₂O co-feed, the authors found the following deN₂O efficiency sequence in all promoted or non-promoted catalysts: $N_2O + CO >> N_2O + O_2 > N_2O + O_2 > N_2O + O_2 + CO$ (Figure 14b). Water co-feed was detrimental to deN₂O performance but in a partially reversible manner after its removal from the feed stream; the catalytic efficiency was totally restored by adding H_2 into $N_2O + H_2O$ feed.

Papista et al. [84] studied the promoting impact of potassium addition to Ir/γ -Al₂O₃ catalyst on N₂O decomposition under both deficient and O₂ excess conditions. Varying K loading in the range of 0–1wt%, the authors found poisoning effect on the deN₂O performance of iridium in the absence of oxygen in the feed. The opposite was true in the case of oxygen excess conditions; a pronounced promotion was recorded showing that K prohibits the oxygen-poisoning of the catalysts; the latter self-poisoning is a typical problem in deN₂O catalytic chemistry, since the removal of the N₂O decomposition-derived adsorbed oxygen on the catalyst surface is a key reaction step of the process. The alkali promotion found in the case of oxygen excess conditions follows volcano behaviour providing therefore an optimal promoter loading of 0.5 wt% K, that resulted in a significant decrease of T₅₀, equal to Δ T₅₀ = -80 °C (Table 5). The effect of alkali on the adsorption properties of Ir involving strengthening of the Ir–O bond of adsorbed O species with concomitant changes on the oxidation state of the metal was considered upon interpretation of the promotional and/or poisoning effects observed.

Finally, Goncalves and Figueiredo [85] studied the effect of potassium on the catalytic activity of Pt supported on activated carbon (ROX 0.8) for the simultaneous reduction of NO + N₂O mixtures. They found a synergistic effect between K and Pt which led to high activity at relatively low temperatures. For instance, at 350 °C, a very stable and close to 100% conversion of both reactants was recorded over a 5wt% K/0.1wt% Pt catalyst, in opposite to the monometallic 0.5wt% K or 0.1wt% Pt counterparts (Table 5). The latter showed significantly lower conversion activity of about 25% for NO and 90% and 0% for N₂O, respectively. The improved redox properties of the bimetallic K/Pt formulation was responsible for this synergy on enhancing both deNO and deN₂O reactions [85].



Figure 14. (a): N₂O conversion profiles as a function of temperature (light-off curves) for bare Pt/ γ -Al₂O₃, structurally promoted Pt/ACL and doubly-promoted Pt(K)/ACL catalysts under direct decomposition of N₂O; Feed: 0.1% N₂O, balance He; GHSV = 10,000 h⁻¹. (b): N₂O (and CO) conversion profiles obtained over the optimally doubly-promoted Pt(2wt%K)/ACL catalyst under direct N₂O decomposition or N₂O + CO reactions in the absence or presence of O₂, that is, under the feeds: 0.1% N₂O, balance He; 0.1% N₂O + 0.1% CO, balance He; 0.1% N₂O + 0.1% CO + 2% O₂, balance He: 0.1% N₂O + 0.1% CO, balance He: (83]; Copyright 2013, Elsevier).

(ii) N₂O reduction by hydrocarbons

Very recently, the effect of alkali modifiers on another very important reaction related both to the TWC chemistry and N₂O abatement, that is, the N₂O reduction by hydrocarbons, was extensively studied by Pekridis et al. [25,86]. In particular, the role of potassium (K) promoter on the surface and catalytic properties of Pd/Al₂O₃ (CCP) catalysts for N₂O reduction by alkanes (CH₄, C₃H₈) or alkenes (C₃H₆) was investigated [86]. Potassium strongly enhanced the N₂O reduction by propane or propene, resulting in notably lower N₂O light off temperatures (~100 °C) compared to the un-promoted catalyst (Table 5). However, a slight inhibition upon K-promotion was obtained when CH₄ was used

as a reducing agent, implying that the extent of K-promotion is strongly related with the nature of the reducing agent. A comprehensive surface characterization study, involving XPS, in situ DRIFT spectroscopy of CO adsorption and FTIR-pyridine adsorption, was undertaken to correlate the de-N₂O performance with the catalyst surface characteristics. The results revealed that the electronic, acidic and structural properties of Pd/Al₂O₃ catalyst can be substantially modified by potassium, which in turn affects the reactants chemisorption bonds and as a consequence the overall surface activity. Specifically, addition of K on Pd/Al₂O₃ decreases the adsorption strength of electron-donor adsorbates, such as C₃H₈ and C₃H₆, enhancing the N₂O adsorption through an electron transfer from metal sites to N₂O antibonding orbitals. Both factors act synergistically towards increasing active sites for N₂O and its dissociation products (O_{ads}), at the expense of weakly bonded CH₄, leading to poisoning by strongly bonded oxygen atoms.

In correlation with the above study, the EPOC concept (through a Pd-film electrocatalyst interfaced with $(K)\beta''Al_2O_3$ solid electrolyte) was also used as a probe technique for exploring in detail the response of Pd metal to alkali (K) promotion in the same reaction [25]. The obtained results matched very well the promotion characteristics of the previous CCP study that involved conventional highly dispersed catalysts, demonstrating experimentally again that electrochemical promotion is an efficient tool for a rapid investigation of the promoter effect on a catalytic system. The obtained information could be further used to design and optimize efficient and "smart" catalysts formulations.

The N₂O reduction by propene was also studied by de Lucas-Consuegra et al. [87,88], over Pt-based electro-catalysts, electrochemically promoted by potassium by the use of a (K) β "Al₂O₃ solid electrolyte. The authors found significant rate enhancement ratios (ρ_{max} ~10) at temperatures between 400–450 °C and a reaction mixture composition of N₂O/C₃H₆/O₂ = 1000 ppm/2000 ppm/2000 ppm (Table 5). Potassium coverages (estimated by appropriate galvanostatic transients) higher than $\theta_{\rm K}$ ~70% were necessary in order to achieve such large rate enhancements. They also found that these pronounced promotional effects are similar for a wide range of O₂ concentration in the reaction mixture (2000–7000 ppm), while at larger O₂ concentrations (>9000 ppm) the promoting phenomena were practically vanished. A very significant observation of this work was also the demonstration that potassium addition strongly decreased the inhibiting effect of water vapor on the Pt-catalysed N₂O/C₃H₆/O₂ catalytic system.

Overviewing the literature findings included in Section 3.1.5 and Table 5 the following general conclusions can emerge:

- Alkali addition on PGM catalysts typically promotes the direct N₂O decomposition reaction (in the absence or presence of O₂ in the feed stream) or its reduction by reducing agents.
- Among Pt, Rh and Ir catalysts, alkali-promotion of the N₂O decomposition reaction was found to be very effective and superior on Pt, less effective on Rh and detrimental (poisoning) on Ir; in the case of Ir, alkali-induced promotion was obtained only under excess oxygen conditions, which was attributed to changes on the oxidation state of the metal.
- Comparative EPOC and CCP studies of alkali promotion on the N₂O reduction by the use of several reducing agents (CH₄, C₃H₈ and C₃H₆) demonstrated the great similarities on the output promotional characteristics of the two methods.

Ref.	[81]	[82]	83]	[84]
Promotion highlights and maximum achievements	Alkali addition enhanced Rh dispersion and thus deN ₂ O catalytic activity. Best promoter was Cs. Volcano type promotion for all alkalis used; optimal promoter loading ~0.078wt% of alkali oxide. $\Delta T_{50} = -60 \degree C$ for Li and Na promotion; $\Delta T_{50} = -70 \degree C$ for K promotion; $\Delta T_{50} = -90 \degree C$ for Cs promotion	Significant promotion by Sr either in the absence or in the presence of O_2 . Small inhibition and marginal promotion by Mg in absence and presence of O_2 , respectively. $\Delta T_{50} = -45$ and $-50 ^{\circ}$ C in the absence and presence of O_2 , respectively, with a Sr-loading = 5wt%. $\Delta T_{50} = +10$ and $-15 ^{\circ}$ C in the absence and presence of O_2 , respectively, with a Mg-loading = 5wt%.	K was found to be an effective promoter of the Pt-catalysed N ₂ O decomposition under both reducing and oxidizing conditions. H ₂ O in the feed caused detrimental but to a great extent reversible (upon removal) influence on deN ₂ O efficiency over both <i>R</i> -free and <i>K</i> -promoted catalysts. Promotion was monotonically enhanced with increasing <i>K</i> -loading. $\Delta T_{100} = -60 ^{\circ}$ C under N ₂ O feed; $\Delta T_{100} = -50 ^{\circ}$ C under N ₂ O + O ₂ feed; $\Delta T_{100} = -45 ^{\circ}$ C under N ₂ O + O ₂ feed	K induces poisoning effects on the Ir-catalysed N ₂ O decomposition at O ₂ -deficient conditions. In opposite, K is an effective promoter under O ₂ excess conditions; volcano type promotion with an optimal K-loading equal to 0.5 wt% K. $\Delta T_{50} = -80 \degree$ C under [N ₂ O] = 0.1% + [O ₂] = 2% and 0.5 wt% K loading.
	> > A	> > A A	> > A	> > A
Reaction conditions	$T = 250-450 °C$ $[N_2O] = 1\%$ Alkali loadings (as alkali oxides): 0.033, 0.066, 0.078, 0.093 wt%	T = 250-450 °C [N_2O] = 0.1%; [N_2O] = 0.1% + [O_2] = 5%. Mg- or Sr-loading: 5 wt%	$T = 200-600 \circ C$ $[N_2O] = 0.1\%; [N_2O] = 0.1\%; [N_2O] = 0.1\%, [N_2O] = 0.1\%, [CO] = 1\%; [N_2O] = 0.1\% + [CO] = 1\%; [N_2O] = 0.1\% + [CO] = 1\%; [N_2O] = 0.1\% + [CO] = 1\%; [N_2O] = 0.0.5, 1.0 and 2.0 wt\%$	T = $300-600 \circ C$ [N ₂ O] = 0.1% ; [N ₂ O] = 0.1% ; [O ₂] = 2% . K-loadings: 0, 0.25, 0.5 and 1.0wt%
Promoter	Li, Na, K, Cs	Mg, Sr	Ж	К
Catalyst and (Promotion method applied)	0.1 wt %kh/y-Al2O3 (CCP)	0.5 wt% Rh/y-Al2O3 (CCP)	0.5 wt% Pt/Al2O3-(CeO2-La2O3) (CCP)	0.5 wt% Ir/γ-Al ₂ O ₃ (CCP)
Reactants	N ₂ O	N ₂ O, presence or absence of O ₂	N ₂ O, presence or absence of O ₂ , CO, H ₂ O	N ₂ O presence or absence of O ₂

Table 5. Electropositive promotion of PGMs-catalysed N_2O decomposition and/or reduction reactions.

Ref.	[85]	[25]	[87]
Promotion highlights and maximum achievements	K addition simultaneously promotes the conversions of NO and N ₂ O at relatively low temperatures. 100% conversion of both N ₂ O and NO over 5 wt% K/0.1 wt% Pt/Al ₂ O ₃ under [N ₂ O] = 0.05% + [NO] = 1% at T = 350 °C.	Both structural and electronic properties of Pd particles were modified by K addition. K addition strongly promotes N ₂ O reduction by both C ₃ H ₈ and C ₃ H ₆ . K addition slightly inhibits N ₂ O reduction by CH ₄ . CCP and EPOC induced K-promotion behaviours were found to match well each other. $\Delta T_{30} = +25 \circ C$ under [N ₂ O] = 0.15% + [CH ₄] = 0.6% and 9 wt% K loadings $\Delta T_{50} = -100 \circ C$ under [N ₂ O] = 0.15% + [C ₃ H ₈] = 0.2% and 9 wt% K loadings $\Delta T_{50} = -135 \circ C$ under [N ₂ O] = 0.15% + [C ₂ H ₆] = 0.2% and 7–9 wt% K-loadings.	Strong K-induced promotion was observed. This was progressively attenuated and finally reversed to poisoning upon increasing O ₂ in the feed stream. K-promotion prevented deN ₂ O activity of Pt from the poisoning effect of H ₂ O. $P_{\rm AO} = 38.5 \ \rho_{\rm C,H_2} = 1.2 \ {\rm under} [N_2O] = 0.1\% + P_{\rm AO} = (C_2 H_6] = 0.2\%, T = 450 \ ^{\circ}C, P_{\rm K} \sim 40\% \ (V_{\rm WR} = -1 \ V)$
	> A	> > >> A A A	> A
Reaction conditions	$T = 50-500 ^{\circ}C$ $[NO] = 0.1\%; [N_2O] = 0.05\%; \text{ and } [N_2O] = 0.05\%, + [NO] = 0.1\%$ $+ [NO] = 0.1\%$ K-loading: 3 and 5 wt%	$T=100-500 \circ C$ $[N_2O] = 0.15\%$ $[CH_4] = 0.6\%$ $[C_3H_8] = 0.2\%$ $[C_2H_6] = 0.2\%$ $([O_2] = 3\%)$ K-loadings: 0, 2.5, 4.5, 7.0, 9.0 wt%	$T = 180-580 \circ C$ $[N_2O] = 0.1\%, [C_2H_6] = 0.2\%,$ $[O_2] = 0.2-1.0\%,$ $\theta_K = 0-90\%$
Promoter	К	Ж	м
Catalyst and (Promotion method applied)	(0.1 and 0.5 wt%)Pt/(ROX 0.8) (CCP)	2 wt% Pd/Y-Al ₂ O ₃ ,and Pd film on (K) \u00c6'Al ₂ O ₃ solid electrolyte (CCP vs EPOC)	Pt film on (K)β" Al2O3 solid electrolyte (EPOC)
Reactants	N2O, NO	$N_2O + CH_4$, $N_2O + G_3H_8$, $N_2O + C_3H_6$, presence or absence of O_2	N2O, C3H6, O2

Table 5. Cont.

3.2. Electropositive Promotion of PGMs Operated Under Simulated Practical Conditions

3.2.1. Simulated TWC Conditions

Given that alkalis and alkaline earths were found to strongly promote reactions related with the three-way catalytic (TWC) chemistry, testing of these new catalyst formulations at conditions that simulate automotive exhaust conditions was a reasonable consequence. To this end, Konsolakis et al. [89] studied the promotional effects of Na on Pt/γ -Al₂O₃ catalysts operated under simulated exhaust conditions at the stoichiometric point (SP), that is, at the value equal to 1 of the stoichiometric number (S), calculated as:

$$S = (2[O_2] + [NO]) / ([CO] + 9[C_3H_6])$$
(8)

using a feed mixture consisting (v/v) 1000 ppm NO + 1067 ppm C_3H_6 + 7000 ppm CO + 7800 ppm O₂.

A series of Na-promoted Pt/γ -Al₂O₃ catalysts were tested under these conditions at temperatures from 200–500 °C typical of TWCs. Typical results of the promoter effect on the overall NO, CO and C₃H₆ conversions and selectivity obtained at 400 °C, are shown in Figure 15. The conversions of all three pollutants, NO, CO and C₃H₆, reach 100% at sodium contents > 4.2wt%. In addition, Na promoter causes a significant decrease in the light off temperature of ~100 °C, together with significant improvements (~75%) in N₂-selectivity (Table 6). The promotional effect was attributed to the modified relative adsorption strengths of reactants and reaction intermediates, as explained above in the previous sections [89].

In a similar motive, Macleod et al. [90,91] studied the Na promotion of Pd/γ -Al₂O₃ and Rh/γ -Al₂O₃ catalysts operated under simulated TWC conditions over a range of stoichiometries from fuel rich to fuel lean conditions (0.90 < S < 1.1). They observed significantly different responses of Na addition in the two cases (Table 6). Na addition to Pd/γ -Al₂O₃ significantly improves its overall performance: although the recorded promotional effects were not as high as those observed in Reference [89] for Na-promoted Pt/ γ -Al₂O₃, results were again very spectacular.

A 7wt% Na loading (optimal promoter loading for the conditions used) offered a ~50 °C lower light off temperature than the unpromoted Pd when tested under stoichiometric feed conditions (1000 ppm NO + 1067 ppm C_3H_6 + 7000 ppm CO + 7800 ppm O_2). Beneficial effects of sodium addition were also recorded under both fuel rich and fuel lean conditions (Table 6). In the former case, Na was found to prevent catalyst poisoning by carbon deposition therefore maintaining high NO conversions, while in the latter case Na reduced the formation of N₂O therefore leading to enhanced selectivities toward N₂. However, the authors found that Na addition to Rh has a detrimental effect causing severe activity poisoning and N₂-selectivity reduction over the greater part of the temperature interval of 200–400 °C investigated [92]. It was attributed to the enhanced oxygen adsorption on the Rh active sites at the expense of the hydrocarbon, whilst in the case of Pd the beneficial effects were considered as a consequence of Na-promoted NO dissociation, together with inhibition of self-poisoning due to excessive propene adsorption [91]. It is worth noting that NO dissociative chemisorption takes place spontaneously on Rh but not on Pd.

Tanaka et al. [92] have also reported the beneficial effects of the addition of a small amount of Na_2O to (1.67wt%Pt)(12.3wt%MoO_3)/SiO_2 catalysts investigated under simulated exhaust from automobile engine and three-way behaviour around the stoichiometric point (SP). They found that the active window of SP for the Na-promoted PtMoNa/SiO_2 catalyst is wider than that of the Na-free counterpart, that is, PtMo/SiO_2 and Pt/SiO_2 catalysts (Table 6). The observed improvements were attributed to the wider redox ratio window and to the depressed (by Mo and Na) oxidation of Pt, even under oxygen excess.



Figure 15. Testing Na-promoted Pt/ γ -Al₂O₃ catalysts at simulated exhaust conditions of conventional stoichiometric gasoline engines (T=400 °C, 1000 ppm NO, 1067 ppm CO and 7800 ppm O₂; F_t = 80 cm³STP/min; m_{cat} = 8 mg). Effect of Na loading on reactants conversions (**a**) and selectivity towards N₂ (**b**). (Reprinted with permission from Ref. [89]; Copyright 2000, Elsevier).

Shinjon et al. [93] have investigated the effect of an alkaline earth (Ba) modifier on the catalytic activity of Pt/γ -Al₂O₃ and Rh/γ -Al₂O₃ catalysts under simulated automotive exhaust conditions at the stoichiometric point (1% CO, 0.3% H₂, 0.1% C₃H₆, 0.1% NO, 0.75% O₂, 3% H₂O, 12% CO₂). They found that the TWC-performance of Pt/γ -Al₂O₃ catalysts was improved by Ba, offering ~30 °C lower light off temperatures compared to those for the un-promoted catalysts (Table 6). However, the overall

performance of Rh catalysts was deteriorated with Ba addition. It was attributed to the fact that Ba addition to Pt catalysts suppressed the strong hydrocarbon chemisorption, enhancing simultaneously the oxygen adsorption. On the other hand, Ba addition to Rh catalysts results in strong adsorption of oxygen species, suppressing hydrocarbon chemisorption and thus the reaction between them.

Kobayashi et al. [94] investigated the influence of Ba and Sr alkaline earth metals on a commercial Pd-only TWC catalyst (Pd/Al₂O₃-based TWC catalyst of N.E. ChemCat Corp.). Improved three-way catalytic performance was obtained regarding CO and NO_x conversions (Table 6). The enhanced basicity and the electron-donation ability of these basic elements were considered to play significant role in the improvement of TWC performance. A higher electron density around Pd^(II) was revealed by XPS measurements in the Ba-promoted Pd/Al₂O₃ catalyst in comparison with the Ba-free counterpart. The stabilization of PdO that may suppress palladium species sintering and the electron enrichment palladium particles that endows Pd to behave like Rh were concluded as the main promoting factors in this study [94].

A NO-free gas mixture of CO, C_3H_6 , H_2 , O_2 , CO_2 and H_2O , simulating two-stroke motorcycle emissions (Table 6), was considered by Lee and Chen [30] for the investigation of the impact of Na₂O and K₂O addition on a 0.4wt% Pt/ γ -Al₂O₃ catalyst operated at 150–450 °C under stoichiometric and oxygen-deficient conditions (by varying the oxygen to reductants stoichiometric number; S = 1, 0.31 and 0.17). Significant enhancements on the CO and C_3H_6 conversions were achieved by the addition of both alkalis under operation at the stoichiometric point (S = SP = 1); superior promotion was the one induced by potassium. Under O₂-deficient conditions alkalis addition enhanced CO conversion but not that of propene, which was reduced. The presence of water in the gas mixture was found to have a strong positive impact on catalyst performance, in particular on K-promoted catalyst and for CO conversion (Table 6).

Yentekakis and co-workers [95,96] developed monolithic type catalytic converters with a washcoat containing only one precious metal (Pt-only TWC), which was optimally promoted by Na. Catalytic performance tests of this novel TWC under simulated exhaust conditions in a wide temperature range (150–500 °C) exhibited similar performance and even better N₂ selectivity (~100% in the whole temperature range) compared to that of a commercial bimetallic (Pt/Rh)-TWC catalyst (Figure 16A, Table 6); notably, the aforementioned TWC had a 4.5-fold lower noble metal loading in comparison to the commercial one. These novel material formulation and design for TWCs is subjected to much lower production cost (use of only one noble metal at much lower loading, without the necessity of scarce Rh in their constitution) and cost-effective recycling.

Most recent tests of these novel TWCs at more severe temperature conditions (>800 °C) have also shown a significant resistance to deactivation (Figure 16B) [97,98], implying that the promoter phase (Na) does not practically escape from the catalyst composite even at very high temperatures, due to its stabilization via the formation of new β' and β'' sodium-alumina phases as verified by TEM and XRD studies [97,98]. These phases, in direct interaction with the active phase (Pt), then act as a spontaneous (thermal diffusion driven) Na promoter source during TWC operation causing permanent promotion [97,98].

Overviewing the literature findings presented in this section and in Table 6, one can readily conclude that electropositive promotion of PGM-catalysts under simulated automotive exhaust conditions follows, in general terms, the promotional characteristics corresponding to the simple (model) reactions of the complex reactions scheme taking place into the converter. However, other possible synergistic or competitive interactions between the reactants and reactions' intermediates on a densely populated surface with all these competitive species can be at work affecting the overall performance.



Figure 16. Three-way catalytic performance at 450 °C of "fresh" (**A**) and "aged" (**B**) TWC samples at simulated stoichiometric gasoline engines exhaust conditions (0.1% NO + 0.7% CO + 0.1067% C₃H₆ + 0.78% O₂, balanced with He at 1 bar; F_t = 3200 cm³ (STP)/min). (Reproduced with permission from Ref. [98]; Copyright 2011, Elsevier).

	Ref.	[68]	[90]	[92]
	Promotion highlights and achievements	Na promotes simultaneously the conversions of all three pollutants, CO, C ₃ H ₆ and CO and enhances the N ₂ -selectivity at all promoter loadings used. CO, C ₃ H ₆ and NO conversions reached 100%, 100% and > 91% at T = 400°C and for Na-loadings >4.18 why, the corresponding values on the unpromoted Pt/Y-Al ₂ O ₃ were only 50, 20 and 11%, ΔT_{50} (CO) = -79° °C, ΔT_{50} (C3H6) = -73° °C and ΔS_{N_2} = +81%.	The studies involve the Na impact on Pd and Rh operated at TWC conditions over a range of the stoichiometric number (5) values from fuel-rich to lean. Na promotion was beneficial under both oxidizing and reducing conditions for Pd catalysts. Optimal loadings $>3.5 \mathrm{wt}^{\circ}/\mathrm{s}$ Na: $\Delta T_{50} (\mathrm{CO}) = -48 ~\mathrm{C}, \Delta T_{50} (\mathrm{C3H6}) = -35 ~\mathrm{C}, \Delta T_{50} (\mathrm{NO}) = -38 ~\mathrm{C}$ (at SP = 1) and $\Delta \mathrm{SN}_{1} = +35\%$ (at S > 1.04; oxidizing conditions). In opposite to Pd, Na addition on Rh was detimental.	Under nearly stoichiometric conditions, a wider redox ratio window of NO _x conversion was obtained with PtMoNa7 SiO ₂ catalyst compared to Pt/SiO ₂ . Under oxidizing conditions PtMoNa7SiO ₂ catalyst resulted to a shift of all CO, C ₃ H ₆ and NO _x conversion profiles to higher temperatures. However, a wider temperature window with higher NOx
		× A A	> > A >	× ×
J	Conditions	T = 200-500 °C (at simulated TWC conditions) [NO] = 1000 ppm, [CO] = 7000 ppm, [C ₃ H ₆] = 1067 ppm, [O ₂] = 7800 ppm; (at $S = SP = 1$) w/F = 6 × 10 ⁻³ g·s/cm ³ Na loadings: 0, 0.52, 1.57, 4.18 and 10.4 wt% Na	$T = 100-340 \ ^{\circ}C \ (at simulated TWC conditions) \\ conditions) \\ [NO] = 1000 \ ppm, [CO] = 7000 \ ppm, [C_3H_6] = 1067 \ ppm, 6970 \le [O_2] \le 8630 \\ ppm \\ ppm \\ ppm \\ w/F = 15 \times 10^{-3} \ g_{-5} / cm^3 \\ Na \ loadings for Pd/Al_2O_3: 0, 1.8, 3.5, 7.0 \\ Na \ loadings for Rh/Al_2O_3: 0, 3.0 \ and 7.5 \\ Wf \ loadings for Rh/Al_2$	T = 100–600 °C (32) w/F = 9.91 × 10 ⁻³ g·s/cm ³ Na ₂ O loadings: 0.05, 0.1, 1.0 wt%. Simulated oxidizing feed, S = 9.91:4.3% O ₂ , 0.12% NO, 800ppm G ₃ H ₆ , 0.12% CO 400ppm H ₂ , 12.3% CO ₂ , 3% H ₂ O. Simulated exhaust feed, S-scan: 0.40-1.21%O ₂ , 0.12° 0.00, 490–62 ppm C ₃ H ₆ , 0.45–1.20% CO, 0.12° 0.50% H ₂ ,
	promoter	a Z	e Z	Na ₂ O
	Catalyst and (promotion method applied)	0.5wt%Pt/y-A1 ₂ O3 (CCP)	0.5 wt% Pd/y-Al ₂ O ₃ and 0.5 wt% Rh/y-Al ₂ O ₃ (CCP)	1.67 wt% Pt/12.3wt% MoO3-SiO2 (CCP)
	Reactants	NO, C ₃ H ₆ , CO, O ₂	NO, C ₃ H ₆ , CO, O ₂	NO, C ₃ H ₆ , CO, O ₂ , H ₂ , CO ₂ , H ₂ O

Table 6. Electropositive promotion of PGM-catalysts operated under simulated automotive exhaust conditions.

Ref.	[93]	[94]	[30]
Promotion highlights and achievements	 [∨] Ba addition significantly promotes Pt/γ-Al₂O₃ activity under simulated exhaust conditions. The opposite is true for Rh/γ-Al₂O₃. [∨] ΔT₃₀(No₃) = −30 °C over 46wt% Ba-promoted Pt/γ-Al₂O₃ [∨] ΔT₃₀(No₃) = +15 °C over 46wt% Ba-promoted Rh/γ-Al₂O₃ 	 The TWC performance of the Pd-based commercial catalyst was significantly improved by the addition of Ba, Sr or La. The performance improvements followed the order: Ba > Sr > La. Alkaline earths pronotion was related with the stabilization of PdO phase and its electron enrichment, which endows Pd to behave like Rh. 	 Significant promotion by both Na₂O and K₂O; superior that of K₂O. Under oxygen-deficient conditions (S = 0.17 and 0.31), H₂O addition enhances both CO and C₃H₆ conversions (in particular X_{CO} and over K-promoted catalyst) MT₅₀(CO) = -90 °C AT₅₀(C3H₆) = -110 °C at S = SP = 1 and absence of H₂O. ΔX_{CO} = 51 and 48 at S = 0.31, without and with H₂O, respectively. ΔX_{CO} = 22 and 48 at S = 0.17, without and with H₂O, respectively.
	Ba % B <	> > >	ke (1) = (1) (1) (1) (1) (1) (1) (1) (1) (1) (1)
Conditions	T = 100-400 °C (at simulated TWC conditions) [NO] = 1000 ppm, [C ₃ H ₆] = 334 ppr [CO] = 1%, [H ₂] = 0.3%, [CO ₂] = 12 ⁹ [H ₂ O] = 3%, [CO ₂] = 0.75% GHSV = 20000 h ⁻¹ Ba-loading: 5.8, 11.5, 23 and 46 wt%	T=300-600 °C Real stoichiometric automotive engi exhaust gas GHSV = 68000 h ⁻¹ Ba, Sr od La-loading: 5wt%	T = 150-450 °C (simulated two-strok motorcycle emissions) [CO] = 1-4.14%, [C ₃ H ₆] = 0.08-0.7% [CD] = 0.96-0.9%, [H ₂ J] = 0.2%, [CO2] 10%, [H ₂ O] = 0 or 10%. Stoichiometric numbers studied S = 0.31, 0.17 Na-loading: 4.5 wt%; K-loading: 7.6 w
promoter	Ba	Ba, Sr, La	Na, K
Catalyst and (promotion method applied)	1.67 wt% Pt/y-Al ₂ O ₃ 0.9 wt% Rh/y-Al ₂ O ₃ (CCP)	5wt% Pd-commercial TWC catalyst supplied by N.E. ChemCat Corp. (CCP)	0.4 wt% Pt/y-Al ₂ O ₃ (CCP)
Reactants	NO, C ₃ H ₆ , CO, O ₂ , H ₂ , CO ₂ , H ₂ O	Exhaust gas from an automotive engine	G ₃ H ₆ , CO, H ₂ , O ₂ , CO ₂ , H ₂ O (NO-free)

Table 6. Cont.

Ref.	[95] [96] [98]
Promotion highlights and achievements	The Na-promoted monometallic (Pt-only) TWC, with 4.5-fold less PGMs loading compared to bimetallic commercial one, exhibits similar performance with the commercial one, exhibits similar performance with the whole temperature range 150-500 °C). Marginal impact of H ₂ O and CO ₂ on the performance of Na/Pt-TWC. A synergy between surface promoter (Na) and structural promoters (CeO ₂ , La ₂ O) was resulted, leading to this attractive promotion. Significant resistant to deactivation, even upon severe aging (>800 °C). Na promotion remains at work even after aging due to the formation of <i>P</i> , <i>P</i> [®] solution alumina phases, which act as permanent spontaneous sources of Na promoter species during TWC operation.
\mid	· · · · · ·
Conditions	T = $150-500 \circ C$ (simulated TWC conditions) [NO] = 0.1% , [Co] = 0.7% , [C ₃ H ₄] = 0.1067% , [O ₂] = 0.7% , [C ₃ H ₂ O] = 10% , [CO ₂] = 10% [CO ₂] = 10% [CO ₂] = 10% CHSV = $5050 h^{-1}$ Na-loading: 1 and 2 wt% Na (i.e., 5 an 10 wt% in the washcoat)
promoter	e Z
Catalyst and (promotion method applied)	0.1 wt% Pt/(γ-Al ₂ O ₃ -CeO ₂ -La ₂ O ₃ -cordierite monolith) versus a Commercial 0.37Pt/0.08Rh-TWC(CCP)
Reactants	NO, C ₃ H ₆ , CO, O ₂ , CO ₂ , H ₂ O

Table 6. Cont.
3.2.2. Oxygen Excess Conditions (Simulated Lean-Burn and Diesel Exhausts Gases)

Some other EPOC studies regarding the electrochemical activation of Pt-based electrochemical catalysts for NO_x reduction by propene under oxygen excess were performed by the groups of Vernoux and de-Lucas Consuegra. In this sense, Vernoux et al. observed a significant promotional effect of Na promoters electrochemically supplied from a NASICON solid electrolyte support [99]. The study was carried out under lean-burn conditions (oxygen excess, typical of Diesel engines). In such a system, electrochemical promotion is shown to strongly enhance both the catalytic activity and the selectivity to N_2 (from 41% to 61%) using very low overpotentials (-100 mV) at 300 °C (Table 7).

In the same line, the group of de Lucas Consuegra performed several EPOC studies by using Na and K promoters by means of $(Na^+)\beta'Al_2O_3$ and $(K^+)\beta Al_2O_3$ solid electrolytes, respectively. First, they studied the effect of reaction temperature and O_2 concentration in the feed stream on Pt catalysts promoted with Na species. Under lean burn conditions, at low reaction temperatures (220 °C), rate enhancement ratios up to 1.4 were observed for the NO reduction rate (Table 7) [100,101]. Nevertheless, as the reaction temperature increased, the promotional effect decreased, even resulting in a transition to a regime where the alkali metal induced poisoning. This progressive suppression of the promotional effect was due to an increase of the oxygen coverage on the Pt surface with the temperature, which led to a C₃H₆ adsorption inhibition. However, at all explored reaction temperatures, the presence of sodium ions induced a large increase of the N₂ selectivity (up to 90%), minimizing the N₂O formation.

Regarding the effect of the O_2 concentration, a study was performed at low temperatures (240 °C) under O_2 concentrations ranging from 0.5% to 5%. The promotional effect of sodium on the overall catalytic activity for NO removal was progressively lowered with increasing oxygen concentrations, as a result of a strong inhibition of propene adsorption and a relative increase of the oxygen coverage. In all cases, the presence of sodium ions induced an increase in the nitrogen selectivity by promoting the NO adsorption and subsequent dissociation.

In a new investigation, de Lucas Consuegra et al. [102] studied for the first time the promotional effect of K (EPOC) under the presence of steam in the feed stream under lean burn conditions, as a new approach to the development of efficient catalysts under real working conditions of combustion engines. In this study, they demonstrated that the presence of K promoters could be practically used to decrease the inhibitory effect of water for the reduction of NO by propene. In addition, as previously demonstrated in other studies, they verified that the K species form stable nitrate species on the Pt surface. This discovery led to a new series of studies performed by the same research team, leading with the NO_x-storage and Reduction technology (NSR) by means of alkali-based Pt electrochemical catalysts.

Briefly, the NSR process was developed in the early 1990s by Toyota. It is based on the ability of alkali and alkali earth elements to store Nox in form of nitrites/nitrates under lean burn conditions (oxygen excess). Then, by the injection of extra fuel, the stored NO_x are released and reduced with hydrocarbons, CO or H₂, to produce N₂, CO₂ and H₂O. The alkali promoters function in two ways in these "smart" materials: first, they act as electronic promoters to enhance the NO oxidation to NO₂ (which has been identified as the rate determining step of the overall process) and to store the NO₂ produced in form of nitrites/nitrates (as previously demonstrated).

In these studies, de-Lucas Consuegra et al. demonstrated, for the very first time, the possibility to perform the NSR process via EPOC, by the use of Pt catalysts supported on K^+ - β Al₂O₃ solid electrolytes (to supply the K promoter species) [103,104]. The studies were performed under lean burn conditions and even under the presence of steam. The 2 main conclusions from these studies were:

- Potassium ions electrochemically transferred to the Pt catalyst play a double role in the NSR
 process, as a promoter for the NO oxidation reaction and as storage sites through the formation of
 potassium nitrates.
- The maximum yield of Pt/(K)βAl₂O₃ to effectively reduce NO_x to N₂ was obtained at T = 300 °C (Figure 17). This temperature is suitable for both, gasoline and diesel engines working under lean burn conditions.



Figure 17. Influence of the reaction temperature on the amount of NO_x stored, potassium transferred and on the NO_x conversion to N₂ during NO_x storage/reduction experiments. Lean phase (NO/C₃H₆/O₂:1000 ppm/1000 ppm/5% O₂), 6 min of duration, V_{cell} = -1.5 V. Rich phase (NO/C₃H₆/O₂:1000 ppm/1000 ppm/0.5% O₂), 5 min of duration, V_{cell} = 3V. Data was acquired in a Pt/(K)βAl₂O₃/Au galvanic cell (Reprinted with permission from Ref. [103]; Copyright 2011, Elsevier).

In respect to conventional highly dispersed catalysts and using CCP, early studies by Burch and Watling that involved the influence of a number of promoters, including alkalis (namely K and Cs) and alkaline earths (namely La, Mg, Ba), on the Pt-catalysed propene-SCR of NO_x under lean-burn conditions have showed that addition of one of the following: 2.42wt% K2O, 7.25wt% Cs2O, 7.93wt% BaO, 2.07wt% MgO and 8.35wt% La₂O₃ in 1wt%Pt/ γ -Al₂O₃ catalyst results in suppression of the catalytic activity of Pt (NO_x conversion efficiency) and to marginal effects on N_2/N_2O selectivity (Table 7) [105]. However, later studies of the de-Nox activity of supported Pt catalysts under lean burn conditions have shown very pronounced promotional effects by alkalis or alkaline earths: Vernoux et al. [106] and Yentekakis et al. [47], systematically investigated the resulted promotional effects of Na addition on supported Pt/γ -Al₂O₃ catalysts (CCP) during NO + propene + excess O₂ as a function of the Na-loading, have shown significant beneficial effects of the alkali on both de-Nox activity and N_2/N_2O selectivity at low and intermediate Na-loadings (Table 7). In both studies, the effects (promoting or poisoning, depending on promoter loading) were understandable in terms of the influence of the alkali promoter on the relative adsorption strengths of reactants and intermediates on a highly populated surface with electrophilic and electrophonic adsorbates. In particular, Yentekakis et al. [47] demonstrated that Na acts beneficially to de-Nox process in a narrow window of Na loading (ca. 0-2.6wt% Na), while overpromotion to the optimal value of ~2.6wt% Na causes significant inhibition of the catalyst performance. The recorded promotional effects can be summarized as follows: Na widened the temperature window of the $C_3H_6 + NO + O_2$ reaction towards lower temperatures by ~50 °C, accompanied by an enhancement in N2-selectivity by ~40 additional percentage points; the promotion is optimized at a sodium loading of 2.6wt% Na on the 0.5wt%Pt/ γ -Al₂O₃ catalyst (Figure 18). Obviously, the promoter loadings chosen by Burch and Watling were out from this optimal window, explaining the poisoning effects observed by the authors [105].

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Ref.	[66]	[100]	[102]	[103] [104]	
Promotion highlights and achievements	Na promotion induced strong enhancements on both activity and N ₂ -selectivity under lean-burn conditions at quite low T. $\rho_{N_2} \sim \rho_{CO_2} \sim 1.7, \rho_{N_2O} \sim 1.1, \Delta S_{N_2} = +20\%$ $(41\% \rightarrow 61\%)$ at T = 295 °C, $\Delta V_{WR} = -200 \text{ mV}.$	 Na significantly promotes the low temperature (T = 220 °C) lean-burn activity and N₂-selectivity of Pt impregnated catalyst on Na⁺ conducting solid electrolyte. At higher T (>240°) Na supply to the catalyst can lead even to poisoning of the activity but not of the selectivity which is increased (reaching -90%). Promotion : <i>P</i>_{N2} ~ <i>P</i>_{CO3} ~ 1.4, ΔS_{N2} = +20% (60 → 75%) at T = 220 °C, [O₂] = 5%, <i>P</i>_{N3} = 7-8% (60 → 75%) at T = 220 °C, [O₂] = 5%, <i>P</i>_{N3} = 7-8% (40 → 90%) at T = 300 °C and <i>P</i>_{Na} = 7-8% 	 K promotion depresses the inhibitory effect of H₂O on the lean NO_x reduction by C₃H₆. K promoter species form stable nitrates on Pt surface, thus can operate as a NO_x-Storage Reduction (NSR) species. 	 The electrochemically assisted NOx-Storage Reduction (NSR) process by a Pt/(K)βAl₂O₃/Pt electrochemical cell is demonstrated. Electrochemically transferred K play a double role: promoter of the lean NO_x reduction and NO_x-storage/release species. Maximum deNO_x to N₂ yield was obtained at T = 300 °C. 	
	5% st		5%,	~ ~ ~	
Conditions	$T = 200-400 \circ C$ [NO] = 0.2%, [C ₃ H ₆] = 0.2%, [C ₃] = 5 Na coverage: non estimated (cataly) potential V _{WR} is given)	$T = 220-300 \ ^{\circ}C$ [NO] = 0.2%, [C ₃ H ₆] = 0.2%, [O ₂] = 0.5, 1 and 5%. $\theta_{Na} = 0.8\%$	T = $180-400 \circ C$ [NO] = 0.2% , $[C_3H_6] = 0.2\%$, $[O_2] = 5$ θ_{K} : varying, non-estimated	$T = 250-370 \ ^{\circ}C$ [NO] = 0.1%, [C ₃ H ₆] = 0.1%, [O ₂] = 5 θ_{K} : varying, non-estimated	
Promoter	Na	Na	Ж	м	
Catalyst and (promotion method applied)	Pt film on NASICON solid electrolyte (EPOC)	Pt impregnated on (Na) <i>f</i> "A1 ₂ O ₃ solid electrolyte (EPOC)	Pt impregnated on (K)βAl ₂ O ₃ solid electrolyte (EPOC)	Pt film on (K)βAl ₂ O ₅ solid electrolyte (EPOC)	
reactants	NO, C ₃ H ₆ , O ₂ excess	NO, C ₃ H ₆ , O2 excess	NO, C ₃ H ₆ , O ₂ excess, H ₂ O	NO, C ₃ H ₆ , O ₂ excess	

 $\mathbf{Table 7}. \ Electropositive \ promotion \ of \ PGMs-catalysed \ lean-NO_X \ reduction \ by \ C_3H_6 \ (the \ so-called \ simulated \ lean-burn \ or \ diesel \ exhaust \ conditions).$

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Ref.	[105]	[106]	[47]	[107]	[108]
Promotion highlights and achievements	 Performance inhibition effects of alkaline materials addition were found at the specific loading used for each alkaline (2.42, 7.25, 7.93, 2.07 and 8.35wt% of K₂O, Cs₂O, BaO, MgO and La₂O₃, respectively). 	 Under C₃H₆/NO/O₂ reaction strong Na-induced promotion was obtained (optimal Na-loading 5wt%). In opposite, under C₃H₈/NO/O₂ reaction severe inhibition was induced by Na addition. Promotion : ΔT₃₀(C_{3H6}) = -39 °C under C₃H₆/NO/O₂ reaction on the 5 wt% Na – model and a modified catalyst, ΔS_{N2} ~ +20% for T > 280 °C. Poisoning: ΔT₅₀(C_{3H8}) = 191 °C under C₃H₈/NO/O₂ reaction on the 5wt% Na-modified catalyst. 	 ✓ Strong promotion by Na with an optimal loading of 2.6wt% Na. Higher Na-loadings caused strong inhibition (volcano behaviour; over-promotion). NO_x conversion window is widened and moved to lower temperatures. > ΔT_{S0}(_{C3H6}) – ΔT_{S0}(_{C0N3}) = −65 °C; ΔS_{N2} = +40% at T > 320 °C for 2.6wt% Na loading 	 Na-loadings > ~1 wt% enhanced the selectivity towards N2 in the whole temperature range (150–450 °C) but not affected the activity. Na-loadings > 3 wt% caused rate inhibition for all CO, NO_x, C₃H₆ conversions. 	 Marginal effects of K addition on the Ir surface at reducing (absence of O₂) conditions. Strong K-induced poisoning at oxidizing conditions for both propene oxidation and NO_x reduction with significant decreases in N₂-selectivity, as well.
	O_2] = 5% olar ratio	: C ₃ H ₆ or wt%	O ₂] = 5% L2 wt%	00 ppm, 0] = 10%,	2%,
Conditions	T = 200-500 °C [NO] = 0.1%, [C ₃ H ₆] = 0.1%, [C Alk-loadings: A/Pt = 10/1 mc	$T = 200-650 \text{ °C}$ $[NO] = 0.2\% [HC] = 0.2\% (HC] = 0.2\% (HC) \text{ C}_{3}H_{8} \text{ (O2)} = 5\% \text{ CHSV} = 20,000 \text{h}^{-1}$ Na-loadings: 0.12, 1 and 5	$T = 200-450 \ ^{\circ}C$ [NO] = 0.1%, [C ₃ H ₆] = 0.1%, [C Na-loadings: 0, 1.6, 2.6 and 4	T = 150-450 °C [NO] = 300 ppm, [C ₃ H ₆] = 18 [CO] = 450 ppm, [O ₂] = 8%, [H ₂ [CO ₂] = 10.7% Na-loadings: 0-10wt%	$\begin{split} T &= 250\text{-}400\ ^{\circ}\text{C} \\ \text{[NO]} &= 0.2\%, \text{ (C}_{3}\text{H}_{6}) = 0.1\\ \text{[O_{2}]} &= 0.5\% \\ \theta_{K} &= 0\text{-}100\% \end{split}$
Promoter	K, Cs, La, Mg, Ba	Na	Na	Na	м
Catalyst and (promotion method applied)	1 wt% Pt/ y-Al2O3 (CCP)	0.9 wt% Pt/y-Al2O3 (CCP)	0.5 wt% Pt/y-Al ₂ O ₃ (CCP)	Ir-black (CCP)	Ir film on (K) β'' Al ₂ O ₃ solid electrolyte (EPOC)
reactants	NO, C ₃ H ₆ , O ₂ excess	NO, C ₃ H ₆ or C ₃ H ₈ , O ₂ excess	NO, C ₃ H ₆ , O ₂ excess	NO, C ₃ H ₆ , CO, H ₂ O, CO ₂ , O ₂ excess	NO, C ₃ H ₆ , (O ₂)

Table 7. Cont.



Figure 18. $C_3H_6 + NO + O_2$ reaction on Na-promoted Pt/ γ -Al₂O₃ catalysts. The effect of temperature on the conversion of NO to various N-containing products and the corresponding N₂-selectivity: (**a**) NO conversion to N₂; (**b**) NO conversion to N₂O; (**c**) NO conversion to NO₂; (**d**) N₂-selectivity. Conditions: 1000 ppm NO, 1000 ppm C₃H₆, 5% O₂; reactant contact time 4 s. (Reprinted with permission from Ref. [47]; Copyright 2005, Elsevier).

The performance of another noble metal, iridium, the three-way catalytic chemistry of which has several promising characteristics, have been studied under Na-promotion by Wogerbauer et al. [107]. A simulated exhaust gas mixture consisted of 300 ppm NO, 0.18% propene, 450 ppm CO, 8% O₂, 10% H₂O, 10.7% CO₂, balance N₂ was used, while the temperature and Na promoter loading were varied between 150–450 °C and 0–10wt%, respectively. They found that under the conditions used and for Na loadings higher than ~3wt%, Na caused an inhibition on the Ir-black catalytic activity for CO, propene and NO_x conversions (Table 7). However, the selectivity towards N₂ was significantly improved up to 100% for samples with Na contend > ~1wt% in the whole temperature range.

More recently, Goula et al. studied the effect of electrochemically supplied potassium on a porous Ir-film catalyst under the NO + C_3H_6 + O_2 reaction at a variety of oxygen concentrations [108]. In this EPOC study the Ir-film was interfaced with a (K) β "Al₂O₃ solid electrolyte on which the catalyst was supported by spattering. It was found that K addition on the Ir surface is detrimental on catalyst performance (propene oxidation, NO reduction efficiency and N₂-selectivity) for all oxygen concentrations investigated. Only at zero O₂ concentration, that is,, at net reducing conditions of the reactive mixture, a slight promoting effect of potassium on the N₂-selectivity but not on the rates was observed (Table 7). This very different effect of alkali-promotion, compared to that observed on Pt

and Pd noble metals under similar conditions (on which strong promotional effects were recorded), was understandable in terms of the electronic influence of co-adsorbed potassium on the adsorption strengths of the neighbour reactants on the Ir surface: the excessive enhancement of oxygen adsorption on Ir sites in the expense of the hydrocarbon adsorption caused O-poisoning of the surface.

Overviewing Section 3.2.2 and the involved literature listed in Table 7, it is apparent that as we have seen in Section 3.1.4, alkali-promotion of the reduction of NO by propene remains very effective even under excess O_2 conditions, at concentrations > ~5% which simulate lean-burn and diesel engines exhaust gases. However, at excess oxygen conditions promotional effects on both activity and selectivity are significantly attenuated and also limited in quite narrows temperature and/or promoter loading windows. Volcano type behaviour of promotion is once again recorded upon increasing alkali loading and EPOC or CCP promotion characteristics were qualitatively similar.

3.3. Mechanistic Implications: The mode of Action of Electropositive Promoters

3.3.1. Main Promotion Characteristics and Mechanistic Implications

The above described results show that electropositive promoters (alkalis and alkaline earths) can markedly affect the catalytic properties of Pt-group metals during CO or hydrocarbons oxidation and NO reduction by CO or hydrocarbons in the absence or presence of oxygen, as well as of other environmentally important reactions (see Tables 1–7). The promotional benefits on the catalytic performance (activity, selectivity) achieved were similar, independently of the promotion method used, that is, the electrochemical promotion of catalysis (EPOC) or the conventional catalyst promotion one (CCP). The intensity of promotion, the optimal promoter loading and other characteristics of the induced promotional effects are subjected to the specific reaction/catalyst system and the individual reaction conditions applied (Tables 1 and 2).

This electropositive promotion (by alkalis and alkaline earths) of PGMs for emissions control catalysis reaction was found to be very pronounced not only when applied to model reaction systems (Tables 1–5) but also under complex reaction systems that mirror practical applications, that is, stoichiometric gasoline, lean-burn and diesel engines exhaust conditions (Tables 6 and 7). The latter boosts the practical importance of the subject and direct implementations on practical systems appear attractive and promising.

Model reactions studies and their individual characteristics (Section 3.1) enabled us to better understand the mechanism of the promoters' action. EPOC has also a significant role on this issue, since it gave us the opportunity of in situ adjustment of the promoter loading, recording at the same time its influence on the catalytic performance. Recall that some studies were explicitly focussed on experimentally establishing that EPOC and CCP are similar in origin and are subjected to the same physicochemical rules; the only practical different is the procedure used for the ptomoter supply [21,22,25]. To this end, Yentekakis and co-workers [21,22] have first shown that the promotional effects of sodium, electrochemically induced (via EPOC) on the NO + C_3H_6 /Pt catalytic system, mirror those obtained on a highly dispersed Pt/ γ -Al₂O₃ catalyst, promoted by means of CCP method at similar conditions. Therefore, EPOC and CCP will be hereafter considered as identical phenomena; the mechanistic model described below that interprets the mode of action of electropositive promoters, may be considered identical valid for both promotion methods. The vast majority of the publications reviewed here have been based on this model to consistently interpret individual promotional phenomena appeared at the specific reaction systems investigated. The main body of the model is described below and several direct spectroscopic evidences supporting its consistency are presented.

Using EPOC (via a galvanic cell as it was described by the formula (1) and Figure 1) one can electrochemically control (Faraday law) the migration of promoting species (e.g., Na⁺, Figure 19) from the solid electrolyte onto the catalyst/gas interface, where an effective and overall neutral double layer is formed through this ions back-spill over imposed electrochemically (Figure 19).

Actually, the electrocatalytic reaction taking place at the three-phase-boundary metal-solid electrolyte-gas that leads to the formation of the effective double layer is [6,7,16,109]:

$$Na^{+}(from \beta''Al_2O_3) + e^{-}(from metal) \rightarrow [Na^{\delta +} - \delta^{-}](on metal)$$
 (9)

where $[Na^{\delta+}-\delta^-]$ denotes the overall neutral Na species present at the metal-gas interface (effective double layer): the Na adatom on the metal surface carries a charge δ^+ and δ^- is the image charge in the metal; the charge δ is about 0.8–0.5, decreasing with increasing Na coverage [7].



Figure 19. Schematic of the effective double layer approach of catalyst promotion [6,7,110]. An electron enriched (lower work function) catalyst surface is developed due to the uniformly dispersed Na^{δ +} adatoms.

It is worth noting that an identical configuration is expected on the metal catalyst surface when Na is introduced to the catalyst by conventional methods (e.g., impregnation, vapor deposition, etc.), as recently demonstrated by Lambert and co-workers [110,111] who showed using XPS that the electrochemically-supplied sodium is identical with gas-supplied (by evaporation) of Na^{δ +} cations on the catalyst surface.

Since alkali ad-atoms are imposed on the catalyst surface, they are spread over the entire metal surface due to strong repulsive dipole-dipole interactions, establishing a homogeneous, neutral, effective double layer. This double layer corresponds to an electrically modified metal surface, practically an electron enriched (lower work function) metal surface due to the δ^- charge located at the metal side interface half-layer. This electrically modified surface interacts with the co-adsorbed reactants, reaction intermediates and products, changing their binding energies (in respect to those on a promoter-free metal surface), producing pronounced alterations in catalytic performance. Therefore, the main feature of the electropositive promotion of PGM by alkalis and alkaline earths is the resulted electron enriched, lower work function, PGM surface.

Therefore, the concomitant alterations on the adsorption characteristics of the co-adsorbed reactants on such a modified metal surface are of course subjected to the electronic properties of the reactants: electrophilic (electron acceptor) or electrophobic (electron donor): both theory [112] and experiment (e.g., [106,113]) have shown that electropositively promoted (by alkalis) Pt-group metal (PGM) surfaces appear strengthening in the bonds of the metal—electron acceptor (electrophilic)

adsorbates, for example, PGM-NO, PGM-O2, PGM-N, PGM-O, PGM-CO and weakening in the bonds of the metal-electron donor (electrophobic) adsorbates, for example, PGM-hydrocarbons and their fragments, PGM-CO. Donation and/or backdonation issues of electron charge between adsorbates and metal surfaces during the formation of a chemisorptive bond and their impact on the bond strength have been thoroughly explained by Vayenas and Brosda in Ref. [114] and in brief summarized in the following statements: "Electron backdonation to bonding singly occupied orbitals of an electron acceptor reactant lying below the metal Fermi level (E_F) results in strengthening of the metal-adsorbate bond, while electron backdonation to antibonding orbitals, leads to weakening of the chemisorptive bond and destabilization of the adsorbate. Also electron donation to a metal from a singly occupied orbital on an electron donor adsorbate lying above Fermi level of the metal leads in general to strengthening of the chemisorptive bond. There are cases, such as the CO chemisorption on transition metals, where both donation of electrons (from the adsorbate to the metal) and backdonation of electronic charge (from the metal to the adsorbate) play an important role in the chemisorptive bond formation (e.g., Blyholder model for CO chemisorption [115])"; (The latter explains why someone can find CO to be considered as an electron-acceptor or even as an electron-donor adsorbate). These modifications on the reactants chemisorptive bonds are accompanied by alterations in the activation energies of the catalytic reactions and in some cases on their mechanism (e.g., on the rate determination step), resulted to dramatic changes on their intrinsic activity and/or selectivity. The specific features of the promotion are subjected to the specific catalytic systems under consideration and to the reaction conditions imposed.

Taking into account that the vast majority of the reactions involved in emission control catalysis processes obey Langmuir-Hinshelwood mechanism (competitive adsorption of the reactants), the following example could be of usefulness for a better understanding of the mechanism of action of the promoters under the concept of the effective double layer approach. Let's suggest a starting rate at point 1 on the rate versus [A] curve I (Figure 20), as determined by the applied reaction conditions on the unpromoted reaction A + B, that are considered to follow a typical Langmuir-Hinshelwood (LH) mechanism with the well-known characteristic rate maximum (point 2) due to the competitive adsorption of the reactants. The reaction rate is proportional to the product of the reactants' coverages, $\theta_A \cdot \theta_B$ (Equation (10)), maximized at equilibrated coverages θ_A and θ_B on the catalyst surface:

$$\mathbf{r} = \mathbf{k}_{o} \exp(-\mathbf{E}_{a}/\mathbf{R}\mathbf{T})\boldsymbol{\theta}_{A} \cdot \boldsymbol{\theta}_{B}$$
(10)

where $k_0 \exp(-E_a/RT) = k$ is the temperature dependence of the rate constant k with a pre-exponential factor k_0 and an apparent activation energy E_a .

Applying (chemically or electrochemically) a promoter, the created effective double layer (Figure 19) causes alterations on the adsorption strengths of the reactants, modifying their coverages θ_A and θ_B thus giving the possibility of their adjustment at the equilibrated values θ_A^* and θ_B^* that minimizes reaction probability and consequently rate (point 2):

$$r^* = k_0 \exp(-E_a / RT) \theta_A^* \cdot \theta_B^*$$
(11)

However, this is the minimum effect that can be offered by this course and actually Equation (11) and consequently path $1\rightarrow 2$ are not real: Due to the induced changes on the chemisorptive bonds of the reactant, alteration on the apparent activation energy of the reaction is also expected from the unpromoted value E_a to a modified value E_a^* , leading the reaction rate to follow a new substantially enhanced behaviour (curve II, Figure 20, for example, point 3), following Equation (12):

$$r^{**} = k_0 \exp(-E_a^*/RT)\theta_A^* \cdot \theta_B^*$$
(12)

Notably, in the cases where reaction mechanism modifications are induced by the promoter as well (e.g., change of the rate determination step), promotion can receive unprecedented enhancements (promoted rate r*** on curve III, Figure 20, e.g., point 4).



Figure 20. Schematic representation of the mechanism of promotion on a Langmuir-Hinshelwood type reaction.

An example which mirrors all the aforementioned issues is the alkali-promoted NO + C_3H_6 reaction on Pt [15,58,61]; reaction rate enhancements up to 40,000% were achieved accompanied with dramatic chances on the selectivity towards N₂, as well (Table 4). The promotion mechanism on this system is described as follow (Figure 21):

First, it is well established that NO + C_3H_6 reaction over Pt obeys Langmuir-Hinshelwood type kinetics with characteristic rate maxima reflecting competitive adsorption of the two NO and C_3H_6 reactants on Pt active sites; these rate maxima occurred at very low C_3H_6/NO ratios (<0.08), that is, very high NO concentrations relative to C_3H_6 are necessary in order to achieve a comparable coverage of both reactants on the Pt surface, thus maximizing the rate [58,62], reflecting a weaker adsorption of NO on Pt sites relative to C_3H_6 [58,62,67,68] mainly due to the presence of π -electrons in the alkene structure. As a consequence, in a wide range of reactants partial pressures and temperatures conditions the un-promoted (alkali-free) Pt surface is predominantly covered by propene and propene-derived fragments (Figure 21a), therefore the un-promoted reaction rate, r_0 , is quite low.

Addition of alkali, which is accompanied by the formation of the [Alkali^{$\delta_+-\delta^-$}] effective double layer, provides an electron enriched (lower work function) metal surface that in effect favours the adsorption of electrophilic co-adsorbates (NO) and inhibit that of electrophobic co-adsorbates (C₃H₆ and its fragments). According to the following Reaction (13)–(18) scheme,

$$NO(g) \rightarrow NO_{ads}$$
 (enhanced by alkali addition) (13)

$$NO_{ads} \rightarrow N_{ads} + O_{ads}$$
 (enhanced by alkali addition) (14)

$$C_3H_6(g) \rightarrow C_3H_{6,ads}$$
 (inhibited by alkali addition) (15)

$$N_{ads} + N_{ads} \rightarrow N_2(g)$$
 (16)

$$N_{ads} + NO_{ads} \rightarrow N_2O(g)$$
 (17)

$$O_{ads}$$
 + hydrocarbonaceous species (or CO) \rightarrow CO₂(g) + H₂O(g) (18)

the resulted effects are (Figure 21b): (i) strengthening of the Pt-NO bond with a concomitant enhancement population of NO molecules on the alkali-modified surface (Reaction (13)) –the surface populations of NO and propene are therefore equilibrated; (ii) strong enhancement of the dissociative adsorption of NO (Reaction (14)) as a result of the Pt–NO bonds strengthening and the concomitant weakening of the N–O bond in the adsorbed NO molecules—the surface is enriched with very active atomic oxygen species; (iii) weakening of the strength of the Pt–C₃H₆ bond—an equilibrated population of more active (due to their weaker adsorption) hydrocarbonaceous species is resulted (Reaction (15), Figure 21b).



Figure 21. Schematic of the mechanism of action of alkalis as promoters on the Pt-catalysed NO + C_3H_6 reaction. Expected distribution of reactants and/or intermediates over the unpromoted surface (**a**) and the Na-promoted surface (**b**).

All these factors operate together in increasing the reaction probability of NO and propene molecules, leading to an enhanced promoted rate, r_p , over the alkali-modified surface. These changes on adsorbed species distribution and/or population (coverage), their bond-strengths and activation energies and on reaction rate-determination-step (rds) were successfully used to convincingly interpret the rate enhancement values, $\rho = r_{CCP}/r_o$, as high as 420 and 280 for N₂ and CO₂ productions have been reported for this catalytic system under electropositive promotion by alkalis or alkaline earths [15,61,62]. Notably, these gains in catalytic activity did not reflect the maxima of the L-H rate curves, *that is*, were not restricted by the optimal balance of the competitive reactants' coverages. These were extraordinarily larger, due to the synergy of the aforementioned factors.

The observed increase in N₂-selectivity upon electropositive promotion can be also readily understood in terms of the above considerations. According to the reaction network (13)–(18), the production rates of N₂ and N₂O depend on the extent of NO dissociation (reaction (14)), which is followed by the elementary Reactions (16) and (17). The observed increase in the N₂-selectivity upon promoter addition is a consequence of increased NO dissociation, that is, less molecular NO and more atomic N on the surface. According to Lang et al. [112], the electrostatic field produced by Na⁺ can shift the π *-orbital energy of NO adsorbed in the vicinity of an alkali ion below the Pt Fermi level. Then, valence electrons from the metal can populate the NO π *-orbital energy, resulting in weakening of the N–O bond and a strengthening of the Pt–N bond. Both factors favour Reaction (16) over Reaction (17), leading to higher N₂ selectivity.

In the case of $CO + O_2$ reaction the strong promotion observed under CO-rich reaction conditions (e.g., [16,27]), was understandable by means of the pronounced strengthening of the O adsorption bond, compared to that of CO: at such conditions where CO coverage predominates and the formation of CO islands suppresses reaction probability, the alkali-induced pronounced adsorption of O species causes destruction of CO islands thus increasing reaction probability [16].

In the case of alkanes oxidation, it is well known that alkanes, in particular CH_4 , have low propensity for adsorption (activation) on PGM surfaces. The adsorption of these electrophobic adsorbates is being even worse on alkali-modified electron enriched surfaces. This makes the generally observed alkali-induced poisoning of alkanes oxidation reactions readily understandable, no matter of what oxidant agent (oxygen or NO) used [48,49,65,66].

The opposite was true for alkenes oxidation (by O_2 or NO), due to their strong propensity to adsorb on PGMs. Thus, weakening their adsorption and at the same time strengthening the adsorption of the electrophilic limiting reactant (O) on the alkali modified PGM surfaces are factors that operate synergistically creating ideal promotion conditions (e.g., [18,20,43–47]).

Comparing the alkali promotion of the NO reduction by alkenes, the Ir~Rh < Pd < Pt increasing order can be observed: marginal promoting effects have been obtained during Rh- or Ir-catalysed NO reduction by propene [59,60,64,107,108], in opposite to the very large on Pd [14] and to the extraordinarily large on Pt (e.g., [15,58,61,62,106]) This is related to the different propensity of NO adsorption on these metals: on Ir and Rh, NO has high propensity of adsorption, which is also dissociative; on Pd less and partly dissociative, while on Pt it is even less and non-dissociative. Therefore, enhancing NO adsorption by alkali-promotion on Ir and Rh surfaces has not further practical value (NO adsorption/dissociation is close to ideal on the unpromoted Ir and Rh), while on Pt the non-dissociative and quite limited NO adsorption can become much stronger and fully dissociative, enabling Pt to behave like Rh or Ir (e.g., [15,47,61,62,89]).

Comparing the amount of an alkali necessary to optimize promotion for a certain reaction on different Pt group metals, the general trend is that only few surface concentrations are needed in the case of Rh and Ir (ca. θ_{Alkali} ~1–3%), quite more for Pd (θ_{Alkali} ~5–10%) and much more for Pt (θ_{Alkali} ~15–40%) (e.g., [14,15,18,22,51,57,58,108]). This is readily understandable in similar terms used above: Rh and Ir electron availability (work function, WF) is by its own close to the optimal value for the reactions under consideration and only few amounts of alkalis are necessary for the optimal

promoter loading adjustment. The opposite is true for Pd and in particular for Pt; these two metals showing a higher WF than Rh and Ir.

For a certain catalytic system and operating conditions, the amount of alkali that is needed to optimize promotion depends on the chemical identity of the alkali: heavier alkalis appear to be more effective that the lighter ones (e.g., [61]). This is fully consistent to the effective double layer approach considered and to the theoretical predictions of Lang et al. [112], which have shown that the larger the alkali cation the greater the effect its electric field has on an electron acceptor adsorbate (e.g., NO).

In the cases of more complex reactions systems, as for instance those simulating automotive exhaust gas mixtures, the explanation of the promotional effects is typically more complicated and may involve all the possible effects, synergistic or competitive, induced through the modification of the metal work function on all electrophilic and/or electrophobic reactant species and intermediates on such a densely populated surface with all these competitive species (e.g., [47,79,89,106]).

It is also of worth noting that: (i) Alkalis and alkaline earths are typically permanent promoters as a result of the fact that they do not participate in the reactions networks. This is of high technological importance: since the optimal alkali amount is incorporated on the catalyst and independently of the method used for its supply (EPOC or CCP), its amount remains practically constant for a very long time offering promotion. This explains why Faradaic Efficiency values as large as $\Lambda > 10^5$ have been measured in EPOC by alkalis studies (e.g., [16–18,42]). (ii) Volcano type behaviour of the promotional effects as a function of the promoter loading was an additional common feature of the titled promotion, that is, outgoing of the optimal amount of the promoter (the so-called over-promotion) the reaction is gradually inhibited rather than promoted (e.g., [15–17,47,58]). Two main factors were considered to be responsible for this behaviour: (a) the extended strengthening of the adsorption bond of the electrophilic adsorbate (electronic effect) and (b) the formation of extended surface complexes of the promoter with the reactant species and/or reaction intermediates that can block active sites (geometric effect). These factors can operate together suppressing the rate in case of over-promoted catalysts. However, it must be noticed that the formation of such 2D or 3D surface alkali complex compounds (alkali nitrites, nitrates, carbonates, oxides or even superoxides, depending on the reaction atmosphere), that can be valid even at low promoter loadings, do not actually cancel the role of alkali as a promoter (e.g., [20,43–46,51,52,57–60]). On the opposite, stabilization on the catalysts surfaces, even at elevated temperatures of catalytic interest, of normally volatile alkali metals is achieved by the formation of such stable surface alkali complex compounds.

3.3.2. Direct Spectroscopic and Other Analytical Technique Evidences

Besides of the numerous kinetic studies demonstrating the usefulness of the electropositive promotion of PGM in emissions control catalysis, particular emphasis to direct evidences for the origin and mechanism of this promotion have been devoted by several research groups using a variety of surface science or other analytical techniques, including: in situ diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) [95–98,113], x-ray photoelectron spectroscopy (XPS) (e.g., [51,52,111]), temperature programmed desorption (TPD) [106,116–120], scanning tunnelling microscopy (STM) [121–123] and work function [51,52] measurements.

To this end, Koukiou et al. [113], studying the interaction of NO with Na-modified Pt surfaces over a conventional, highly dispersed, Pt catalyst, namely $Pt(Na)/\gamma$ -Al₂O₃, by means of in situ DRIFTS, have shown that increasing sodium loading causes a pronounced and progressive red shift of the N–O stretching frequency (Figure 19) associated with molecular NO adsorbed on the Pt component of the supported catalyst.

At the highest sodium loading (10wt% Na; corresponding to an upper limit of about $\theta_{\text{Na}} = 0.5$ nominal sodium fractional coverage on Pt particles [102]) a 1680 cm⁻¹ species is observed that corresponds to an activated NO^{δ -} species with bond order 2, that is, a negatively charged adsorbate with increased electron density in the π^* antibonding orbital. The rigorous model developed by Lang et al. [112] was considered for understanding the effect of alkali on Pt surface. They have shown

that the inhomogeneous electric field associated with adsorbed alkali ions acts to depress the energy of the antibonding π^* orbital of electron-accepting (electrophilic) co-adsopbates (e.g., CO, NO) bellow the metal Fermi level. Therefore, the red shifts showed in Figure 22, as authors argue [113], are due to the resulting metal \rightarrow adsorbate charge transfer that acts to weaken the N–O bond, red-shifting its vibration frequency and promoting its dissociation, rather than to alkali-induced changes in the relative populations of various forms of adsorbed NO.



Figure 22. DRIFT spectra (expanded scale for the region 2000–1600 cm⁻¹ and subtracting spectrum of pure Al₂O₃) obtained with 10wt%Na/ γ -Al₂O₃ (**B**), Pt/ γ -Al₂O₃ (**C**), Pt(Na5)/ γ -Al₂O₃ (**D**) and Pt(Na10)/ γ -Al₂O₃ (**E**) samples after NO adsorption for 1 min at 27 °C. (Reprinted with permission from Ref. [113]; Copyright 2007, Elsevier).

The chemisorption characteristics of NO on Na-dosed Pt{111} surface has been also studied by Harkness and Lambert [116] by means of temperature-programmed desorption (TPD). It was clearly demonstrated that the presence of Na on the Pt{111} single crystal causes an increase in the adsorption energy of NO (i.e., strengthening of the metal-NO bond) accompanied by a monotonically increased

NO dissociation (i.e., N_2 desorption) as a function of Na coverage. In particular, NO-TPD spectra on Na-dosed Pt{111} (from 0 to 1.2ML of Na coverage) have showed two characteristic peaks, indicated as features (a) and (b); sodium caused an increase in the desorption peak (a) temperature from 360 K on the clean Pt{111} to 380 K at 0.25 ML Na-dosed Pt{111}. Even at very low Na coverages (>0.1 ML), the second NO desorption feature (b) from sodium affected sites is appeared at ~520K which is grew at the expense of the feature (a) and progressively shifted to higher desorption temperature, up to 610 K, upon increasing Na coverage to 1.2 ML. At the same time, the feature (a) is progressively attenuated to zero (practically vanished at~>1ML of Na); A progressively increased N₂ peak is concomitant to the aforementioned TPD experiment, with the N₂ desorption peak to remain constant at 540K up to 1ML of Na, after which it is shifted to 540 K for 1.2 ML (i.e., 100% dissociation of adsorbed NO can be obtained at sufficiently high, ca 1 ML, sodium coverage) [116].

Similar TPD studies by Garfungel et al. [117], involving the interaction of NO with K-dosed Pt{111}, yielded the same evidences: an alkali-induced increase in the Pt–NO bond strength accompanied by alkali-promoted NO dissociation was demonstrated.

Temperature-programmed desorption (TPD) studies after adsorption of oxygen on Na-dosed Pt/Al_2O_3 catalysts where performed by Vernoux et al. [106] in order to elucidate possible influences of the Na addition on the Pt–O bond strength. They demonstrated that the adsorption strength of the atomic oxygen (an electron acceptor adsorbate) on the Pt component of the catalyst is enhanced by Na addition; their data clearly showed that the temperature of the oxygen desorption peak correspond to Pt–O interaction increases with increasing Na loading of the catalyst, directly evidencing a strengthening of the Pt–O bond [106].

A number of early and recent studies concern the chemisorption of CO on alkali-modified PGM surfaces [95–97,117–120]. To this end, Bertolini et al. [118] studied the chemisorption of CO on K-modified Pt(100) surfaces by means TPD and Auger Electron Spectroscopy (AES). The progressive increase on the CO desorption peak temperature (i.e., strengthening of the Pt–CO) as a function of K coverage, observed at low K coverages $\theta_{\rm K}$ (ca. 0–0.4), was attributed to long range uniform electronic modifications, that is, work function changes, due to charge transfer; K^{δ +} ($\delta \approx 1$) adspecies appear on the K-dosed surface, uniformly distributed due to repulsion between these charged species. Such an electron enriched (lower work function) Pt surface strengthens the adsorption of CO via substantial charge donation form the K-modified Pt surface into the $2\pi^*$ orbitals of CO [118]. At high K coverages (ca. 0.45–1) short range interactions between the CO and K adspecies, that is, an even more direct bonding of CO with K (formation of CO-K-Pt surface complexes), were considered.

Pitchon et al. studied the effect of (Li, Na, K and Cs)-dosed Pd/SiO_2 on CO adsorption characteristics [119]. They demonstrated drastic changes in the infrared spectrum of CO adsorbed on Pd sites; a significant weakening on the strength of the C–O bond of adsorbed CO molecules caused the appearance of a new ν CO band at low infrared frequency—the extent of the interaction was depended by the nature of the alkali. These influences were assigned to a localized interaction "Pd–CO–alkali" rather than to long range electronic modifications.

Using FTIR spectroscopy, Liotta et al. studied the effect of sodium on the adsorption of CO on Pd-based catalysts supported on two different supports (SiO₂ and model or natural pumices) [120]. Depending on the support used, the appearance of both electronic and geometric effects—that attributed to the different localization of sodium ions in the catalysts- were evidenced; the geometric effects were predominant in the Pd/SiO₂ catalysts, whereas electronic effects were most important than the geometric to Pd is the common effect of the presence of alkali either it is subsequently added to the catalysts or is present as a structural component of the support, although, the authors observed important geometric effects in the former case. The electronic effect, that produces a red shift (towards lower frequencies) of the CO bands of the IR spectra of chemisorbed CO, owing to an enhanced transfer of electron density from the metal to the π^* molecular orbitals of CO, was again invoked [120].

Electron enriched Pt surfaces resulted by combined application of alkali-induced surface promotion and support-mediated promotion by Ce-based mixed oxides were also recently demonstrated via in situ DRIFTS studies by Yentekakis and co-workers [95–97]. Over doubly promoted Pt(Na)/Al₂O₃-CeO₂-La₂O₃ catalysts operated under TWC conditions, the authors demonstrated a substantial population of adsorbed species at 2060 cm⁻¹ attributed to CO on reduced Pt⁰ sites when CeO₂-La₂O₃ was incorporated into the Al₂O₃ support. This feature was substantially exacerbated when Na promoter was also incorporated in the catalyst formulation, while at the same time the formation of active Pt–NCO intermediates at about 2180 cm⁻¹, which resulted from an enhanced NO decomposition, was facilitated by alkali-promotion on the Pt(Na)/Al₂O₃-CeLa catalyst (Figure 21). On the un-modified Pt/Al₂O₃ catalyst counterpart, CO species adsorbed on positively charged Pt sites (Pt^{δ+}-CO) were assigned (Figure 23). These features, fully compatible with the way of action of alkali promotion analysed so far, were considered as a convincing explanation of the substantially enhanced TWC performance of this complex reactions/catalyst system.



Figure 23. In-situ DRIFT spectra during simulated TWC reaction conditions (1000 ppm NO + 1067 ppm C_3H_6 + 7000 ppm CO + 7800 ppm O₂; $F_t = 80 \text{ cm}^3/\text{min}$; T = 200 °C). (Reproduced with permission from Ref. [96]. Copyright 2008, Elsevier).

Finally, a number of studies involving x-ray photoelectron spectroscopy (XPS) published by the groups of Lambert and Yentekakis (e.g., [20,51,52,57–60]) have doubtlessly shown that in EPOC studies that concern PGMs films interfaced with an alkali conducting solid electrolyte, the promoting species are indeed alkaline ions reversibly supplied by the external bias onto the catalyst surface.

The authors have also shown that under reaction conditions the electrochemically supplied alkali ions form stable surface compounds (e.g., alkali nitrites, nitrates and carbonates) that can act as promoters even under such formulations and which, in excess (over-promotion), induce geometric, active sites blocking, phenomena and thus poisoning effects on PGMs activity [20,51,52,57–60]. Combining XPS and ultraviolet photoelectron spectroscopy (UPS), the authors demonstrated a linear relationship of the catalyst work function ($\Delta \Phi$) and the alkali coverage (θ_{Na}) with the catalyst potential overpotential (ΔV_{WR}) (Figure 24) [51,52] and confirmed that electrochemically supplied alkali ions on the catalyst surface are identical in behaviour (and chemical state) with the alkali supplied by vacuum deposition [21,22]. This makes the close similarities found between EPOC and CCP methodologies of catalyst promotion readily understandable [15,21,22,25,58] and prompts for the use of EPOC as an effective and rapid method for exploring the effects of a range of promoters and for assessing the response of the reaction rate/selectivity to promoter coverage (finding out its optimal loadings at any set of conditions) before applying them to the design of efficient conventional catalyst formulations.



Figure 24. Na 1s XPS spectra versus catalyst potential (V_{WR}) for Rh film interfaced with (Na) β'' Al₂O₃ Na⁺-ions conductor, at 307 °C and UHV conditions. The inset shows the integrated Na1s XPS intensity due to sodium on the Rh surface and associated work function change of the Rh film as a function of V_{WR} . Data was acquired using a Rh/(Na) β'' Al₂O₃/Au galvanic cell (Reproduced with permission from Ref. [51]. Copyright 2001, American Chemical Society).

A recent review of the literature using surface analysis techniques to shed light on the electropositive promotion via the EPOC concept has recently provided by Gonzalez-Cobos and de Lucas-Consuegra [124].

All the aforementioned in this chapter studies, unambiguously lead to the main conclusion that alkalis adatoms on PGM surfaces provide electron enriched metal sites (lower work function surfaces), which favour the adsorption of electron donor (electrophilic) co-adsorbates and inhibit the adsorption of electron acceptor (electrophobic) co-adsorbates. The formation of alkali-reactants-metal surface complexes and therefore geometric type influences, besides the above electronic ones, can be at work

in cases where the conditions assist such a course. These issues were considered in the vast majority of electropositive promotion studies by alkalis or alkaline earths to explain promotion phenomena observed on PGM-catalysed emissions control catalysis reactions.

4. Conclusions and Perspectives

The electropositive promotion, by alkalis or alkaline earths, has been found to be an effective tool for promoting the catalytic performance of Platinum Group Metals for the vast majority of important reactions involved in emissions control catalysis systems, such as CO and alkenes oxidations, NO reduction by CO, alkenes or H_2 , N_2O decomposition and/or reduction; the only exceptions were the alkanes oxidation by dioxygen or NO.

Up to two orders of magnitude rate increases were achievable, in particular with electropositive-promoted Pt, while in de-Nox reactions the selectivity towards N_2 was simultaneously improved towards values approaching 100% over the "optimally promoted" catalysts.

Promotional achievements were large not only for model reactions but also for gas mixtures that simulate practical systems, such as stoichiometric gasoline, lean-burn or diesel engines exhaust streams.

The intensity of promotion was generally followed the order Pt > Pd > Rh~Ir, thus providing Pt or Pd-based catalysts to operate like Rh or even better in emissions control catalytic reactions. These achievements were lead to the synthesis of novel simple in synthesis and cost-effective monometallic (Pt or Pd-only) with low precious metal loading, extremely active, stable and readily recycling, three-way catalyst formulations that were compared well with commercial bi- or three-metallic (Rh, Pt, Pd)-TWCs for the control of stoichiometric gasoline engines exhaust emissions. The absence of rare and expensive Rh in the formulations of three-way converters has therefore been introduced with the concomitant environmental and economic gains.

Electropositive promotion of PGM was also found to be effective at excess oxygen conditions (e.g., lean-burn of diesel engines exhaust gases), without however in that case to obtain solutions fully satisfying the requirements.

Latterly, doubly promoted catalyst formulations, combing electropositive promotion by alkalis and support-mediated promotion by mixed oxide supports (including combinations of Al_2O_3 with CeO_2 , ZrO_2 , La_2O_3 or TiO_2) have been found to provide even better promotion and stability for emissions control implementations. Such synergistic promotion catalysts' designs could lead to control systems readily adapted on a number of specific emissions of stationary sources, that have high environmental footprint and their control is an urgent issue.

From the theoretical point of view, electropositive promotion of PGM was successfully understood and interpreted in terms of the effective double layer approach, that is, a homogeneous, neutral, effective double layer established by the alkali (or alkaline earths) adatoms that are spread over the entire PGM surface providing an electron enriched (lower work function) metal surface due to the δ^- charge located at the metal side interface half-layer, which interacts with the co-adsorbed reactants, reaction intermediates and products, changing their binding energies producing pronounced alterations in catalytic performance. To this end, strengthening in the bonds of the metal—electron acceptor adsorbates and weakening in the bonds of the metal—electron donor adsorbates are resulted, which are accompanied by alterations in the activation energies of the reactions between them, leading to dramatic changes of the intrinsic catalytic performance (activity and/or selectivity). In particular, in emissions control catalysis processes involving NO_x reduction, the promotion of the dissociative chemisorption of the NO molecules, which is a key reaction initiating step in deNO_x processes, facilitated by the alkali-induced strengthening of the PGM–NO and the accompanied weakening of the N–O bond of the adsorbed NO molecules, convincingly interpret the enormous effective promotion of both deNO_x activity and N₂-selectivity offered via electropositive promotion of these catalytic systems.

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in the field, publishing a number of papers that are included and analysed in the present review. All authors read and approved the final version of the manuscript.

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A Review of Low Temperature NH₃-SCR for Removal of NO_x

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Abstract: The importance of the low-temperature selective catalytic reduction (LT-SCR) of NO_x by NH₃ is increasing due to the recent severe pollution regulations being imposed around the world. Supported and mixed transition metal oxides have been widely investigated for LT-SCR technology. However, these catalytic materials have some drawbacks, especially in terms of catalyst poisoning by H₂O or/and SO₂. Hence, the development of catalysts for the LT-SCR process is still under active investigation throughout seeking better performance. Extensive research efforts have been made to develop new advanced materials for this technology. This article critically reviews the recent research progress on supported transition and mixed transition metal oxide catalysts for the LT-SCR reaction. The review covered the description of the influence of operating conditions and promoters on the LT-SCR performance. The reaction mechanism, reaction intermediates, and active sites are also discussed in detail using isotopic labelling and in situ FT-IR studies.

Keywords: low-temperature selective catalytic reduction; NH_3 -SCR; de-NO_x catalysis; SO_2/H_2O tolerance; transition metal-based catalysts

1. Introduction

The non-renewable fossil fuels are continuing to remain the dominant energy source in power plants and automobiles to satisfy the ever-growing energy demands. However, the combustion of fossil fuels mainly generates nitrogen oxide (NO_x) pollutants (NO, NO₂, and N₂O and their derivatives) which can cause acid rain, photochemical smog, ozone depletion, and eutrophication problems [1–4]. Due to the negative impacts of NO_x, the mitigation of NO_x emissions is of paramount importance for environmental protection. Several technologies are available to reduce NO_x emissions by using catalytic materials and among them, selective catalytic reduction of NO_x with NH₃ (NH₃-SCR) has been widely applied due to its high NO_x removal efficiency [5–7]. Usually, the flue gas temperature of the industrial process is as low as 300 °C and, thus, the SCR catalyst must be active in the low-temperature regime (100–300 °C). V₂O₅–WO₃(MoO₃)/TiO₂ is the typical and efficient catalyst and has been commercialized for NH₃-SCR technology for medium temperature process [8,9]. However, this catalyst has some intrinsic drawbacks such as narrow and high working temperature window (350–400 °C), and low N₂ selectivity in the high-temperature range [3,10,11]. Therefore, many researchers continue to develop highly active catalysts for low-temperature NH₃-SCR in a wide temperature window.

With this perspective, several transition metal oxide-based catalysts have been extensively investigated for low-temperature NH_3 -SCR reaction due to their excellent redox properties, low price, and high thermodynamic stability. Especially, the easy gain and loss of electrons in the *d* shell of

the transition metal ions could be responsible for the facile redox properties [12–14]. For example, the Cr/TiO₂ [15], Cr-MnO_x [8], Fe-MnO_x [16], Mn/TiO₂ [17,18], Fe_xTiO_y [19], MnO_x/CeO₂ [20], and Cu/TiO₂ [21] catalysts were shown to exhibit good SCR activity in the low temperature range. In our earlier work, we investigated the low-temperature NH₃-SCR in the presence of excess O₂ on the TiO₂ supported V, Cr, Mn, Fe, Co, Ni, and Cu oxides and found the catalytic performance decreased in the following order of Mn > Cu \geq Cr \gg Co > Fe \gg V \gg Ni [16]. Particularly, manganese-containing catalysts have attracted much attention due to its variable valence states and excellent redox ability [2,5]. In the recent past, we published a series of papers on Mn-based SCR catalysts that showed a highly promising deNO_x potential in the low-temperature region [5,22–26]. However, these catalysts are very sensitive to the presence of SO₂ in the feed and exhibit lower N₂ selectivity [8,27-29]. Hence, the development of catalysts that show both good low-temperature activity and high SO₂/H₂O durability is of great importance for the NH₃-SCR reaction. In general, there are two plausible strategies available to enhance low-temperature NH₃-SCR performance. One strategy is to modify the transition metal oxide with one or multiple metal oxides, which could enhance the active sites for the reaction by inducing the synergistic effect [30-33]. The other approach is to synthesize the supported materials to disperse the transition metal-based oxides which can increase the activity by metal-support interactions [26,34–37]. Recently, many supported and mixed transition metal catalyst formulations have been studied to improve the low-temperature SCR performance, as well as resistance to SO_2/H_2O .

In this study, we systematically reviewed the recent advancements in developing the transition metal-based catalysts for low-temperature NH₃-SCR reaction. This review also demonstrated the action of different promoters and supports on the catalytic performance and SO₂/H₂O tolerance of the transition metal-based catalysts in NH₃-SCR of NO_x. The reported catalysts were divided into four categories, such as binary, ternary/multi, supported single, and supported binary/multi-transition metal-based catalysts.

2. Binary Transition Metal-Based Catalysts

Various transition-metal oxides have been proved to be active for the NH₃-SCR at low-temperature. However, the catalytic performance on single transition metal oxide is far from satisfactory due to their low specific surface area and thermal instability [38–41]. The addition of dopants is a common method to improve the drawbacks associated with pure transition metal oxide. Hence, much progress has done to improve the low-temperature SCR activity of transition metal oxides by mixing or doping with other metal oxides. In recent years, Mn, Fe, Co, Ni, and Cu-based binary oxide catalysts have been extensively studied for low-temperature NH₃-SCR reaction due to their attractive catalytic performance [19,33,41-49]. Particularly, Mn-based binary oxides are popular and proven to be effective catalysts for low-temperature NH₃-SCR reaction [42,50,51]. Recently, Xin et al. [52] designed bifunctional V_a -MnO_x (where a represents the molar ratios of V / (V + Mn)) catalysts composed of Mn_2O_3 and $Mn_2V_2O_7$ phases that significantly improved both NO_x conversion and N_2 selectivity in comparison with Mn_2O_3 at low-temperature (Figure 1). Although $Mn_2V_2O_7$ showed an excellent N_2 selectivity, the NO_x conversion is much lower on it. Especially, above 90% NO_x conversion and 80% N_2 selectivity was observed in the temperature region of 120–240 °C over the $V_{0.05}$ -MnO_x catalyst. The $V_{0.05}$ -MnO_x catalyst also found to be exhibit higher NO_x conversion to N₂ as compared to the mechanically mixed $Mn_2O_3 + Mn_2V_2O_7$ sample which has the same component content to $V_{0.05}$ -MnO_x (Figure 1). This finding indicated that the synergism between Mn_2O_3 and $Mn_2V_2O_7$ exists in the chemically prepared $V_{0.05}$ -MnO_x rather than the mechanically mixed Mn₂O₃ + Mn₂V₂O₇ sample. Moreover, the mechanically mixed $Mn_2O_3 + Mn_2V_2O_7$ sample showed higher activity in comparison to mechanically mixed $MoO_3 + Mn_2V_2O_7$ sample, suggesting that the presence of Mn_2O_3 phase in the catalyst is necessary for NH₃-SCR reaction. In conjunction with in situ IR characterization and DFT (density functional theory) calculation results, the authors concluded that the Mn₂O₃ phase of the catalyst could activate NH3 into NH2 intermediate, which then transferred to the Mn2V2O7 phase of

the catalyst and reacted with gaseous NO into NH₂NO. Finally, the generated NH₂NO intermediate on the $Mn_2V_2O_7$ phase exclusively decomposed to the N_2 rather than the undesired byproduct, N_2O , which is formed due to the deep oxidation of adsorbed NH₃ on Mn_2O_3 .



Figure 1. (a) NO_x conversion and (b) N_2 selectivity for V_a -MnO_x, Mn₂O₃, Mn₂V₂O₇, and reference samples. Reprinted from Reference [52]. Copyright 2018, with Permission from American Chemical Society.

Han and co-workers [53] fabricated triple-shelled NiMn₂O₄ hollow spheres (Figure 2a,b) by using a solvothermal method and tested their ability for low-temperature NH₃-SCR reaction. As shown in Figure 2c, the prepared NiMn₂O₄ hollow spheres (NiMn₂O₄-S) showed the best catalytic activity with NO_x conversion of above 90% over a wide temperature range from 100 °C to 225 °C as compared to the NiMn₂O₄ nanoparticles (NiMn₂O₄-P). The triple-layer shell structure of the NiMn₂O₄-S catalyst generates a larger surface area (165.3 m² g⁻¹) that exposes more active sites (such as surface Mn⁴⁺ and surface adsorbed oxygen species), which are responsible for its superior activity. Additionally, the NiMn₂O₄-S catalyst displayed outstanding stability and good tolerance to H₂O and SO₂ (Figure 2d).



Figure 2. (a) Scheme of the preparation of triple shelled NiMn₂O₄ hollow spheres and (b) TEM image of NiMn₂O₄-S; (c) NO_x conversion over the NiMn₂O₄ catalysts and (d) durability tests of the NiMn₂O₄-S catalyst at 150 °C. Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol%, [SO₂] = 100 ppm (when used), [H₂O] = 5 vol% (when used), balanced with Ar, GHSV = 68,000 h⁻¹. Adapted from Reference [53]. Copyright 2018, with Permission from Royal Society of Chemistry.

Gao et al. [54] investigated the low-temperature NH₃-SCR reaction over the hydroxyl-containing Me-Mn binary oxides (Me = Co, Ni) prepared by a combined complexation–esterification method. It was found that the NO_x conversion decreased in the order of Mn₃O₄-Co₃O₄-OH (Co-MnO_x binary oxide) > Mn₂O₃-NiMnO₃-OH (Ni-MnO_x binary oxide) > Mn₂O₃-OH, while the N₂ selectivity increased in the sequence of Mn₃O₄-Co₃O₄-OH < Mn₂O₃-OH < Mn₂O₃-NiMnO₃-OH. Although the Co and Ni elements in the catalysts delay the poisoning of SO₂ as compared to MnO_x sample, the Co-MnO_x and Ni-MnO_x binary oxides are deactivated by SO₂ over the postponement due to the formation of metal sulfate and ammonia hydrogensulfite species. In another study, Sun and co-workers [55] prepared Mn_{0.66}M_{0.33}O_x catalysts (M = Fe, Zn, Cu) and a series of Fe_αMn_{1-α}O_x ($\alpha = 1, 0.25, 0.33, 0.50, 0$ mol%) catalysts and examined for NH₃-SCR at low-temperatures. The results demonstrated that the Fe_{0.33}Mn_{0.66}O_x catalyst displayed the superior NH₃-SCR activity (NO_x removal efficiency > 90%) in a wide temperature range (75-225 °C) among the Cu_{0.33}Mn_{0.66}O_x, Zn_{0.33}Mn_{0.66}O_x, and Fe_αMn_{1-α}O_x ($\alpha = 1, 0.25, 0.50, 0$ mol%) catalysts. The authors proposed that the distortion of the catalyst structure by Fe doping could play a key role in improving the NH₃-SCR performance over the Fe_{0.33}Mn_{0.66}O_x catalyst.

Rare-earth metal oxides have been frequently adopted to modify the MnO_x as an efficient low-temperature NH₃-SCR catalyst due to their incomplete 4f and empty 5d orbitals [50,51,56]. Fan et al. [57] synthesized Gd-modified MnO_x catalysts with Gd/Mn molar ratio of 0.05, 0.1, and 0.3 to improve the catalytic performance and sulfur resistance in the NH₃-SCR reaction at low-temperature. The MnGdO-2 catalyst (the mole ratio of Gd/Mn = 0.1) found to show the optimal NO conversion and N₂ selectivity among the investigated catalysts. The addition of a proper amount of Gd into MnO_x could enhance the concentrations of surface Mn^{4+} and chemisorbed oxygen species, and increase the amount and the strength of surface acid sites, which lead to better low-temperature catalytic performance than the others. Furthermore, the MnGdO-2 catalyst had an excellent tolerance to SO₂/H₂O as compared to pure MnO_x sample (Figure 3). Their results demonstrate that the doping of Gd could restrains the transformation of MnO₂ to Mn₂O₃ and the generation of MnSO₄, obstructs the decrease in Lewis acid sites and the increase in Brønsted acid sites, and eases the competitive adsorption between the NO and SO₂ and, thus, improves the resistance to SO₂.



Figure 3. The (**a**) resistance to water vapor poisoning test and (**b**) resistance to sulfur poisoning test. (Reaction condition: $[NO] = [NH_3] = 500 \text{ ppm}$, $[O_2] = 5 \text{ vol}\%$, balanced with N_2 , $[H_2O] = 5 \text{ vol}\%$, $[SO_2] = 100 \text{ ppm}$, and $GHSV = 36,000 \text{ h}^{-1}$). Reprinted from Reference [57]. Copyright 2018, with Permission from Elsevier.

Li et al. [58] developed hollow MnO_x -CeO₂ binary nanotubes as efficient low-temperature NH₃-SCR catalysts via an interfacial oxidation-reduction process using KMnO₄ aqueous solution and Ce(OH)CO₃ nanorod as both template and reducing agent without any other intermediate. They reported that the MnO_x-CeO₂ hollow nanotube catalyst with 3.75 g of Ce(OH)CO₃ template (denoted it as MnO_x-CeO₂-B) exhibited outstanding performance with more than 96% NO_x conversion in the temperature range of 100–180 °C. The best activity of the MnO_x-CeO₂-B catalyst was due to its ample number of surface Mn⁴⁺ and O species, and hollow and porous structures that provide abundant Lewis acid sites and large surface area. Additionally, MnO_x-CeO₂-B catalyst showed an excellent resistance to H₂O and SO₂ (Figure 4) and especially, the great SO₂ tolerance was ascribed to the hierarchically porous and hollow structure that inhibits the deposition of ammonium sulfate species, and the doping of ceria that acts as an SO₂ trap to limit sulfation of the main active phase.



Figure 4. H₂O tolerance and SO₂ tolerance of the MnO_x-CeO₂-B hollow nanotube. Reaction conditions: $[NO_x] = [NH_3] = 1000 \text{ ppm}, [O_2] = 5\%$, N₂ as balance gas, and GHSV = 30,000 h⁻¹. Reprinted from Reference [58]. Copyright 2018, with Permission from Elsevier.

Fe-based binary catalysts have also been studied as NH₃-SCR catalysts due to their high activity, excellent resistance to H₂O and SO₂, outstanding environmentally friendly performance, lower cost, and higher abundance [59-62]. Mu et al. [63] synthesized a series of vanadium-doped Fe₂O₃ catalysts and evaluated the effect of V on the low-temperature NH₃-SCR activity of hematite. The NH₃-SCR activity and N₂ selectivity are greatly enhanced after the incorporation of vanadium into Fe₂O₃ and the Fe_{0.75}V_{0.25}O_{δ} catalyst with a Fe/V mole ratio of 3/1 showed the best catalytic performance over a wide temperature window and strong resistance to H_2O and SO_2 . They found that the charge transfer from Fe to V due to the electron inductive effect between Fe and V which could enhance the redox ability and surface acidity thereby superior NH₃-SCR activity at low-temperature. The in situ DRIFTS and kinetic studies suggested that the SCR reaction followed the Langmuir-Hinshelwood mechanism below 200 °C, while an Eley-Rideal mechanism dominated at and above 200 °C. Li and co-workers [64] reported novel iron titanium (CT-FeTi) catalyst, prepared by a CTAB-assisted process, showing good deNO_x efficiency and H₂O resistance at low-temperature as compared to the FeTi catalyst that prepared without adding CTAB. The addition of CTAB during the CT-FeTi catalyst preparation not only promotes to form the uniform mesoporous structure to avoid being excessively enlarged in the presence of H₂O, but also enhances the adsorption of bridging nitrate and NH₃ species on Lewis acid sites. Thus, the authors concluded that the CTAB acted as a "structural" and "chemical" promoter in improving the NH₃-SCR activity and H₂O resistance at low-temperature.

Recently, Co-based spinel catalysts have shown to exhibit a remarkable low-temperature NH₃-SCR activity, N₂ selectivity, and tolerance to SO₂/H₂O [30,65–67]. Meng et al. [48] synthesized a highly efficient Co_aMn_bO_x (where a/b is the molar ratio of Co/Mn) mixed oxide catalysts and investigated the effects of the Co/Mn molar ratio on the low-temperature NH₃-SCR reaction. The Co_aMn_bO_x mixed oxides showed higher NH₃-SCR activity than either MnO_x or CoO_x alone due to their improved redox properties and surface acid sites by the synergistic effects between the Co and Mn species. Particularly, the catalyst with Co/Mn molar ratio of 7:3 (Co₇Mn₃O_x) exhibited the greatest activity (>80% NO_x conversion) in a temperature window of 116–285 °C as compared to the catalysts with Co/Mn molar ratio of 5:5 (Co₅Mn₅O_x) and 3:7 (Co₃Mn₇O_x). They considered that the high NO + O₂ adsorption ability and enhanced redox properties of the Co₇Mn₃O_x catalyst, emerging from its MnCo₂O_{4.5} spinel phase and higher surface area, were beneficial to augment the NH₃-SCR performance by forming

nitrate species on the catalyst surface. Although $Co_7Mn_3O_x$ catalyst had better resistance to H_2O/SO_2 than the $Co_3Mn_7O_x$ and MnO_x , the tolerance to SO_2 poisoning still need to be improved for practical use. Nevertheless, it was found that the deactivated $Co_7Mn_3O_x$, $Co_3Mn_7O_x$, and MnO_x catalysts in SO_2 stream can be regenerated simply by washing with water. Based on their results, the authors also proposed the NH_3 -SCR reaction mechanism over the $Co_7Mn_3O_x$ catalyst, which is shown in Scheme 1. The reaction was initiated by adsorption and activation of gaseous oxygen on oxygen vacancies (symbol \Box), which was then transformed into lattice oxygen O^{2-} (Step 1); This lattice oxygen was diffused to the catalyst surface and then it had become surface active oxygen (O*) (Step 2); Gaseous NO was adsorbed and subsequently reacted with O* to form NO_2/NO_3^- intermediates (Step 3); Meanwhile, NH₃ was activated to $-NH_2$ and NH_4^+ species by Mn^{4+} (Step 4); Finally, NO_2/NO_3^- intermediates reacted with the NH species to produce reaction products, N_2 , and H_2O (Step 5); By the electron transfer from Mn^{3+} to Co^{3+} (Step 6); the catalyst was recovered to its original state (Step 7); Thus, the synergistic effect between the Co and Mn plays a key role in improving the NH_3 -SCR activity over $Co_7Mn_3O_x$ catalyst.



Scheme 1. The proposed mechanism of the NH₃-SCR reaction over the $Co_7Mn_3O_x$ catalyst and the synergetic catalytic effect between Mn and Co cations. Reprinted from Reference [48]. Copyright 2018, with Permission from Elsevier.

Mesoporous materials have been proved as promising catalysts for NH₃-SCR reaction since they can facilitate to promote effective diffusion of reactants towards the active sites [30,65,66,68]. With this perspective, Hu et al. [47] developed mesoporous 3D nanosphere-like Mn-Co-O catalysts through a template-free approach and evaluated for low-temperature NH₃-SCR reaction. It was found that the synthesized Mn-Co-O samples showed excellent NH₃-SCR activity in a broad working temperature window of 75 to 325 °C (NO_x conversion above 80%). They ascribed this outstanding performance to the strong and abundant acid sites, the strong adsorption of NO_x, robust redox properties, the formation of more oxygen vacancies and metal-metal interactions between the cobalt and manganese species.

Besides Mn, Fe, and Co oxides, CuO_x has also been considered in the bimetallic catalyst formulations for low-temperature NH₃-SCR reaction [69,70]. For instance, Ali et al. [71] reported the Cu_x -Nb_{1.1-x} (x = 0.45, 0.35, 0.25, 0.15) bimetal oxides catalysts and found that the Cu/Nb ratio was crucial in enhancing the NH₃-SCR activity. As shown in Figure 5a,b, all binary Cu_x -Nb_{1.1-x} oxides exhibited significantly higher activity than the CuO_x and Nb₂O₅, and among the Cu_x -Nb_{1.1-x} samples, $Cu_{0.25}$ -Nb_{0.85} catalyst displayed the best performance in a wide temperature window of 180–330 °C (>90% NO conversion). Even at a high GHSV of 105,000 h⁻¹, the optimal $Cu_{0.25}$ -Nb_{0.85} catalyst

showed a good NO removal efficiency (above 90% NO conversion) from 210 °C to 360 °C (Figure 5c). Although the SO₂/H₂O streams in the feed gas have some adverse impact on Cu_{0.25}-Nb_{0.85}, still the catalyst showed excellent resistance to SO₂/H₂O with reversible deactivation (Figure 5d). The superior NH₃-SCR performance and SO₂/H₂O tolerance of Cu_{0.25}-Nb_{0.85} catalyst were attributed to its high acid amount and NO adsorption capacity.



Figure 5. (a) NO conversion; (b) N₂ selectivity over Cu_x -Nb_{1.1-x} (x = 0.45, 0.35, 0.25, 0.15) as a function of temperature under a GHSV of 35,000 h⁻¹; (c) effect of GHSV on NO conversion over $Cu_{0.25}$ -Nb_{0.85}, and (d) effect of SO₂, H₂O, and SO₂ + H₂O on NO conversion over $Cu_{0.25}$ -Nb_{0.85} at 200 °C under a GHSV of 35,000 h⁻¹. (Reaction conditions: [NO] = [NH₃] = 1000 ppm, [O₂] = 3% and N₂ balance), and Effect of SO₂, H₂O, and SO₂ + H₂O on NO conversion over $Cu_{0.25}$ -Nb_{0.85} at 200 °C under a GHSV of 35,000 h⁻¹. Reprinted from Reference [71]. Copyright 2018, with Permission from Elsevier.

3. Ternary and Multi-Transition Metal-Based Catalysts

The catalytic performance of single transition metal oxides can also be improved by mixing with two or multi other metal oxides. Thus, transition metal oxides are widely reported to fabricate ternaryor multi-metal-based low-temperature NH₃-SCR catalysts, which could improve the catalytic activity by the enlarged synergetic interactions [14,72–87]. Fang et al. [88] investigated the low-temperature NH₃-SCR reaction over the Fe_{0.3}Mn_{0.5}Zr_{0.2} catalyst and found that it showed an excellent deNO_x activity with 100% NO conversion in the temperature range of 200–360 °C as compared to the Fe_{0.5}Zr_{0.5} and Mn_{0.5}Zr_{0.5} samples. Moreover, the Fe_{0.3}Mn_{0.5}Zr_{0.2} catalyst had outstanding stability and good tolerance to SO₂ (Figure 6), which they attributed to the strong interactions among Fe, Mn, and Zr species. However, the durability of the catalyst in the presence of both SO₂ and H₂O need to be tested to investigate its feasibility in practical use.



Figure 6. (a) NO removal efficiency of $Fe_{0.3}Mn_{0.5}Zr_{0.2}$ as a function of time at 200 °C and (b) effect of SO₂ on the NO removal over $Fe_{0.3}Mn_{0.5}Zr_{0.2}$ at 200 °C. Reprinted from Reference [88]. Copyright 2017, with Permission from Elsevier.

Guo and co-workers [89] studied the effect of Sb doping on the activity of MnTiO_x catalyst for NH₃-SCR reaction. The results showed that Sb modification has greatly improved the NH₃-SCR performance of MnSbTiO_x catalysts in comparison to the MnTiO_x and SbTiO_x samples. Particularly, the MnSbTiO_x-0.2 (Sb/Mn molar ratio = 0.2) catalyst exhibited the best activity with above 90% NO_x conversion in the temperature range of 138–367 °C as it had good adsorption and activation properties for NH₃ and NO_x reactants in SCR. It can be seen from Figure 7 that the addition of Sb dramatically improved the SO₂ and H₂O resistance of MnTiO_x catalyst. Although the NO_x conversion over the MnSbTiO_x-0.2 catalyst slightly decreased in the presence of SO₂ and H₂O, it recovered to almost the original level after stopping the SO₂/H₂O supply.



Figure 7. SO₂ resistance of MnTiO_x and MnSbTiO_x-0.2 catalysts at 150 °C. Reaction conditions: 600 ppm NO, 600 ppm NH₃, 5% O₂, 5% H₂O, 100 ppm SO₂, balance Ar, GHSV = 108,000 h⁻¹. Reprinted from Reference [89]. Copyright 2018, with Permission from Elsevier.

Shi et al. [90] synthesized a series of $Ni_yCo_{1-y}Mn_2O_x$ microspheres (MSs) (y = 0.1, 0.3, 0.5, 0.7, 0.9) for NH₃-SCR using a hydrothermal method. It was observed that the activity of all ternary MSs was greater than binary CoMn₂O_x and NiMn₂O_x, and Ni_{0.7}Co_{0.3}Mn₂O_x showed the best NH₃-SCR performance among the Ni_yCo_{1-y}Mn₂O_x catalysts. Although the Ni_{0.7}Co_{0.3}Mn₂O_x catalyst exhibited

good resistance to H_2O , it had poor SO_2 tolerance which needs to be improved. Wu and co-workers [91] compared the DeNO_x performance of MnO₂/CoAl-LDO and CoMnAl-LDO mixed metal oxides prepared from CoAl-MnO₂-LDH and CoMnAl-LDH templates by ion-exchange/redox reaction and hexamethylenetetramine (HMT) hydrolysis methods, respectively. The CoAl-MnO₂-LDH showed higher NH₃-SCR activity in a broad temperature window of 90–300 °C (Figure 8a) as well as better stability and SO₂/H₂O resistance (Figure 8b) than the CoMnAl-LDO, which was attributed to its larger specific surface area, stronger redox ability, more quantitative acid sites, and abundant active components.



Figure 8. (a) NO_x conversion and N₂ selectivity over MnO₂/CoAl-LDO and CoMnAl-LDO catalysts prepared by calcination at 500 °C; and (b) the stability and SO₂/H₂O resistance test (inset) of MnO₂/CoAl-LDO and CoMnAl-LDO catalysts at 240 °C. Reaction conditions: [NH₃] = 600 ppm, [NO] = 600 ppm, [O₂] = 5 vol%, 100 ppm SO₂ (when used), 10 vol% H₂O. (when used) balanced by N₂. Reprinted from Reference [91]. Copyright 2019, with Permission from Elsevier.

Leng and collaborators [92] synthesized $Mn_{0.2}TiO_x$, $Ce_{0.3}TiO_x$ and series of $Mn_aCe_{0.3}TiO_x$ (a = 0.1, 0.2, 0.3) catalysts and investigated their applicability for low-temperature NH₃-SCR reaction. They demonstrated that the low-temperature NH₃-SCR activity of $Mn_aCe_{0.3}TiO_x$ was greatly improved after incorporation of Mn, and the $Mn_{0.1}Ce_{0.3}TiO_x$ catalyst displayed the best performance (with 100% NO conversion and above 90% N₂ selectivity) in the temperature range of 175–400 °C even at high GHSV of 80,000 h⁻¹. The outstanding performance of $Mn_{0.1}Ce_{0.3}TiO_x$ catalyst in NH₃-SCR resulted from its enhanced acidity and chemisorbed oxygen, and suitable redox property derived from Ce^{3+} + $Mn^{4+} \leftrightarrow Ce^{4+}$ + Mn^{3+} reaction. Furthermore, the NO conversion over the $Mn_{0.1}Ce_{0.3}TiO_x$ decreased and stabilized at 82% after the introduction of 100 ppm SO₂ and 6% H₂O and restored to almost 100% NO conversion after stopping the supply of SO₂ and H₂O (Figure 9), suggesting that the catalyst had excellent resistance to SO₂/H₂O and the effects were reversible.



Figure 9. Effect of H₂O and SO₂ on NO conversion over the $Mn_{0.1}Ce_{0.3}TiO_x$ catalyst at 200 °C (1000 ppm NO, 1000 ppm NH₃, 3% O₂, balance N₂, GHSV = 40,000 h⁻¹). Reprinted from Reference [92]. Copyright 2018, with Permission from Elsevier.

Ali et al. [93] developed a series of Nb-promoted $Fe_x-Nb_{0.5-x}-Ce_{0.5}$ (x = 0.45, 0.4, 0.35) oxides for NH₃-SCR. The best activity (>90% NO conversion and near 100% N₂ selectivity) in the broad temperature window of 180–400 °C as well as excellent SO₂/H₂O resistance (Figure 10) was observed for Fe_{0.4}-Nb_{0.1}-Ce_{0.5} catalyst. The authors considered the strong interaction among Nb, Fe, and Ce oxides leading to the enhancement of BET surface area, redox ability, acid amount, and NO adsorption capacity, which could be responsible for the outstanding performance of the catalyst.



Figure 10. Effect of SO₂, H₂O, and SO₂ + H₂O on NO conversion over the $Fe_{0.4}$ -Nb_{0.1}-Ce_{0.5} catalyst at 220 °C. Reprinted from Reference [93]. Copyright 2018, with Permission from Elsevier.

Sun and co-workers [94] reported the multimetallic Sm- and/or Zr-doped MnO_x-TiO₂ catalysts for NH₃-SCR reaction. As shown in Figure 11, the Sm and Zr co-doped MnO_x-TiO₂ (MSZTO_x) catalyst had better activity (\approx 100% NO conversion and >95% N₂ selectivity) in a wide temperature range (125–275 °C) with an excellent H₂O/SO₂ tolerance than the MSTO_x (MnO_x-SmO_x-TiO₂), MZTO_x (MnO_x-ZrO_x-TiO₂) and MTO_x (MnO_x-TiO₂) catalysts. The authors claimed that the enhanced redox properties and acidic sites play a crucial role in improving the NH₃-SCR performance of MSZTO_x catalyst. Yan and co-workers [95] fabricated Cu_{0.5}Mg_{1.5}Mn_{0.5}Al_{0.5}O_x catalyst from layered double hydroxides and found that it showed better activity in a wide temperature range together with superior SO₂ and H₂O tolerance than conventional Mn/ γ -Al₂O₃. The improved performance of Cu_{0.5}Mg_{1.5}Mn_{0.5}Al_{0.5}O_x was attributed to the high specific surface area, high reducibility of MnO₂ and CuO species, an abundance of acid sites, and the good dispersion of MnO₂ and CuO species.

Chen et al. [96] investigated the NH₃-SCR reaction over a series of $Co_{0.2}Ce_xMn_{0.8-x}Ti_{10}$ (x = 0, 0.05, 0.15, 0.25, 0.35, and 0.40) oxides catalysts and observed that the $Co_{0.2}Ce_{0.35}Mn_{0.45}Ti_{10}$ catalyst exhibited the best catalytic performance with 100% NO conversion and over 91% N₂ selectivity in a broad temperature window of 180–390 °C. Although NO_x conversion decreased to some extent after introducing SO₂ and H₂O, the $Co_{0.2}Ce_{0.35}Mn_{0.45}Ti_{10}$ catalyst showed excellent resistance to SO₂/H₂O with reversible inhibition effect (Figure 12). It was concluded that the interactions among Ce, Co, Mn, and Ti oxides led to more surface Brønsted acid and Lewis acid sites, NO_x adsorption sites and modest redox ability which could play a crucial role to improve the NH₃-SCR activity of $Co_{0.2}Ce_{0.35}Mn_{0.45}Ti_{10}$.



Figure 11. (a) NO conversions and (b) N₂ selectivities of the catalysts in the NH₃-SCR reaction as a function of temperature; (c) SO₂ (100 ppm) resistance tests, and (d) H₂O + SO₂ (2.5 vol%, 100 ppm) resistance tests at 200 °C over the catalysts. Reproduced from Reference [94]. Copyright 2018, with Permission from Elsevier.



Figure 12. Effects of H_2O and/or SO_2 on NO_x conversion over the $Co_{0.2}Ce_{0.35}Mn_{0.45}Ti_{10}$ catalyst. Reproduced from Reference [96]. Copyright 2018, with Permission from Elsevier.
4. Supported Single Transition Metal-Based Catalysts

Support materials have proved to be highly beneficial for enhancing the activity and durability of catalysts as they possess high surface area and good thermal stability. In virtue of the fine dispersion of the active component on the surface of the support and the synergistic effect between active component and support, supported catalysts exhibits improved NH₃-SCR performance than the unsupported transition metal oxide. Therefore, a lot of attention has been focused to increase the de-NO_x efficiency by dispersing the transition metal oxides over different support materials such as TiO₂, Al₂O₃, SiO₂, carbon nanotubes (CNTs), etc. TiO₂ supported transition metal oxides, especially manganese oxides, have been widely reported as promising catalysts for NH₃-SCR reaction at low temperature [17,24,97–105]. Smirniotis and co-workers [17,106,107] first reported the transition metal oxides (V, Cr, Mn, Fe, Co, Ni, and Cu) supported on Hombikat TiO2 for NH3-SCR reaction at low-temperature. Among the investigated samples, the Mn/Hombikat TiO₂ catalyst found to exhibit the highest activity even in the presence of water. They also studied the effect of different supports on the NH₃-SCR performance and observed that the Mn/Hombikat TiO₂ (anatase, high surface area) had the best activity as compared to the Kemira TiO₂ (rutile), Degussa P25 TiO₂ (anatase, rutile), Aldrich TiO₂ (anatase, low surface area), Puralox γ -Al₂O₃, Aldrich SiO₂ supported Mn catalysts. It was concluded that the Lewis acidity, redox behavior, and a high surface concentration of MnO2 could play a key role in improving the NH₃-SCR activity. Later, they investigated the effect of Mn loading on the NH₃-SCR performance of Mn/Hombikat TiO₂ and reported that the catalyst with 16.7 wt% Mn had optimal activity and excellent tolerance to H_2O during 10 days of the reaction [108]. In another work, they also proposed the NH₃-SCR reaction mechanism over the Mn/TiO₂ catalyst using transient isotopic labeled and in-situ FT-IR studies. As shown in Figure 13, the reaction proceeds via a Mars-van-Krevelen-like mechanism, in which NH3 and NO species were first adsorbed onto the Mn⁴⁺ sites (Lewis acid sites), followed by the formation of nitrosamide and azoxy intermediate species. Finally, these intermediates converted into N₂ and H₂O products [24].



Figure 13. Plausible SCR mechanism over the surface of Mn/TiO_2 catalyst. Adapted from Reference [24]. Copyright 2012, with Permission from Elsevier.

In the aspect of catalyst structure design, Smirniotis and co-workers [26] developed a series of manganese confined titania nanotube (Mn/TNT-X) catalysts using different TiO₂ precursors (X = Ishihara (I), Kemira (K), Degussa P25 (P25), Sigma–Aldrich (SA), Hombikat (H), TiO₂ synthesized from titanium oxysulfate (TOS)) for low-temperature NH₃-SCR reaction. As can be observed from the Figure 14, all Mn/TNT-X catalysts exhibited an excellent NO_x conversion in broad temperature window, and especially, Mn(0.25)/TNT-H sample obtained the superior activity in the temperature range of 100–300 °C as compared to other catalysts. They believed that the better performance of Mn(0.25)/TNT-H catalyst was due to the high surface area (421 m²/g) of the TNT-H support, and high dispersion of active components. The Mn(0.25)/TNT-H catalyst also showed greater catalytic performance than the conventional Mn-loaded titania nanoparticles (Mn/TiO₂), suggesting that the unique multiwall nanotube with open-ended structure could be advantageous to promote the reaction. Besides, the Mn (0.25)/TNT-H displayed outstanding tolerance to 10 vol% H₂O in the feed (Figure 15), which might be attributed to the preferential existence of highly active and redox potential pairs of Mn⁴⁺ and Mn³⁺ in the tubular framework.



Figure 14. Catalytic evaluation of the Mn(0.25)/TNT-X (X = Hombikat, Ishihara, P25 Degussa, Kemira, Sigma–Aldrich, and Titania oxysulfate) family of catalyst for the SCR of NO_x by NH₃, in the presence of 900 ppm NO, 100 ppm NO₂, 1000 ppm NH₃, 10 vol% O₂ with He balance under a GHSV of 50,000 h⁻¹ in the temperature range from 100–300 °C. Reproduced from Reference [26]. Copyright 2016, with Permission from Elsevier.



Figure 15. Influence of inlet water concentrations (10 vol%) on NO_x conversion in the SCR reaction over Mn(0.25)/TNT-H catalyst at 140 °C; feed: NO = 900 ppm, NO₂ = 100 ppm, NH₃/NO_x (ANR) = 1.0, O₂ = 10 vol%, He carrier gas, GHSV = 50,000 h⁻¹. Reproduced from Reference [26]. Copyright 2016, with Permission from Elsevier.

Recently, Boningari et al. [109] extended this work by comparing the NH_3 -SCR activity of various metal oxide confined titania nanotubes M/TNT (M = Mn, Cu, Ce, Fe, V, Cr, and Co) based on the Hombikat TiO₂ support. As shown in Figure 16, the Mn-, V-, Cr-, and Cu-oxide confined titania nanotubes had excellent low-temperature activity and meanwhile, vanadium oxide confined titania nanotubes showed a broad operation temperature window for the NH_3 -SCR reaction.

Sheng et al. [100] synthesized core-shell MnO_x/TiO_2 nanorod catalyst, showed high activity, stability, and N_2 selectivity in NH₃-SCR. They concluded that the abundant mesopores, Lewis-acid sites, and high redox capability could be beneficial to improve catalytic performance. Although the MnO_x/TiO_2 catalyst exhibited excellent resistance to H_2O , it was deactivated in the presence of SO₂ and SO₂/H₂O. Jia et al. [110] reported the low-temperature NH₃-SCR efficiency of MnO_x/TiO_2 , MnO_x/ZrO_2 , and MnO_x/ZrO_2 -TiO₂ catalysts, and found that MnO_x/ZrO_2 -TiO₂ obtained good activity at a temperature of 80-360 °C and excellent resistance to H_2O at 200 °C. However, all the catalysts showed poor tolerance to SO₂ and SO₂/H₂O that caused irreversible deactivation. Similar findings were also observed by Zhang et al. [111] over the Mn/Ti, Mn/Zr, and Mn/Ti-Zr catalysts, in which Mn/Ti-Zr sample exhibited an excellent NH₃-SCR performance in a wide temperature range due to its high surface area, Lewis acid sites, and surface Mn^{4+} ions.



Figure 16. Catalytic activity evaluation of metal oxides confined titania (made of Hombikat titania) nanotube catalytic formulations M/TNT where M = Mn, Cu, Ce, Fe, V, Cr, and Co for the selective catalytic reduction of NO_x by NH₃ in the presence of 900 ppm NO, 100 ppm NO₂, 1000 ppm NH₃, 10 vol% O₂ in He balance, under a GHSV = 50,000 h⁻¹. Reprinted from Reference [109]. Copyright 2018, with Permission from Elsevier.

Carbon nanotubes (CNTs) have been reported as promising catalyst support for NH₃-SCR catalysis due to their excellent stability and unique electronic and structural properties [36,112–115]. Qu and co-workers [116] reported that the NH₃-SCR performance of Fe₂O₃ was dramatically enhanced when it supported on CNTs (Figure 17a). It was concluded that the large surface area, fine dispersion of Fe₂O₃, and interaction between Fe₂O₃ and CNTs were important factors to improve the NH₃-SCR activity. In addition, the Fe₂O₃/CNTs catalyst showed an excellent tolerance to H₂O/SO₂. Interestingly, SO₂ stream in the feed had promoting effect on the NO conversion (Figure 17b), which could be attributed to the increased acid sites for NH₃ adsorption and activation on the catalyst surface in presence of SO₂. Bai et al. [117] developed CNTs supported copper oxide catalysts, and found that the 10 wt% CuO/CNTs showed good NH₃-SCR activity and excellent stability at 200 °C. The 10 wt% CuO/CNTs also had greater performance in comparison to 10 wt% CuO/TiO₂. However, it exhibited poor resistance to SO₂ and moderate tolerance to H₂O.



Figure 17. (a) NO_x conversion as a function of temperature over different catalysts and (b) SO_2/H_2O tolerance of the Fe₂O₃/CNTs catalyst. Reprinted from Reference [116]. Copyright 2015, with Permission from Royal Society of Chemistry.

5. Supported Binary and Multi Transition Metal-Based Catalysts

Given that the dispersion of two active components on support enhances the active sites further, researchers have been widely reported the supported binary transition metal-based oxides to improve the performance and SO_2/H_2O tolerance in NH₃-SCR reaction. With this perspective, several composites, such as MnCe/CNTs [118], Mn-Fe/TiO₂ [119], MnO_x-CeO₂/graphene [120], Mg-MnO_x/TiO₂ [121], CeO_x-MnO_x/TiO₂-graphene [122], Fe-Mn/Al₂O₃ [123], Mn-Fe/W-Ti [124], MnO_x-CeO₂/TiO₂-1%NG (NG = N-doped grapheme) [125], Mn-Ce/CeAPSO-34 [126], etc., were investigated for the NH₃-SCR reaction at low-temperature. Smirniotis et al. [22] studied the promotional effect of co-doped metals (Cr, Fe, Co, Ni, Cu, Zn, Ce, and Zr) on the NH₃-SCR performance of Mn/TiO₂. As shown in Figure 18, except Zn and Zr, all other co-doped metals had a positive impact on the activity of Mn/TiO₂, and particularly, the Mn-Ni/TiO₂ exhibited the highest NO conversion and N₂ selectivity among the other titania-supported bimetallic catalysts.



Figure 18. N₂ selectivity and catalytic performance of Mn-M'/TiO₂ anatage (M' = Cr, Fe, Co, Ni, Cu, Zn, Zr, and Ce) catalysts: NH₃ = 400 ppm; NO = 400 ppm; O₂ = 2.0 vol%; GHSV = 50,000 h⁻¹; catalyst wt. = 0.1 g; reaction temperature = 200 °C. Reproduced from Reference [22]. Copyright 2011, with Permission from Elsevier.

They also investigated the influence of Ni loading on the activity of Mn/TiO₂ catalyst, and found that the 5wt%Mn-2wt%Ni/TiO₂ (Mn-Ni(0.4)/TiO₂, where Ni/Mn = 0.4) had the optimal activity with complete NO conversion at the temperature range of 200–250 °C (Figure 19a) and outstanding stability even in the presence of 10 vol% water (Figure 19b,c) [5,23]. The enhanced reducibility of manganese oxide and dominant phase of MnO₂ claimed to be responsible for the best activity and stability of Mn–Ni/TiO₂ catalyst [5,22,23]. In another study, they compared the de-NO_x performance of high surface texture hydrated titania and Hombikat TiO₂ supported Mn-Ce bimetallic catalysts, and observed that the Mn–Ce/TiO₂ (Hombikat) showed the better activity and excellent resistance to H₂O (Figure 20). The superior performance could be attributed to the enhancement in reduction potential of active components, broadening of acid sites distribution, and the promotion of Mn⁴⁺/Mn³⁺, Ce³⁺/Ce⁴⁺ ratios including surface labile oxygen and small pore openings [25].



Figure 19. (a) Influence of Ni/Mn atomic ratio on NO conversion in the SCR reaction at a temperature range (160–240 °C) over Mn-Ni/TiO₂ catalysts (X_{NO} % = conversion of NO at 6 h on stream); (b) SCR of NO with NH₃ at 200 °C temperature over Mn/TiO₂ and Mn-Ni/TiO₂ catalysts; (c) Influence of inlet water concentrations (10 vol%) on NO conversion in the SCR reaction over Mn–Ni(0.4)/TiO₂ catalyst at 200 °C (GHSV = 50,000 h⁻¹; feed: NO = 400 ppm, NH₃ = 400 ppm, O₂ = 2 vol%, He carrier gas, catalyst = 0.1 g, total flow = 140 mL min⁻¹). Reprinted from Reference [5]. Copyright 2012, with Permission from Elsevier.



Figure 20. Influence of inlet water concentrations (7 vol%) on NO_x conversion in the SCR reaction over Mn–Ce(5.1)/TiO₂-Hk (Hombikat) catalyst at 175 °C; feed: NO = 900 ppm, NO₂ = 100 ppm NH₃/NO_x (ANR) = 1.0, O₂ = 10 vol%, He carrier gas, catalyst. 0.08 g, GHSV. 80,000 h⁻¹. Reprinted from Reference [25]. Copyright 2015, with Permission from Elsevier.

Xu and co-workers [127] reported Ce-Mn/TiO₂ catalysts with different Ce loadings, and the Ce(20)-Mn/TiO₂ found to show high activity with >90% NO conversion in the temperature range of 140–260 °C (Figure 21a). Their SO₂ tolerance results showed that the resistance ability was decreased in the order of Ce(20)-Mn/TiO₂ > Ce(30)-Mn/TiO₂ > Ce(10)-Mn/TiO₂ (Figure 21b). Although the Ce(20)-Mn/TiO₂ catalyst had reasonable resistance to 100 ppm SO₂ at different reaction temperatures (Figure 21c), it exhibited moderate tolerance to SO₂ poisoning when added higher than 100 ppm SO₂ to the reaction feed (Figure 21d). They ascribed the good SO₂ resistance of Ce(20)-Mn/TiO₂ to the widely distributed elements of Mn and Ce which in turn led to the inability of the sulfate material to remain on the surface.



Figure 21. (a) Catalytic activity of Ce-Mn/TiO₂ catalyst for NH₃–SCR. Catalysts were loaded with 10%, 20% and 30% Ce and denoted as Ce(10), Ce(20) and Ce(30), respectively. Pure TiO₂ was also used for comparison; (b) The effect of various Ce concentrations using a Ce-Mn/TiO₂ catalyst on SO₂ resistance; (c) The effects of reaction temperature on NO conversion of the Ce(20)-Mn/TiO₂ catalyst in the presence of SO₂. The above three types of reactions were performed at: 500 ppm NO, 500 ppm NH₃, SO₂ 100 ppm, 3% O₂, N₂ balance. gas, GHSV = 10 000 h⁻¹; and (d) the effects of SO₂ concentration on NO conversion of Ce(20)-Mn/TiO₂ catalysts (T = 180°C, 500 ppm NO, 500 ppm NH₃, 3%O₂, N₂ balance gas, GHSV = 10 000 h⁻¹). Adapted from Reference [127]. We thank the Royal Society Open Science for this Contribution.

Lin et al. [128] synthesized Me-Fe/TiO₂ (SD) catalyst via an aerosol-assisted deposition method, showing an excellent NH₃-SCR performance and good tolerance to SO_2/H_2O as compared to its counterparts prepared by co-precipitation and wet impregnation methods. The authors concluded that the enhanced surface reducibility and adsorption ability of NH₃/NO_x of Mn-Fe/TiO₂ (SD) catalyst could be responsible for its superior activity. Lee and co-workers [129] investigated the poisoning effect of SO₂ as metal sulfate and/or ammonium sulfate deposits on the low-temperature

NH₃-SCR activity of MnFe/TiO₂ catalysts. They found that the metal sulfates had a more serious deactivation effect than that of ammonium salts on the MnFe/TiO₂ catalysts. Their results showed that metal sulfates poisoning resulted in lower crystallinity, lower specific surface area, a lower ratio of Mn⁴⁺/Mn³⁺, higher surface acidity, and more chemisorbed oxygen, which in turn led to an adverse effect on the NH₃-SCR activity of the catalyst. Mu et al. [130] prepared Fe-Mn/Ti catalyst by ethylene glycol-assisted impregnation method, showing high NH₃-SCR efficiency over a broad temperature window (100–325 °C) and outstanding tolerance to sulfur poisoning. The formation of the Fe-O-Ti structure with strong interaction strengthened the electronic inductive effect and increased the ratio of surface chemisorption oxygen, thereby the enhancement of NO_x adsorption capacity and NO oxidation performance, which could be beneficial to improve the NH₃-SCR activity.

Liu and co-workers [131] reported that the addition of Eu had noticeably improved the NH₃-SCR performance of Mn/TiO₂ catalyst even after sulfation process under SCR conditions (Figure 22a). However, both the Mn/TiO₂ and MnEu/TiO₂ catalysts showed poor activity when they sulfated only with SO₂ + O₂ (Figure 22a). Further, the MnEu/TiO₂ catalyst found to show better SO₂ tolerance as compared to the Mn/TiO₂ (Figure 22b). Their results revealed that Eu modification could inhibit the formation of surface sulfate species on the Mn/TiO₂ catalyst during the NH₃-SCR in the presence of SO₂, which could be the reason for improved SO₂ resistance.



Figure 22. (a) NH₃-SCR activities of the fresh and sulfated catalysts; and (b) SO₂ tolerances of Mn/TiO_2 and $MnEu/TiO_2$ in NH₃-SCR reaction. Reprinted from Reference [131]. Copyright 2018, with Permission from Elsevier.

Sun et al. [132] investigated the NH₃-SCR activity over the Nb-doped Mn/TiO₂ catalysts with different Nb/Mn molar ratios, and found that the MnNb/TiO₂-0.12 (where Nb/Mn = 0.12) catalyst had optimal NO_x conversion and N₂ selectivity in the temperature range of 100–400 °C (Figure 23a,b). The optimal MnNb/TiO₂-0.12 catalyst also exhibited greater SO₂ resistance than Mn/TiO₂ catalyst (Figure 23c). The incorporation of Nb into Mn/TiO₂ catalyst led to increase surface acidity and reducibility as well as generate more surface Mn⁴⁺ and chemisorbed oxygen species along with more NO₂, which results in the better NH₃-SCR activity. In situ DRIFT studies over the Mn/TiO₂ and MnNb/TiO₂-0.12 catalysts disclosed that the NH₃-SCR took place through Eley–Rideal mechanism even in presence of SO₂, in which the reaction mainly occurred between adsorbed NO₂ and gaseous NH₃. Hence, it was concluded that the higher SO₂ tolerance of the MnNb/TiO₂-0.12 catalyst could be due to the existence of more adsorbed NO₂ on its surface.



Figure 23. (a) SCR activities and (b) N₂ selectivities over different catalyst samples as a function of reaction temperature. Reaction conditions: $[NO] = [NH_3] = 600 \text{ ppm}$, $[O_2] = [H_2O] = 5\%$, balance Ar, GHSV = 108,000 h⁻¹; (c) Effect of SO₂ on the SCR activities over Mn/TiO₂ and MnNb/TiO₂-0.12 catalyst samples. Reaction conditions: $[NO] = [NH_3] = 600 \text{ ppm}$, $[O_2] = [H_2O] = 5\%$, $[SO_2] = 100 \text{ ppm}$, balance Ar, GHSV = 108,000 h⁻¹, reaction temperature = 150 °C. Reproduced from Reference [132]. Copyright 2018, with Permission from Elsevier.

In another study, they reported the Mo-modified Mn/TiO_2 catalysts, exhibiting improved NH₃-SCR activity from 50 to 400 °C in comparison to Mn/TiO_2 catalyst. Particularly, the optimal MnMo/TiO₂-0.04 (where molar ratio of Mo/Mn = 0.04) catalyst better tolerance to SO₂ poisoning compared with Mn/TiO₂ (Figure 24) [133].

Fan and co-workers [134] fabricated ordered mesoporous titania supported CuO and MnO₂ composites (CuO/MnO₂-mTiO₂) through a facile acetic acid-assisted one-pot synthesis approach, showing high deNO_x efficiency (>90% NO conversion) and N₂ selectivity (>95%) in a wide operating temperature range of 120–300 °C. They considered that the superior NH₃-SCR performance could be attributed to the unique structure and highly integrated mesoporous TiO₂ supported by the multicomponent system with high surface areas, accessible and homogenously dispersed CuO and MnO₂ with multivalent nature and good redox activity. Although the CuO/MnO₂-mTiO₂ catalyst had good tolerance to H₂O, the resistance to SO₂ and H₂O/SO₂ poisoning, as well as high space velocity (GHSV), still need to be enhanced for practical use. Li et al. [135] synthesized fly ash-derived SBA-15 mesoporous molecular sieves supported Fe and/or Mn catalysts, and reported that Fe-Mn/SBA-15

catalyst showed notably greater NH₃-SCR activity than Mn/SBA-15 or Fe/SBA-15 in the temperature range of 150–250 °C. Moreover, the Fe-Mn/SBA-15 catalyst exhibited good time-on-stream stability (200 h) and water tolerance at 200 °C. The high metal dispersion, Mn^{4+}/Mn^{3+} ratio, the concentration of adsorbed oxygen, and the redox activity are important features to enhance the NH₃-SCR performance of the Fe-Mn/SBA-15 catalyst. In their subsequent study, the authors investigated the mechanisms of NO reduction and N₂O formation using in-situ DRIFT and transient reaction studies and proposed a possible denitration mechanism over the Fe-Mn/SBA-15 catalyst which is shown in Figure 25. The NH₃-SCR reaction over the Fe-Mn/SBA-15 catalyst proceeded through Langmuir-Hinshelwood, Eley-Rideal, and Mars-van Krevelen mechanisms. Their results also revealed that a large amount of nitrate thereby N₂O being produced over the Fe-Mn/SBA-15 during the reaction due to its strong oxidation ability, low acidity, and high basicity, which resulted in the lower N₂ selectivity [136].



Figure 24. SO₂ tolerance of Mn/TiO₂ and MnMo/TiO₂-0.04 catalysts at 150 °C, Reaction conditions: 600 ppm NO, 600 ppm NH₃, 100 ppm SO₂, 5% O₂, balance Ar, GHSV = 108,000 h⁻¹. Reprinted from Reference [133]. Copyright 2018, with Permission from Elsevier.



Figure 25. Low-temperature NH₃-SCR reaction mechanism on Fe-Mn/SBA-15 catalyst. Adapted from Reference [136]. Copyright 2018, with Permission from American Chemical Society.

Tang and co-workers [137] reported that Mn_2CoO_4 /reduced graphene oxide (Mn_2CoO_4 /rGO) catalyst with an optimal amount of $CoCl_2 \cdot 6H_2O$ of 0.3 (millimole) showed excellent NH₃-SCR activity and stability at low-temperature due to its large specific surface area, abundant Lewis acid sites,

and special three-dimensional architecture. When 100 ppm SO₂ added to reaction feed, the NO_x conversion over the optimal catalyst decreased significantly (96% to 53%), and the NO_x conversion was recovered to original level by water-washing after stopping the supply of SO₂. However, the decreased activity (100% to 82% NO_x conversion) in the presence of H₂O was restored to the original level after removing H₂O from the feed gas. Wang et al. [138] investigated the honeycomb cordierite-based Mn-Ce/Al₂O₃ catalyst for NH₃-SCR reaction and found that it showed good activity and reasonable resistance to SO₂/H₂O. The catalyst deactivation in the presence of SO₂ was ascribed to the deposition of ammonium hydrogen sulfate and sulfated CeO₂ on the catalyst surface during the NH₃-SCR process.

Meng et al. [139] synthesized a novel $CuAlO_x/CNTs$ (CNTs = carbon nanotubes) catalyst by facile one-step carbothermal reduction decomposition method for low-temperature NH₃-SCR. The $CuAlO_x/CNTs$ catalyst was found to exhibit higher NO_x conversion (>80%) and N₂ selectivity (>90%) than the $CuAlO_x$ in the temperature range of 180–300 °C (Figure 26a). They concluded that more favorable formation of Cu^+ active sites, better dispersion of active CuO species and higher surface adsorbed oxygen were beneficial to enhance the NH₃-SCR activity of $CuAlO_x/CNTs$ catalyst. As shown in Figure 26b, the $CuAlO_x/CNTs$ catalyst displayed excellent resistance to SO_2/H_2O at 240 °C during the NH₃-SCR. The authors attributed this outstanding SO_2/H_2O tolerance to the presence of CNTs that could promote the reaction of NH₄HSO₄ and NO continuously to avoid the formation and accumulation of excess ammonium sulfate salts on the catalyst surface. Li group [140] reported a series of ultra-low content copper-modified TiO₂/CeO₂ catalysts and observed that the catalyst with a Cu/Ce molar ratio of 0.005 exhibited the high NH₃-SCR performance and good tolerance to SO_2 . Their characterization results disclosed that the addition of Cu into TiO₂/CeO₂ lead to enhance the Brønsted acid sites, amount of surface adsorbed oxygen and Ce³⁺ species, redox, and surface acidic properties, which in turn improve the NH₃-SCR activity.



Figure 26. (a) NH₃-SCR activity and N₂ selectivity as a function of temperature from 150 °C to 330 °C; and (b) SO₂/H₂O resistance test of CuAlO_x/CNTs catalyst at 240 °C. Reaction conditions: 600 ppm NH₃, 600 ppm NO, 5.0 vol% O₂, 100 ppm SO₂ (when used), 10 vol% H₂O (when used) balanced by N₂ with a GHSV was 45,000 h⁻¹. Reprinted from Reference [139]. Copyright 2019, with Permission from Elsevier.

Recently, supported multi-metal oxide catalysts have been considered as the very promising candidates for low-temperature NH₃-SCR reaction because of the enlarged synergetic catalysis effects of different components as well as improved metal-support interactions [141–145]. Wang and co-workers [146] reported a series of Nb modified Cu-Ce-Ti mixed oxide (Nb_yCCT, where y represented the atomic ratio of Nb to Ti) catalysts for low-temperature NH₃-SCR reaction. It was found that Nb_yCCT catalysts demonstrated a better activity than the Cu-Ce-Ti (CCT) and Ce-Ti (CT) samples (Figure 27a). Among all the Nb_yCCT catalysts, Nb_{0.05}CCT showed a higher NO conversion (>90%) in a broad temperature range of 180–360 °C under the GHSV of 40,000 h⁻¹ (Figure 27a). Results indicated that the incorporation of Nb to Cu-Ce-Ti led to strong interactions among the active phases that increased the oxygen vacancies and inhibited the over-oxidation of NH₃, which in turn improved the NH₃-SCR activity and N₂ selectivity in a wide temperature window. DRIFTS studies revealed

that the introduction of Nb promoted the generation of NO₂, which could improve the activity via "fast" SCR reaction process (Langmuir–Hinshelwood reaction pathway). As shown in Figure 27b, the optimal Nb_{0.05}CCT catalyst exhibited higher resistance to SO₂/H₂O as compared to the Nb free catalyst. Li et al. [147] investigated the effect of Ho doping on the NH₃-SCR performance and the SO₂/H₂O resistance of Mn-Ce/TiO₂ catalyst. Among the catalysts tested, the catalyst with Ho/Ti of 0.1 (Mn_{0.4}Ce_{0.07}Ho_{0.1}/TiO₂) showed the best performance with >90% NO conversion in the temperature range of 150–220 °C, which was attributed to high concentration of chemisorbed oxygen, surface Mn⁴⁺/Mn³⁺ ratio, and acidity, as well as large specific surface area. Although the Mn_{0.4}Ce_{0.07}Ho_{0.1}/TiO₂ showed higher resistance to SO₂ and H₂O than the Mn_{0.4}Ce_{0.07}/TiO₂ catalyst, it was deactivated some extent in presence of SO₂/H₂O which is irreversible.



Figure 27. (a) NH₃-SCR activities of different catalysts {Reaction conditions: $[NH_3] = [NO] = 600 \text{ ppm}$, $[O_2] = 3\%$, $[H_2O] = 5 \text{ vol}\%$, N_2 as balance. GHSV = 40 000 h⁻¹}; and (b) SO₂ and H₂O resistance of the catalysts at 250 °C in SCR reaction process {Reaction conditions: $[NH_3] = [NO] = 600 \text{ ppm}$, $[O_2] = 3\%$, $[SO_2] = 50 \text{ ppm}$, $[H_2O] = 5 \text{ vol}\%$, N_2 balance, GHSV = 40 000 h⁻¹}. Reproduced from Reference [146]. Copyright 2018, with Permission from Elsevier.

Lu group [148] synthesized a series of activated coke (AC) supported Fe_xCo_yCe_zO_m catalysts for low-temperature NH₃-SCR, and found that the 3%Fe_{0.6}Co_{0.2}Ce_{0.2}O_{1.57}/AC catalyst had the best activity at 250–350 $^\circ\text{C}$ and good tolerance to $H_2\text{O}/\text{SO}_2$ at 250 $^\circ\text{C}.$ The superior performance of the catalyst was ascribed to the co-participation of Fe, Co, and Ce species with different valence states, high concentration of chemisorbed oxygen, well dispersed active components, increase of weak acid sites, good redox properties of metallic oxides, and abundant functional groups on the catalyst surface. Their mechanistic and kinetic studies also indicated that the enhanced active sites for the adsorption of NO and NH₃, and the redox cycle among Fe, Co and Ce were responsible for the improved activity. Zhao et al. [149] reported a series of Mn-Ce-V-WO_x/TiO₂ composite oxide catalysts, exhibiting greater NH₃-SCR activity than the TiO₂ supported single-component catalysts (Figure 28a,b). Particularly, the catalyst with a molar ratio of active components/ $TiO_2 = 0.2$ showed the best performance (>90% NO conversion) from 150 to 400 °C (Figure 28a). As shown in Figure 28c, the optimal Mn-Ce-V-WO_x/TiO₂ (molar ratio of Mn-Ce-V-WO_x/TiO₂ = 0.2) showed excellent stability and outstanding tolerance to H₂O/SO₂ at 250 °C. The authors concluded that the better performance of Mn-Ce-V-WO_x/TiO₂ mainly attributed to the variety of valence states of the four active components and their high oxidation-reduction ability.



Figure 28. (A) Selective catalytic reduction (SCR) activity of Mn-Ce-V-WO_x/TiO₂ composite catalysts with molar ratio of active components/TiO₂ at different values; (a) 0.1; (b) 0.2; (c) 0.3; (d) 0.6; (**B**) SCR activity of V₂O₅/TiO₂, WO₃/TiO₂, MnO₂/TiO₂, CeO₂/TiO₂ and TiO₂. Reaction conditions: [NO] = [NH₃] = 1500 ppm, O₂ = 3%, gas hourly space velocity (GHSV) = 40,000 h⁻¹; and (**C**) the lifetime of Mn-Ce-V-WO_x/TiO₂ catalyst with molar ratio of 0.2 at 250 °C: inset (a-c) H₂O and SO₂ resistance at 250 °C. Reaction conditions: [NO] = [NH₃] = 1500 ppm, [O₂] = 3%, [SO₂] = 100 ppm, GHSV = 40,000 h⁻¹. Adapted from Reference [149].

6. Conclusions

Since the emission standards for NO_x are becoming more stringent to keep our atmosphere clean, the widespread use of fossil fuel in automobiles and industries require advanced catalytic materials for NO_x emission control. The low-temperature SCR of NO_x with NH₃ would be a promising solution to mitigate the NO_x emissions from mobile and stationary sources. Hence, the development of efficient catalysts for the low-temperature NH₃-SCR with high deNO_x activity, N₂ selectivity, and high resistance toward SO_2/H_2O poisonings is the subject of increasing interest in the field of environmental catalysis. Transition metal-based oxide catalysts have drawn much attention for low-temperature NH₃-SCR due to their excellent redox properties, high activity, durability, and relatively low manufacturing costs. In this review, we have summarized the recent progress in the low-temperature NH₃-SCR technology over the various transition metal-based catalysts. Over the past decades, significant research efforts have been made to improve the de-NO_x efficiency and SO₂/H₂O tolerance of transition metal-based oxides in NH₃-SCR at low-temperatures. Various transition metal-based mixed oxides with and without support have been extensively studied for NH₃-SCR reaction and, particularly, MnOx-based catalyst formulations have caught much attention because of their excellent de-NO_x efficiency at low-temperatures. The modification of transition metal oxides by doping with other metal oxides led to high redox ability and acidic sites, and consequently, better NH₃-SCR performance at low-temperature. The loading of single and multi-transition metal-based oxides on the surface of supports (TiO₂, TiO₂ nanotubes, carbon nanotubes, etc.) could also enhance the NO_x conversion and N₂ selectivity in NH₃-SCR reaction by the fine dispersion of active component/s and its/their strong interaction with the support. The choice of metal loading and the support

could play a key role in the catalytic function of the supported transition metal-based catalysts. The synergistic redox interaction between the active components of mixed metal oxide/supported metal oxide catalysts was also found to be an important factor to design the efficient denitration catalysts. In spite of the significant progress on the SO_2/H_2O tolerance of the catalysts, the durability of catalysts in the presence of both SO₂ and H₂O still needs to be improved. Most transition metal-based catalysts suffered from low resistance when the reaction feed contains both SO_2 and H_2O streams simultaneously. Hence, researchers have continuously explored the different options of transition metal-based mixed oxides and active transition metal/s-support combinations in order to develop the better NH_3 -SCR catalysts in terms of SO_2/H_2O tolerance at low-temperature. The understanding of the inhibition mechanism of SO_2 and H_2O could be a promising strategy to develop high SO_2/H_2O resistance catalysts for NH₃-SCR reaction. However, the SO₂/H₂O inhibition mechanism was not very clear that needs to be investigated deeply. Especially, the design of transition metal-based catalysts with a combination of high NOx conversion and N2 selectivity in a wide operation temperature window and good resistance to SO₂/H₂O have attracted paramount attention, but it is still challenging task. The scope of NH₃-SCR research is quite vast and a large number of improvements need to be achieved in the near future.

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Review



Hydrogenation of Carbon Dioxide to Value-Added Chemicals by Heterogeneous Catalysis and Plasma Catalysis

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Abstract: Due to the increasing emission of carbon dioxide (CO_2), greenhouse effects are becoming more and more severe, causing global climate change. The conversion and utilization of CO_2 is one of the possible solutions to reduce CO_2 concentrations. This can be accomplished, among other methods, by direct hydrogenation of CO_2 , producing value-added products. In this review, the progress of mainly the last five years in direct hydrogenation of CO_2 to value-added chemicals (e.g., CO, CH_4 , CH_3OH , DME, olefins, and higher hydrocarbons) by heterogeneous catalysis and plasma catalysis is summarized, and research priorities for CO_2 hydrogenation are proposed.

Keywords: carbon dioxide; hydrogenation; heterogeneous catalysis; plasma catalysis; value-added chemicals; methanol synthesis; methanation

1. Introduction

Climate changes are mostly induced by the greenhouse effect, and carbon dioxide (CO_2) accounts for a dominant proportion of this greenhouse effect. Indeed, one can barely ignore the connection between the emission of CO_2 and climate changes [1]. The CO_2 concentration in the atmosphere has actually climbed to 405 ppm in 2017, as shown in Figure 1a. As the global energy consumption is still mainly based on burning coal, oil, and natural gas, and this situation will last until the middle of the century (see Figure 1b), experts predict that the CO_2 concentration in the atmosphere will continue to rise to ~570 ppm by the end of the century if no measures are taken [2]. Hence, it is urgent to control the CO_2 emissions by taking effective measures to capture and utilize CO_2 .

In principle, there are three strategies to reduce CO_2 emissions, i.e., reducing the amount of CO_2 produced, storage of CO_2 , and utilization of CO_2 [3]. Recently, some comprehensive reviews have discussed the technological state-of-the-art of carbon capture and storage (CCS) [4–6]. Direct air capture technology (DAC) draws people's attention to mitigate climate change by taking advantage of chemical sorbents (e.g., basic solvents, supported amine and ammonium materials, etc.) [4]. In addition, Bui et al. also considered the economic and political obstacles in terms of the large-scale deployment of CCS [6]. Significantly reducing the amount of CO_2 produced is unrealistic in view of the current energy structure dominated by fossil energies, as shown in Figure 1b. CO_2 storage seems to be a potential approach, but there are some challenges, such as efficiency of capture and sequestration of CO_2 , cutting down the operation costs for capture and separation of CO_2 , and the long-term stability of underground storage [7,8]. Utilization of CO_2 is a promising approach, since CO_2 is a cheap and

attractive carbon source, which can be used to yield a variety of industrial raw materials, which can be further converted into value-added chemicals and fuels. Interestingly, CO_2 can also be used as a desired trigger for stimuli-responsive materials. Darabi et al. summarized the synthesis, self-assembly, and applications of CO_2 -responsive polymeric materials [9].



Figure 1. (a) Trends in atmospheric CO₂ concentrations (ppm). (b) Projected global energy consumption (Mt) from 1990 to 2040. Data from International Energy Agency.

One of the options to convert CO_2 , is catalytic hydrogenation, as illustrated in Figure 2. The production of H_2 , however, is also a vital problem to realize direct hydrogenation of CO_2 to oxygenates and hydrocarbons. Hydrogen production can be either based on renewable or non-renewable sources, such as electrical, thermal, photonic, and hybrid [10,11]. The main methods of hydrogen production include electrolysis, thermolysis, photo-electrolysis, and hybrid thermochemical cycles [10]. According to some evaluation criteria—such as global warming potential (GWP), social cost of carbon (SCC), acidification potential (AP), energy and exergy efficiencies, and production cost—the hybrid hydrogen production methods seems a promising route. On the other hand, the production cost evaluation shows that coal gasification ($\$0.92/kg H_2$) and fossil fuel reforming ($\$0.75/kg H_2$) are relatively low-cost methods compared to early R&D phase methods (e.g., photo-electrochemical: $\$10.36/kg H_2$ from water dissociation). However, fossil fuel is non-renewable, being a major drawback. Therefore, reducing the cost of hydrogen production is an urgent and challenging issue, since we have to balance both the economy and sustainability.



Figure 2. Catalytic hydrogenation of carbon dioxide.

Various methods have been adopted—including photo-catalysis, electro-catalysis, heterogeneous catalysis, and plasma catalysis [12,13]—to realize hydrogenation of CO₂. Dalle et al. systematically

presented the activity of the first-row transition metal complexes (i.e., Sc, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn) in CO₂ reduction by electro-catalysis, photo-catalysis and photoelectron-catalysis [14]. Li et al. discussed the important roles of co-catalysts (e.g., biomimetic, metal-based, metal-free, and multifunctional) in selective photo-catalytic CO₂ reduction [15]. Besides thermochemical approaches, Mota et al. made a detailed retrospective based on electrochemical and photo-chemical approaches for CO₂ hydrogenation to oxygenates and hydrocarbons [16]. In this review, we focus on the latter two methods. Heterogeneous catalysis has been widely studied, while plasma catalysis is still an emerging technology. Some excellent reviews were recently published for heterogeneous catalytic CO₂ hydrogenation [12,17,18]. Jadhav et al. focused on methanol production [17], Porosoff et al. on the synthesis of CO, CH₃OH, and hydrocarbons [18], while Alvarez et al. discussed the greener preparation of formates (formic acid), CH₃OH, and DME [12]. With regard to plasma catalysis for CO₂ hydrogenation, there are only a handful reports (as discussed below). Nevertheless, it exhibits great potential since plasma can operate at ambient temperature and atmospheric pressure.

In this review, we summarize the progress of mainly the last five years in CO_2 hydrogenation to value-added chemicals (e.g., CO, CH₄, CH₃OH, DME, olefins, and higher hydrocarbons) driven by both heterogeneous catalysis and plasma catalysis. The literature bibliography range is shown in Figure 3. Indeed, on the one hand, the insights obtained by heterogeneous catalysis can be useful for the further development of the emerging field of plasma catalysis. On the other hand, we also want to pinpoint the differences between heterogeneous and plasma catalysis, and thus the need for dedicated design of catalytic system tailored to the plasma environment.



Figure 3. Literature bibliography included in this review.

2. Heterogeneous Catalysis

In heterogeneous catalysis, support materials, active metals, promoters, and the preparation methods of catalysts are major adjustable factors determining the catalytic activity. As far as the support materials are concerned, the catalytic performance is influenced by the metal-support interaction since it usually induces specific physicochemical properties for catalysis, e.g., active cluster, active oxygen vacancy, and acid-based property. Metal oxides and zeolites with special channel structures are usually selected. For the active metal components, research focuses on searching cheap and available metals to replace precious metals or to reduce the amount of precious metals in industrial heterogeneous catalysis. Promoters (structural-type and electron-type) can also significantly influence the catalytic activity by regulating the adsorption and desorption behavior of molecules (reactant, intermediate, and product) on the catalyst surface. Preparation methods usually determine the catalyst morphology including metal dispersion (particle size distribution), specific surface area, and channel structure,

which also influences the catalytic performance [12]. In this section, we summarize recent progress of CO_2 hydrogenation to CO, CH_4 , CH_3OH , and some other products in terms of rational design of heterogeneous catalysts.

2.1. CO₂ to CO

CO can be used as feedstock to produce liquid fuels and useful chemicals by Fischer–Tropsch (F–T) synthesis reaction. Therefore, the catalytic hydrogenation of CO_2 to CO via the reverse water–gas shift (RWGS) reaction, reaction (1), is promising to account for the shortage of energy. Accordingly, catalyst design for the RWGS reaction have attracted extensive attention.

$$CO_2 + H_2 \leftrightarrow CO + H_2O_{\ell} \bigtriangleup H_{298K} = 41.2 \text{ kJ mol}^{-1}$$
(1)

Generally, noble metal catalysts (e.g., Pt, Ru, and Rh) exhibit effective ability towards H₂ dissociation, and thus precious metals have been investigated extensively for the RWGS reaction. However, these catalysts yield, methane as dominant product, mainly caused by the higher rate of C-H bond formation than CO desorption [19]. To change the product distribution, Bando et al. applied Li-promoted Rh ion-exchanged zeolites for CO₂ hydrogenation, achieving 87% CO selectivity for an atomic ratio of Li/Rh higher than 10/1 [19]. Although the catalytic performance of noble metals can be improved via promoters, the high price and instability of noble metals (aggregation of nano-particles) limit the industrial applicability. In order to properly decrease the operation cost and improve the life cycle of catalysts, non-noble metal carbides have been developed. Porosoff et al. synthesized a molybdenum carbide (Mo₂C) catalyst for the RWGS reaction [20], yielding 8.7% CO₂ conversion and 93.9% CO selectivity (at 573 K reaction temperature). The catalytic performance was much better than those of bimetallic catalysts (e.g., Pt-Co, Pt-Ni, Pd-Co, Pd-Ni) supported on CeO₂ (~5% CO₂ conversion and 83.3% CO selectivity at the same reaction temperature). The catalytic mechanism of Mo₂C in the RWGS reaction was further studied by ambient pressure X-ray photoelectron spectroscopy (AP-XPS) and in situ X-ray absorption near edge spectroscopy (XANES), which demonstrate that Mo₂C not only broke the C=O bond, but also dissociated hydrogen. Hence, Mo₂C has a dual functional and is an ideal catalytic material for the RWGS reaction. Interestingly, the authors also found that by modifying the catalyst with Co, the catalytic activity of the Co-Mo₂C catalyst was further improved (9.5% CO₂ conversion, 99% CO selectivity at 573 K), probably attributed to the existence of the $CoMoC_vO_z$ phase confirmed by X-ray Diffraction (XRD) (see Figure 4).



Figure 4. XRD pattern of fresh and reduced Co-Mo₂C with Mo₂C as a reference. The symbols correspond to the following: $\bigtriangledown -\beta$ -Mo₂C, \diamond —CoMoC_yO_z, \blacklozenge —MoO₂, reproduced with permission from [20]. Copyright Wiley-VCH, 2014.

The support plays an important role in CO₂ hydrogenation to CO through the RWGS reaction. Kattel et al. prepared Pt, Pt/SiO₂ and Pt/TiO₂ materials as catalysts for the RWGS reaction [21], and showed that Pt nanoparticles alone cannot catalyze the RWGS reaction, while using SiO₂ and TiO₂ as support, the overall CO₂ conversion can be significantly improved, pointing towards a synergy effect between Pt and the oxide support. To reveal the synergy effect, they combined density functional theory (DFT), kinetic Monte Carlo (KMC) simulations and other experimental measurements. They found that the hydrogenation of CO₂ to CO is promoted once CO₂ is stabilized by the Pt-oxide interface. In the case of Pt nanoparticles (NP) alone, the conversion was close to 0, since the ability of Pt NP in binding CO₂ is weak. When SiO₂ and defected TiO₂ with oxygen vacancies served as supports, however, the CO₂ conversion was enhanced (to 3.35% and 4.51%, respectively) on the Pt-oxide interface. The synergy effect between Pt and oxide supports in activating and hydrogenating CO₂ is shown in Figure 5, and possible reaction pathway [21] for the RWGS reaction is illustrated in Scheme 1.



Figure 5. DFT optimized geometries. (1) (a) Pt_{25} /hydroxylated SiO₂ (111), and (b) *CO₂ species, (c) *CO species, (d) *HCO species, (e) *H₂COH species, (f) *CH₃OH species, (g) *CH₂ species, and (h) *OH species adsorbed on Pt/SiO₂ (111). The dashed lines show hydrogen bonds. (2) (a) Pt_{25} /TiO₂ (110) with oxygen vacancy, and side (top) and top (bottom) views of (b) *CO₂ species, (c) *CO species, (d) *HCO species, (e) *H₂COH species, (f) *CH₃OH species, (g) *CH₂ species, and (h) *OH species adsorbed on Pt/TiO₂ (110). The black circle in (a) depicts the position of oxygen vacancy on TiO₂ (110). Note: Si: green, Ti: light blue, Pt: light gray, C: dark gray, O: red, and H: blue, reprinted with permission from [21]. Copyright Elsevier, 2016.

RWGS: $CO_2(g) \longrightarrow ^{\circ}CO_2 \xrightarrow{+H} ^{\circ}HOCO \longrightarrow ^{\circ}CO + ^{\circ}OH \longrightarrow ^{\circ}CO + ^{*}H_2O$

Scheme 1. Reaction pathways for the RWGS reaction, where '*X' represents species X adsorbed on a surface site, reproduced with permission from [21]. Copyright Elsevier, 2016.

Yan et al. investigated the effect of the Ru-Al₂O₃ interfaces on the catalytic activity of the RWGS reaction, and proposed Ru₃₅/Al₂O₃ and Ru₉/Al₂O₃ catalyst models to explain the experimental observations [22]. The product selectivity switched between CH₄ and CO over the Ru/Al₂O₃ catalyst, i.e., monolayer Ru sites favored the production of CO, while Ru nano-clusters preferred the production of CH₄, during CO₂ reduction reaction. Confirmed by kinetic analysis, characterization of the surface structures and real-time monitoring of the active intermediate species, the product selectivity of CH₄ and CO was regulated by the Ru sites and Ru-Al₂O₃ interfacial sites. Furthermore, based on the combination of theoretical calculations and isotope-exchange experimental results, the authors found that the O* species derived from the dissociative adsorption of CO₂ at interfacial Ru sites easily bridge with the Al sites from the γ -Al₂O₃ support, and new Ru-O-Al bonds was responsible for the CO₂ activation via oxygen-exchange with the O atoms of CO₂. Therefore, this experimental work is a good inspiration to further explore the influence of metal-support interfaces for the effective activation of CO₂.

Besides the widely used impregnation method, Yan et al. also reported a doping-segregation method for the preparation of Rh-doped SrTiO₃ [23]. Precursors with a molar ratio of Sr:Ti:Rh = 1.10:0.98:0.02. First, TiO₂ was suspended in deionized water, and then Sr(OH)₂·8H₂O and Rh(NO₃)₃ were introduced. Subsequently, the sample was poured in a stain steel acid digestion vessel, which was kept at 473 K for 24–48 h. Finally, the reaction product was dried at 343 K overnight after it was centrifuged, and washed via deionized water. As confirmed by in situ X-Ray-Diffraction (XRD) and X-ray Absorption Fine Structure (XAFS) measurements, the Rh-doped SrTiO₃ catalysts produce sub-nanometer Rh clusters, which are highly active for the conversion of CO₂ compared to the supported Rh/SrTiO₃ prepared by wetness impregnation. The better catalytic performance (7.9% CO₂ conversion and 95% CO selectivity at 573 K) could be ascribed to the cooperative effect between sub-nanometer Rh clusters and the reconstructed SrTiO₃ which is active for dissociation of H₂ and is favorable for adsorption/activation of CO₂. Therefore, the novel approach, doping-segregation method, maybe a novel strategy to tune the size of active metals and the physicochemical properties of supports for rational design of catalysts for the RWGS reaction.

Dai et al. studied CeO₂ catalysts which were prepared by the hard-template method (Ce-HT), the typical complex method (Ce-CA), and the typical precipitation method (Ce-PC) for the RWGS reaction [24]. The experimental results show that catalysts prepared by Ce-CA, Ce-HT, and Ce-PC methods exhibit a 100% CO selectivity and the CO₂ conversions were 9.3, 15.9, and 12.7% respectively at 853 K. Obviously, the hard-template (Ce-HT) method is beneficial for the RWGS reaction in view of the catalytic performance. The Ce-HT method comprises the following steps: (1) Ce(NO₃)₃· $6H_2O$ was dissolved in ethanol, and KIT-6 mesoporous silica was added; (2) The mixture was stirred until a dry power was obtained; (3) The powder was calcined; (4) The obtained samples were treated with NaOH to remove the template. To reveal the relationship between the preparation method and the catalytic activity, the authors carried out XRD, Transmission Electron Microscopy (TEM) and Brunauer-Emmett-Teller (BET) characterization, and the characterization results show that CeO₂ catalysts which were prepared by the Ce-HT method have a porous structure and a high specific surface area, while CeO₂ catalysts prepared by the other methods (Ce-CA, Ce-PC) have an agglomerated structure (Ce-CA) and overlapped bulk structure (Ce-PC) with low porosity. Moreover, in the CeO₂ catalysts, oxygen vacancies as active sites were formed by H₂ reduction at 673 K, which were confirmed by in situ X-ray photoelectron spectroscopy (XPS) and H₂-temperature-programmed reduction (H₂-TPR). Therefore, the improvement of the catalytic activity for the RWGS reaction can be ascribed to the change of catalyst structure and oxygen vacancies.

Except for the above monometallic catalysts, some bimetallic catalysts—i.e., Pt-Co, Fe-Mo and Ni-Mo [25–27]—were also prepared and tested in the RWGS reaction. In general, the preparation method of bimetallic catalysts generally can be divided into two categories: (1) the bi-metal was first prepared and then supported on the carrier [25]; (2) the bi-metal as formed in the preparation progress [26,27]. Compared with pure Co catalyst, Pt-Co catalyst showed a better catalytic activity (mainly CO, close to 100%) for the RWGS reaction. Ambient pressure X-ray photoelectron spectroscopy (AP-XPS) and environmental transmission electron microscopy (ETEM) showed that Pt migrates on the catalyst surface, and Pt aids the reduction of Co to its metallic state under appropriate reaction conditions confirmed by near edge X-ray absorption fine structure (NEXAFS) spectroscopy [25]. Abolfazl et al. investigated Mo/Al₂O₃ and Ni-Mo/Al₂O₃ catalysts prepared by the impregnation method used for the RWGS reaction. The experimental results showed that Ni-Mo/Al₂O₃ catalyst is a promising catalyst (35% CO₂ conversion, i.e., close to the 38% equilibrium conversion) compared with Mo/Al₂O₃ catalyst (15% CO₂ conversion). As demonstrated by XPS, the electronic effect, which transfers electrons from Ni to Mo and leads to an electron-deficient state of the Ni species, is beneficial for CO₂ adsorption, and thus improves the CO₂ conversion. Furthermore, the XRD and H₂-TPR profiles indicated that the Ni–O–Mo structure crystallizes into the NiMoO₄ phase which can improves the adsorption and dissociation of H₂ on the Ni-Mo/Al₂O₃ catalyst [27]. Similarly, they also found that Fe-Mo/Al₂O₃ catalyst synthesized by the impregnation method is an efficient catalyst with high CO yield, almost no by-products and relatively stable (60 h) for the RWGS reaction. The enhancement of the catalytic activity may be attributed to better Fe dispersion with the addition of Mo and smaller particle size of active Fe species, which was confirmed by the BET method and scanning electron microscopy (SEM) [26].

Overall, to some extent, the reaction activity and stability of bimetallic catalysts for the direct hydrogenation of CO₂ to CO are excellent compared to mono-metallic catalysts. The reduction of the active metal, the formation of alloy metal, the dispersion and particle size of the catalyst are possible factors for the improvement of the catalytic performance in the RWGS reaction. Details of the conversion and product selectivity, along with the reaction conditions of several representative RWGS catalytic systems, are compared in Table 1. Obviously, precious metal catalysts (e.g., Pt, Rh, Ni) are still advantageous for the formation of CO compared to non-noble metal catalysts (e.g., Fe, Co, Mo), and upon increasing the gas hourly space velocity (GHSV), the selectivity of CO slightly decreases to some extent.

Catalyst	H ₂ :CO ₂	GHSV	Temperature (°C)	Pressure (MPa)	CO ₂ Conversion (%)	CO Selectivity (%)
Mo ₂ C [20]	3	36,000 ^a	300	0.1	8.7	93.9
Co-Mo ₂ C [20]	3	36,000 ^a	300	0.1	9.5	~99.0
Pt-TiO ₂ [21]	1	119.7 ^b	300	0.1	4.5	99.1
Pt-SiO ₂ [21]	1	24.7 ^b	300	0.1	3.3	100
Rh-SrTiO ₃ [23]	1	12,000 ^a	300	N/A	7.9	95.4
Co-MCF-17 [25]	3	60,000 ^b	200-300	0.5	~5.0	~90.0
Pt-Co-MCF-17 [25]	3	60,000 ^b	200-300	0.5	~5.0	~99.0
Fe-Mo-Al ₂ O ₃ [26]	1	30,000 ^a	600	1	~45.0	~100.0
Mo-Al ₂ O ₃ [27]	1	30,000 ^a	600	0.1	16	N/A
Ni-Mo-Al ₂ O ₃ [27]	1	30,000 ^a	600	0.1	34	N/A
La-Fe-Ni [28]	2	24,000 ^a	350	N/A	16.3	96.6

Table 1. Catalytic performance of several catalytic systems for CO_2 hydrogenation into CO, in terms of CO_2 conversion and CO selectivity, along with the reaction conditions

^a mL g_{cat}^{-1} h⁻¹; ^b h⁻¹; N/A: not available.

2.2. CO₂ to CH₄

Methane, a high value carbon source, is used to produce syngas via steam reforming, and subsequently the syngas is usually converted into chemicals and/or fuels through F–T synthesis.

The direct hydrogenation of CO_2 to CH_4 (also called CO_2 methanation), reaction (2), is a feasible approach, if the production technology of H_2 becomes widespread and at low-cost.

$$CO_2 + 4H_2 \leftrightarrow CH_4 + 2H_2O_1 \bigtriangleup H_{298K} = -252.9 \text{ kJ mol}^{-1}$$
 (2)

Ni-based [29], Co-based [30], and Ru-based [31] catalysts have been used in CO₂ methanation, and Ni-based catalysts are considered to be the most effective and the lowest cost alternative. However, coke formation is a serious problem for Ni-based catalytic systems [2]. Therefore, researchers are trying to seek appropriate promoters to improve the activity and stability of Ni-based catalytic systems for CO₂ methanation. Yuan et al. investigated the effect of Re on the catalytic activity of Ni-based catalysts in CO₂ methanation [32]. Based on DFT calculations, they found that, attributed to the strong affinity of Re to O (see Figure 6), the presence of Re markedly lowered the energy barrier of C-O bond cleavage, which benefits the activation of CO₂. Moreover, CH₄ selectivity can be enhanced owe to the presence of Re, which was confirmed by analysis of surface coverage of the adsorbed species on Ni (111) and Re@Ni (111). In addition, micro-kinetic analysis showed that, in addition to CO* and H*, a suitable amount of O_{ad} atoms were present on Re@Ni (111), and thus a possible reaction network of CO₂ methanation was proposed as shown in Scheme 2, in which a red line represents the preferable steps in each pathway.



Figure 6. Adsorption sites existed in the (**a**) Re@Ni(111) and (**b**) Ni(111) surface. Ni: blue, Re: green, reprinted with permission from [32]. Copyright Elsevier, 2018.



Scheme 2. Reaction network for CO_2 methanation. The preferable steps in each pathway are marked with a red line, reproduced with permission from [32]. Copyright Elsevier, 2018.

Besides Re, La is also an excellent promoter of Ni-based catalysts for CO_2 methanation. Quindimil et al. [29] applied Ni-La₂O₃/Na-BETA catalysts for CO₂ methanation. They found that the presence of La₂O₃ created more CO₂ adsorption sites and more hydrogenation sites, mainly attributed to the improved surface basicity and Ni dispersion by the co-catalyst effect of La. Under the optimized reaction conditions, 65% CO₂ conversion and nearly 100% CH₄ selectivity were achieved over a Ni-10%La₂O₃/Na-BETA catalyst with a good stability for more than 24 h at 593 K.

CeO₂, TiO₂, and SiO₂ have been used as the supports of methanation catalysts [33]. Reactions over Ni/CeO₂ catalyst performed full selectivity to CH₄ with higher TOF (up to forty-fold) compared to TiO₂ and SiO₂ supported Ni nanoparticles, and in CO₂ methanation, the catalytic stability of Ni/CeO₂ catalyst lasted for 50 h at 523 K. HRTEM analysis indicated that different supports induced distinctive crystal structure. Ni/CeO₂ catalyst presented hexagonal Ni nanocrystallites, while TiO₂ and SiO₂ favored the formation of pseudo-spherical Ni nanoparticles. Demonstrated by TPR, XPS, and UV Raman analysis, characterization results revealed partial reduction of the CeO₂ surface, and the partial reduction of the CeO₂ surface contributed to the generation of oxygen vacancies, which is beneficial for the formation of a strong metal-support interaction (SMSI) between Ni and CeO₂, while no SMSI was observed over Ni/SiO₂ and Ni-TiO₂ catalyst. Furthermore, pulse reaction by temporal analysis products (TAP) demonstrated the capacity of CO₂ adsorption following the order: Ni/SiO₂ < Ni/TiO₂ < Ni/CeO₂. Therefore, metal particle morphology and surface oxygen vacancies were used to anchor/stabilize Ni nanoparticles, and SMSI of Ni/CeO₂ catalyst contributed to the remarkable catalytic activity for CO₂ methanation.

Lin et al. investigated the influence of TiO₂ phase structure on the degree of dispersion of Ru nanoparticles [31]. Experiments showed that Ru/r-TiO₂ (rutile-type TiO₂) catalysts have a fast rate of CO₂ conversion, more than twice as fast as Ru/a-TiO₂ (anatase-type TiO₂) for CO₂ methanation. Meanwhile, compared to Ru/a-TiO₂ catalysts, Ru/r-TiO₂ catalysts exhibited a much higher thermal stability. High-angle annular dark-field scanning transmission electron microscopy (HAADF-STEM) images showed that r-TiO₂ supported Ru nanoparticles displayed a narrower particle size distribution (1.1 ± 0.2 nm) compared with a-TiO₂ supported Ru nanoparticles (4.0 ± 2.4 nm). As confirmed by XRD measurements and H₂-TPR experiments, a strong interaction existed in RuO₂ and r-TiO₂ contributed to the formation of the Ru–O–Ti bond. Experimental tests, revealed that the strong interaction between RuO₂ and r-TiO₂, not only promotes the highly dispersion of Ru nanoparticles, but also prevents nanoparticles' aggregation, which is responsible for the enhancement of the catalytic activity and thermal stability.

Furthermore, the influence of Al₂O₃, ZrO₂, SiO₂, KIT-6, and GO supports on Ni-based catalysts [30,34,35] has also been reported recently (Table 2). The specific area of support, the metal-support interaction and particle size of active sites are still mainly adjustable factors. Interestingly, Ni-SiO₂/GO-Ni-foam catalyst which was synthesized via intercalation of graphene oxide (GO) exhibited excellent activity compared with Ni-SiO₂/Ni-foam catalyst (see Figure 7) at 743 K for CO₂ methanation. Furthermore, the formation of nickel silicates on GO is responsible for the uniform dispersion of Ni active sites, which is favorable for inhibition of catalyst sintering.

Catalyst	H ₂ :CO ₂	GHSV	Temperature (°C)	CO ₂ Conversion (%)	CH ₄ Selectivity (%)
Ni-La/Na-BETA [29]	4	10,000 ^b	350	65	100
Co/meso-SiO ₂ [30]	4.6	60,000 ^a	280	40	94.1
Co/KIT-6 [30]	4.6	22,000 ^a	280	48.9	100
Ru/BF ₄ /SiO ₂ [30]	4	2400 ^b	250	70.5	N/A
Re-Ni(111) [32]	4	N/A	250	N/A	100
Ni/Al ₂ O ₃ -ZrO ₂ [34]	4	40,000 ^a	400	~70.0	N/A
Ni-SiO ₂ /GO [35]	4	500 ^b	470	54.3	88
Ni-ZrO ₂ [36]	4	75 ^a	350	~40.0	~95.0
Ni-Ce/USY [37]	4	N/A	350	65	95
Co/KIT-6 [38]	4	60,000 ^a	340	40	86.7
Co/SiO ₂ [39]	4	60,000 ^a	360	44.3	86.5
Ni-Nb ₂ O ₅ [40]	4	750 ^b	325	81	~99.0

Table 2. Catalytic performance of several catalysts for CO_2 methanation, in terms of CO_2 conversion and CH_4 selectivity, along with the reaction conditions

^a mL g_{cat}⁻¹ h⁻¹; ^b h⁻¹; N/A: not available; pressure: 0.1 MPa.



Figure 7. TOF of Ni-SiO₂/Ni-foam and Ni-SiO₂/GO-Ni-foam catalysts for CO₂ methanation, reprinted with permission from [35]. Copyright Elsevier, 2019.

Besides the traditional methods (co-impregnation method and deposition–precipitation method), some novel preparing methods, such as the sequential impregnation and a change of the preparation progress (e.g., the metal incorporation order, the reduction temperature, and the selection of different types of precursors), have been recently reported for preparing methanation catalysts.

Romero-Sáez et al. synthesized Ni-ZrO₂ catalysts supported on CNTs via the sequential and co-impregnation methods for CO₂ methanation [36]. The catalyst prepared by co-impregnation was apparently less active and selective to CH₄, compared with the catalyst synthesized by the sequential impregnation method. The preparation approach of the sequential impregnation method is as follows: (1) the appropriate amount of $ZrO(NO_3)_2$ xH₂O was dissolved in acetone; (2) the CNTs were added to the solution; (3) removal of the solvent, drying and heat treatment at 623 K; (4) in a rotatory evaporator, the same procedure for Ni(NO₃)₂ impregnation, followed drying and heat treatment was applied. As characterized by TEM analysis, NiO nanoparticles surrounded by ZrO₂ in core–shell structures were formed via co-impregnation method. The existence of core–shell structure reduced reactant access to Ni and Ni–ZrO₂ interface. However, when the catalyst was prepared via a sequential impregnation method, NiO nanoparticles were available and deposited either on the surface or next to the ZrO₂ nanoparticles, which improved the extent of the Ni–ZrO₂ interface. The ratio of Ni–O–Zr exposed species thus increases in the sequential impregnation method, and the interaction between H atoms (produced upon H₂ dissociation on the Ni surface) and the CO₂ molecule (activated by ZrO₂), can also be enhanced. The schematic representation is shown in Figure 8.



Figure 8. Schematic representation of the disposition of NiO, ZrO₂ and interface Ni-O-Zr at the surface of CNTs for (**A**) Ni-Zr-CNT-COI catalytic system, and (**B**) Ni-Zr-CNT-SEQ catalytic system, reproduced with permission from [36]. Copyright Elsevier, 2018.

Interestingly, Bacariza et al. investigated the influence of the metal incorporation order in the preparation of Ni-Ce/Y(USY) zeolite catalysts [37]. In their experiments, three ways were used for the preparation of catalysts: Ni before Ce (Ce/Ni), Ce before Ni (Ni/Ce), and co-impregnation (Ni-Ce). Experimental tests showed that the catalytic activity follows the order: Ce/Ni \approx Ni/Ce < Ni-Ce. TEM and H-TPR characterizations demonstrated that the Ni⁰ average size decreases to approximately 2.5 nm, and in terms of Ni/Ce or Ni-Ce catalysts, stronger interactions between Ni and Ce species are established. However, the CO₂ adsorption capacity is smaller for Ni/Ce catalyst. In contrast, even

though larger Ni^0 particles (13.3 nm) are formed for Ce/Ni catalyst, CO₂ adsorption capacity can be enhanced. Finally, Ce-Ni was found as the best preparation method for CO₂ methanation using Ni-based catalysts supported on CeO₂.

In addition, it has been reported that the reduction temperature and a variety of different precursors (e.g., nitrate, chlorate, and oxalate) affect the number of active centers for CO_2 methanation [30,38]. As discussed above, the preparation methods showed beneficial influences on the catalytic performance, suggesting that reasonable adjustment of the preparation process is a strategy to optimize the experiments for the conversion of CO_2 to CH_4 . Details of CO_2 conversion and product selectivity, along with the reaction conditions of several representative catalytic systems, are compared in Table 2. From Table 2, we can see that Ni-based catalysts are the main catalytic systems for the conversion of CO_2 to CH_4 . Furthermore, the optimal temperature for CH_4 production is 300–400 °C.

2.3. CO₂ to CH₃OH

Methanol is not only an important industrial raw material, but also a stable hydrogen storage compound, which is convenient for transport and reserve. Recently, the concept of "methanol economy" advocated by the Nobel Laureate George Olah revealed the importance of methanol in energy structure [41]. Hence, a large amount of heterogeneous catalysts for the direct hydrogenation of CO_2 to CH_3OH , reaction (3), have emerged in recent years. Herein, we summarize some catalytic systems for the direct hydrogenation of CO_2 to CH_3OH (see Figure 9).

$$CO_2 + 3H_2 \leftrightarrow CH_3OH + H_2O, \ \triangle H_{298K} = -49.5 \text{ kJ mol}^{-1}$$
(3)



Figure 9. Catalytic system for the direct hydrogenation of CO₂ to CH₃OH in this review.

Cu-ZnO-based catalysts, developed by ICI (Imperial Chemical Industries) in the 1960s, are still widely used in CH₃OH synthesis from syngas (CO and H₂), and Cu⁰ is considered to be the active site. Li et al. reported the synthesis of Cu/ZnO catalysts using a unique method, i.e., facile solid-phase grinding, using mixture of oxalic acid, copper nitrate, and zinc nitrate as raw materials [42]. Appropriate amount of these compounds were physically mixed, and then manually ground for 0.5 h. Subsequently, at 393 K, the obtained precursor was dried for 12 h, and at 623 K, calcined for 3 h in N₂ flow, followed by passivation in 1% O_2/N_2 flow for 5 h. In contrast, by the following steps: (1) at 623 K, calcining the dried precursor at for 3 h in air; (2) at 503 K, reducing the oxide for 10 h in 5% H₂/N₂ flow; and (3) at room temperature, passivating the catalyst in 1% O_2/N_2 flow for 5 h, the catalyst can be obtained via H₂ reduction. As demonstrated by XRD, H₂-TPR and thermal-gravity-differential thermal analysis (TG-DTA), in the facile solid-phase grinding process, the decomposition of oxalate complexes and the reduction of CuO took place simultaneously when the samples were calcined in N₂, which prevented the growth of active Cu⁰ species and the aggregation of catalyst particles. In contrast, in the

conventional H₂ reduction process, the growth of active species or the aggregation of catalyst particles were inevitable. Notably, Cu/ZnO catalysts prepared by a facile solid-phase grinding method achieved 29.2% CO₂ conversion (at 523 K and 3 MPa), which is even higher than the thermodynamic equilibrium conversion. Indeed, the equilibrium conversion of CO₂ at 473 K and 3 MPa is 25.78% according to the Benedict–Webb–Rubin equation [43], and it is even lower at 523 K, since CO₂ hydrogenation to CH₃OH is an exothermic reaction. Generally, tandem reaction can break the limit of thermodynamic on reaction result. Inspiring from which, the above unusual experimental results (higher than thermal equilibrium value) could be most likely achieved through tandem reactions, in which every stepwise reaction was accelerated by different metal sites. Therefore, the simple and solvent-free method based on solid-phase grinding, opened a new approach to synthesize bimetallic or multimetallic catalysts without further reduction.

To improve the catalytic activity of Cu/ZnO/Al₂O₃ catalysts, ZnO was replaced by g-C₃N₄-ZnO hybrid material, as reported by Deng et al. [44]. The experimental results showed that, at 12 bar and 523 K, the methanol space time yield (STY) reached 5.73 mmol h⁻¹ g_{Cu}⁻¹ for Cu- g-C₃N₄-ZnO/Al₂O₃, which is superior to the methanol yield (5.45 mmol h⁻¹ g_{Cu}⁻¹) of industrial catalyst (Cu-ZnO-Al₂O₃) under the same reaction pressure. As confirmed by time-resolved photoluminescence (TRPL) and electronic spin resonance (ESR), the electron-richness of ZnO was enhanced via the formation of type-II hetero-junction between g-C₃N₄ and ZnO. In addition, in the TPR curve, enhanced SMSI was observed, and the SMSI between electron-rich ZnO and Cu could boost the catalytic performance in CH₃OH production.

For CuO-ZnO-ZrO₂ catalysts, doping graphene oxide (GO) is a good strategy for improving the activity of CO₂ to CH₃OH [45]. The highest methanol selectivity reached up to 75.88% (473 K, 20 bar) at 1 wt % GO content, while under the same conditions, the methanol selectivity was found to be 68% over the GO-free catalyst. Furthermore, CuO-ZnO-ZrO₂-GO catalyst exhibited an almost constant space-time yield (STY) of methanol after 96 h time-on stream experiment. As demonstrated by H₂-TPD and CO₂-TPD, the CO₂ and H₂ adsorption capacity is enhanced over the CuO-ZnO-ZrO₂-GO catalysts. Additionally, to explain the improvement of catalytic activity, the authors proposed that GO nano-sheet can serve as a bridge between mixed metal oxides, which strengthens a hydrogen spillover (see Figure 10). That is, H species migrating from the copper surface to the carbon species are adsorbed on the isolated metal oxide particles.



Figure 10. Graphene oxide (GO) nanosheet as a bridge promoting hydrogen spillover from the surface of copper to the surface of other metal oxides, reproduced with permission from [45]. Copyright Elsevier, 2018.

It is also reported that modification by small amounts (2 and 5 at %) of WO₃ can improve the CO₂ conversion and CH₃OH selectivity of the CuO-ZnO-ZrO₂ catalyst [46]. The optimal catalytic performance can be attributed to the specific surface area of metallic Cu, basic sites, and the reducibility of catalysts. However, if we want to effectively tune the catalytic reaction, a better understanding of the reaction mechanisms is essential. Up to now, the possible reaction pathways for the hydrogenation of CO₂ to methanol over Cu-ZnO-based catalysts are shown in Scheme 3.

RWGS + CO-Hydro pathways



Formate pathways

Scheme 3. Possible reaction pathways for CO_2 hydrogenation to methanol, where '*X' represents species X adsorbed on a surface site, reprinted with permission from [21]. Copyright Elsevier, 2016.

Besides Cu-based catalysts, In₂O₃ catalyst with surface oxygen vacancies has attracted more and more attention from researchers. Using density functional theory (DFT) calculations, Ye et al. investigated a In_2O_3 catalyst for the hydrogenation of CO_2 to CH_3OH [47]. On a perfect In_2O_3 (110) surface, six possible surface oxygen vacancies (O_{v1} to O_{v6}) were investigated (see Figure 11a), and the D4 surface with the O_{v4} defective site was found to be most beneficial for CO_2 activation and further hydrogenation. Potential energy profiles of CO_2 hydrogenation and protonation on the D4 defective In₂O₃ (110) surfaces are shown in Figure 11b. In addition, the simulation results showed that the formation of CH₃OH replenishes the oxygen vacancy sites, while H₂ contributes to generate the vacancies, and this cycle between perfect and defective states of the surface is responsible for the formation of CH₃OH from CO₂ hydrogenation. To demonstrate this hypothesis, In₂O₃ catalyst was used for the hydrogenation of CO₂ to CH₃OH in practice, which showed the superior catalytic activity (7.1% CO₂ conversion, 39.7% CH₃OH selectivity at 603 K). This experimental result is better than for many other reported catalytic systems, which generally show low selectivity of CH₃OH at 603 K. Confirmed by thermo-gravimetric analysis (TGA), XRD, and HR-TEM, the authors found that In₂O₃ catalyst had satisfactory thermal and structural stability for CO₂ conversion to CH₃OH below 773 K [48]. Furthermore, to reveal the effect of oxygen vacancies, Martin et al. synthesized bulk In₂O₃ catalyst, and at a wide range of reaction conditions, the CH₃OH selectivity could be tuned up to 100%. XRD, H₂-TPR, CO₂-TPD, XPS, operando diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS), and electron paramagnetic resonance (EPR) characterization confirmed that oxygen vacancies on the In₂O₃ catalyst are active sites for the reduction of CO₂. Additionally, to further improve the stability of In₂O₃ catalyst for the production of CH₃OH, the authors investigated a variety of supports (i.e., ZrO₂, TiO₂, ZnO₂, SiO₂, Al₂O₃, C, SnO₂, MgO). ZrO₂ showed the best catalytic performance since it prevented the sintering of the In2O3 phase, which was demonstrated by the enduring stability of In_2O_3/ZrO_2 catalyst over 1000 h [49]. Overall, In_2O_3 is a potential catalyst for CO₂ hydrogenation to CH₃OH with high selectivity.


Figure 11. (a) Optimized structure of the In_2O_3 (110) surface. Red: O atoms; brown: In atoms. (b) Potential energy profiles of CO_2 hydrogenation and protonation on the D4 defective In_2O_3 (110) surfaces. Red line: hydrogenation; black line: protonation. A* represents the adsorption state of A on the surface, while [A+B]* represents the co-adsorption state of A and B on the surface, reproduced with permission from [47]. Copyright American Chemical Society, 2013.

Interestingly, Wang et al. synthesized a series of x% ZnO-ZrO₂ solid solution catalysts (x% represents the molar ratio of Zn) for CO₂ direct hydrogenation to CH₃OH [50]. Under the specified reaction conditions (5.0 MPa, H₂/CO₂ = 3:1 to 4:1, 593 K to 588 K), ZnO-ZrO₂ catalyst can achieve 86–91% methanol selectivity with single-pass CO₂ conversion more than 10%, better than the results reported by other researchers. Moreover, in the presence of 50 ppm SO₂ or H₂S in the reaction stream,

no deactivation was observed. Experimental results confirmed that ZrO_2 and ZnO alone showed little activity in methanol synthesis, while the catalytic performance was significantly enhanced and CO_2 conversion reached the maximum value when the Zn/(Zn + Zr) molar percentage is close to 13%. Demonstrated by CO_2 -TPD, most of the CO_2 adsorbed on the Zr sites of ZnO-ZrO₂ catalyst, and the H_2 - D_2 exchange experiment indicated that ZnO had much higher activity than ZrO_2 . Therefore, in ZnO-ZrO₂ solid solution catalyst, the synergetic effect between the Zn and Zr sites markedly promotes the activation of H_2 and CO_2 . To explain the reaction mechanism on the solid solution catalyst (ZnO-ZrO₂), in situ diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) and density functional theory (DFT) calculations were conducted. It was concluded that CO_2 direct hydrogenation to CH₃OH is dominated via the formate pathway, and CH₃OH was formed by H₃CO* protonation on the surface of ZnO-ZrO₂ catalyst.

Although Cu-ZnO-based catalysts are highly selective for CO_2 conversion to methanol, there are still some serious problems during industrial operation, such as deactivation of active sites at high temperature and agglomeration of catalytic particles under industrial reaction conditions. To overcome these issues, researchers designed a series of bimetallic catalysts for the direct hydrogenation of CO₂ to CH_3OH . Jiang et al. prepared a series of Pd-Cu bimetallic catalysts supported on SiO_2 with a wide range of total metal loading (i.e., 2.4–18.7 wt %) and evaluated the catalytic performance for the reduction of CO₂ to CH₃OH. With a decrease in metal loadings, the CO₂ conversion dropped stepwise, namely from 6.6 to 3.7%. However, the CH₃OH STY was slightly enhanced by 31% from 0.16 to $0.21 \text{ mmol mol}^{-1} \text{ s}^{-1}$ [51]. To unclose the reaction mechanisms, the authors carried out in situ diffuse reflectance FT-IR (DRIFTS) measurements on Pd(0.34)-Cu/SiO₂ catalyst [52]. The resulting spectra identified that the dominant species on a bimetallic surface were formate and carbonyl species on a bimetallic surface, which were dependent on the catalyst composition. Therefore, they proposed that on Pd-Cu catalysts, the surface coverage of formate species was correlated to the methanol promotion, implying its vital role in CH₃OH synthesis. Bahruji et al. prepared PdZn/TiO₂ bimetallic catalysts by a solvent-free chemical vapor impregnation method [53]. According to the order in which the metals were impregnated, the catalysts could be classified as $^{2}Pd^{-1}Zn^{-1}O_{2}$, $^{2}Zn^{-1}Pd^{-1}TiO_{2}$, and $PdZn/TiO_{2}$ (1 and 2 represent the order of sequential metal impregnations). The experimental results showed that PdZn/TiO₂ catalyst achieved optimal catalytic performance, 10.1% CO₂ conversion and 40% CH₃OH selectivity. The formation of ZnTiO₃ and PdZn nanoparticles was confirmed via XPS, XRD, and TEM. Combining the experimental results, the authors proposed that PdZn nanoparticles were beneficial for methanol formation and catalytic stability. Although it is hard to assign the formation of the PdZn alloy to the differences between the preparation methods, an interaction between $Pd(acac)_2$ and Zn(acac)₂ precursors might be responsible for the production of the PdZn active site in terms of PdZn/TiO₂ catalyst. Additionally, Yin et al. reported the preparation of Pd@zeolitic imidazolate framework-8 (ZIF-8) catalyst for the conversion of CO2 to CH3OH [54]. The optimal methanol yield can reach 0.65 g g_{cat}^{-1} h⁻¹ over a PdZn catalyst. As demonstrated by XPS, XRD, TEM and electron paramagnetic resonance (EPR), after H₂ reduction, PdZn alloy particles formed, and abundant oxygen defects existed on the ZnO surface. The experimental results revealed that the active site is a PdZn alloy rather than metallic Pd, in terms of CH₃OH formation.

Numerous studies suggested that promoters have an indispensable role in the catalytic performance of CuZn catalysts [55], although the mechanism of action still remains unclear. Pd-doped CuZn catalysts were prepared to evaluate the surface modification effect. An interesting observation was made from the volcano-shaped relationship between CH₃OH STY and Pd loading, which implied that an appropriate amount of Pd loading is beneficial for methanol synthesis. The chemisorption analyses further revealed a strong interaction between Pd and Cu surface, where a hydrogen spillover has taken place, generating more activated Cu sites. Hence, the CH₃OH STY and the methanol turnover frequency (TOF) improved a lot with Pd loading of 1 wt % [55]. In addition, a variety of bimetallic catalysts (e.g., Cu-Fe, Cu-Co, Cu-Ni, Co-Ga, Ni-Ga) have been evaluated by many researchers [56–59]. A possible catalytic reaction mechanism of CO₂ hydrogenation over Cu–Ni/CeO₂–NT is shown in

Figure 12. Details of the CO_2 conversion and product selectivity, along with reaction conditions of several representative catalytic systems are compared in Table 3. In terms of hydrogenation of CO_2 to CH_3OH , In-based and Pd-based catalysts have gradually drawn researchers' attention, besides traditional Cu-based catalysts. Additionally, many researchers demonstrated that 250 °C is the optimal temperature to produce CH_3OH with the high selectivity. According to the current experimental results, increasing the reaction pressure is also a good strategy to improve CO_2 conversion, but a high pressure means a high cost of operation and equipment.



Cu-Ni/CeO₂-NT

Figure 12. Possible catalytic mechanism of CO₂ hydrogenation over Cu–Ni/CeO₂–NT, reproduced with permission from [59]. Copyright American Chemical Society, 2018.

Table 3. Catalytic performance of several catalytic systems for CO_2 hydrogenation into CH_3OH , in terms of CO_2 conversion and CH_3OH selectivity, along with the reaction conditions.

Catalyst	H ₂ :CO ₂	GHSV	Temperature (°C)	Pressure (MPa)	CO ₂ Conversion (%)	CH ₃ OH Selectivity (%)
Cu-ZnO [42]	2.9	2160 ^a	250	3	29.2	83.6
Cu/g-C ₃ N ₄ -Zn/Al ₂ O ₃ [44]	3	6800 ^a	250	1.2	~7.0	~55.0
CuO-ZnO-ZrO2-GO [45]	3	15,600 ^a	240	2	N/A	75.8
W-Cu-Zn-Zr [46]	2.7	2400 ^a	240	3	19.7	49.3
In ₂ O ₃ [48]	3	15,000 ^a	270	4	1.1	54.9
In ₂ O ₃ [48]	3	15,000 ^a	330	4	7.1	3
In ₂ O ₃ [49]	4	16,000–48,000 ^b	300	5	N/A	100
ZnO-ZrO ₂ [50]	3-4	24,000 ^a	315-320	5	>10	86-91
Pd-Zn-TiO ₂ [53]	3	916 ^b	250	2	10.3	61
Pd-Zn-ZIF-8 [54]	3	21,600 ^a	270	4.5	~22.0	~50.0
Pd-Cu-Zn [55]	3	10,800 ^a	270	4.5	~8.0	~65.0
Co ₅ Ga ₃ [58]	3	N/A	250	3	1	63
CuNi2/CeO2-NT [59]	3	6000 ^b	260	3	17.8	78.8
Cu-Zn-SiO ₂ [60]	3	2000 ^a	220	3	14.1	57.3
Ni ₅ Ga ₃ /SiO ₂ /Al ₂ O ₃ /Al [61]	3	3000 ^a	210	0.1	~1.0	86.7
Cu-Zr-SiO ₂ [62]	3	N/A	230	5	N/A	77
Cu/Mg/Al [63]	2.8	2000 ^b	200	2	3.6	31
Cu-Ce-Zr [64]	3	7500 ^a	250	3	14.3	53.8
Cu-TiO ₂ [65]	3	3600 ^a	260	3	N/A	64.7
Cu-Zn-Mn-KIT-6 [66]	3	120,000 ^a	180	4	8.2	>99.0
Cu-SBA-15 [67]	3	N/A	210	2.2	13.9	91.3
Au-CuO/SBA-15 [68]	3	3600 ^b	250	3	24.2	13.5
Cu-Zr-SBA-15 [69]	3	N/A	250	3.3	15	N/A
Pd/In ₂ O ₃ [70]	4	>21,000 a	300	5	>20.0	>70.0

^a mL g_{cat}^{-1} h⁻¹; ^b h⁻¹; N/A: not available.

2.4. CO₂ to Other Products

Besides CO, CH₄, and CH₃OH, dimethyl ether (DME), light olefins [71–74], alcohol [75], isoparaffins [76], and aromatics [77,78] have also been produced by CO₂ hydrogenation. Clearly, in terms of CO₂ conversion, the coupling of C-C bond to produce higher hydrocarbons and oxygenates is a technological barrier. However, taking advantage of tandem catalysis to realize one-step synthesis of hydrocarbons via hydrogenation of CO₂ is feasible. The extensive studies can be mainly categorized into two categories: methanol-mediated and non-methanol-mediated reactions [79,80]. In the methanol-mediated approach, DME is usually produced as main product, while for the non-methanol-mediated approach, alkenes and alkanes are generally produced as main products. The possible reaction mechanism for CO₂ hydrogenation to DME and light olefins is shown in Scheme 4.



Scheme 4. Possible reaction pathways for CO₂ hydrogenation to DME and light olefins. Note: ---H, ---O derived from H-ZSM5 zeolite, reprinted with permission from [21]. Copyright Elsevier, 2016.

Cu-based catalysts are in practice used for methanol synthesis, and HZSM-5 zeolites are widely employed for methanol dehydration due to its solid acid catalysis. Therefore, a variety of studies have been reported with regard to the bi-functional catalysts consisting of Cu and HZSM-5 used for the direct hydrogenation of CO₂.

Zhang et al. reported a Cu-ZrO₂/HZSM-5 catalyst promoted by Pd/CNT for direct synthesis of DME from CO_2/H_2 [79]. Although a minor amount of the Pd-decorated CNTs into the CuZr/HZSM-5 catalytic system caused little change in the activation energy for CO_2 conversion compared with the CuZr/HZSM-5 catalyst, the former created a micro-environment including higher concentration of active H species and adsorbed CO_2 species on the surface of the catalyst system. Owing to the increase

of hydrogenation reactions, under reaction conditions of 5 MPa and 523 K, the specific rate of CO₂ hydrogenation-conversion was 1.22 times higher than the pure CuZr/HZSM-5 catalyst. Furthermore, Zhang et al. reported a series of V-modified CuO-ZnO-ZrO₂/HZSM-5 catalysts which were prepared via an oxalate co-precipitation method for the synthesis of DME from CO₂/H₂ [81]. The catalytic performances of the catalysts were strongly dependent on the content of V, which was confirmed by XRD, N₂O chemisorption, and X-ray photoelectron spectroscopy (XPS). Similarly, the influence of the promoter—i.e., La and W—has also been examined. The optimal catalytic activity was obtained (43.8% CO₂ conversion and 71.2% DME selectivity) when the amount of La was 2 wt %. In contrast, the Cu-ZnO-ZrO₂ catalyst admixed with WO_x obtained 18.9% CO₂ conversion and 15.3% DME selectivity. Obviously, La is a better promoter compared with W, and this can be explained by the fact that the reducibility and dispersion of bi-functional catalysts (CuO-ZnO-Al₂O₃-La₂O₃/HZSM-5) were largely dependent on the modification of La, while the hybrid catalyst (Cu-ZnO-ZrO₂-WO_x/Al₂O₃) modified by W strongly adsorbed water molecules, resulting in a lower catalytic performance. Moreover, as shown in Figure 13, the STY of DME of different WO_x/Al_2O_3 catalysts (i.e., on supports with different pore sizes) as a function of W surface density shows a volcanic curve relation. Details of CO₂ conversion and DME selectivity reported recently are shown in Table 4 [82,83]. The above studies indicate that the current catalysts used for DME preparation from CO₂/H₂ are still dominated by Cu-based-HZSM-5 catalysts. Some other zeolite catalysts (SBA-15, SAPO-5, SUZ-4) with similar acid properties (acid content and acid strength) as HZSM-5 zeolite may be a direction of future research.



Figure 13. STY of DME of different WO_x/Al_2O_3 catalysts, with different support pore size as a function of W surface density. AS, AM, and AL refer to the mean pore size of alumina supports with small, medium and large pores, respectively, reproduced with permission from [83]. Copyright Elsevier, 2018.

Among the large-scale technologies with industrial potential, the conversion of CO₂ to DME is promising and relatively mature [84]. Korea Gas Corporation (KOGAS) has developed a DME plant, with CO₂ as a raw material. The schematic diagram of the KOGAS tri-reforming process is shown in Figure 14 [85,86]. Kansai Electric Power Co. and Mitsubishi Heavy Industries have realized a bench-scale (100 cm³ catalyst loading) experiment for DME synthesis [87,88]. However, during the reaction process, water formation decreased the yield of DME [89–91]. Catizzone et al. and Bonura et al. evidenced that zeolites or ferrierite could effectively mitigate the influence of water and avoid catalyst sintering [89,90]. In a fixed bed reactor, kinetic modeling confirmed the negative

effect of water (formed during the reaction process) on DME production, which is consistent with the experimental results. Therefore, Falco et al. suggested that hydrophilic membranes could be promising for industrial production of DME [92]. Recently, Fang et al. advocated CO_2 capture and conversion by using a membrane reactor system, where a high-temperature mixed electronic and carbonate-ion conductor (MECC) membrane was used for CO_2 capture and a solid oxide electrolysis cell (SOEC) was used for CO_2 reduction [93]. Furthermore, based on membrane technology, Sofia et al. carried out a techno-economic analysis project of power and hydrogen co-production via an integrated gasification combined cycle (IGCC) plant with CO_2 capture [94]. The development of membrane technology would be an advantageous factor for the capture and utilization of CO_2 .



Figure 14. Schematic diagram of KOGAS tri-reforming process, reproduced with permission from [86]. Copyright Elsevier, 2009.

Li et al. synthesized a novel ZnZrO/SAPO tandem catalyst, combined by a ZnO-ZrO₂ solid solution and a Zn-modified SAPO-34 zeolite, which achieved 80–90% olefin selectivity among the hydrocarbon products [95]. Based on the surface reaction kinetics, they proposed that the tandem reaction process proceeded as follows: (1) generation of CH_xO species on ZnZrO via CO₂ reduction; (2) olefins production from the derived CH_xO species which migrate/transfer onto SAPO zeolite pore structure. Moreover, experimental results confirmed that the excellent selectivity can be ascribed to the effective synergy between ZnZrO and SAPO for the tandem catalyst. In addition, this catalyst (ZnZrO and SAPO-34) shows an excellent stability toward thermal and sulfur treatments, indicating the potential value for industrial application.

Recently, CO₂ was converted to aromatics with a selectivity up to 73%, at 14% CO₂ conversion over ZnZrO/HZSM-5 catalyst [77]. Demonstrated by operando infrared (IR) characterization, Li et al. proposed that CH_xO, as an intermediate species, transformed from the ZnZrO surface into the pore structure of HZSM-5, which is responsible for C-C bond formation for aromatics production. The reaction scheme is shown in Figure 15. Interestingly, the presence of H₂O and CO₂ markedly suppressed the generation of polycyclic aromatics, consequently, enhanced the stability (100 h in the reaction stream) of the tandem catalyst, which showed potential industrial application prospects. Similarly, Ni et al. synthesized a tandem catalyst of ZnAlO_x and HZSM-5, which yields 73.9% aromatics selectivity and 0.4% CH₄ selectivity among the carbon products without CO [78]. Confirmed by XRD, SEM, TEM, and element distribution analysis, ZnAlOx was formed and uniformly dispersed in the tandem catalyst. Furthermore, demonstrated by 2,6-di-tert-butyl-pyridine absorption (DTBPy-FTIR), the external Brønsted acid of HZSM-5 can be shielded by ZnAlO_x, which is beneficial to aromatization. According to the operando diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) study, they proposed the following possible reaction mechanism: MeOH and DME, produced by hydrogenation of formate species, are transmitted to HZSM-5 and then converted into olefins and finally aromatics.



Figure 15. Reaction scheme of CO₂ hydrogenation to aromatics, reproduced with permission from [77]. Copyright Elsevier, 2019.

Non-methanol-mediated reactions, i.e., the direct hydrogenation of CO_2 to light olefins is even more significant than CH_3OH synthesis from CO_2 hydrogenation, since a large proportion of methanol is used for the synthesis of olefins in industry, through the methanol-to-olefins (MTO) reaction by using SAPO-34 zeolite catalysts. CO_2 to lower olefins can be realized by coupling of the RWGS reaction and F–T synthesis, as shown in Schemes 1 and 3.

Iron-based catalysts has been considered as an excellent option for the synthesis of light olefins from CO₂/H₂, mainly because of their excellent activity, high selectivity, and low price. Using honeycomb-structure graphene (HSG) as the support and K as a promoter, Wu et al. prepared an iron-based catalyst [80]. They found out that the confinement effect of the porous HSG was beneficial for the sintering of the Fe active sites, and within 120 h stability test, no significant deactivation occurred. In addition, as confirmed by CO₂ temperature programmed desorption (CO₂-TPD) and H₂-TPD, they found out that potassium as a promoter effectively enhanced the chemisorption and activation of the reactants CO_2 and H_2 . Moreover, as revealed by 5^7 Fe Mossbauer absorption spectroscopy, for the Fe/HSG catalyst, there were two doublets with isomer shift (IS) of 0.96 mm s⁻¹ and guadrupole splitting (QS) of 0.28 mm s⁻¹ and IS of 1.21 mm s⁻¹ and QS of 1.82 mm s⁻¹, which corresponded to the Fe (II) species in low- and high-coordination environments. Instead, for the FeK1.5/HSG catalyst, only one doublet with IS of 0.31 mm s⁻¹ and QS of 1.05 mm s⁻¹ ascribed to the Fe (III) species, which implies that K is capable of stabilizing high valence-state iron (Fe (III)) during CO₂ hydrogenation to light olefins (CO_2 -FTO). The above mentioned three factors contributed to the excellent catalytic performance, i.e., 56% olefins selectivity and a 120-h stability testing experiment (as shown in Figure 16).



Figure 16. CO_2 – FTO results over the FeK1.5/HSG catalyst during 120 h on stream (Fe time yield to hydrocarbons, termed as FTY). Reaction conditions: 0.15 g catalyst, T = 613 K, P = 20 bar, H₂/CO₂ = 3 by volume, and the space velocity of 26 L h⁻¹ g⁻¹. The CO selectivity is in the range of 39–43%, reproduced with permission from [80]. Copyright American Chemical Society, 2018.

Zhang et al. used an impregnation method to prepare Fe-Zn-K catalysts, in which K was used as a promoter [71]. The experimental results showed that the Fe-Zn-K catalyst with H_2/CO reduction showed the best catalytic activity, with 51.03% CO₂ conversion and 53.58% C₂–C₄ olefins selectivity, at the reaction conditions of 593 K and 0.5 MPa. XRD, H₂-TPR, and XPS characterization results revealed that, in the Fe-Zn-K catalyst, ZnFe₂O₄ spinel phase and ZnO phase were formed. Among them, ZnFe₂O₄ spinel phase strengthens the interaction between iron and zinc, and changed the reduction and CO₂ adsorption behaviors. In addition, the H₂-TPR profiles show that the catalyst modified by K contributed to a slight shift of the initial reduction peak to higher temperature, indicating the reduction of Fe₂O₃ and formation of Fe phase was inhibited by K modification.

You et al. investigated the catalytic activity of non-supported Fe catalysts (bulk Fe catalysts) modified by alkali metal ions (i.e., Li, Na, K, Rb, Cs) for the conversion of CO₂ to light olefins [72]. By calcining ammonium ferric citrate, non-supported Fe catalyst was prepared. Compared with Fe catalyst without modification (5.6% CO₂ conversion, 0% olefins), the modification of the Fe catalyst with an alkali metal ion markedly enhanced the catalytic activity for CO₂ conversion to light olefins. For Fe catalyst modified by K and Rb, the conversion of CO₂ and the olefin selectivity (based on only the hydrocarbon compounds, without CO) increased to about 40% and 50%, respectively, and the yield of light (C₂–C₄) olefins can reach 10–12%. Further investigation via XRD showed that over the alkali-metal-ion-modified Fe catalysts, Fe₅C₂ was formed, while in the unmodified catalyst, iron carbide species were not observed after the reaction. Accordingly, they proposed that in the presence of an alkali metal ion, the generation of iron carbide species was one possible reason for the enhanced catalytic activity. Therefore, modification by alkali metal ions remains a good strategy to tune the product distribution.

Additionally, Wei et al. reported a highly efficient, stable and multifunctional Na–Fe₃O₄/HZSM-5 catalytic system [76], for direct hydrogenation of CO₂ to gasoline-range (C₅–C₁₁), in which the selectivity of C₅–C₁₁ hydrocarbons reached 78% (based on total hydrocarbons), while under industrial conditions only reached 4% methane selectivity at a 22% CO₂ conversion. Characterization by high resolution transmission electron microscopy (HRTEM), X-ray diffraction (XRD), and Mossbauer spectroscopy showed that two different types of iron phase—i.e., Fe₃O₄ and x-Fe₅C₂—were discerned in the spent Na–Fe₃O₄ catalyst, which cooperatively catalyzed a tandem reaction (RWGS and FT).

The possible reaction scheme for CO₂ hydrogenation to gasoline-range hydrocarbons is shown in Figure 17. Furthermore, acid sites existing in the HZSM-5 zeolite were favorable for acid-catalyzed reactions (oligomerization, isomerization, and aromatization). Notably, the multifunctional catalyst exhibited a significant stability for 1000 h on stream, showing the potential as promising industrial catalyst material for CO₂ conversion to liquid fuels. Similarly, Gao et al. investigated a bi-functional catalytic system composed of In₂O₃ and HZSM-5, which can achieve 78.6% C₅₊ selectivity with only 1% CH₄ at a 13.1% CO₂ conversion [73]. As demonstrated by Ye et al. [47], CO₂ and H₂ can be activated in the oxygen vacancies on the In₂O₃ surfaces, and catalyzed CH₃OH formation. Subsequently, C–C coupling occurred inside the zeolite pores structure (HZSM-5) to synthesize gasoline-range hydrocarbons with a relatively high octane number (C₅₊).



Figure 17. Reaction scheme for CO₂ hydrogenation to gasoline-range hydrocarbons, reproduced with permission from [76]. Copyright Nature Publishing Group, 2017.

Details of the conversion and selectivity for the hydrogenation of CO₂ into other products, along with the reaction conditions of several representative catalytic systems, are compared in Table 4. Fe-based catalysts are beneficial for the production of olefins, while Cu-based + HZSM-5 is still a dominant catalytic system for CO₂ conversion to DME. Similarly, 250 °C is the optimal temperature for the production of DME based on the current experimental results, and 300–400 °C, which is close to industrial production conditions, seems to be better for the direct conversion of CO₂ to olefins. On the other hand, the selectivity of DME can be maintained upon increasing GHSV (>10,000 mL g_{cat}⁻¹ h⁻¹), but the high selectivity of olefins generally relies on a relatively low GHSV (<5000 mL g_{cat}⁻¹ h⁻¹).

Table 4. Catalytic activity of several catalysts for CO_2 hydrogenation into other products (e.g., DME, olefins, alcohol, isoparaffins, gasoline, aromatics), in terms of CO_2 conversion and product selectivity, along with the reaction conditions.

Catalyst	H ₂ :CO ₂	GHSV	Temperature (°C)	Pressure (MPa)	CO ₂ Conversion (%)	Selectivity (%)
Fe-Zn-K [71]	3	1000 ^b	320	0.5	51.03	olefins (53.58)
K-Fe [72]	3	1200 ^a	340	2	38	light olefins (78)
In/HZSM-5 [73]	3	9000 a	340	3	13.1	liquid fuels (78.6)
Ce-Pt@mSi-Co [74]	3	N/A	250	N/A	~3.0	$C_2 - C_4$ (60)
K/Cu-Zn-Fe [75]	3	5000 ^b	300	6	42.3	alcohol (56.43)
Na-Fe/HMCM-22 [76]	2	4000 a	320	3	26	isoparaffins (74)
ZnZrO-HZSM-5 [77]	N/A	1200 a	320	4	14	aromatic (73)
ZnAlOx-HZSM-5 [78]	3	2000 a	320	3	9.1	aromatics (74)
Cu-Zr-Pd/HZSM-5 [79]	3	25,000 a	250	5	18.9	DME (51.8)
Fe-K/HSG [80]	3	26,000 ^a	340	20	N/A	olefins (59)

Catalyst	H ₂ :CO ₂	GHSV	Temperature (°C)	Pressure (MPa)	CO ₂ Conversion (%)	Selectivity (%)
V-Cu-Zn-Zr/HZSM-5 [81]	3	4200 ^b	270	3	32.5	DME (58.8)
Cu-Zn-Al-La/HZSM-5 [82]	3	3000 ^b	250	3	43.8	DME (71.2)
ZnO-ZrO2-SAPO-34 [95]	N/A	3600 ^a	380	2	12.6	olefins (80-90)
Cu-ZnO-Al ₂ O ₃ [96]	3	10 ^b	270	5	9	DME (31)
In-Zr/SAPO-34 [97]	3	9000 ^a	380	3	26.2	C ₂₊ (36.1)
Cu-Zn-kaolin-SAPO-34 [98]	3	1800 ^a	400	3	50.4	C2-C4 (65.3)
Fe/C-Bio [99]	3	N/A	320	1	31	C ₄₋₁₈ alkenes (50.3)
Cu-Zn-Zr/HZSM-5 [100]	3	10,000 ^a	220	3	9.6	DME (46.6)
Cu-Zn-Zr/HZSM-5 [101]	3	9000 ^a	240	3	~30	DME (~35)
Cu-Zn-Al/HZSM-5 [102]	3	4200 ^b	270	3	30.6	DME (49.02)
Cu-Zn-Al/HZSM-5 [103]	3	1800 ^a	262	3	46.2	DME (45.2)
Cu-Zn-Zr-zeolite [104]	3	10,000 ^b	240	3	24	DME (38.5)
Cu-Zn-Al/HZSM-5 [105]	3	1800 ^a	270	3	48.3	DME (48.5)
Fe-K/HPCMs-1 [106]	3	3600 ^a	400	3	33.4	olefins (47.6)
Na-Fe/HZSM-5 [107]	3	4000 ^a	320	3	22	C ₅ -C ₁₁ (78)
Fe ₅ C ₂ -/a-Al ₂ O ₃ [108]	3	3600 ^a	400	3	31.5	C ₂₊ (69.2)
Fe-Zr-Ce-K [109]	3	1000 ^b	320	2	57.34	C2-C4 (55.67)
Fe/C+K [110]	3	24,000 ^a	320	30	24	C2-C6 (36)
Co-Cu/TiO ₂ [111]	3	3000 ^a	250	3	18.4	C ₅₊ (42.1)

Table 4. Cont.

^a mL g_{cat}^{-1} h⁻¹; ^b h⁻¹; N/A: not available.

2.5. Opportunities of Heterogeneous Catalysis for CO2 Conversion

The relationship between products selectivity and CO_2 conversion by heterogeneous catalytic hydrogenation is shown in Figure 18. It is obvious that, in the RWGS reaction, although the CO selectivity reaches up to nearly 100%, the CO_2 conversion is low. In the methanation reaction of CO_2 , the CH₄ selectivity is high enough, and the CO₂ conversion also exceeds 50%. With regard to the synthesis of CH₃OH, DME, and light olefins, the relationship between conversion and selectivity is less clear.



Figure 18. Selectivity and conversion distribution of different products according to recent reports in heterogeneous catalysis.

In the RWGS reaction, CO is mainly produced over Fe-, Co-, Mo-, and Pt-based catalysts (Table 1), while CO_2 methanation is typically carried out on Co- and Ni-based catalysts (Table 2). Although CO is a main component of syngas and CH₄ is an important energy resource, neither is the best product for CO_2 conversion from a relatively economic point of view. Furthermore, the conversion and utilization of CO and CH₄ is also an important research field, which also implies that CO and CH₄ are not the

optimal choice as end-products of CO_2 conversion. Therefore, the coupling reaction is a potential direction for CO_2 conversion to some higher value products, i.e., CH_3OH and light olefins.

Cu-ZnO-based catalyst is still the main catalytic material for direct conversion of CO₂ to CH₃OH under industrial reaction condition, while highly innovative methodologies are awaited to achieve low-pressure and low-temperature CH₃OH synthesis processes. Considering the industrial application value of methanol, exploring novel catalytic systems and designing rational reactors to improve the catalytic activity should be targeted in the future. DME, up to now, is synthesized via Cu-based and H-ZSM5 catalyst (Table 4), and this process essentially remains a two-step tandem reaction. Hence, the bottleneck of DME production could be the deactivation of CH₃OH dehydration catalyst, due to water poisoning. The formation of coke is also the major reason of catalyst for the production of DME, water- and coke-resistant catalysts need to be developed for CH₃OH dehydration.

As shown in Table 4, the direct hydrogenation of CO₂ to light olefins is mainly catalyzed by Fe-based catalysts. However, the starting point of current research work is mainly the coupling of RWGS, F–T and/or methanol-to-olefins (MTO) reactions. Therefore, suitable catalytic materials are sought for the coupling of C-C bonds, which is an effective strategy for improving the catalytic activity.

3. Plasma Catalysis

Plasma, the 'fourth state of matter', consists of electrons, neutral species (i.e., molecules, radicals, and excited species) and ions. Plasma can be in so-called thermal equilibrium or not, based on which it is subdivided into 'thermal plasma' and 'non-thermal plasma' (NTP) [13,112,113]. In non-thermal plasma, the gas temperature remains near room temperature, while electrons temperature is extremely high, usually in the range of 1–10 eV (~10,000–100,000 K). The latter is enough to activate stable gas molecules into reactive species (e.g., radicals, excited atoms, molecules, and ions). These reactive species, especially the radicals, can trigger reactions at low temperature. That is, NTP offers a unique approach to enable thermodynamically unfavorable chemical reactions to proceed at low temperature by breaking thermodynamic limits. Nevertheless, the control of selectivity of desired products in plasma is extremely difficult, since the reactions in plasma are mainly triggered through nonselective collision between active species (radicals, molecules, atoms, and ions).

To improve the desired product selectivity of the reactions in plasma, the combination of catalysts with plasma technology (so-called plasma catalysis) is a promising strategy, since catalysts usually have a special feature of regulating product distribution.

NTP can be generated through various types of discharges—i.e., microwave discharges, glow discharges, gliding arc discharges, dielectric barrier discharges (DBD), etc.—but DBD are the best option to be used in plasma catalysis. Indeed, DBDs are usually operated at atmospheric pressure and ambient temperature, and the integration of DBD plasma with catalysts has the advantages of simple operation and low cost. Thus plasma-driven direct hydrogenation of CO₂ is mostly based on DBD plasmas. The possible plasma/catalyst synergism is illustrated in Figure 19 [13].

A lot of research has been performed for pure CO₂ splitting, in various types of plasma reactors, including DBD, microwave discharge and gliding arc discharge, without catalysts [114–117]. A typical experimental set-up of a DBD plasma for CO₂ decomposition is shown in Figure 20. Paulussen et al. studied the conversion of CO₂ to CO and oxygen in DBD [114], and they found that the gas flow rate is the most crucial parameter affecting the CO₂ conversion. At 0.05 L min⁻¹ flow rate, 14.75 W cm⁻³ power density and 60 kHz discharge frequency, 30% CO₂ conversion was achieved. The performance might be further enhanced by optimizing the discharge parameters (i.e., power, frequency, dielectric material) or by implementing parallel reactors.



Figure 19. Overview of the possible effects of the catalyst on the plasma and vice versa, possibly leading to synergism in plasma-catalysis, reproduced with permission from [13]. Copyright Royal Society of Chemistry, 2017.



Figure 20. Schematic diagram of the experimental set-up used in the decomposition of CO₂, reproduced with permission from [116]. Copyright IOP Publishing, 2016.

Dry reforming of methane (DRM), i.e., the combined conversion of CH_4 and CO_2 by plasma and/or plasma catalysis, has attracted extensive attention in recent years. For instance, Li et al. studied CO_2 reforming of CH_4 by taking advantage of atmospheric pressure glow discharge plasma [118]. Liu et al. reported high-efficient conversion of CO_2 and CH_4 in AC-pulsed tornado gliding arc plasma [119]. Kolb et al. investigated DRM in a DBD reactor [120]. In the above-mentioned studies [118–120], syngas (CO and H₂) was produced as the main product. Recently, Wang et al. reported a novel one-step reforming of CO_2 and CH_4 into liquid oxygenate products, dominated by acetic acid, at room temperature by the coupling of DBD plasma and catalysts [121]. They examined the effect of CH_4/CO_2 molar ratio and of various catalysts (γ -Al₂O₃, Cu- γ -Al₂O₃, Au- γ -Al₂O₃). Interestingly, compared with plasma-only mode, the coupling of plasma and catalysts tuned the selectivity of liquid chemicals, and oxygenates selectivity was achieved up to approximately 60% when Cu catalyst was used. Details about the effects of operating mode and catalysts on the CO_2 conversion reaction results are shown in Figure 21. Although much efforts need to be made to reveal the unknown mechanisms, the results are attractive and show an excellent application prospect.



Figure 21. Effect of operating modes and catalysts on the reaction: (**a**) selectivity of liquid oxygenates, (**b**) selectivity of gaseous products, (**c**) conversion of CH_4 and CO_2 (total flow rate 40 mL min⁻¹, discharge power 10 W, catalyst ca. 2 g), reproduced with permission from [121]. Copyright Wiley online library, 2017.

Besides the above-mentioned metal catalysts, zeolites have also been used in plasma catalytic DRM. Zhang et al. studied the catalytic performance of zeolite catalysts (i.e., NaA, NaY, and HY) for the direct conversion of CH_4 and CO_2 at relatively low temperature range and ambient pressure via DBD plasma [122], and the products were dominated by syngas and C_4 hydrocarbons. Form the investigated catalysts (NaA, NaY, and HY), HY zeolite catalyst exhibited the best performance (26.7% CO_2 conversion and 52.1% C_4 hydrocarbon selectivity), which was mainly attributed to the appropriate pore size and electrostatic properties of HY zeolite.

Although direct hydrogenation of CO_2 to CH_3OH is an exothermic reaction, studies of heterogeneous catalysis for CO_2 hydrogenation to CH_3OH were usually operated at high temperature and high pressure, mainly caused by the high stability of CO_2 and the low equilibrium constant at atmospheric pressure. Plasma catalysis, however, is a promising approach to enable CO_2 conversion to CH_3OH at ambient conditions, and has gradually attracted more and more interest. For instance, Eliasson et al. reported the direct hydrogenation of CO_2 to CH_3OH by coupling of DBD plasma and a discharge-activated catalyst (CuO-ZnO-Al₂O₃) [123]. By comparison of the experiments, with catalyst only, discharge only, and discharge + catalyst, they found that DBD plasma effectively lowered the

optimal reaction temperature corresponding to the best catalytic performance. Indeed, the optimal reaction temperature was 493 K for catalyst only, while in terms of discharge only and discharge + catalyst, the optimal reaction temperature for CO₂ conversion was 373 K. The maximum selectivity for the methanol formation (10%) was achieved at a temperature of 373 K, which implies that the plasma improved the catalytic activity of CuO-ZnO-Al₂O₃ catalyst at low temperature. Additionally, the authors found that low input power and high pressure are beneficial for the improvement of the methanol selectivity.

Zeng et al. studied CO₂ hydrogenation by combining various catalysts (i.e., Cu/γ -Al₂O₃, Mn/γ -Al₂O₃, and Cu-Mn/ γ -Al₂O₃) with DBD plasma in a coaxial packed-bed [124]. The experimental results showed that the addition of catalysts in the reactor improved the conversion of CO₂. At the same time, with the increase of the H₂/CO₂ molar ratio, the CO₂ conversion was improved, and the CO yield was also enhanced. It is also worth mentioning that, compared with plasma only experiments, the energy efficiency was enhanced by adding catalysts, although the synergetic mechanism between catalysts and plasma is still unknown.

Recently, Wang et al. examined the influence of plasma reactor structure and catalysts on CO₂ conversion and CH₃OH selectivity for plasma catalytic CO₂ hydrogenation [125], and a schematic diagram of the experimental setup and images of the H₂/CO₂ discharge are shown in Figure 22. They investigated three kinds of reactors, i.e., a cylindrical reactor (aluminum foil sheet as ground electrode), a double dielectric barrier discharge reactor (water as ground electrode), and a single dielectric barrier discharge reactor (water as a ground electrode). The single DBD reactor equipped with a special water-electrode showed the optimal reaction performance (21.2% CO₂ conversion and 53.7% CH₃OH selectivity). In addition, they tested the catalytic performance of Pt/ γ -Al₂O₃ catalysts and Cu/ γ -Al₂O₃ catalysts in the optimized reactor for direct conversion of CO₂ to CH₃OH. As shown in Figure 23, both Pt/ γ -Al₂O₃ catalysts and Cu/ γ -Al₂O₃ catalysts showed a better catalytic performance (21.2% CO₂ conversion, but also the CH₃OH selectivity, and Cu/ γ -Al₂O₃ catalysts showed a better catalytic performance (21.2% CO₂ conversion, but also the CH₃OH selectivity, and Cu/ γ -Al₂O₃ catalysts showed a better catalytic performance (21.2% CO₂ conversion, but also the CH₃OH selectivity, and Cu/ γ -Al₂O₃ catalysts showed a better catalytic performance (21.2% CO₂ conversion and 53.7% methanol selectivity). That is, a strong synergistic effect between the plasma and Cu/ γ -Al₂O₃ catalysts promoted the hydrogenation of CO₂ to methanol, although the reaction temperature in the optimized reactor remained near room temperature.



Figure 22. (a) Schematic diagram of the experimental setup of a DBD plasma catalytic reactor. (b) Images of H_2/CO_2 discharge generated in DBD reactor without catalyst, reprinted with permission from [125]. Copyright American Chemical Society, 2017.



Figure 23. Effect of H_2/CO_2 molar ratio and catalysts on the reaction performance of the plasma hydrogenation process, reprinted with permission from [125]. Copyright American Chemical Society, 2017.

It is obvious that the combination of the plasma and catalysts can enhance the catalytic reaction at room temperature and atmospheric pressure. The maximum methanol selectivity of 53.7% was achieved with a CO_2 conversion of 21.2% via the plasma-catalysis process (see Figure 23). All the above studies show a strong synergistic effect between plasma and catalysts.

Using one-dimensional fluid modeling [126], De Bie et al. proposed a chemical reaction network in the plasma (i.e., without catalysts) for conversion of CO₂ to value-added chemicals, i.e., CO, CH₄, CH₂O, CH₃OH, and hydrocarbons. The simulation results indicated the dominant reaction pathways for the conversion of CO_2 and H_2 , as illustrated in Scheme 5. According to the model, the combination between H atoms and CHO radicals is the most important reaction to form CO, while this reaction is counterbalanced by the reorganization of H with CO into CHO radicals. Therefore, the most effective net CO formation reaction is dissociation of CO_2 influenced by electrons. The production of CH_4 was generally driven by two reactions, i.e., three-body recombination reaction between CH₃ and H radicals, and charge transfer reaction between CH_5^+ and H_2O . However, the latter reaction is partly balanced by the loss of CH_4 , resulting from a charge transfer reaction with H_3^+ . The production of CH₂O is closely related to the initial CO₂ fraction in the gas mixture. At a low initial fraction of CO₂, the reaction between CO₂ and CH₂ radicals seem to be the most important channel for the formation of CH₂O, while at higher initial CO₂ fractions, CH₂O is also produced out of two CHO radicals to some extent. In addition, as predicted by the model, the most important channel for the formation of CH₃OH is the three-body reaction between CH₃ and OH radicals, while the three-body reaction between CH₂OH and H radicals is also an effective production channel for CH₃OH. Furthermore, they also found that a higher density of CH₃ and CH₂ radicals would be essential to tune the distribution of end products. Therefore, it can be predicted that the degree of hydrogenation in the reaction has a significant influence for the targeted products.

Figure 24 summarizes the selectivity of CO, CH₄, or CH₃OH, as a function of the CO₂ conversion, obtained from the recent reports discussed above. Several main trends are clear. Firstly, the main products are CO and CH₄ for direct hydrogenation of CO₂ in plasma catalysis, while the selectivity of CH₃OH is relatively low. Moreover, it is obvious that researchers have paid more attention to the reduction of CO₂ to CO and CH₄ up to now. However, the production of other hydrocarbons, such as olefins and gasoline hydrocarbons, would also be a promising direction, to achieve the maximum utilization of CO₂ by plasma catalysis. Therefore, some insights from heterogeneous catalysis, especially the combination of metal catalysts, metal oxide catalysts, and zeolite catalysts could be helpful to develop a methodology for plasma catalytic hydrogenation of CO₂. On the other hand, there is no guarantee that good catalysts in thermal processes would also perform well in plasma

catalysis, because of the clearly different operating conditions (e.g., lower temperature, abundance of reactive species, excited species, charges, and electric field present in plasma).



Scheme 5. Dominant reaction pathways for the plasma-based conversion (without catalysts) of CO_2 and H_2 into various products, in a 50/50 CO_2/H_2 gas mixture. The thickness of the arrow lines is proportional to the rates of the net reactions. The stable molecules are indicated with black rectangles, reprinted with permission from [126]. Copyright American Chemical Society, 2016.



Figure 24. Overview of selectivity into CO, CH_4 , and CH_3OH , as a function of CO_2 conversion, based on all reports available in literature about plasma catalysis (all in DBD reactors). The references where these data are adopted from are discussed in the text.

Up to now, in view of the complexity of this interdisciplinary field, the development of plasma catalysis still needs major research efforts. On one hand, plasma catalysis has potential for industrial application, because it drives the CO_2 conversion reaction at ambient temperature and atmospheric pressure, breaking thermodynamic equilibrium to make full use of feedstock. On the other hand,

the excessive energy consumption—caused by high input power and heat loss—is an important disadvantage, but it can be mitigated with the further development of renewable energy (i.e., wind, solar, and tidal energy). Additionally, to improve the reaction activity for CO_2 conversion, it is crucial to explore the reaction mechanisms by in situ characterization and computer modeling, to improve the synergistic effect between plasma and catalysts. Finally, we need to search for suitable catalytic materials to strength the reaction performance and decrease the production costs.

4. Outlook and Conclusions

Currently, fuels and base chemicals are nearly all produced from non-renewable fossil energy (oil, natural gas, and coal), and CO_2 is generally the end product (e.g., upon burning fossil fuels) or a waste product in chemical industry. This indicates that we should use CO_2 as the main carbon source when fossil energy would get depleted in the future. Thus, in the long term, hydrogenation of CO_2 (as well as CO_2 conversion with other H-sources) into value-added chemicals and fuels is very significant, since it can close the carbon cycle, as shown in Figure 25. However, some crucial issues should be addressed in advance for application of CO_2 hydrogenation in industry.



Figure 25. CO₂ conversion into fuels, which release CO₂ again upon burning, aiming a closed carbon cycle.

From a relatively economical point of view, the direct hydrogenation of CO_2 is not yet a viable approach. On one hand, CO_2 of a certain purity generally depends on the development of capture and separation techniques. On the other hand, compared with the price of the feedstock (H₂, 10,000 \$/ton), the price of the main products (i.e., liquefied natural gas, 770 \$/ton, and CH₃OH, 340 \$/ton) obtained by CO_2 hydrogenation is too low to make this an economically viable process. However, once the production of H₂ can be realized with fully-fledged solar power technology, the production cost for CO_2 hydrogenation will be greatly reduced. Meanwhile, the CO_2 conversion driven by solar energy (artificial photosynthesis) is also a promising routing.

Technologically, the direct hydrogenation of CO_2 to CO is an endothermic reaction ($\Delta H = 41.2 \text{ kJ mol}^{-1}$), and a higher reaction temperature is beneficial for the production of CO according to Le Chatelier's principle. Based on the available experimental results (Table 1), the CO selectivity can reach nearly 100% under conditions of 873 K and 1 MPa. On the other hand, the production of other products (i.e., CH₄, CH₃OH, DME, and light olefins) from CO₂ is exothermic, and in theory a low temperature favors the equilibrium conversion. However, the inert CO₂ molecule generally needs relatively high reaction temperature to be activated. The competition between these two factors makes the reaction performance not very satisfactory. Up to now, from the perspective of technology

readiness level, methanol synthesis is a far more advanced production process, with a high selectivity up to 80–90% with Cu-based catalyst (Table 3). However, there are some limitations in heterogeneous catalysis for direct hydrogenation of CO_2 , such as catalyst sintering at high temperature, the influence of water produced in the reaction process and low CO_2 conversion caused by thermo-dynamical restriction in terms of the formation of CH_3OH and hydrocarbons. Hence, we should explore novel catalytic materials and reactors to break the thermodynamic equilibrium, as well as design bi-functional and/or multi-functional catalysts (metal, metal oxides, and zeolites) to couple the chemical reactions (RWGS, methanation, methanol synthesis, and/or MTO).

Plasma catalysis, a new field of catalysis, has attracted sufficient attention in recent years, due to its simple operating conditions (ambient temperature and atmospheric pressure) and unique advantages in activating inert molecules. However, the energy efficiency is still too high for commercial exploitation. On the other hand, renewable energy (e.g., wind, solar, and tidal energy) will further develop in the near future, and it will be a perfect match with plasma catalysis (because plasma is generated by electricity and can simply be switched on/off-allowing storage of fluctuating energy), hence making the problem of limited energy efficiency less dramatic. Nevertheless, significant efforts must be devoted to elucidate the reaction mechanisms in plasma catalysis, to improve not only the energy efficiency, but certainly also the selectivity towards value-added products. Indeed, plasma catalysis is complicated from chemistry and physics point of view, but the potential of plasma technology for industrial applications deserves major research efforts. A combination of computer simulations with experiments will be needed for an in-depth understanding of the reaction mechanisms, responsible for the synergy between plasma and catalysts. Although the development process in plasma catalysis may be slow due to its complex character, it would bring great benefits to human society, once it is developed mature enough, and once appropriate catalysts for plasma catalysis can be designed. Therefore, future research should definitely emphasize on a better understanding and rational screening of highly active catalysts.

Compared with the numerous studies on CO_2 hydrogenation by heterogeneous catalysis, much more research should be carried out in the field of plasma catalysis, to improve the CO_2 conversion and target products selectivity (and even to exploit new products, such as olefins, gasoline hydrocarbons, and aromatics), since the results achieved by plasma catalysis (Figure 24) are far from those of heterogeneous catalysis (Figure 18). To achieve this goal, the advantage of plasma should be exploited in full, and at the same time, insights from heterogeneous catalysis (e.g., catalyst combination, reaction combination, active sites design, etc.) can help to further improve the potential of this promising field of catalysis.

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Article H₂O and/or SO₂ Tolerance of Cu-Mn/SAPO-34 Catalyst for NO Reduction with NH₃ at Low Temperature

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Abstract: A series of molecular sieve catalysts (Cu-Mn/SAPO-34) with different loadings of Cu and Mn components were prepared by the impregnation method. The deNO_x activity of the catalyst was investigated during the selective catalytic reduction (SCR) of NO with NH₃ in the temperature range of 120 °C to 330 °C, including the effects of H₂O vapors and SO₂. In order to understand the poisoning mechanism by the injection of H_2O and/or SO_2 into the feeding gas, the characteristics of the fresh and spent catalyst were identified by means of Brunner-Emmet-Teller (BET), X-ray Diffraction (XRD), Scanning Electronic Microscopy (SEM) and Thermal Gravity- Differential Thermal Gravity (TG-DTG). The conversion of NO by the catalyst can achieve at 72% under the reaction temperature of 120 $^\circ$ C, while the value reached more than 90% under the temperature between 180 °C and 330 °C. The deNO_x activity test shows that the H₂O has a reversible negative effect on NO conversion, which is mainly due to the competitive adsorption of H₂O and NH₃ on Lewis acid sites. When the reaction temperature increases to 300 $^{\circ}$ C, the poisoning effect of H₂O can be negligible. The poisoning effect of SO₂ on deNO_x activity is dependent on the reaction temperature. At low temperature, the poisoning effect of SO_2 is permanent with no recovery of deNO_x activity after the elimination of SO_2 . The formation of $(NH4)_2SO_4$, which results in the plug of active sites and a decrease of surface area, and the competitive adsorption of SO2 and NO should be responsible for the loss of deNO_x activity over Cu/SAPO-34.

Keywords: SCR; Catalyst; (NH₄)₂SO₄; deNO_x; H₂O and SO₂ poisoning

1. Introduction

Nitrogen oxides were estimated as one of the major air pollutants released from the combustion of fossil fuels (especially coal), being hazardous for the ecological and environmental system [1–4]. Currently, selective catalytic reduction (SCR) was regarded as a widely-used deNO_x technology for the purification of flue gas, where the performance of the catalyst plays a significant role in the process [5]. Most of the commercial catalysts (e.g., V-W-Ti system) exhibited the effective activity with temperature located in a narrow and relatively high window as 300–400 °C, accelerating the deactivation of catalyst through sintering and occlusion of salts produced from H₂O or SO₂. The high deNO_x activity of SCR catalyst at relatively low temperature is highly required without the formation of salts from H₂O and

 SO_2 in the flue gas. The stability of the air-preheating system can be improved along with the secure low-temperature SCR system, leading to a full-time deNO_x for the power plant under different power loadings. Therefore, it is of great significance to develop efficient and stable low-temperature SCR catalyst [6].

Among all these catalysts, transition metal loaded on zeolite materials with CHA structure have been widely focused due to the broad operating temperature range and the high deNO_x activity [7–9]. A number of research works have been conducted on the deNO_x performance of molecular sieve catalysts, such as the ZSM-5, BEA, USY, SSZ-13 and so on [10–12]. Kim [13] prepared Mn–Fe/ZSM-5 by the impregnation method, exhibiting the NO_x conversion ratio as high as 95% at 175 °C, and the conversion of NO_x nearly to 100% during the temperature between 200 and 350 °C. SAPO-34 possesses pear-shaped cages with 8-membered ring (8MR) openings, and double 6-membered ring (D6MR) units linked by 4-membered ring (4MR) units, while most of the P is replaced by Si to generate a Si–O–Al linkage, resulting in remarkable SCR performance. Compared with Fe-zeolites and vanadia-zeolites, Cu-zeolites exhibited a superior deNO_x activity and high N₂ selectivity [14,15]. Wang [16,17], Ye [18] and Deka [19] prepared Cu/SAPO-34 by different methods, exhibiting an excellent low-temperature SCR activity in the temperature range from 150 °C to 500 °C. Among all these researches, Cu-SAPO-34 based catalysts showed a remarkable deNO_x activity compared to other zeolites [8,20].

However, Cu-SAPO-34 catalyst is proved to be sensitive to SO₂ poisoning due to the strong chemical binding strength and the oxidative conditions in the gas and the deactivation effect is more pronounced at low temperatures [21]. Zhang et al. [7] have used DRIFT and Temperature Programmed Desorption (TPD) method to study the poisoning effect of SO₂ over Cu/SAPO-34 catalyst, they reported that low-temperature deactivation is caused by the formation of ammonium sulfates. Shen et al. [22] observed no obvious sulfur species on the Cu/SAPO-34 catalyst and they concluded that the reduction of the number of isolated Cu²⁺ caused by SO₂ might induce the loss of SCR activity. Wijayanti et al. [23] studied the SO₂ poisoning effects and found that the main reason for the deactivation is the formation of copper sulfates, resulting in the loss of redox properties. Jangjou et al. [24] reported the S species formed on Cu²⁺ at 6MR by DRIFT study. On the basis of such observations, the formation of ammonium sulfates in a complex with Cu is claimed as the main mechanism for the loss of low-temperature deNO_x activity. In general, different SO₂ poisoning mechanisms have been proposed by different researchers.

In addition, H_2O is also one of the main components in the flue gas and often causes the loss of low-temperature deNO_x activity [3]. In this work, for a better and specific understanding of the SO₂ poisoning mechanism and the synergistic effect of H_2O and SO₂ over Cu/SAPO-34 catalyst, the influence of SO₂ or/and H_2O with different concentrations at different reaction temperatures on deNO_x activity and physicochemical properties over Cu/SAPO-34 catalyst was studied. A series of Cu-Mn/SAPO-34 catalysts were prepared through the impregnation method. The low-temperature deNO_x activity of the catalysts was estimated by means of a self-designed apparatus, where the effects of H_2O and SO₂ on the reaction activity were also investigated. BET, XRD, SEM and TG-DTG were employed to determine the characteristics of the fresh and spent catalyst, in order to understand the poisoning mechanism of the catalyst by H_2O and SO₂ during the low-temperature SCR process.

2. Results and Discussion

2.1. DeNOx Activity of the Catalysts Without H₂O and SO₂

Figure 1a illustrates the deNO_x activity of Cu/SAPO-34, Mn/SAPO-34, Cu–Mn/SAPO-34 catalysts under different temperatures without H_2O and SO_2 . The deNO_x activity of Cu–Mn/SAPO-34 catalyst (bimetallic composite molecular sieve) was much higher than that of Cu/SAPO-34 and Mn/SAPO-34 catalysts (monomeric molecular sieve). Even when the reaction temperature is lower than 100 °C the NO conversion by Cu–Mn/SAPO-34 catalyst can be achieved at about 60%, while for the other two catalysts was around 20%.



Figure 1. $DeNO_x$ performance of mono-component and multi-component catalysts (a); $deNO_x$ performance of molecular catalysts with different *Mn* loadings (b); outlet N₂O concentration over Cu(2)-Mn(6)/SAPO-34 catalyst (c); outlet NO₂ concentration over Cu(2)-Mn(6)/SAPO-34 catalyst (d).

The NO conversion over Cu–Mn/SAPO-34 catalyst could reach 90% at around 180 °C, compared to that of 220 °C for Cu/SAPO-34 and 270 °C for Mn/SAPO-34. It can be concluded that the Cu–Mn/SAPO-34 catalyst gives a better deNO_x performance under the relatively low-temperature range from 100 to 200 °C. This can be attributed to the promotion of NH₃ adsorption on the surface of the catalyst by the bimetallic system on SAPO-34 [25–27]. In addition, the deNO_x reaction energy might be declined by the bimetallic interaction on the catalyst, improving the low-temperature activity of the catalyst and broadening its SCR temperature range.

The effect of metal loading on the activity of Cu-Mn/SAPO-34 catalyst is shown in Figure 1b. The deNO_x activity of Cu(2)–Mn(6)/SAPO-34 performed as the best one among the catalysts at the temperature from 120 to 210 °C and from 270 to 330 °C, while Cu(2)–Mn(8)/SAPO-34 performs the best one at the temperature from 210 to 270 °C. It needs to be noted that the NO conversion over Cu–Mn/SAPO-34 with the Mn content more than 8% was notably declined when the reaction temperature is higher than 270 °C. Under high temperatures, the oxidation of the catalyst can be enhanced with the increased loading of Mn [28]. This leads to the oxidation of NH₃ to NO and then the decline of the NO conversion (Figure 1b). In addition, the nonselective catalytic reduction (NSCR) reaction and catalytic oxidation reaction (i.e., the C–O reaction) happens simultaneously during the NH₃-SCR reaction. The outlet N₂O and NO₂ concentration over Cu(2)–Mn(6)/SAPO-34 catalyst are shown in Figure 1c,d. It can be seen that the N₂O concentration increases with the increase of reaction temperature, which may contribute to the decrease of deNO_x activity over Cu(2)–Mn(6)/SAPO-34 catalyst at 330 °C. With the increase of reaction temperature, the NO₂ concentration decreases. These results demonstrate that N_2 is the main product of NH_3 -SCR reaction over Cu(2)–Mn(6)/SAPO-34 catalyst according to the "standard SCR" reaction [29].

2.2. Effect of H_2O on the deNO_x Activity of the Catalyst

The effect of the injected H_2O concentration into feeding gas on the deNO_x activity of Cu(2)–Mn(6)/SAPO-34 catalyst at a reaction temperature of 240 °C was shown in Figure 2a. The NO conversion decreased with an increased H_2O concentration. When the volume concentration of 2% water vapor was injected into the original feeding gas after 8 h, the deNO_x activity of the catalyst was declined from 93% to 91%. While the volume concentration of the vapor injected into the feeding gas was increased to 10%, after 8 h, the activity of the catalyst decreased to 86%. It needs to be noted that the activity of catalyst after the 8-hour injection of vapor was recovered to its original level after a while of the vapor cut-off in spite of the concentration between H_2O vapor and NH_3 or NO, which is reversible. When the water vapor concentration was increased, the activity of the spent catalyst (after the cut-off of vapor) was a little bit lower than that of the original catalyst. It is demonstrated that the hydroxyl may be created due to the adsorption and decomposition of H_2O on the surface of Cu/SAPO-34 with the increase of H_2O concentration, resulting in the irreversible loss of deNO_x activity [3].



Figure 2. Effect of the injected H₂O concentration on the deNO_x activity of the catalyst (240 $^{\circ}$ C) (**a**); effect of H₂O on the deNO_x activity of catalyst under different temperatures (10% H₂O) (**b**).

Figure 2b shows the effect of H_2O on the deNO_x activity of the catalyst Cu(2)–Mn(6)/SAPO-34 under different temperatures with the concentration of water vapor as 10%. The deNO_x activity of the catalyst was significantly influenced at low reaction temperature with the presence of H_2O . A sharp decline of the deNO_x activity of catalyst (from 86% to 68%) can be observed under the reaction temperature of 180 °C after the 4-hour injection of vapor into the feeding gas. However, no significant change in the activity of the catalyst can be found for the reaction under 300 °C in the presence of vapor. After 8 h injection of the vapor under 180 °C, the activity of the catalyst was all recovered to its original level in 2 h, while the activity recovery duration for the catalyst was decreased with the increased reaction temperature after the cut-off of vapor. The adsorption capacity of H_2O on the surface of the catalyst can be enhanced under the lower temperatures, occupying more active sites than that of NH₃, NO and other reaction gases [30]. It can be also concluded that poisoning performance of water on the activity of the catalyst can be ignored while the reaction temperature is increased over 300 °C.

2.3. Effect of SO_2 on the deNO_x Activity of the Catalyst

The effect of the injected SO₂ concentration (from 500 ppm to 1500 ppm) on the deNO_x activity of the catalyst Cu(2)–Mn(6)/SAPO-34 under the reaction temperature of 240 $^{\circ}$ C was shown in Figure 3a.

It is obvious that the deNO_x activity of the catalyst was remarkably and abruptly decreased with the injection of SO₂ for the three different concentrations. After 2 h injection of SO₂, the deNO_x activity of catalyst was reduced from 97% to 72%, 68%, 63% for the SO₂ concentration of 500, 1000 and 1500 ppm. Compared to that of the injected H₂O, the high concentration of the injected SO₂ could accelerate the decline of the activity of the catalyst. The formation of SO₃ can be promoted by the high concentration of the injected SO₂, consequently enhancing the formation of ammonium sulfate with NH₃ covering the catalyst active sites on the surface. After the cut-off (8 h) of the injected SO₂, the activity of the catalyst was increased, but much lower than its initial activity. It is indicated that the competitive adsorption between SO₂ and NH₃ or NO is not the main reason for the loss of deNO_x activity in the presence of SO₂. Part of active sites on the catalyst was occupied by SO₂ over NH₃ and NO leading to the temporary poisoning, while a great number of active sites was covered by the formed sulfate salts (such as ammonium sulfate) for the permanent deactivation of the catalyst [31,32].



Figure 3. Effect of the injected SO₂ concentration of on deNO_x activity of catalyst (240 °C) (**a**); Effect of the injected SO₂ (500 ppm) on deNO_x activity of catalyst under different temperatures (**b**).

The effect of the injected SO₂ on the deNO_x activity of the catalyst Cu(2)–Mn(6)/SAPO-34 under different reaction temperatures was shown in Figure 3b. The effect of the injected SO₂ on the activity of the catalyst was greatly influenced by the reaction temperature. It needs to be noted that the activity of the catalyst was decreased from 90% to 29% after 8-hour injection of 500 ppm SO₂ under 180 °C. Comparatively, under the reaction temperature of 300 °C, the activity of the catalyst was reduced from 99% to 88% after 8-hour injection of 500 ppm SO₂. This result demonstrates that the effect of SO₂ on deNO_x activity is greatly related to the reaction temperature and a much more rapid decrease of deNO_x activity happened with the addition of SO₂ at a lower temperature. It should be noted that no obvious recovery of deNO_x activity at the reaction temperature of 180 °C can be observed. The more detailed mechanism of SO₂ poisoning over Cu/SAPO-34 at different reaction temperatures will be further discussed.

2.4. Effect of Both H_2O and SO_2 Injection on the deNO_x Activity of the Catalyst

Figure 4 shows the effect of injection of H_2O and SO_2 on the catalytic activity of Cu(2)–Mn(6)/SAPO-34. NO conversion was decreased at the presence of H_2O and SO_2 , compared to the performance the individual injection of H_2O or SO_2 . Moreover, the co-existence of H_2O and SO_2 in the feeding gas gave a much more serious decline on the deNO_x activity of the catalyst compared to the sum of the effect of H_2O and SO_2 respectively, which is labeled as the estimated value as shown in Figure 4. The existence of H_2O could enhance the deactivation of the catalyst by SO_2 through two ways: (1) SO_2 in gas phase reacted with the H_2O adsorbed on active site of catalyst to generate sulfuric

acid or sulfurous acid which is easily reacted with NH_3 ; (2) the thermal decomposition of the formed ammonium sulfate on the active sites was confined at the presence of H_2O .



Figure 4. Effect of the injected H₂O and SO₂ on the deNO_x activity of catalyst (240 °C).

2.5. Mechanism of the Catalyst Poisoning by H₂O and/or SO₂

In order to identify the change of crystalline phases for the fresh and spent Cu(2)-Mn(6)/SAPO-34 catalyst after SO₂ and/or H₂O poisoning. Powder X-ray diffraction (XRD) measurements were carried out and the patterns are shown in Figure 5. For the fresh Cu/SAPO-34 catalyst, sharp diffraction peaks corresponding to CHA phases are obtained. After poisoned with SO₂ or H₂O, the intensity of each peak decreases, especially for the catalyst after $SO_2 + H_2O$ treatment, indicating the obvious skeleton damage of SAPO-34 after adding $SO_2 + H_2O$ for 8 h [33]. No obvious crystalline change can be observed in the presence of 10% H₂O, this indicates that the phase poisoned by H₂O existed either in amorphous form or in the particle beyond the limited size of XRD detection [34]. For the catalyst after SO₂ poisoning for 8 h, A number of new diffraction peaks at $2\theta = 11.9^{\circ}$, 29.8° and 40.7° can be observed. The peaks at 11.9° and 29.8° can be attributed to NH₄HSO₄ and (NH₄)₂SO₄, while the latter peak at 40.7° is assigned to MnSO₄. For the catalyst after SO₂ + H₂O poisoning for 8h, new diffraction peaks at $2\theta = 29.8^{\circ}$ and 34.1° can be observed, which can be assigned to $(NH_4)_2SO_4$. In addition, the new peak at $2\theta = 21.1^{\circ}$ can be assigned to the formation of CuSO₄. It can be deduced that the newly formed NH_4HSO_4 , $(NH_4)_2SO_4$ and $MnSO_4$ result in the loss of $deNO_x$ activity in the presence of SO₂. When H₂O and SO₂ were added simultaneously, SO₂ in gas phase may react with adsorbed H₂O to generate sulfuric acid or sulfurous acid, which is easily reacted with NH₃ to form (NH₄)₂SO₄. Thus, the formation of $(NH_4)_2SO_4$ and $CuSO_4$ may cause the loss of deNO_x activity in the presence of $SO_2 + H_2O$.

In order to determine the morphology before and after SO_2 or/and H_2O poisoning, the fresh Cu/SAPO-34 and poisoned catalyst by SO_2 or/and H_2O were characterized by SEM method shown in Figure 6. No significant change of the surface morphology can be found between the catalysts before and after the H_2O poisoning. The active particles on the surface of fresh catalyst were replaced by the bulks of ammonium sulfate after the poisoning by individual SO_2 , inhibiting the deNO_x activity of the catalyst. Similar phenomena can be observed for the spent catalyst in presence of H_2O and SO_2 , where the agglomeration of ammonium sulfate lead to the decline in the surface active sites of the catalyst and thus the activity of the catalyst.



Figure 5. X-ray diffraction spectra of the fresh and spent Cu(2)-Mn(6)/SAPO-34.



Figure 6. SEM micrographs of fresh and spent Cu(2)-Mn(6)/SAPO-34: (a) fresh; (b) 10% H₂O 240 $^{\circ}$ C 8 h; (c) 500 ppm SO₂ 240 $^{\circ}$ C 8 h; (d) 10% H₂O and 500 ppm SO₂ 240 $^{\circ}$ C 8 h.

The TG curves of the poisoned catalyst in the presence of individual SO₂ and both H₂O and SO₂ were shown in Figure 7. Three mass loss stages can be observed for the two kinds of the poisoned catalyst: the first mass loss in the temperature between 50 °C and 100 °C was attributed to the dehydration of the catalyst; the second stage between 200 °C and 400 °C can be designated to the thermal decomposition of NH₄HSO₄ and (NH₄)₂SO₄ is 200 °C [35]; the third stage between 600 °C and 800 °C can be attributed to the decomposition of the geometry in the presence of H₂O and SO₂ is notably higher than that of the poisoned catalyst. This confirms that the formation of ammonium sulfate can be facilitated and promoted with the injection of H₂O in the feeding gas, which is consistent with the XRD and TG-DTG analysis. While the peak value of the third mass loss stage of the poisoned catalyst in the presence of SO₂. It might be because the formation of (NH₄)SO₄ can inhibit the formation of sulfate metal salts due to the consumption of SO₂.

The specific surface area, pore volume and pore size of fresh and spent Cu(2)–Mn(6)/SAPO-34 were determined by N₂ adsorption and summarized in Table 1. The specific surface area and pore volume of Cu(2)–Mn(6)/SAPO-34 catalyst was decreased after the all different NO reduction experiments with or without the injection of H₂O and/or SO₂. It can be seen that the damage of the surface area is related to the concentration of SO₂ or H₂O and the reaction temperature. With the increase of reaction temperature, the damage of the surface area caused by SO₂ is lightening. The low reaction temperature could confine the thermal decomposition of the formed NH₄HSO₄ and (NH₄)₂SO₄ on the surface, resulting in a decline in surface area and the blockage of pore channel [37]. With the increase of the concentration of SO₂, a change of which (from 331 m²/g to 320 m²/g) can be observed for different SO₂ concentrations (from 500 ppm to 1500 ppm) under the temperature of

240 °C. For the poisoning effect of H₂O, with the increase of reaction temperature, the existence of H₂O has a more severe effect on the surface area of Cu/SAPO-34, which may result from the inhibited adsorption capacity of H₂O. With the increase of the concentration of H₂O, the surface area decreases from 425 m²/g to 408 m²/g as the concentration of H₂O increased from 2% to 10% as well as the pore volume of catalyst. When H₂O and SO₂ added into the gas stream simultaneously for 8 h, the surface area decreases from 457 m²/g to 276 m²/g, which is a dramatically decrease compared to the injection of SO₂ for 8 h at 240 °C (from 457 m²/g to 331 m²/g) and the injection of H₂O for 8 h at 240 °C (from 457 m²/g). It is demonstrated that the synergistic poisoning effect of SO₂ and H₂O is enhanced, ascribed to the large amount of deposited (NH₄)₂SO₄ on the pore channel of Cu/SAPO-34, which is consistent with the TG-DTG results.



Figure 7. TG-DTG curves of the thermal decomposition for the Cu(2)-Mn(6)/SAPO-34 catalyst after 8 h poisoning in presence of 500 ppm SO₂ and/or 10% H₂O.

Catalyst	BET Surface Area (m ² g ⁻¹)	Pore Volume (cm ³ g ⁻¹)	Pore Diameter (nm)
Fresh catalyst	457	0.23	2.08
Catalyst exposed to 500 ppm SO2 at 180 °C for 8 h	307	0.14	2.42
Catalyst exposed to 500 ppm SO ₂ , at 210 °C for 8 h	314	0.16	2.29
Catalyst exposed to 500 ppm SO2, at 240 °C for 8 h	331	0.16	2.15
Catalyst exposed to 500 ppm SO2, at 270 °C for 8 h	367	0.19	2.09
Catalyst exposed to 500 ppm SO2, at 300 °C for 8 h	372	0.18	2.09
Catalyst exposed to 1000 ppm SO ₂ , at 240 °C for 8 h	321	0.16	2.14
Catalyst exposed to 1500 ppm SO ₂ , at 240 °C for 8 h	320	0.16	2.28
Catalyst exposed to 10% H ₂ O, at 180 °C for 8 h	383	0.18	2.03
Catalyst exposed to 10% H ₂ O, at 240 °C for 8 h	408	0.19	2.02
Catalyst exposed to 10% H ₂ O, at 300 °C for 8 h	441	0.21	2.04
Catalyst exposed to 2% H2O, at 240 °C for 8 h	425	0.20	2.03
Catalyst exposed to 6% H2O, at 240 °C for 8 h	413	0.19	1.95
Catalyst exposed to 10% H2O+500 ppmSO2, at 240 °C for 8 h	276	0.17	2.62

Table 1. Brunner–Emmet–Teller (BET) analysis of the fresh and spent Cu(2)–Mn(6)/SAPO-34 catalyst.

Based on the activity tests and characterizations, the poisoning effect of SO_2 or/and H_2O can be described as follows. The poisoning effect of H_2O , especially at low temperatures, can be ascribed to the competitive adsorption between H_2O and NH_3 on Lewis acid sites by occupying the metal sits [38], which can be recovered to nearly the original activity after H_2O was removed. The XRD patterns of the catalyst after H_2O poisoning did not show ant obvious changes, indicating no change with the active metal sites.

While for the poisoning effect of SO₂ over Cu/SAPO-34 catalyst, the deNO_x activity decreases from 90% to 29% at 180 °C in a short time of SO₂ injection. In addition, the deNO_x activity could not be

recovered after the elimination of SO₂, indicating the permanent deactivation of SO₂ on Cu/SAPO-34 at 180 °C. The poisoning mechanism could be summarized as three aspects. Firstly, SO₂ in the gas may be oxidized to SO₃ and further react with NH₃ to form NH₄HSO₄, which is a drying powdery decomposed at 280 °C [39]. The formation of NH₄HSO₄ causes the plug of active sites of the catalyst and the decline in surface area. Ammonium sulfate crystallite is observed on the XRD spectra and the weight loss peak ascribed to the decomposition of ammonium sulfates is also observed from the TG curves. Secondly, the active sites (i.e., MnO₂ or CuO species) may react with SO₂ or SO₃ to form MnSO₄ or CuSO₄, which inhibited the redox properties [23]. The diffraction peaks assigned to MnSO₄ and $CuSO_4$ are observed in XRD spectra. Thirdly, the competitive adsorption of SO_2 and NO on metal sites may be a part of the reason for the loss of $deNO_x$ activity at the reaction temperature of above 180 °C [40]. With regards to the synergistic effect of SO₂ + H₂O, the deNO_x activity tests and characterizations show that the existence of H₂O could enhance the deactivation of the catalyst by SO₂ through two ways: (1) SO₂ in gas phase reacted with the H₂O adsorbed on active site of catalyst to generate sulfuric acid or sulfurous acid which is easily reacted with NH3 to form large amount of (NH₄)₂SO₄; (2) the thermal decomposition of the formed ammonium sulfate on the active sites was confined at the presence of H_2O . Both of the two explanations can be assigned to the deposition of (NH₄)₂SO₄, which further plug the pore channel of catalyst and cause the rapid decrease of surface area as shown in Table 1.

3. Materials and Methods

3.1. Catalyst Preparation

The molecular sieve was modified by impregnation method. A certain amount of zeolite molecular sieve H-SAPO-34 (n(P_2O_5):n(SiO_2):n(Al₂O₃) = 1:1:1), provided by the catalyst factory of Nankai University, was weighed and dried in a drying oven at 105 °C for 30 min. A certain amount of Cu(NO₃)₂•3H₂O powder was mixed with manganese nitrate solution with the 50 wt.% in the beaker of 200 mL, and then added 50 mL of deionized water into the immersion liquid. The beaker with a magnetic stirrer inside was immersed in a water bath at a constant temperature of 40 °C. After 12 h of immersion, the solution was thoroughly mixed and heated until the moisture was completely evaporated. The powder was then put into a dry oven at about 100 °C for 12 h. The dried powder was ground and sieved by the 40 to 60 mesh. The obtained powders were placed in a tube furnace and calcined at 450 °C for 6 h to obtain the catalyst sample for the experiment. The catalyst sample Cu(2)-Mn(6)/SAPO-34(450) indicated that the 2 wt.% Cu and 6 wt.% Mn were loaded on the molecular sieve SAPO-34 with the calcination temperature of 450 °C.

3.2. DeNOx Activity Measurements

DeNO_x activity measurements using NH₃ were carried out in a fixed-bed stainless steel tubular flow reactor (Inner diameter: 16 mm) (Figure 8). The tube furnace of the reactor can be heated from the room temperature to 800 °C The gas feeding system was composed of five gas feeding pipes controlled by the mass flow meter (0–1.5 L) and one liquid feeder a microinjection pump (0.001 μ L/min–127 mL/min), equipped with the reactor for adjusting the composition of the feeding gas. The reaction temperature in the experiments was set to be changed from 90 to 330 °C.

Two milliliter catalyst was placed on the holder of the reactor, and the feeding gas consisted of 350 ppm NO, 350 ppm NH₃, 3 vol.% O₂ and N₂ as balanced. H₂O (0–10 vol.%) and/or 500–1500 ppm SO₂ would be injected into the feeding gas for investigating the effect of H₂O and/or SO₂ on the catalyst poisoning. The GHSV was set to be 15,000h⁻¹. The NO and NO₂ concentration of the reactor inlet and outlet was collected by an airbag and analyzed by the flue gas analyzer (Testo 350, Testo, Inc., Lenzkirch, Germany). The outlet N₂O concentration is collected by the N₂O analyzer (Medi-Gas G200,

Bedfont Scientific Ltd., Bedfont, United Kingdom). The NO conversion was obtained from the equation as follows:

$$\eta = \frac{C_{NO}^{in} - C_{NO}^{out}}{C_{NO}^{in}} \times 100\%$$

where η , C_{NO}^{in} , C_{NO}^{out} represented the NO conversion, inlet and outlet NO concentration.



Figure 8. Schematics for the fixed-bed tubular flow reactor.

3.3. Catalyst Characterization

A micromeritics ASAP 2010M micropore size analyzer was used to measure the N₂ adsorption isotherms of the catalyst sample at liquid N₂ temperature (-196 °C. Specific surface area, pore volume and pore diameter can be determined by N₂ adsorption using the BET and BJH methods.

The XRD measurement for the catalyst was carried out on a Rigaku D/Smartlab(III) system (Rigaku, Neu-Isenburg, Germany) with Cu Ka radiation. The X-ray source was operated at 40kV and 40mA. The diffraction patterns were taken in the 2θ range of 5–50° at a scan speed of 10° min⁻¹ and a resolution of 0.02° .

SEM was performed using a SIRION-50 scanning electron microscope from Field Electron and Ion Company, the Netherlands, with a resolution of 150 eV.

Thermal gravimetric analysis of the catalyst samples was performed using TGA-101 type produced by the Nanjing Exhibition Electrical and Mechanical Technology Company (Nanjing, China). The accuracy of the instrument is 0.2 μ g. For the TG experiments, the spent catalyst samples were measured under the temperature from room temperature to 900 °C at the heating rate of 15 °C min⁻¹.

4. Conclusions

Cu-Mn/SAPO-34 with the loading of 2 wt.% Cu and 6 wt.% Mn exhibited remarkable low-temperature NO reduction activity among the prepared Cu-Mn/SAPO-34 catalysts. The NO conversion could achieve as high as 72% under the reaction temperature of 120 °C, while the value reached more than 90% under the temperature between 180 °C and 330 °C. The reversibly poisoning effect of H₂O is mainly due to the competitive adsorption between H₂O and NH₃ on Lewis acid sites by occupying the metal sits. With the increase of reaction temperature, the poisoning effect is less important. The poisoning effect of SO₂ on deNO_x activity is dependent on the reaction temperature. At low temperature, the poisoning effect of SO₂ is permanent with no recovery of deNO_x activity after the elimination of SO₂. XRD, SEM and BET analysis suggested that the deposition of (NH₄)₂SO₄ on active sites may be the main reason for the loss of deNO_x activity. TG-DTG analysis shows that some metal sulfates are formed on the surface of Cu/SAPO-34 catalyst, which may inhibit the redox properties and cut off the redox cycle during the low-temperature SCR reaction. The addition of H_2O into the SO₂-containing atmosphere promotes the formation of $(NH_4)_2SO_4$ on the surface of the catalyst, causing the damage of surface area and the rapid decrease of deNO_x activity. At a higher reaction temperature, the formation of ammonium sulfate might be inhibited by the high reaction temperature, or the formed ammonium sulfate from SO₂ and NH₃ can be easily decomposed under the high temperature.

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Article

Comprehensive Comparison between Nanocatalysts of Mn–Co/TiO₂ and Mn–Fe/TiO₂ for NO Catalytic Conversion: An Insight from Nanostructure, Performance, Kinetics, and Thermodynamics

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The nanocatalysts of Mn-Co/TiO₂ and Mn-Fe/TiO₂ were synthesized by Abstract: hydrothermal method and comprehensively compared from nanostructures, catalytic performance, kinetics, and thermodynamics. The physicochemical properties of the nanocatalysts were analyzed by N₂ adsorption, transmission electron microscope (TEM), X-ray diffraction (XRD), H2-temperature-programmed reduction (TPR), NH3-temperature-programmed desorption (TPD), and X-ray photoelectron spectroscopy (XPS). Based on the multiple characterizations performed on $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts, it can be confirmed that the catalytic properties were decidedly dependent on the phase compositions of the nanocatalysts. The $Mn-Co/TiO_2$ sample presented superior structure characteristics than $Mn-Fe/TiO_2$, with the increased surface area, the promoted active components distribution, the diminished crystallinity, and the reduced nanoparticle size. Meanwhile, the Mn⁴⁺/Mnⁿ⁺ ratios in the Mn–Co/TiO₂ nanocatalyst were higher than Mn-Fe/TiO₂, which further confirmed the better oxidation ability and the larger amount of Lewis acid sites and Bronsted acid sites on the sample surface. Compared to Mn-Fe/TiO₂ nanocatalyst, Mn–Co/TiO₂ nanocatalyst displayed the preferable catalytic property with higher catalytic activity and stronger selectivity in the temperature range of 75-250 °C. The results of mechanism and kinetic study showed that both Eley-Rideal mechanism and Langmuir-Hinshelwood mechanism reactions contributed to selective catalytic reduction of NO with NH₃ (NH₃-SCR) over $Mn-Fe/TiO_2$ and $Mn-Co/TiO_2$ nanocatalysts. In this test condition, the NO conversion rate of Mn–Co/TiO₂ nanocatalyst was always higher than that of Mn–Fe/TiO₂. Furthermore, comparing the reaction between doping transition metal oxides and NH₃, the order of temperature-Gibbs free energy under the same reaction temperature is as follows: $Co_3O_4 < CoO < Fe_2O_3 < Fe_3O_4$, which was exactly consistent with nanostructure characterization and NH₃-SCR performance. Meanwhile, the activity difference of MnO_x exhibited in reducibility properties and Ellingham Diagrams manifested the promotion effects of cobalt and iron dopings. Generally, it might offer a theoretical method to select superior doping metal oxides for NO conversion by comprehensive comparing the catalytic performance with the insight from nanostructure, catalytic performance, reaction kinetics, and thermodynamics.

Keywords: NH₃-SCR; nanostructure; kinetics; thermodynamics; manganese oxides

MDPI

1. Introduction

Nitrogen oxides (NO_x) generated from fossil fuels are regarded as the primary reason for acid rain, ozone depletion, photochemical smog, and greenhouse effects [1]. Selective catalytic reduction of NO_x with NH₃ or urea as a reductant (NH₃-SCR) is proposed to be the most efficient method for eliminating NO_x from stationary source and mobile source [2]. In recent decades, the commercial catalyst of V₂O₅-WO₃(MoO₃)/TiO₂ used for NH₃-SCR process exhibited an excellent catalytic property in the typical standard SCR reaction within the temperature range of 300–400 °C [3,4].

$$4NO + 4NH_3 + O_2 \rightarrow 4N_2 + 6H_2O$$
 (1)

Nowadays, for the purpose of reducing the inhibiting effects of ash and SO₂ over the SCR catalysts, there is an intense interest in developing catalysts for NO reduction at lower temperature [5]. However, the V₂O₅-WO₃(MoO₃)/TiO₂ catalysts demand a strict temperature window, which limits their arrangement flexibility. The vanadium-based catalysts cannot reach satisfactory efficiency of eliminating NO_x as the reaction temperature lower than 250 $^{\circ}$ C. Hence, the catalysts appropriate to low temperature SCR are strongly desired, which could be placed at the downstream of electrostatic precipitator and desulfurizer [6]. A great deal of catalysts comprised of different transition metal oxides on various supports were analyzed to improve the catalytic ability for low-temperature deNO_x. The typical transition metals, such as Mn [7], Co [8], Fe [9], Ni [10], Zn [11], and Cr [9], were well-reported to display satisfactory properties at the low temperature. Among the various transition metal elements applied in the catalysts for NO_x reduction, manganese oxides display superior activity especially at the low temperature, which can be attributed to the various types of labile oxygen and high mobility of valence states [12]. Meanwhile, it was found that cobalt and iron species can combine with manganese to produce mixed nanoparticle-oxides and exhibit high SCR activity and excellent N₂ selectivity with a wide temperature window from 100–300 °C [13,14]. These mixed nanoparticle-oxides contain abundant oxygen vacancies on the catalyst surface, forming strong interaction bands at atomic scale, such as Mn-O-Fe [15,16] and Mn-O-Co [17]. Meanwhile, the active metal species of CoO_x and FeO_x are also regarded as the typical promoters for NO_x conversion, which serve as core catalyst components of active metal oxides, supplying surface oxygen to accelerate NO_x elimination [17,18].

Another crucial field of the investigation on NH₃-SCR process is the reaction mechanism on the active sites with a better understanding of the surface chemistry and developed kinetic models to evaluate the reactions. It was widely accepted that the main reaction of low temperature SCR complies with both the Eley-Rideal mechanism and Langmuir-Hinshelwood mechanism, which is further depended on the reaction temperature and the catalyst components [19,20]. According to the Eley-Rideal mechanism, the reaction occurred between NO and NH₃ (adsorbed) to generate an activated transition state and further decomposed into H₂O and N₂ [20]. Based on the Langmuir-Hinshelwood mechanism, the reaction happened between adsorbed NO and adsorbed NH₃ on the adjacent sites to form H₂O and N₂ [21]. Meanwhile, it was widely accepted that the first step in SCR was an oxidative abstraction of the hydrogen from adsorbed ammonia [22]. Therefore, oxidative NH₃ played an important part in the mechanism of NH₃-SCR, which could be evaluated by thermodynamics in details [7]. However, the comprehensive comparison of NO elimination over the nanocatalysts of MnCoO_x and MnFeO_x on TiO₂ support, especially with an insight from structure, performance, kinetics, and thermodynamics, has not been explored clearly.

In this study, the nanocatalysts of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ were synthesized by hydrothermal method. The physicochemical properties of the obtained samples were researched by SEM, BET, XRD, H2-TPR, NH3-TPD, and XPS. In the meantime, the catalytic performance comparation of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts was investigated using kinetic and thermodynamic analysis. The purpose of this work was mean to explore a comprehensive perspective for optimizing multimetals SCR catalysts, especially for the Mn-based bimetals nanocatalysts.

2. Results and Discussions

2.1. Physicochemical Properties of the Nanocatalysts

2.1.1. TEM Analysis

The topographic characteristics and the porous structures of the $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts were collected by TEM test. As shown in Figure 1a, the most part of $Mn-Co/TiO_2$ nanocatalyst was formed with uniform elliptic nanoparticles with glabrous surfaces and well-proportioned size distribution and without apparent agglomeration, although there was a spot of tightly aggregated TiO₂ nanoparticles interfused into the smaller regular particles. Therefore, a distinct and unbroken mesh structure of micropore was form in $Mn-Co/TiO_2$ sample. However, according to the TEM images of $Mn-Fe/TiO_2$ nanocatalyst from Figure 1b, a noticeable augment in the particle size was observed. The nanoparticles were irregular, lots of which stacked on the nanocatalyst surface. According to Figure 1(c)(l), captured at a larger scope, it could be found that all active elements contained in $Mn-Co/TiO_2$ nanocatalyst were well dispersed without obvious irregular stacked particles. On the surface of $Mn-Fe/TiO_2$ nanocatalyst, the element of manganese appeared slight regional accumulation, which might be caused by the abundant micropore structure collapsing [23].



Figure 1. Cont.



Figure 1. TEM, SEM, and Mapping of Mn–Co/TiO₂ and Mn–Co/TiO₂ nanocatalysts. (a) TEM of Mn–Co/TiO₂; (b) TEM of Mn–Fe/TiO₂; (c) SEM of Mn–Co/TiO₂; (d) SEM of Mn–Fe/TiO₂; (e) Mapping of Mn on Mn–Co/TiO₂; (f) Mapping of Mn on Mn–Fe/TiO₂; (g) Mapping of Co on Mn–Co/TiO₂; (h) Mapping of Fe on Mn–Fe/TiO₂; (i) Mapping of Ti on Mn–Co/TiO₂; (j) Mapping of Ti on Mn–Fe/TiO₂; (k) Mapping of O on Mn–Co/TiO₂; (l) Mapping of O on Mn–Fe/TiO₂.

2.1.2. BET Analysis

The structure parameters of $Mn-Co/TiO_2$ and $Mn-Co/TiO_2$ nanocatalysts, such as the specific surface areas, total pore volumes, and average pore diameters, were investigated by nitrogen-adsorption-desorption, and the test data was summarized in Table 1. $Mn-Co/TiO_2$ nanocatalyst attained the larger specific surface areas than $Mn-Fe/TiO_2$, which was possible due to the improved dispersion of $MnCoO_x$ species on the nanocatalyst surface. In the meantime, the average pore diameters increased obviously, from 33.06 nm in the $Mn-Co/TiO_2$ sample to 54.85 nm in the $Mn-Fe/TiO_2$ sample. Therefore, the $MnCoO_x$ species were more likely to promote nanocatalyst to form micropores compared with $MnFeO_x$ species [24,25]. Nevertheless, the difference of total pore volumes between $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts was not conspicuous. The total pore volumes reduced slightly from 0.531 cm³·g⁻¹ in $Mn-Co/TiO_2$ sample to 0.424 cm³·g⁻¹ in $Mn-Fe/TiO_2$ sample, which was probable due to the mesoporosity generation, which blocked the micropore formation, resulting in a small decrease of total pore volume. Considering the comprehensive test

results, the $Mn-Co/TiO_2$ nanocatalyst exhibited superior physical properties than the $Mn-Fe/TiO_2$ sample with higher specific surface area, larger total pore volume, and smaller average pore diameter, which were contributed to abundant micropores structures.

Samples	Specific Surface Aarea ($m^2 \cdot g^{-1}$)	Total Pore Volume (cm ³ ⋅g ⁻¹)	Average Pore Diameter (nm)
Mn–Co/TiO ₂	189.9	0.531	33.06
Mn–Fe/TiO ₂	104.6	0.424	54.85

Table 1. Physical properties of Mn-Co/TiO₂ and Mn-Fe/TiO₂ nanocatalysts.

2.1.3. Components Analysis

The XRD patterns of Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts were displayed in Figure 2. The diffraction peaks for TiO₂ support contained in the nanocatalysts were preserved entirely, with the strong and distinguished reflections appearing at $2\theta = 25.3^{\circ}$, 37.8° , 48.0° , 53.9° , 62.7° , 68.8° , 70.3° , 75.1° , and 82.7° . These accorded with the XRD pattern of anatase TiO₂ (ICDD PDF card # 71-1166) [26], with the diffraction angles of the matching peaks shifted at tiny degrees. For Mn–Co/TiO₂ nanocatalyst, each corresponding peak of anatase TiO₂ appeared at the lower diffraction angle, which indicated the interaction between MnCoO_x and anatase TiO₂ was stronger than that of MnFeO_x and anatase TiO₂. Meanwhile, the diffraction peaks of anatase TiO₂ in the Mn–Co/TiO₂ sample were broader and weaker than that in Mn–Fe/TiO₂ sample, which demonstrated the crystalline of anatase TiO₂ reduced with MnCoO_x loading.



Figure 2. XRD patterns of Mn-Co/TiO₂ and Mn-Fe/TiO₂ nanocatalysts.

The diffraction peaks corresponding to MnO_x contained in $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts were very complex due to the transformation among the incomplete crystallization of manganese oxides, such as MnO_2 , Mn_2O_3 , Mn_3O_4 , and MnO. The characterization reflections for MnO_x in the $Mn-Co/TiO_2$ sample were obviously weaker than that in $Mn-Fe/TiO_2$ sample, which manifested the active species in the $Mn-Co/TiO_2$ nanocatalyst, which were superior dispersed on the nanocatalyst surface or the active species incorporated into TiO₂ lattice better [27]. The diffraction peaks fitted MnO_2 (ICDD PDF card # 82-2169) precisely at about 20 values of 22.10°, 35.19°, 36.96°, 38.72°, 47.86°, and 57.166°, which were coinciding with the crystallographic plane reflections of (110), (310), (201), (111), (311), and (420), respectively [12]. Meanwhile, both $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts exhibited the diffraction peaks corresponding to Mn_2O_3 and Mn_3O_4 , apparently. The Mn_2O_3 (ICDD PDF card # 78-0390) was well-matched with intensive distinct signals at $2\theta = 23.08^\circ$, 26.72°, 32.87°, and 56.89°, attributed to the crystallographic plane reflections of (211), (220), (222), and (433), respectively. The Mn_3O_4 (ICDD PDF card # 75-0765) was fitted to characteristic peaks at 36.28°, 40.67°, 41.80°, 57.73° and 64.17°, ascribed to the crystallographic plane reflections of (112), (130), (131), (115), and (063), respectively [23].

Comparing the XRD patterns of the Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts, it could be observed that the diffraction peaks of both Mn₂O₃ and Mn₃O₄ were visibly weakened in the Mn–Co/TiO₂ sample. In the meantime, the diffraction peaks of anatase TiO₂ in the Mn–Co/TiO₂ sample were also notably lower than that in Mn–Fe/TiO₂ sample. It was believed that the doping of cobalt into MnO_x obtained superior enhancement than iron on reducing the crystallization of MnO_x and TiO₂ simultaneously. Moreover, in the XRD patterns of the Mn–Co/TiO₂ nanocatalyst, there were no distinguished diffraction peaks for CoO_x, which manifested the doping of cobalt not only improved MnO_x dispersion, but also enhanced CoO_x dispersing on the nanocatalyst surface completely. The similar results were received over Mn–Fe/TiO₂ sample. Based on the phase characteristics comparation between Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts, it was believed that the Mn–Co/TiO₂ sample exhibited more excellent properties with smaller crystallinities of chemical compounds and better distribution of the active species, which were conducive to catalytic reactions.

2.1.4. Reducibility Properties

The oxidation states of the active compounds contained in nanocatalysts of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ were tested by H₂-TPR. The H₂ consumption curve was fitted by reduction peak separation with Gaussian function, as shown in Figure 3. The details of H₂ consumptions and reduction temperatures for each peak were listed in Table 2. In $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts, the anatase TiO₂ support did not bring remarkable reduction peaks within the test temperature range. Hence, all H₂ consumption could be attributed to the reduction process of MnO_x , CoO_x , and FeO_x. It was proposed that the typical reduction process of MnO_x contained in the nanocatalysts of Mn–Co/TiO₂ and Mn–Fe/TiO₂ complied with the following order: MnO₂ \rightarrow Mn₂O₃ (Mn₃O₄) \rightarrow MnO [28]. Mn-Co/TiO₂ nanocatalyst exhibited four reduction peaks with temperature rising from 50—850 °C, as shown in Figure 3b. The first reduction peak centered, at about 218 °C, was attributed to the high oxidation state of manganese ion converting from MnO₂ to Mn₂O₃ [29]. The dominant reduction peak (Peak 3) was caused by two sequential processes of Mn₂O₃ reducing to Mn₃O₄ and Mn_2O_3 reducing to MnO, as reported in previous literatures [12,28]. The reduction process of Mn_2O_3 to Mn_3O_4 was more liable to happen over the original amorphous Mn_2O_3 [30], while reduction process of Mn₂O₃ to MnO preferred to occur at relatively higher temperature [31]. For the Mn–Co/TiO₂ sample, the typical reduction process of cobalt oxides commonly displayed two separated peaks, the one of Co₂O₃ reducing to CoO appeared at around 327 °C, presenting as an asymmetrical reduction peak (Peak 2), the other one of CoO reducing to Co^0 occurred at about 418 °C and overlapped with MnO_x reduction peaks in whole or partly (Peak 3) [8,32]. Therefore, the fourth medium reduction peak in the $Mn-Co/TiO_2$ nanocatalyst was related to the reduction processes of Mn_3O_4 to MnO.

Samples	Temperature (°C) / H ₂ Consumption (g ⁻¹)
	Peak 1	Peak 2	Peak 3	Peak 4	Total
Mn-Co/TiO ₂ Mn-Fe/TiO ₂	218/1.14 276/0.83	327/0.21 387/0.41	418/2.12 436/0.54	517/0.96 501/1.59	- /4.43 - /3.37

Table 2. H₂-TPR quantitative analysis of Mn-Co/TiO₂ and Mn-Fe/TiO₂ nanocatalysts.



Figure 3. H₂-TPR profiles of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts: (a) Total H₂-TPR curves; (b) Multi-peaks Gaussian fitting for $Mn-Co/TiO_2$ nanocatalyst; (c) Multi-peaks Gaussian fitting for $Mn-Fe/TiO_2$ nanocatalyst.

In comparison, the peak intensions and the reduction temperatures of the $Mn-Fe/TiO_2$ nanocatalyst were evidently different from that of the $Mn-Co/TiO_2$ sample. The peak of MnO_2 reducing to Mn_2O_3 weakened remarkably and shifted to higher temperature (276 °C). There were wide joint peaks (Peak 2 and Peak 3) at temperature of 330–530 °C, which were ascribed to the coinstantaneous reduction processes of Mn_2O_3 to Mn_3O_4 and Fe_2O_3 to Fe_3O_4 . It was proposed that the majority of Fe_2O_3 was reduced at around 380 °C, which located at effortlessly reducible sites in the form of oligomeric clusters, nanoparticles or isolated ions [33]. After the majority of Fe_2O_3 reduction, there was a small quantity of residual Fe_2O_3 reducing to Fe_3O_4 at the higher temperature [34]. The dominating peak (Peak 4), at around 501 °C, was assigned to the overlapped peaks of Mn_3O_4 to MnO and Fe_3O_4 to FeO. Figure 4 provided a graphical representation of the reduction process, in which each active component embodied qualitatively [1,35]. In these two kinds of nanocatalysts, $Mn-Co/TiO_2$ sample displayed the higher low-temperature reducibility and exhibited a noticeable medium temperature reduction peak at the same time, which manifested the higher oxidation states of manganese ion (Mn^{4+} and Mn^{3+}) constituted the dominating phase [31].



Figure 4. The qualitative component states of active elements during reduction process.

According to Table 2, the total H₂ consumption of $Mn-Co/TiO_2$ nanocatalyst was 4.43 mmol·g⁻¹ much larger than that of $Mn-Fe/TiO_2$ sample. Meanwhile, the starting reduction peak temperature of $Mn-Co/TiO_2$ nanocatalyst was at 218 °C lower than that of $Mn-Fe/TiO_2$, which was regarded as an important factor influencing the reducibility. It was believed that the catalyst performed superior catalytic ability at low temperatures as the reduction peaks arising in lower temperature regions [1].

2.1.5. Ammonia Adsorption Properties

The acid capacity of the Mn-Co/TiO₂ and Mn-Fe/TiO₂ nanocatalysts was measured by NH₃-TPD test, which was another crucial factor effecting the catalyst performance in SCR process [36]. The results were exhibited in Figure 5 and Table 3, respectively. The NH₃-TPD curves for the Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts were ascribed to four desorption peaks of chemisorbed NH₃ with the temperature rising from 150 °C to 750 °C. For Mn-Fe/TiO₂ nanocatalyst, the first NH₃ desorption peak occurred at around 209 °C, attributed to NH₃ desorption from weak acid sites, which was too feeble to bound NH₃ steadily in the gas mixture during SCR process [36]. Peak 2 and peak 3 were combined from 398 °C to 585 °C, which were attributed to the medium strong acid sites on the nanocatalyst surface. Peak 4, at around 649 °C, was assigned to the strong acid sites, which were regarded as abundant Lewis acid sites adsorbing a great deal of strongly bound NH₃ [37]. In comparation, the NH₃ desorption results of the Mn–Co/TiO₂ nanocatalyst exhibited better acidity capacity at medium and high temperatures. However, there was an undesired temperature shift to higher temperature occurred in the meantime. For $Mn-Co/TiO_2$ sample, the desorption temperature of weak acid sites, medium strong acid sites and strong acid sites were 276 °C (Peak 1), 402 °C (Peak 2), 587 °C (Peak 3), and 653 °C (Peak 4), respectively. It was obvious that the medium strong acid sites and the strong acid sites were enriched in the $Mn-Co/TiO_2$ sample, which were positive to form more abundant Brønsted acid sites and Lewis acid sites promoting NH₃ adsorption on the nanocatalyst surface [38,39].

Table 3. Quantitative analysis of NH₃-TPD profiles.

Samples	Temperature (°C) / NH ₃ composition (mmol·g ⁻¹)				
	Peak 1	Peak 2	Peak 3	Peak 4	Total
Mn-Co/TiO ₂ Mn-Fe/TiO ₂	276/0.11 209/0.14	402/0.49 398/0.24	587/0.52 585/0.39	653/0.32 649/0.20	-/1.44 -/0.97



Figure 5. NH₃-TPD profiles of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts: (a) Total NH₃-TPD curves; (b) Multi-peaks Gaussian fitting for $Mn-Co/TiO_2$ nanocatalyst; (c) Multi-peaks Gaussian fitting for $Mn-Fe/TiO_2$ nanocatalyst.

The quantitative comparation of the total acid capacity between the $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts was performed and summarized in Table 3. The $Mn-Co/TiO_2$ sample achieved larger total NH₃ desorption of 1.44 mmol·g⁻¹ than $Mn-Fe/TiO_2$, which further confirmed the superior enhancement of $MnCoO_x$ on the surface acidity than $MnFeO_x$. Although the $Mn-Fe/TiO_2$ sample generated weak acid sites at the lower temperature, the NH₃ desorption on the weak acid sites was only 0.14 mmol·g⁻¹, which was too little to offer enough NH₃ for SCR process. However, it was reported that NH₃ could restrain the adsorption and activation of NO onto the active sites on catalyst surface via the undesirable electron transfer between the adsorbed NH₃ and the metal sites [40]. Consequently, the excessive adsorbed NH₃ could inhibit SCR reactions. Therefore, the surface acidity properties of the $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts were bound up with their reducibility properties.

2.1.6. Oxidation States of Active Species

In order to better understand the oxidation states and the atomic concentrations of active species on the surface of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts, the XPS spectra of Ti 2*p*, Mn 2*p*, Co 2*p*, Fe 2*p*, and O 1 *s* were tested and numerically analyzed by Gaussian fitting, respectively, as shown in Figure 6. The binding energy and atomic concentration of each element was summarized in Table 4.



Figure 6. XPS analysis of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts: (a) XPS spectra for Ti 2*p*; (b) XPS spectra for Mn 2*p*; (c) XPS spectra for Co 2*p*; (d) XPS spectra for Fe 2*p*; (e) XPS spectra for O 1s.

Samples	Binding Energy (eV) / Atomic Composition (%)								
	Mn		Mn Fe		C	Со		0	
	Mn ²⁺	Mn ³⁺	Mn ⁴⁺	Fe ²⁺	Fe ³⁺	Co ²⁺	Co ³⁺	Oα	O_{β}
Mn-Co/TiO ₂	641.2/ 13.8	642.6/ 39.4	644.1/ 46.8	-/-	-/-	779.6 39.3	782.1/ 60.7	531.4/ 33.7	530.3/ 66.3
Mn-Fe/TiO ₂	641.4/ 21.3	642.8/ 38.5	644.6/ 40.2	709.6/ 43.4	711.7/ 56.6	-/-	-/-	531.6/ 28.2	530.3/ 71.8

As shown in Figure 8a, the XPS spectrum of Ti 2*p* was consisted of two characteristic peaks, attributed to Ti $2p_{1/2}$ at around 464.3 eV and Ti $2p_{3/2}$ at about 458.7 eV, respectively [41]. It was obvious that Ti⁴⁺ ion was stabilized on the catalyst surface and appeared as the dominating state in the nanocatalysts of Mn–Co/TiO₂ and Mn–Fe/TiO₂. The XPS spectra of Mn 2*p* was composed of Mn $2p_{1/2}$ peak at about 653 eV and Mn $2p_{3/2}$ peak at around 642 eV [42], as exhibited in Figure 6b. The Mn $2p_{3/2}$ peak could be further split into three peaks for Mn²⁺ at 641.2 ± 0.3 eV, Mn³⁺ at 642.6 ± 0.2 eV,

and Mn^{4+} at 644.1 \pm 0.5 eV, respectively, according to the multi-peaks Gaussian fitting results [43]. It is evidently difficult to distinguish the three valence states of manganese in MnO_x with the binding energy difference value less than 3.7 eV. For accurate comparison of the atomic concentration of Mn^{n+} on the surface of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts, a quantitative analysis was performed according to the integral area under every divided peak, as exhibited in Table 4. It was widely accepted that the catalytic ability of MnO_x ranked as: $MnO_2 > Mn_2O_3 > Mn_3O_4$ [12,22,30]. The abundant MnO_2 generated on the nanocatalyst surface was beneficial to SCR reactions [5]. As listed in Table 4, the major amount of manganese in both $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts was Mn^{4+} . Meanwhile, the Mn^{4+}/Mn^{n+} atomic composition was 46.8 % in $Mn-Co/TiO_2$ sample higher than that of 40.2 % in $Mn-Fe/TiO_2$ sample, generating more lattice oxygen and plenty of oxygen vacancy on the catalyst surface [1], which was regarded as the main reason for higher reducibility properties of $Mn-Co/TiO_2$ sample as discussed above.

In Co 2*p* spectrum of Mn–Co/TiO₂ nanocatalyst, the two individual peaks were attributed to Co $2p_{1/2}$ at about 796.5 eV and Co $2p_{3/2}$ at around 780.6 eV, respectively. Both of these two main peaks had satellite peaks at 803.1 eV and 786.8 eV, correspondingly, as shown in Figure 6c. The gentler and broader satellite peaks appeared at relatively higher binding energy, which were introduced by the metal-to-ligand electron transfer or the shakeup process of cobalt in its high spin state. However, this shakeup process was only observed with the high spin state of Co²⁺ ion, but did not appear with the diamagnetic low-spin Co³⁺ ion [13]. The intense peak of Co $2p_{3/2}$ was composed of two overlapped peaks, one attributed to Co³⁺ seated at about 780.0 eV and the other ascribed Co²⁺ and located at around 781.6 eV. These two distinguishing peaks indicated the co-occurrence of cobalt in +2 and +3 valence states on the Mn–Co/TiO₂ nanocatalyst surface. There was a dynamic equilibrium sustained on the nanocatalyst surface with the electron transfer between Mn and Co ions during the catalytic oxidation process, expressed as Co³⁺ + Mn³⁺ \leftrightarrow Co²⁺ + Mn⁴⁺ [44]. Moreover, Co³⁺ species were exhibited as the major atomic concentration of 60.7%, which were at the comparatively higher valence state of Coⁿ⁺ generating vast anionic defects, forming abundant surface oxygen and facilitating adsorption and oxidation during SCR process [45].

In Fe 2*p* spectra of the Mn–Fe/TiO₂ nanocatalyst, as shown in Figure 6d, the two main peaks were ascribed to Fe 2*p*_{1/2} at about 724 eV and Fe 2*p*_{3/2} at around 710 eV. There was a satellite peak at 718.3 eV assigned to Fe³⁺ in Fe₂O₃ [46]. The XPS spectra of Fe 2*p*_{3/2} was composed of two characteristic peaks, ascribed to Fe³⁺ at around 711.6 eV and Fe²⁺ at about 709.6 eV, which meant these two kinds of Feⁿ⁺ coexisted on the surface of the Mn–Fe/TiO₂ nanocatalyst. It was proposed that the promotion effect between manganese and iron was attributed to the electron transfer in the redox reaction: Mn³⁺ + Fe³⁺ \leftrightarrow Mn⁴⁺ + Fe²⁺ [47].

The O 1 s spectrum of Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts were compared in Figure 6e, which was composed of the chemisorbed oxygen peak at 531.2–531.6 eV (O_{α}) and the lattice oxygen peak at 530.2–530.3 eV (O_{β}). According to the curve-fitting results, the binding energy of O_{α} shifted to higher values from 531.4 eV in Mn–Co/TiO₂ sample to 531.6 eV in Mn–Fe/TiO₂ sample. Meanwhile, the similar changes happened to the binding energy of O_{β}. It was believed that the chemisorbed oxygen (O_{α}) was the most active oxygen species owing to its high mobility [48]. On the surface of Mn–Co/TiO₂ nanocatalyst, the atomic concentration of O_{α} reached 33.7% much higher than that of Mn–Fe/TiO₂ sample, which was regarded as another reason for its superior redox ability.

2.2. Catalytic Performance

Figure 7 exhibited the NO conversion and catalytic selectivity over $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts. It can be seen that the NO conversion improved conspicuously with the temperature increasing from 25 °C to 250 °C. For both $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts, the satisfactory conversion (>90%) was obtained above 150 °C. Meanwhile, the $Mn-Ce/TiO_2$ nanocatalyst exhibited higher catalytic activity than the $Mn-Fe/TiO_2$ sample within the temperature range of 50–175 °C. Furthermore, the catalytic selectivity of $Mn-Ce/TiO_2$ nanocatalyst was higher

than that of $Mn-Fe/TiO_2$ within 75–250 °C. Therefore, both the catalytic activity and the catalytic selectivity over $Mn-Co/TiO_2$ were remarkably improved compared to $Mn-Fe/TiO_2$, potentially due to the interaction between manganese and cobalt stronger than that between manganese and iron, which enhanced the generation of Brønsted acid sites and Lewis acid sites on the nanocatalyst surface, resulting in promoting Langmuir–Hinshelwood mechanism reactions and Eley-Rideal mechanism reactions, simultaneously [49].



Figure 7. Catalytic performance of Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatsalysts.

2.3. Reaction Kinetic Study

It was proposed that both the Langmuir–Hinshelwood mechanism and the Eley–Rideal mechanism contributed to the SCR reactions over nanostructured Fe-Mn oxides [36]. According to Langmuir-Hinshelwood mechanism, the SCR reactions over $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatsalysts could be approximately described as:

$$M^{n+} = O + NO(ad) \rightarrow M^{(n-1)+} - O - NO$$
⁽²⁾

$$M^{(n-1)+} - O - NO + NH_3(ad) \rightarrow M^{(n-1)+} - O - NO - NH_3 \rightarrow M^{(n-1)+} - OH + N_2 + H_2O$$
(3)

$$M^{(n-1)+} - O - NO + O_2 \rightarrow M^{(n-1)+} - O - NO_2$$
 (4)

$$M^{(n-1)+} - O - NO + NH_3(ad) \rightarrow M^{(n-1)+} - O - NO_2 - NH_3 \rightarrow M^{(n-1)+} - OH + N_2O + H_2O$$
(5)

$$M^{(n-1)+} - OH + O_2 \rightarrow M^{n+} = O + H_2O$$
 (6)

The adsorbed NO(ad) was oxidized by M^{n+} (Mn^{4+} , Mn^{3+} , Co^{3+} , Fe^{3+}) on the nanocatalyst surface to form NO₂⁻ (Reaction 2). The NO₂⁻ reacted with adsorbed NH₃(ad) to generate NH₄NO₂ and decomposed to N₂ and H₂O, subsequently (Reaction 2). In the meantime, an undesired oxidization reaction occurred on NO₂⁻ which was oxidized to NO₃⁻ under O₂ (Reaction 4). The formed NO₃⁻ reacted with adsorbed NH₃(ad) to generate NH₄NO₃ and further decomposed to N₂O and H₂O (Reaction 5). Finally, the active species $M^{(n-1)+}$ – OH was oxidized by O₂ to regenerate M^{n+} = O (Reaction 6).

According to the Eley-Rideal mechanism, the SCR reactions over $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatsalysts could be described as:

$$M^{n+} = O + NH_3(ad) \rightarrow M^{(n-1)+} - OH + NH_2(ad)$$
 (7)

$$NH_2(ad) + NO \rightarrow N_2 + H_2O \tag{8}$$

The adsorbed $NH_3(ad)$ on the nanocatalyst surface was activated by M^{n+} to generate NH_2 (Reaction 7). Then, the NH_2 reacted with NO to form N_2 and H_2O (Reaction 8).

However, in addition to the main Eley-Rideal mechanism reactions, undesirable side reactions occurred on the active species of $M^{n+}=O$.

$$M^{n+} = O + NH_2(ad) \rightarrow M^{(n-1)+} - OH + NH(ad)$$
(9)

$$M^{n+} = O + NH(ad) + NO \rightarrow M^{(n-1)+} - OH + N_2O$$
 (10)

$$M^{n+} = O + NH(ad) + O_2 \rightarrow M^{(n-1)+} - OH + NO$$
 (11)

The activated $NH_2(ad)$ formed in Reaction 7 was further oxidized to NH(ad) over $M^{n+} = O$. Then NH(ad) reacted with NO to form N_2O or react with O_2 to generate NO.

According to Reactions 3 and 8, the kinetic equations of N_2 formation rates via the Langmuir-Hinshelwood mechanism and the Eley-Rideal mechanism could be described as:

$$k_{\rm scr} = \frac{d[N_2]}{dt} = \frac{d[N_2]}{dt} \Big|_{\rm LH} + \frac{d[N_2]}{dt} \Big|_{\rm ER} = -\frac{d[M^{(n-1)+} - O - NO - NH_3]}{dt} \Big|_{\rm LH} - \frac{d[NH_2(ad)]}{dt} \Big|_{\rm ER}$$
(12)
= $k_3 \Big[M^{(n-1)+} - O - NO - NH_3 \Big] + k_8 [NH_2(ad)] [NO]$

where k_3 and k_8 were the kinetic constants of Reactions 3 and 8, $\left[M^{(n-1)+}-O-NO-NH_3\right]$, $[NH_2(ad)]$, and [NO] were the concentrations of NH_4NO_2 , $NH_2(ad)$, and NO, respectively. When the reactions above reached stable conditions, the concentrations of NH_4NO_2 and $NH_2(ad)$ on the surface of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts could be regarded as constants without connecting with the gaseous concentrations of NH_3 and NO [50]. Therefore, the SCR reaction rate was in a linear relationship with NO concentration as exhibited in Equation (13), the slope and the intercept could be obtained from the linear regression in Figure 8, as listed in Table 5.

$$k_{\rm scr} = k_{\rm SCR-LH} + k_{\rm SCR-ER}[\rm NO]$$
(13)

$$k_{\rm SCR-LH} = k_3 \left[M^{(n-1)+} - O - NO - NH_3 \right]$$
 (14)

$$k_{\rm SCR-ER} = k_8[\rm NH_2(ad)] \tag{15}$$



Figure 8. Effect of NO concentration on selective catalytic reduction (SCR) reaction rate.

Sample	Temperature (°C)	$k_{\rm SCR-ER}/10^6$	$k_{\rm SCR-LH}$	R ²	$k_{\rm SCR}^{a}$
Mn-Co/TiO ₂	100	0.029	5.75	0.993	23.5
	150	0.046	13.25	0.994	41.1
	200	0.064	20.50	0.988	58.2
	250	0.075	25.01	0.984	68.4
Mn-Fe/TiO ₂	100	0.018	2.35	0.989	13.5
	150	0.027	8.25	0.990	24.4
	200	0.033	16.01	0.991	36.0
	250	0.062	14.50	0.998	52.1

Table 5. Reaction kinetic constants of NO reduction.

^a [NO] = [NH₃] = 600 ppm.

2.4. Thermodynamic Calculation

The standard thermodynamic characteristics of the chemical individual species of $M_x O_y$ and M in crystalline states, as well as NH_3 , N_2 and H_2O in gaseous states, were listed in Table 6. The oxidative capacities of $M_x O_y$ were evident according to the reliable standard thermodynamic parameters, which was a convincing support for investigating the roles of $M_x O_y$ in NH₃-SCR process. The Gibbs free energies for the reactions between $M_x O_y$ species and ammonia were calculated for estimating their oxidative capacities according to Equation (16), displayed in Table 7. In order to discuss the reaction mechanism conveniently, simplifying assumptions were introduced as follows. Only the kind of main oxidation state was considered in the thermodynamic calculation, although there were several kinds of oxidation states for each $M_x O_y$ [51]. Meanwhile, it was assumed that every $M_x O_y$ was reduced into metal, although the metallic oxide in the high-oxidation state was reduced step-by-step in the actual process. It had been proved that the entropy of metals would vary drastically as the metal physical states transforming [52]. Furthermore, the transformation could bring about the changes of the straight slope of the temperature–Gibbs free energy curve which was named Ellingham diagrams. The Ellingham diagrams were a particular graphical form of the principle that the thermodynamic feasibility of a reaction depends on the Gibbs free energy change ($\Delta_f G^0$), which could be calculated using Equation (16), where $\Delta_f H^0$ was the enthalpy change and ΔS^0 was the entropy change under standard state.

$$\Delta_{\rm f} G^0 = \Delta_{\rm f} H^0 - T \Delta S^0 \tag{16}$$

	C _p J mol ⁻¹ K	$ riangle_{ m f} { m H^0}$ kJ mol $^{-1}$	S ⁰ J mol ⁻¹ K	$ riangle_{ m f} { m G}^0$ kJ mol $^{-1}$
Mn	26.3	0.0	32.0	-
MnO	45.4	-385.2	59.7	-362.9
Mn ₃ O ₄	139.7	-1387.8	155.6	-1283.2
Mn_2O_3	107.7	-959.0	110.5	-881.1
MnO ₂	54.1	-520.0	53.1	-465.1
Co	24.8	0.0	30.0	-
CoO	55.2	-237.9	53.0	-214.2
Co ₃ O ₄	123.4	-891.0	102.5	-774.0
Fe	25.1	0.0	27.3	-
FeO	-	-272.0	-	-
Fe ₃ O ₄	143.4	-1118.4	146.4	-1015.4
Fe ₂ O ₃	103.9	-824.2	87.4	-742.2
NH ₃	-	45.9	192.8	-
N_2	29.1	0.0	191.6	-
H ₂ O (g)	33.6	-241.8	188.8	-228.6

Table 6. Standard thermodynamic properties of M_xO_y [7,53–55].

Substance	$ riangle_{ m f} { m G} / { m k} { m J}$						
	298 K	323 K	373 K	423 K	473 K	523 K	
MnO ₂	-46.25	-49.09	-54.77	-60.45	-66.12	-71.80	
Mn_3O_4	42.00	39.27	33.81	28.34	22.88	17.42	
MnO	84.06	81.65	76.83	72.01	67.19	62.37	
Co ₃ O ₄ ^a	-85.71	-88.74	-94.79	-100.84	-106.89	-112.94	
CoO	-64.64	-67.17	-72.22	-77.28	-82.34	-87.39	
Fe ₂ O ₃	-31.40	-34.23	-39.89	-45.55	-51.21	-56.87	
Fe ₃ O ₄	-24.99	-27.69	-33.09	-38.49	-43.89	-49.29	

Table 7. Gibbs free energy for the reaction between $M_x O_y$ and NH_3 under different temperatures.

^a Co₃O₄ was regard as the compound of Co₂O₃ and CoO.

The Ellingham diagram plotted $\Delta_f G$ for each oxidation reaction as a function of temperature. In order to compare different reactions, all $\Delta_f G$ values referred to the same quantity of oxygen, chosen as one mole O (1/2 mol O₂) [53], as shown in Equation (17)

$$\frac{1}{y}M_{x}O_{y} + \frac{2}{3}NH_{3} \to \frac{x}{y}M + \frac{1}{3}N_{2} + H_{2}O$$
(17)

According to Equations (16) and (17), the redox reactions that happened between M_xO_y and NH_3 were discussed under thermodynamic calculation, assuming that the enthalpy, entropy, and Gibbs free energy remained unchanged with temperature. The $\Delta_f G$ in Table 7 for each kind of M_xO_y was calculated in order to estimate its oxidative capacities and gain semiquantitative conclusions. Meanwhile, this research just focused on the catalytic performance and reaction mechanism of cobalt and iron doped in the Mn/TiO_2 nanocatalysts, so that only one M_xO_y was different at a time. The negative $\Delta_f G$ values for Co_3O_4 (the compound of Co_2O_3 and CoO) and Fe_2O_3 indicated M_xO_y was effective for the oxidative abstraction of the hydrogen from adsorbed ammonia [7].

The Ellingham diagrams for the Gibbs free energy of reactions between M_xO_y and NH_3 with the reaction temperature rising was displayed in Figure 9. As the $\triangle_f G$ was negative ($\triangle_f G < 0$), the reaction took place more easily. On the contrary, the reaction would happen at higher temperature or would not happen if the $\triangle_f G$ was positive ($\triangle_f G > 0$). Under the same reaction temperature, the order of temperature–Gibbs free energy for the reaction between M_xO_y and NH_3 is as follows: Co_3O_4 (compound of Co_2O_3 and CoO) < CoO < Fe_2O_3 < Fe_3O_4 (only considering the doping transition metals). These results were exactly consistent with the NH₃-SCR performance and almost coincided with the data of the nanostructure characterization, as discussed above, which demonstrated the simplifying assumptions used in the thermodynamic calculations was reasonable. However, the activity of MnO_x exhibited in Ellingham Diagrams was not in well accord with the reducibility properties, which might be on account of positive kinetics or simplifying assumptions during the thermodynamic calculations [51,53]. It manifested that MnO_x , as the dominate phase in nonacatalysts of $Mn-Co/TiO_2$ and Mn-Fe/TiO₂, was activated by the doping of cobalt or iron. A large number of researches had been carried out focusing on optimizing catalysts via doping transition metals in order to enhance NO catalytic conversion remarkably. Based on the thermodynamic study, Ellingham diagrams would provide a novel insight for developing catalysts doped proper metal oxides with higher NH₃-SCR activity.



Figure 9. Temperature–Gibbs free energy curve for the reaction between M_xO_y and NH₃.

3. Materials and Methods

3.1. Catalysts Preparation

The noncatalysts of Mn–Co/TiO₂ and Mn–Fe/TiO₂ were prepared by hydrothermal method. Co(CH₃COO)₂·4H₂O (analytical pure 99.9%, Kermel, Tianjin, China) and Fe(NO₃)₃·9H₂O (analytical pure 99.9%, Sinopharm, Shanghai, China) were used as the precursors of CoO_x and FeO_x, respectively. The aqueous solution of Mn(NO₃)₂ (analytical pure 50.0%, Sinopharm, Shanghai, China) was used as the precursors of MnO_x. The tetrabutyl titanate was introduced as the precursors of anatase TiO₂ to support the bimetals oxides. The precursors were added into deionized water at room temperature. One-hundred fifty mL glycol was added into the above mixture with magnetic stirring continuously to regulate the reaction rate of hydrothermal reaction. Then the homogeneous solution was heat at 180 °C for 8 h in a Teflon-lined stainless steel autoclave. After that, tetrabutyl titanate was added into the solution and reheated in the autoclave at 180 °C for 3 h. The precipitate was obtained by reduplicative centrifugation and wash. In the end, the nanocatalysts were dried at 150 °C for 12 h and calcined in air at 500 °C for 4 h. The nanocatalyst was denoted as Mn–Co/TiO₂ and Mn–Fe/TiO₂ with the molar ratios of Mn:M:Ti = 2:1:7.

3.2. Catalysts Characterization

The advanced microstructural image data and the surface element contents of the nanocatalysts was achieved by a high resolution transmission electron microscope JEOL JEM-2010 (Japan electronics corporation, Tokyo, Japan) combined with EDS. The Maxon Tristar II 3020 micropore-size analyzer (Maxon, Chicago, IL, USA) was used for testing N₂ adsorption isotherms of the prepared nanocatalysts at -196 °C. The surface areas and the pore-size distributions of the nanocatalysts were measured after the nanocatalysts degassing in vacuum at 350 °C for 10 h. BET plot linear portion was used to determine the nanocatalysts specific surface areas, and the desorption branch with Barrett-Joyner-Halenda (BJH) formula was introduced to calculate the pore-size distributions. The XRD data was captured by a Bruker D8 advance analyzer (Bruker, Frankfurt, Germany) with Mo K_{α} radiation, diffraction intensity from 10° to 90° , point counting time of 1 s and point counting step of 0.02° . The element phases contained in the nanocatalysts were distinguished by comparing characteristic peaks presented in the XRD patterns with the International Center for Diffraction Data (ICDD). H₂-TPR and NH₃-TPD tests were performed with a Micromeritics Autochem II 2920 chemical adsorption instrument (Micromeritics, Houston, TX, USA). During H₂-TPR experiment, nanocatalysts were pretreatment in He at 400 °C for 1 h, and then cooled to environment temperature in H₂ and He gas mixture at 30 mL/min. The test temperature range of H₂ consumptions was from 50 °C to 850 °C with the heating rate of 10 °C/min. The operating process of NH₃-TPD test was similar to that of H₂-TPR test with NH₃ replacing H₂. XPS analysis was performed on a Thermo ESCALAB 250XI (Thermo Fisher, Boston, MA, USA) with

pass energy 46.95 eV, Al K_{α} radiation 1486.6 eV, X-ray source 150 W, and binding energy precision \pm 0.3 eV. The C 1 *s* line at 284.6 eV was measured as a reference.

3.3. Catalytic Performance Tests

The catalytic properties of the Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts were tested on a fixed-bed quartz tube reactor including a tube furnace and a temperature control unit, as shown in Figure 10. The mass of the catalyst was 250 mg with the particle diameter of 40–60 mesh. During NO conversion test, 5 mL nanocatalyst with the particle diameter of 40–60 mesh was used under the total flow rate of 2000 mL/min, corresponding to the gas hourly space velocity (GHSV) of 24,000 h⁻¹. The feed gas contained 300 ppm NO, 300 ppm NH₃, 5% O₂, and N₂ as balance gas. During reaction kinetic study, the reaction conditions were NO 150–600 ppm, NH₃ 600 ppm, O₂ 5%, catalyst mass of 200 mg, total flow rate at 4000 mL/min, and GHSV at 1 200 000 cm³·g⁻¹·h⁻¹. The concentrations of NO, N₂O, and NH₃ in the outlet were continually monitored using German MRU MGA-5 analyzer (MRU, Berlin, Germany) joint with an external special detector for N₂O and NH₃. The NO conversion was calculated according to Equation (18). The N₂ selectivity was calculated by the concentrations of N₂O and NO, as show in Equation (19). Each experiment was repeated three times to confirm the results accuracy.

NO conversion =
$$\left(\frac{[NO]_{in} - [NO]_{out}}{[NO]_{in}}\right) \times 100\%$$
 (18)

$$N_{2} \text{ selectivity} = 1 - \frac{2[N_{2}O]_{out}}{[NO]_{in} - [NO]_{out}} \times 100\%$$
(19)



Figure 10. The schematic diagram of SCR experiments. 1, standard gas; 2, mass flowmeter; 3, gas mixer; 4, shutdown valve; 5, resistance furnace; 6, temperature controller; 7, nanocatalysts; 8, flue gas analyzer; 9, record system; 10, gas washing bottle; 11, induced draft fan.

4. Conclusions

The multiple characterizations performed on the Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts indicated that the Mn–Co/TiO₂ sample obtained the superior structure characteristics than Mn–Fe/TiO₂ with the surface area increased by 85.3 m²·g⁻¹, the nanoparticle size reduced by 21.79 nm, the active components distribution improved, and the crystallinity decreased. Meanwhile, these results further confirmed the catalytic property was highly dependent on the phase compositions of the catalysts. The ratios of Mn⁴⁺/Mnⁿ⁺ in the Mn–Co/TiO₂ sample was 46.8% higher than that of 40.2% in the Mn–Fe/TiO₂ sample, which demonstrated the better oxidation ability and the larger amount of Bronsted acid sites and Lewis acid sites on the sample surface. Both Mn–Co/TiO₂ and Mn–Fe/TiO₂ nanocatalysts exhibited the satisfactory conversion (>90%) above 150 °C. When compared to the

 $Mn-Fe/TiO_2$ nanocatalyst, the $Mn-Ce/TiO_2$ sample displayed the preferable catalytic property with higher catalytic activity and stronger selectivity in the temperature range of 75—250 °C. In the meantime, the mechanism and kinetic study showed that Langmuir-Hinshelwood mechanism and Eley-Rideal mechanism reactions contributed to NH_3 -SCR catalysis, simultaneously. Under the same test conditions, the NO conversion rate of $Mn-Co/TiO_2$ nanocatalyst was always higher than that of $Mn-Fe/TiO_2$ sample. Furthermore, Ellingham Diagrams of $Mn-Co/TiO_2$ and $Mn-Fe/TiO_2$ nanocatalysts were in accordance with their catalytic performances, based on the Gibbs free energy calculation of the reactions between ammonia and each kind of active metal oxide. Generally, $Mn-Co/TiO_2$ nanocatalyst exhibited the better catalytic properties than $Mn-Fe/TiO_2$ sample according to the comprehensive comparison in this research. The comparison with the insight from structure, performance, kinetics, and thermodynamics might supply a theoretical method to select superior metal oxides for NH_3 -SCR.

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Article

Mechanism and Performance of the SCR of NO with NH₃ over Sulfated Sintered Ore Catalyst

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Abstract: A sulfated sintered ore catalyst (SSOC) was prepared to improve the denitration performance of the sintered ore catalyst (SOC). The catalysts were characterized by X-ray Fluorescence Spectrometry (XRF), Brunauer–Emmett–Teller (BET) analyzer, X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS) and diffuse reflectance infrared spectroscopy (DRIFTS) to understand the NH₃-selective catalytic reduction (SCR) reaction mechanism. Moreover, the denitration performance and stability of SSOC were also investigated. The experimental results indicated that there were more Brønsted acid sites at the surface of SSOC after the treatment by sulfuric acid, which lead to the enhancement of the adsorption capacity of NH₃ and NO. Meanwhile, Lewis acid sites were also observed at the SSOC surface. The reaction between $-NH_2$, NH_4^+ and NO (E-R mechanism) and the reaction of the coordinated ammonia with the adsorbed NO₂ (L-H mechanism) were attributed to NO_x reduction. The maximum denitration efficiency over the SSOC, which was about 92%, occurred at 300 °C, with a 1.0 NH₃/NO ratio, and 5000 h⁻¹ gas hourly space velocity (GHSV).

Keywords: sintered ore catalyst; sulfate; In-situ DRIFTS; SCR

1. Introduction

Nitrogen oxide (NO_x) is one of the major atmospheric pollutants and is mainly generated from the combustion of fossil fuel, which has serious harmful effects on human health and the ecological environment. In 2016, the total emission of NOx from the iron and steel industry was about 1.04 million tonnes [1]. The sintering process in iron-making plants uses coal or coke as fuel, which is a major emission source of NOx. About 35–50% of the total NOx emission from the iron and steel industry is attributed to the sintering process [2–4]. The new ultra-low emission standard of air pollutants for the iron and steel industry will be issued by Chinese government and require that NOx concentration in the sintering flue gas should be below 50 mg/m³. Hence, it is urgent to address the treatment of sintering flue gas.

Selective catalytic reduction (SCR) over V_2O_5 -WO₃ (MoO₃)/TiO₂ catalyst has been widely applied in power stations because of its high denitration efficiency [5–7]. However, there are still shortcomings such as the toxicity of the catalyst, and the high operation costs [8]. In particular, the reaction temperature of the current commercial catalysts is higher than the temperature of the sintering flue gas [9]. Therefore, the sintering flue gas must be heated to the reaction temperature of the catalyst by using additional fuels. Many researchers paid more attention to developing the low temperature catalysts or the catalysts prepared by the non-noble metals for sintering flue gas. Zhang et al. reported that the Fe–based catalysts exhibited a high catalytic activity for NO reduction [10]. Wang [11] also stated that the Fe₂O₃ particles had a good performance during NO_x elimination, where the highest NO_x conversion reached 95%. Yang et al. [12] claimed that α –Fe₂O₃ had a poor SCR activity, while γ –Fe₂O₃ had an excellent SCR activity at 200–350 °C. Meanwhile, it was also found that the further increase reaction temperature from 350 to 500 °C would suppress NOx conversion over γ –Fe₂O₃ [13].

It was well known that the sintered ore was one of the raw materials in the sintering plant, its main component was Fe_2O_3 . In our previous study, Han et al. and Chen et al. [14,15] proposed that the hot sintered ore was used as catalysts for NO_x removal from the sintering flue gas. The experimental results illustrated that the hot sintered ore had a good denitration performance, the denitration efficiency was about 60% at 300 °C, with a 1.0 NH₃/NO ratio, and 1000 h⁻¹ GHSV. However, the denitration efficiency was too low and the NOx concentration from the sintering flue gas could reach the limit after SCR over the sintered ore.

Ciambelli [16] found that introduction of SO_4^{2-} into SCR catalysts can promote the surface acidity of the catalysts. Thus, the catalytic activity was promoted. Khodayari [17] and Xu [18] thought that the sulfation of the catalysts would decrease the oxidization ability of Fe³⁺, and separated the active sites of adsorbing $-NH_2$ and the active sites of oxidizing $-NH_2$. Therefore, the catalytic oxidization of NH_3 over γ -Fe₂O₃ was suppressed. This process resulted in an obvious promotion of NOx conversion. Zhang [19] also investigated the sulfation of CeO₂-ZrO₂, and their experiment results demonstrated that the sulfated CeO₂-ZrO₂ provided more surface acidities and acidic sites, and Brønsted acid sites were also increased.

In order to improve the denitration efficiency, SSOC was prepared and the influence of the acidification on SCR performance was investigated. At the same time, the reaction mechanisms of SCR over SSOC were also discussed.

2. Results and Discussion

2.1. Characterizations

The main components of the two catalysts were shown in Table 1. It presented that the main components of both two catalysts were Fe_2O_3 and CaO. After the introduction of sulfuric acid, the proportion of iron oxide in the catalyst was decreased and sulphate content was significantly increased.

	Fe ₂ O ₃	CaO	SiO ₂	Al ₂ O ₃	MgO	MnO ₂	TiO ₂	P ₂ O ₅	SO _x
SOC	75.42	12.54	4.22	1.19	1.02	0.19	0.20	0.09	0.14
SSOC	48.92	11.52	2.82	0.79	0.85	0.13	0.09	0.05	29.92

Table 1. Chemical compositions of catalysts (%).

The BET surface area, pore-size and pore volume of the catalysts were presented in Table 2. The specific surface area of the SOC was $3.684 \text{ m}^2/g$, the total pore volume was $0.00714 \text{ cm}^3/g$, and the average pore diameter was 6.3 nm. The BET surface area, pore size and pore volume of the SSOC were $4.734 \text{ m}^2/g$, 5.9 nm and $0.00964 \text{ cm}^3/g$, respectively. It was indicated that the introduction of sulfuric acid would open some micro-pores, which lead to the increase of the total pore volume. At the same time, the sulphate would block some pores. The BET of SSOC was higher than that of SCO, which meant that the sulfation had a positive effect on the pore structure of SOC.

Catalysts	BET Surface Area (m ² ·g ⁻¹)	Total Pore Volume (cm ³ ·g ⁻¹)	Average Pore Diameter (nm)
SOC	3.684	0.00714	6.3
SSOC	4.734	0.00964	5.9

Table 2. Textural properties of the catalysts.

The XRD patterns of the catalysts were demonstrated in Figure 1. Several sharp diffraction peaks at 24.1°, 33.2°, 35.6°, 49.6°, 54.2°, 57.1° and 62.6° were observed, which were assigned to α -Fe₂O₃ (JCPDS NO. 33-0664), and the characteristic peaks of α -Fe₃O₄ at 30.1°, 33.2°, 49.6°, 54.2° and 57.1° were also detected in Figure 1. The results indicated that the main components of SOC were α -Fe₂O₃ and a small amount of α -Fe₃O₄. After the acidification with sulfuric acid, it could be seen that the peak of α -Fe₂O₃ was decreased. Meanwhile, the characteristic peaks (25.4°, 31.4°, 38.7°, 52.2°) of Fe₂(SO₄)₃ were detected, and a small amount of FeSO₄ and CaSO₄ also appeared.



Figure 1. XRD patterns of the catalysts: SOC (a), SSOC (b) and SSOC after tested (c).

X-ray photoelectron spectroscopy (XPS) was used to study the valence states of the catalysts, as shown in Figure 2. It could be seen from Figure 2 (I) that there were obvious peaks of C (1s), O (1s), and Fe (2p) in both of the two catalysts. Moreover, there were obvious S (2p) peaks after the introduction of sulfuric acid. As shown in Figure 2 (II), the binding energies of Fe 2p 2/3 and Fe 2p 1/2 on SOC were mainly centered at about 710.8 and 724.3 eV, which were indications that the iron species were mainly in the form of Fe³⁺ in the SOC [20]. At the same time, a small number of Fe²⁺ existed, which was also consistent with the XRD results. The binding energies of Fe 2p 2/3 and Fe 2p 1/2 of SSOC were higher than those of SOC. The peak of Fe 2p 2/3 appeared at 711.8 eV and the peak of Fe 2p 1/2 appeared at 724.9 eV. After the acidification with sulfuric acid, both Fe₂(SO₄)₃ and FeSO₄ appeared in the SSOC, which contributed to Fe 2p peaks shifting to the higher position. In addition, in Figure 2 (III), the peak of S 2p appeared at 169.3 eV, indicating that sulfur compounds existed in the form of SO²₄⁻ [21].



Figure 2. XPS spectra of the catalysts: SOC (a) and SSOC (b).

2.2. In-Situ DRIFTS Studies

2.2.1. Adsorption of NH₃

Figure 3 showed the FTIR spectra of NH₃ adsorption over the SSOC with 1000 ppm NH₃/Ar under different temperatures (250–350 $^{\circ}$ C). Thirty minutes after NH₃/Ar introduction, the bands at

1690, 1525, 1460 and 1408 cm⁻¹ were observed. The bands at 1690, 1460 and 1408 cm⁻¹ were symmetric bending vibration of NH_4^+ species at Brønsted acid sites [20,22]. The bands at 1525 and 3434 cm⁻¹ were assigned to the intermediate species, such as ammonium or amide species ($-NH_2$) [20]. Meanwhile, the bands at, 3242 and 3120 cm⁻¹ could be ascribed to the N–H stretching vibration of coordination NH₃, and the bands at 1259 cm⁻¹ (symmetric bending vibration of NH₃ on Lewis acid sites) and 1102 cm⁻¹ (NH₃ species adsorbed on Lewis acid sites) also appeared [20,23,24]. The intensities of the peaks of NH₄⁴ species (1690, 1460 and 1408 cm⁻¹) and $-NH_2$ species (1525 cm⁻¹) first increased and then decreased in the temperature range of 250–350 °C. There were peaks at 960 and 930 cm⁻¹ (NH₃ of gaseous state or weak adsorption state) under 300 °C. Yu et al. [25] stated that the following reactions probably took place in the reaction:

$$NH_3(g) \rightarrow NH_3(a)$$
 (1)

$$NH_3(a) + O(a)/O^{2-} \rightarrow NH_2(a) + OH(a)$$
⁽²⁾

$$NH_3(g) + H^+ \rightarrow NH_4^+ \tag{3}$$

According to Equations (1)–(3), the more functional groups formed at the surface of the catalyst, the higher the reaction activity. Thus, the optimum adsorption activity of NH_3 over the SSOC was 300 °C, as shown in Figure 3.

The dependence of FTIR spectra over the SOC and SSOC on the reaction time at 1000 ppm NH₃/Ar and 300 °C was presented in Figure 4. Figure 4a demonstrated that there were 6 bands in the range of 1690–1100 cm⁻¹. The bands at 1690, 1405 and 1454 cm⁻¹ were related to the symmetric bending vibration of NH⁺₄ species, and the band at 1525 cm⁻¹ was ascribed to the intermediate products of ammonium or -NH2 species. Moreover, the band at 3242 cm⁻¹ (N-H stretching vibration of coordinated NH_3) appeared at 10 min. With the feed of NH_3 , the adsorption band at 1259 cm⁻¹ (symmetric bending vibration of NH_3 on Lewis acid sites) appeared at 30 min, and the band at 1107 cm⁻¹ (NH_3 of gaseous state or weak adsorption state) appeared at 60 min. Hence, there were both Lewis acid sites and Brønsted acid sites on the surface of SOC. The FTIR spectra over SSOC dependence of the reaction time was presented in Figure 4b, and the bands of NH₃ adsorption were basically the same as those of SOC. Moreover, the bands at 3434 cm⁻¹ (–NH₂ groups) and 3128 cm⁻¹ (coordination NH₃ on Lewis acid sites) were observed because the surface acidity of SSOC was strengthened. There were two new weak bands that appeared at 965 and 927 cm^{-1} , where NH_3 was in a gaseous state or weak adsorption state. These weakly adsorbed or gaseous NH₃ could rapidly adsorbed on the acid sites once the adsorbed NH₃ species was consumed. It could also be seen that the strength of NH₃ adsorption peak was enhanced after SOC acidification with sulfuric acid, which was the reason that the surface acidity of SSOC was enhanced after the acidification. The influence of the acidity of SSCO on the dentiration efficiency is presented in Figure 5. The different acidity of SSOC was achieved by sulfating with 1, 3 and 5 mol/L sulfuric acid. In this experimental run, the mass of catalyst and the volume of sulfuric acid were the same. Only the concentration of sulfuric acid was varied. It was demonstrated the maximum denitration efficiency occurred at the catalyst treated by 5 mol/L sulfuric acid, and its denitration efficiency was 92.3% at 300 °C. At the same time, the denitration efficiency of the catalysts sulfated by 1 and 3 mol/L was 56.6% and 68.5% at the same reaction temperature. Moreover, the experimental results proved that the optimum reaction temperature for all catalysts was 300 °C. These sulfates on the surface of SSOC provided more Brønsted acid sites (Peaks at 1690, 1525 cm⁻¹ were increased), which promoted the adsorption capacity for NH₃. Hence, the catalyst denitration performance was also improved [26,27].



Figure 3. DRIFT spectra of SSOC in the condition of 1000 ppm NH₃ at 250–350 °C for 30 min.



Figure 4. DRIFT spectra of SOC (a) and SSOC (b) exposed to 1000 ppm NH₃ at 300 °C at different time.



Figure 5. The influence of acidity of SSOC on SCR performance.

2.2.2. Co-Adsorption of NO and O₂

Figure 6 indicated the DRIFTS spectra of co–adsorption of NO and O_2 under 250–350 °C, and 1000 ppm NO + 15% O_2 /Ar. After 30 min, there were 4 bands appeared in the range of 1620–1290 cm⁻¹ and a weak band appeared at 1014 cm⁻¹. The band at 1607 cm⁻¹ was ascribed to bridged nitrate species and adsorbed NO₂ molecules [28,29]. The bands at 1485 and 1417 cm⁻¹ were assigned to bidentate nitrates and monodentate nitrates respectively [30–32]. In addition, the bands at 1293 and 1014 cm⁻¹ were related to nitro compounds [31]. The variation trend of the intensities of these bands were same as NH₃ adsorption in the temperature range of 250–350 °C. The adsorption process can be explained by the following formula:

$$NO(g)NO + O(a) \leftrightarrow NO_2(a) - Bridgednitrites$$
 (4)

$$NO(g) + O(a) \leftrightarrow NO_2(a) - Monodentated nnitrites$$
 (5)

$$NO_2(g) + O(a) \leftrightarrow NO_3(a) - Bidentatenitrites$$
 (6)

$$NH_2(a) + NO(g) \rightarrow NH_2NO \rightarrow N_2 + H_2O$$
 (7)

$$NH_4^+(a) + NO(g) \rightarrow \{NH_3NO \rightarrow NH_2NO + H_2O\} \rightarrow N_2 + H_2O$$
(8)

$$NO_2(a) + NH_3(a) \rightarrow NO_2[NH_3]_2(a) + NO \rightarrow 2N_2 + 3H_2O$$
(9)

$$NO_2(a) + 2NH_4^+(a) \rightarrow NO_2NH_4^+(a) + NO \rightarrow N_2 + 2H_2$$
(10)

According to Equations (4)–(10), all the peaks in Figure 6 were important to the SCR reaction. The intensity of peaks at 300 $^{\circ}$ C was the highest, which mean that the optimum adsorption temperature for NO was also 300 $^{\circ}$ C.

Figure 7 showed the DRIFTS spectra over SOC and SSOC at different reaction times under 300 °C, 1000 ppm NO + 15% O₂/Ar. Similar with the spectra in Figure 6, after 10 min, there were 4 bands that appeared in the range of 1620–1300 cm⁻¹ and a weak band appeared at 1022 cm⁻¹. The bands intensities were gradually increased with the adsorption time. It could be seen that in Figure 7b, the adsorption intensity of SSOC was obviously higher than that of SOC, especially for the nitro compounds (1290 cm⁻¹) and the nitrate species (1490 cm⁻¹). The results indicated that the adsorption capacity of NO was improved after the acidification with sulfuric acid. It had been shown that the introduction of SO_4^{2-} enhanced the adsorption of NO [26].



Figure 6. DRIFT spectra of SSOC in the condition of 1000 ppm NO and 15% O₂ at 250–350 °C for 30 min.



Figure 7. DRIFT spectra of SOC (a) and SSOC (b) exposed to 1000 ppm NO and 15% O_2 at 300 °C for a different time.

The results showed that the peak intensities of the adsorbed species related to NO_2 molecules and bridged nitrate species (1607 and 1618 cm⁻¹) were decreased obviously with purging by Ar, which indicated NO_2 molecules and bridged nitrate species absorbed at the surface of catalysts were unstable.

At the same time, it was found that the intensities of bidentate nitrates, monodentate nitrates and nitro compounds were stable even after an Ar purge.

2.2.3. Reaction between NH₃ and NO

The reaction at the surface of catalyst by introducing NH_3 and $NO + O_2$ into an in-situ reactor was also studied. Figure 8 shows that NH_4^+ species (1695 cm⁻¹), amide species(-NH₂) or ammonium (1533 cm^{-1}) , and several weak adsorption bands at 1252, and 3242 cm⁻¹ (NH₃ adsorbed on Lewis acid site) were detected on the surface of SOC after NH₃ introduction at 30 min. On the surface of SSOC catalyst, NH_4^+ species at 1695, 1427 and 1454 cm⁻¹ and weakly adsorbed NH_3 or gaseous NH_3 $(966, 925 \text{ cm}^{-1})$ were also detected. There were two bands at 3434 cm⁻¹ (-NH₂ groups) and 3127 cm⁻¹ (coordination NH_3 on Lewis acid sites) that appeared at the same time. After switching to $NO + O_2$, the adsorbed species of NH₃ over SOC and SSOC gradually disappeared and the adsorbed species of NO_x appeared. These bands were ascribed to NO₂ molecules and bridged nitrate species (1602 and 1618 cm⁻¹), bidentate nitrates (1488 and 1498), monodentate nitrates (1413 and 1419 cm⁻¹) and nitro compounds (1290, 1295, 1020 and 1011 cm⁻¹). Comparing Figure 8a with Figure 8b, it could be seen that there were more nitrite species and NH₃ species adsorbed at SSOC than those at SOC. This could be explained that the SSOC contained more active sites, which resulted from the sulfates on SSOC, and Equations (7)–(10) occurred [33]. The reaction of amide species $(-NH_2)$ and NH_4^+ species on the surface of catalysts with the gaseous NO was E-R mechanism, and the reaction between adsorbed state NO_2 and adsorbed NH_3 followed L-H mechanism. Therefore, the reaction between NO and NH_3 over SOC and SSOC had two mechanisms.



Figure 8. DRIFT spectra of SOC (a) and SSOC (b) successively exposed to 1000 ppm NH₃, 1000 ppm NO + 15% O₂ for different time under 300 °C.

2.2.4. SCR Performance

SCR performance were carried out in a fixed-bed reactor, which was made of quartz with an inner diameter of 20 mm and a length of 1000 mm. In this experiment, the mass of the catalysts samples was 13.49 g, the flow rate of the flue gas which simulated sintering flue gas was about 600 mL/min, the simulated sintering flue gas contained 300 ppm NO, 15% O₂ and balance N₂. The temperature in the reactor was kept at 100 °C–350 °C, with the condition of a 0.5–1.0 NH₃/NO ratio and 5000 h⁻¹ GHSV. The NO_x concentrations in simulated flue gas at the inlet and outlet of the reactor were continuously recorded by a gas analyzer (PG-350, Horiba, Kyoto, Japan) with an accuracy of ±1.0%. The NO_x conversion was calculated according to the following equations:

$$NO_{x} \text{ conversion} = \frac{[NO_{x}]_{in} - [NO_{x}]_{out}}{[NO_{x}]_{in}} \times 100\%$$
(11)

Figure 9 presented the denitration performance of SOC and SSOC in the temperature range of 100 °C–350 °C. It was found that the reaction temperature had a great effect on the NH₃-SCR denitration performance of SOC and SSOC. The optimum reaction temperature was 300 °C, which conformed well with the in-situ DRIFTS results. The NO_x conversion of SOC was only 27% at 300 °C, 1.0 NH₃/NO ratio and 5000 h⁻¹ GHSV, and the NO_x conversion of SSOC was 92% at the same condition. It was found that the denitration performance of SSOC was greatly improved and the optimum reaction temperature was 300 °C.



Figure 9. Effect of temperature on SCR activity on SOC and SSOC. 300 ppm NO, 15% O_2 , Ar as balance gas, NH₃/NO ratio: 1.0, GHSV: 5000 h⁻¹.

The stability test of SSOC was shown at Figure 10. The reaction continued for 24 h at 300 °C, 1.0 NH₃/NO ratio, GHSV = 5000 h⁻¹. It could be seen that the NO_x conversion was stable at about 92%, which indicated that SSOC has a good denitration stability. Compared with the SOC, the denitration performance had been greatly improved after the acidification with sulfuric acid. The adsorption of NH₃ and NO on SSOC was obviously improved, which subsequently promoted the NH₃-SCR reaction.



Figure 10. The stability test of SSOC. 300 ppm NO, 15% O₂, Ar as balance gas, NH₃/NO ratio: 1.0, GHSV: $5000 h^{-1}$.

3. Experimental Process

3.1. Catalyst Preparation

In this experiment, the sintered ore was sampled from a sintering workshop in Wuhan. The sintered ore was dried, milled and sieved to 0.15-0.25 mm, which was denoted as SOC. The SSOC was prepared by using an impregnation method. Firstly, 100 g sintered ore (0.15-0.25 mm) was weighed and put into a beaker, then 50 mL sulfuric acid solution with a concentration of 5 mol/L was added and stirred simultaneously for 30 min. After filtration and washing with the deionized water, the mixture was dried at 105 °C and then calcined at 500 °C for 3 h in the air atmosphere. Finally, the prepared SSOC were naturally cooled to the room temperature, then crushed and sieved to 0.15-0.25 mm.

3.2. Catalyst Characterization

The main chemical composition of the SOC and SSOC were analyzed by X-ray fluorescence spectroscopy (XRF) (ARL SMS-XY, Thermo Fisher Scientific Corp., Waltham, MA, USA). The specific surface area, pore volume and pore size distribution of the catalysts were measured by an automated adsorption analyzer (Micromeritics ASAP 2020, Micromeritics Corp., Norcross, GA, USA). The catalysts samples were firstly degassed at 240 °C for 4 h before the test, and the adsorption medium was liquid nitrogen. The nitrogen adsorption and desorption were analyzed at –196 °C using BET analyzer (Micromeritics ASAP 2020, Micromeritics Corp., Norcross, GA, USA). The surface area was calculated by using the BET method according to nitrogen adsorption data in the relative pressure (P/P0) range of 0.01–1. X-ray diffraction (XRD; Rigaku RINT2000, Tokyo, Japan) was performed using CuK α radiation ($\lambda = 1.54056$ Å) to detect the crystalline phases of the samples. The analysis of XRD was referred to International Centre for Diffraction Data (ICDD). X-ray photoelectron spectroscopy (XPS, AXIS Ultra DLD, Shimazu Corp., Kyoto, Japan) was used to determine the valence states of the surface atoms of the catalysts with Al K α radiation.

In-situ DRIFTS experiments were carried out in a FTIR spectrometer (FT-IR, Bruker Tensor II, Bruker optics Corp., Karlsruhe, Germany) equipped with an in-situ cell and a mercury-cadmium-telluride detector [34–37]. For the adsorption of NH₃ (or NO + O₂), the catalyst was exposed to a 20 mL/min NH₃ (or NO + O₂), which resulted in the variation with adsorption time of the DRIFT spectra, and argon purging was subsequently performed. In the reaction mechanism studies, the catalyst was pretreated in a flow of 20 mL/min NH₃ for 40 min, then was shifted NO + O₂ at 300 °C to get the DRIFT spectra. All spectra were recorded by accumulating 100 scans at a spectra resolution of 4 cm⁻¹.
4. Conclusions

In this paper, SSOC displayed an excellent catalytic denitration activity. There were some sulfates that appeared in SSOC after acidification with sulfuric acid solution, which provided more Brønsted acid sites. In-situ DRIFTS results demonstrated that there were Lewis and Brønsted acid sites simultaneously on the surfaces of SOC and SSOC. The acidification contributed to the increase of the Brønsted acid sites at SSOC, which improved the adsorption capacity of NH₃ and NO. NH₃ and NO were adsorbed on the surface of catalysts to form amide species (–NH₂), NH₄⁺ species, NO₂ molecules in gaseous or weakly adsorbed state and nitrate species. Meanwhile, it could be seen that the NH₃-SCR process of SOC and SSOC followed E-R and L-H mechanisms. Moreover, the optimum reaction temperature of catalysts was 300 °C and maximum NO_x conversion fraction over SSOC was 92% at 1.0 NH₃/NO ratio and 5000 h⁻¹ GHSV. It was observed that the NO_x conversion could be steadily maintained.

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Article Promotional Effect of Cerium and/or Zirconium Doping on Cu/ZSM-5 Catalysts for Selective Catalytic Reduction of NO by NH₃

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Abstract: The cerium and/or zirconium-doped Cu/ZSM-5 catalysts (CuCe_xZr_{1-x}O_y/ZSM-5) were prepared by ion exchange and characterized by X-ray diffraction (XRD), transmission electron microscopy (TEM), X-ray photoelectron spectroscopy (XPS), temperature-programmed reduction by hydrogen (H₂-TPR). Activities of the catalysts obtained on the selective catalytic reduction (SCR) of nitric oxide (NO) by ammonia were measured using temperature programmed reactions. Among all the catalysts tested, the CuCe_{0.75}Zr_{0.25}O_y/ZSM-5 catalyst presented the highest catalytic activity for the removal of NO, corresponding to the broadest active window of 175–468 °C. The cerium and zirconium addition enhanced the activity of catalysts, and the cerium-rich catalysts exhibited more excellent SCR activities as compared to the zirconium-rich catalysts. XRD and TEM results indicated that zirconium additions improved the copper dispersion and prevented copper crystallization. According to XPS and H₂-TPR analysis, copper species were enriched on the ZSM-5 grain surfaces, and part of the copper ions were incorporated into the zirconium and/or cerium lattice. The strong interaction between copper species and cerium/zirconium improved the redox abilities of catalysts. Furthermore, the introduction of zirconium abates N₂O formation in the tested temperature range.

Keywords: selective catalytic reduction; nitric oxide; ammonia; Cu/ZSM-5; cerium; zirconium

1. Introduction

Metal-based ZSM-5 catalyst has received considerable attention due to its excellent performance in the low-temperature selective catalytic reduction (SCR) reaction [????]. As a crystalline inorganic polymer, ZSM-5 consists of a three-dimensional network of SiO₄ and AlO₄ tetrahedra linked by interconnecting oxygen ions. When transition metal is introduced in ZSM-5, the catalytic performance can be enhanced through different types of active species, including isolated ions on the exchange sites within the ZSM-5 structure, metal oxide clusters, and metal oxide particles in the appearance of ZSM-5 [??]. Among all the relevant catalysts studied, Cu/ZSM-5 is of particular interest because it exhibits excellent SCR activity in NO abatement [??]. However, it has been reported that the catalytic activity of Cu/ZSM-5 catalysts is decreased, after an initial increase, with increasing copper content [?], which suggests that it is difficult to improve catalytic performance by merely increasing the copper content.

Fabrication of copper catalysts can be implemented through incorporating more reducible promoters, such as ceria, lanthana, and zirconia [???]. Cerium-containing materials are promising candidates for the application of NO abatement due to their high oxygen storage capacity and unique redox properties [??]. Nevertheless, a major drawback of cerium lies in its poor thermal and hydrothermal stability, especially under a practical diesel exhaust atmosphere [??]. It is well known that the incorporation of zirconia in cerium-based materials via high-temperature calcination can

enhance the thermal stability and dispersion of the active component on the support surface [? ?]. Furthermore, the addition of zirconia to ceria introduces structural defects through substitution of Ce⁴⁺ by Zr⁴⁺, which further enhances the oxygen storage capacity of ceria, the oxygen mobility in the lattice, the redox property, and thermal resistance [? ?]. Consequently, cerium and zirconium materials have the potential to improve the catalytic activity of Cu/ZSM-5 catalyst. Based on this background, the present work attempts to address the effects of the cerium and/or zirconium addition on Cu/ZSM-5 catalysts for SCR of NO by NH₃. A series of CuCe_xZr_{1-x}O_y/ZSM-5 catalysts (x = 0, 0.25, 0.50, 0.75 and 1) were synthesized using a conventional ion-exchange method, and the effects of adding cerium/zirconium metal ions into Cu/ZSM-5 catalysts on the SCR reaction were investigated using X-ray diffraction (XRD), transmission electron microscopy (TEM), X-ray photoelectron spectroscopy (XPS), and temperature-programmed reduction by hydrogen (H₂-TPR). The correlation between the structural characteristics, dispersion, reduction, and activity for the SCR process are discussed. The purpose of this work is to establish the relationships between structure and catalytic performance, which will be beneficial for the design and rationalization of the practical diesel catalysts.

2. Results and Discussion

2.1. Structure and Morphology

Figure **??** shows the XRD patterns of ZSM-5, Cu/ZSM-5, and CuCe_xZr_{1-x}/ZSM-5 (x = 0, 0.25, 0.5, 0.75 and 1). As expected, the catalysts still retain the ordered microstructure after the addition of copper, and cerium and/or zirconium because the inherent MFI structure of ZSM-5 ($2\theta = 7.8^{\circ}, 8.7^{\circ}$, 24.5°, 24.9°, PDF 44-003) can be observed for all the catalysts. However, the incorporation of copper, and cerium and/or zirconium leads to a decrease of the intensity of the ZSM-5 principal diffraction peaks, which can be attributed to the higher absorption coefficient of metal compounds for X-ray radiation [**?**]. Crystallized CuO ($2\theta = 35.5^{\circ}, 38.6^{\circ}, PDF = 02-1041$) nanoparticles are detected for the Cu/ZSM-5 catalysts. For the CuZr₁/ZSM-5, CuCe_{0.25}Zr_{0.25}/ZSM-5, CuCe_{0.5}Zr_{0.5}/ZSM-5, and CuCe_{0.75}Zr_{0.25}/ZSM-5 catalysts, no diffraction peaks for metal or metal oxide clusters were observed, indicating that the copper, zirconium, and cerium oxides are well dispersed as amorphous metal species, or aggregated into mini-crystals that are too small (<4 nm) to be detected by XRD [**?**]. For CuCe₁/ZSM-5, the observation of a small peak of CeO₂ ($2\theta = 28.2^{\circ}$, PDF = 34-0394) suggests that the extra-framework cerium is prone to agglomerating into cerium oxide clusters.



Figure 1. XRD patterns of $CuCe_xZr_{1-x}/ZSM$ -5 catalysts with different loading ratios: (a) ZSM-5, (b) Cu/ZSM-5, (c) $CuZr_1/ZSM$ -5, (d) $CuCe_{0.25}Zr_{0.75}/ZSM$ -5, (e) $CuCe_{0.5}Zr_{0.5}/ZSM$ -5, (f) $CuCe_{0.75}Zr_{0.25}/ZSM$ -5, and (g) $CuCe_1/ZSM$ -5.

To validate the XRD results, for each catalyst sample, more than 200 metal oxide particles from different TEM images were randomly chosen to determine the size distribution of the metal oxide particles. The size distributions of the metal oxide particles for $CuCe_{0.25}Zr_{0.75}/ZSM-5$, $CuCe_{0.5}/ZSM-5$, $CuCe_{0.75}Zr_{0.25}/ZSM-5$, and $CuCe_1/ZSM-5$ are shown in Figure ??. The metal oxide particles in the d, e and f samples are less than 4 nm in size, while most metal oxide particles in the g sample are larger than 4 nm in size. These results are consistent with the XRD data.



Figure 2. Size distributions of metal oxide particles for (d) CuCe_{0.25}Zr_{0.75}/ZSM-5, (e) CuCe_{0.5}Zr_{0.5}/ZSM5, (f) CuCe_{0.75}Zr_{0.25}/ZSM-5, and (g) CuCe₁/ZSM-5.

The typical TEM images of pure ZSM-5 and $CuCe_xZr_{1-x}/ZSM-5$ catalysts (x = 0, 0.5 and 1) are shown in Figure ??. The typical interference fringes of the ZSM-5 crystal structure are found in Figure ??a. Upon zirconium addition, small aggregates are observed for the $CuZr_1/ZSM-5$ catalyst (see Figure ??b), and most of them are less than 2 nm in size. The subsequent energy-dispersive X-ray (EDX) analysis for Point 1 in Figure ??b indicates a copper, zirconium, and oxygen-rich phase for these small aggregates. With increasing cerium content, the sizes of aggregates become larger, as shown in Figure ??c for $CuCe_{0.5}Zr_{0.5}/ZSM-5$ and Figure ??d for $CuCe_1/ZSM-5$. These results reveal that the addition of cerium decreases the dispersion of metal ions. The EDX analysis for Point 2 in Figure ??d indicates copper and cerium enrichment for the large black aggregates.



Figure 3. TEM images and energy-dispersive X-ray (EDX) quantitative analyses for (**a**) ZSM-5, (**b**) CuZr₁/ZSM-5, (**c**) CuCe_{0.5}Zr_{0.5}/ZSM-5, and (**d**) CuCe₁/ZSM-5.

2.2. XPS Analysis

The chemical state and surface composition of the elements in the catalysts were characterized by XPS. The Cu 2p spectra of the $CuCe_xZr_{1-x}/ZSM$ -5 catalysts (x = 0, 0.5 and 1) in Figure ??a show two main peaks: a peak at 933.3 eV that can be attributed to Cu $2p_{3/2}$, and a peak at about 953 eV that can be attributed to Cu $2p_{1/2}$. The intense satellite peak located at approximately 943 eV confirms the existence of divalent copper. Peak deconvolution and fitting to experimental data indicate that the Cu $2p_{3/2}$ peak can be well fitted by two peaks at 932.5 and 933.6 eV, corresponding to the Cu⁺ and Cu²⁺ ions, respectively [???].

The XPS spectra of O 1s for the CuCe_xZr_{1-x}/ZSM-5 catalysts (x = 0, 0.5, and 1) in Figure ??b show two primary peaks. The peak at a higher binding energy (BE) of 531.9 eV can be assigned to regular lattice oxygen from the ZSM-5 zeolite structure [??], while the shoulder peak at about 529.6 eV corresponds to characteristic lattice oxygen bound to metal (copper, cerium and zirconium) cations. As the cerium content increases by 6.5% in CuCe₁/ZSM-5, the shoulder peak corresponding to the lattice oxygen slightly shifts to a lower BE value (529.3 eV). Considering the findings from XRD that no metal oxide crystals are observed except for the CuCe₁/ZSM-5 catalyst, the existence of lattice oxygen implies that the oxides are well dispersed on the ZSM-5 support as the mini-crystal form.

The XPS doublet spectra of Zr 3d obtained from the $CuCe_xZr_{1-x}/ZSM$ -5 catalysts (x = 0 and 0.5) are shown in Figure ??c. The peaks centered at 181.9 and 184.2 eV correspond to Zr $3d_{5/2}$ and Zr $3d_{3/2}$, respectively. The BE of Zr $3d_{5/2}$ is about 0.9 eV, higher than that reported for zirconium metal and 0.5 eV lower than that reported for zirconia [?]. This phenomenon probably originates from the contribution of electrons from the surrounding species. Wang et al. [?] also found that the zirconium cations in Cu/ZrO₂ catalysts possess a lower BE of Zr $3d_{5/2}$ than zirconia, and they ascribed it to copper oxides located near oxygen vacancies on the exterior of the zirconia.

Figure **?**?d shows the XPS spectra of Ce 3d for the CuCe_xZr_{1-x}/ZSM-5 catalysts (x = 0.5 and 1). The complex spectrum of Ce 3d with respect to Ce $3d_{3/2}$ and Ce $3d_{5/2}$ ionization features are deconvoluted into eight components. The peaks at about 882.5, 889.3, and 898.5 eV are ascribed to the peak of Ce $3d_{5/2}$ and two "shake-up" satellite peaks, respectively. The peaks at approximately 901.1, 908.1, and 916.6 eV can be assigned to Ce $3d_{3/2}$ and two shake-up satellite peaks, respectively [**?**]. These features are viewed as the fingerprints for the existence of Ce⁴⁺. The inconspicuous peaks at about 885.5 and 904.3 eV represent the initial state of Ce³⁺ [**?**]. It is evident from Figure **??** d that cerium is mostly in a quadrivalent oxidation state, and a small quantity of Ce³⁺ co-exists. The presence of Ce³⁺ can be attributed to the relative homogeneous Ce_xZr_{1-x}O₂ or the substitution of Ce⁴⁺ by Zr⁴⁺ ions [**?**].



Figure 4. X-ray photoelectron spectra for (a) Cu 2p, (b) O 1s, (c) Zr 3d, and (d) Ce 3d over (1) CuZr₁/ZSM-5, (2) CuCe_{0.5}Zr_{0.5}/ZSM-5 and (3) CuCe₁/ZSM-5.

For the catalysts, the atomic ratios of Cu/Si, Ce/Si, and Zr/Si obtained from AAS and XPS are listed in Table **??**. The atomic ratios determined from XPS are larger than those obtained from AAS, suggesting that the metal oxides become enriched on the surface of ZSM-5 grains. The atomic ratios of Cu⁺/Cu²⁺ and Ce³⁺/Ce⁴⁺ calculated from XPS spectra are also listed in Table **??**. As the cerium content increases from 0% for CuZr₁/ZSM-5 to 3.4% for CuCe_{0.5}Zr_{0.5}/ZSM-5, the Cu⁺/Cu²⁺ ratio decreases from 0.64 to 0.35, while the Ce³⁺/Ce⁴⁺ ratio increases from 0 to 0.198. Upon further increasing the cerium content, the Cu⁺/Cu²⁺ ratio increases to 0.71 for the CuCe₁/ZSM-5 catalyst, while the Ce³⁺/Ce⁴⁺ ratio decreases after an initial increase.

Sampla	0 +10 2+	C 3+/C 4+	Cu/Si (Atomic)		Ce/Si (Atomic)		Zr/Si (Atomic)	
Sample	Cu ⁺ /Cu ⁻⁺	Ce ^s /Ce ^s	AAS	XPS	AAS	XPS	AAS	XPS
CuZr ₁ /ZSM-5	0.64	/	0.025	0.118	0	0	0.026	0.804
CuCe _{0.25} Zr _{0.75} /ZSM-5	0.498	0.187	0.026	0.116	0.005	0.015	0.018	0.723
CuCe _{0.5} Zr _{0.5} /ZSM-5	0.35	0.198	0.025	0.119	0.013	0.029	0.014	0.643
CuCe _{0.75} Zr _{0.25} /ZSM-5	0.672	0.298	0.024	0.117	0.020	0.034	0.007	0.312
CuCe ₁ /ZSM-5	0.71	0.247	0.026	0.116	0.025	0.066	0	0

Table 1. Surface compositions of $CuCe_xZr_{1-x}/ZSM-5$ catalysts (x = 0, 0.25, 0.5, 0.75, 1) obtained from X-ray photoelectron spectroscopy (XPS) analysis.

2.3. H₂-TPR

Figure ?? shows the H₂-TPR profiles of ZrO₂, CeO₂, and the Cu/ZSM-5 and CuCe_{0.75}Zr_{0.25}/ZSM-5 catalysts, and the relevant data are listed in Table ??. Three main reduction peaks are evident for the catalysts: the α -peak (182–213 °C) is ascribed to the reduction of the copper species dispersed on the ZSM-5 support, the β -peak (260–264 °C) is associated with the reduction of the copper oxide adhering to the external surface of zeolite crystallites, and the γ -peak (406 °C) is generally considered to be due to the reduction of bulk and unsupported copper oxide. Some of the copper species have incorporated into the vacant sites of cerium and/or zirconium oxides to form a coordinated surface structure with capping oxygen [?], corresponding to the δ peak (164–178 °C) and the ϵ peak (315–342 °C). As the cerium content increases, the δ peak shifts from 171 °C for CuZr₁/ZSM-5 to 164 °C for CuCe₁/ZSM-5, indicating that the composite oxides of copper and cerium are more readily reduced than those of copper and zirconium.



Figure 5. Temperature-programmed reduction by hydrogen (H₂-TPR) profiles of (a) ZrO_2 , (b) CeO_2 , (c) Cu/ZSM-5 and (d) $CuCe_{0.75}Zr_{0.25}/ZSM$ -5.

Total H2 Consumption (μmol/g _{cat})		188.59	190.42	192.82	197.18	202.39	202.23
' Peak	H ₂ Consumption (μmol/g _{cat})	63.96	28.53	23.54	18.66	14.09	23.33
	T (°C)	393	398	407	399	402	405
e Peak	H ₂ Consumption (μmol/g _{cat})	/	25.12	17.51	3.24	14.16	10.37
~	T (°C)	/	342	332	338	315	322
3 Peak	H ₂ Consumption (μmol/g _{cat})	68.23	76.22	68.49	61.40	58.10	82.83
	T (°C)	278	261	259	263	255	260
α Peak	H ₂ Consumption (μmol/g _{cat})	56.4	45.06	66.36	91.39	91.07	67.15
	T (°C)	236	198	209	213	192	182
ó Peak	H ₂ Consumption (μmol/g _{cat})	/	15.49	16.92	22.49	24.97	18.55
	T (°C)	/	171	175	178	170	164
Sample		Cu/ZSM-5	CuZr ₁ /ZSM-5	CuCe _{0.25} Zr _{0.75} /ZSM-5	CuCe _{0.5} Zr _{0.5} /ZSM-5	CuCe _{0.75} Z _{0.25} /ZSM-5	CuCe ₁ /ZSM-5

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From the H₂-TPR quantitative data in Table **??**, it can be seen that as the cerium content increases from 0% for CuZr₁/ZSM-5 to 5.2% for CuCe_{0.75}Zr_{0.25}/ZSM-5, the hydrogen consumption for the δ and α peak monotonously increases, while the hydrogen consumption for the β and γ peak gradually decreases. As mentioned above, the δ and α peak are related to highly dispersed copper species. Therefore, this increase in hydrogen consumption for the δ and α peak indicates that the addition of cerium improves the copper dispersion. For the CuCe₁/ZSM-5 catalyst, although the reduction temperatures of the δ and α peaks are lower than those for the other catalysts, the H₂ consumption of the δ and α peaks is not the maximum. This can be attributed to the formation of ceria, which is evidenced by the above XRD and TEM results. The ceria formed over the catalyst leads to the reduction not only extending deeply into the bulk of crystalline ceria but also being confined to its surface, so that the CuCe₁/ZSM-5 catalyst may consume less hydrogen than the CuCe_{0.75}Zr_{0.25}/ZSM-5 catalyst. Moreover, because the total H₂ consumption shows an increase with the cerium content increasing from 0% (CuZr₁/ZSM-5) to 5.2% (CuCe_{0.75}Zr_{0.25}/ZSM-5), the hydrogen uptake can be attributed not only to the copper reduction but also to a partial reduction of the cerium surface for the cerium-containing catalysts.

2.4. Catalytic Activity Test

Figure **??** shows the NO conversion in the SCR reaction on all the catalysts from 50 to 600 °C. All of the catalysts exhibit excellent SCR activity. For the Cu/ZSM-5 catalyst, the temperature range for 95% NO conversion is 209–406 °C. After the addition of cerium and/or zirconium to Cu/ZSM-5, the temperature range for 95% NO conversion extends toward both lower and higher temperatures, corresponding to active window broadening. The temperature ranges for 95% NO conversion are 190–436 °C for the CuZr₁/ZSM-5, 202–456 °C for CuCe_{0.25}Zr_{0.75}/ZSM-5, 203–460 °C for CuCe_{0.55}Zr_{0.5}/ZSM-5, 175–468 °C for CuCe_{0.75}Zr_{0.25}/ZSM-5, and 179–435 °C for CuCe₁/ZSM-5. Among the catalysts tested, the CuCe_{0.75}Zr_{0.25}/ZSM-5 catalyst has the highest catalytic activity for NO conversion. The improvements in the SCR activity after the addition of cerium and/or zirconium can be accounted for by three factors:

- The active copper species are well dispersed on the surface of the catalysts. The XRD and TEM results show that the introduction of zirconium promotes surface copper enrichment and prevents copper crystallization.
- (2) The addition of cerium results in high Cu⁺/Cu²⁺ ratios, as evidenced from the XPS results. In the SCR reaction, NO preferentially chemisorbs on Cu⁺, and the formation of intermediate Cu⁺–NO adsorbed sites is crucial for the activation of the NO molecule [??]. Thus, a high percentage of Cu⁺ species favors catalytic NO reduction.
- (3) Partial substitution of cerium and zirconium ions by copper ions contributes to the formation of copper/cerium or copper/zirconium composite oxides. As shown in the H₂-TPR experiment, these composite oxides greatly increase the reactive lattice oxygen content and reducibility, and thus improve the low-temperature activity [?].

Furthermore, as the temperature increases from 406 to 600 °C, a clear decrease of NO conversion for all the tested catalysts is observed in Figure ??. This decrease is mainly due to the occurrence of non-selective NH₃ oxidation by the reaction $4NH_3 + 5O_2 \rightarrow 4NO + 6H_2O$ at high temperatures [? ? ?]. The consumption of NH₃, together with the production of new NO, limits NO conversion.

The SCR process induced by all the catalysts inevitably produces N₂O and NO₂ as byproducts. In the whole temperature range considered in this study, the NO₂ concentration is below 3 ppm for all the catalysts (not shown). Figure **??** shows the yield of N₂O plotted against reaction temperature. For all the catalysts, the yield of N₂O increases with increasing temperature up until a certain temperature and then starts to decrease. The N₂O produced in the low temperature range mainly arises from side reactions between NO and NH₃: 4NO + 4NH₃ + 3O₂ \rightarrow 4N₂O + 6H₂O [**?**]. Upon further increase of the reaction temperature, the formation of N₂O is mainly from NH₃ oxidation via the

reaction $2NH_3 + 2O_2 \rightarrow N_2O + 3H_2O$ [??], and the amount of N_2O rapidly increases with the maximum yield at reaction temperatures <450 °C. In the high temperature range, non-selective NH_3 oxidation takes place by the reaction $4NH_3 + 5O_2 \rightarrow 4NO + 6H_2O$ [???], and thus the amount of N_2O rapidly decreases as the temperature increases. Moreover, close inspection of Figure ?? shows that the zirconium-containing catalysts yield less N_2O than the Cu/ZSM-5 and $CuCe_1/ZSM$ -5 catalysts, which means that the introduction of zirconium abates N_2O formation in the tested temperature range.



Figure 6. Catalytic activities for NO_x reduction by NH₃ for Cu/ZSM-5 and CuCe_xZr_{1-x}/ZSM-5 catalysts (x = 0, 0.25, 0.5, 0.75 and 1).



Figure 7. Yield of N₂O during the selective catalytic reduction (SCR) process as a function of reaction temperature for Cu/ZSM-5 andCuCe_xZr_{1-x}/ZSM-5 catalysts (x = 0, 0.25, 0.5, 0.75 and 1).

3. Experimental

3.1. Synthesis of Catalysts

H/ZSM-5 with an atomic Si/Al ratio of 25 was provided by Nankai University, Tianjin, China. A series of $CuCe_xZr_{1-x}O_y/ZSM$ -5 catalysts (x = 0, 0.25, 0.5, 0.75, and 1) were synthesized by an aqueous ion-exchange technique. A desired amount of copper acetate, and zirconium and cerium nitrate was added to deionized water and mixed with 20 g of H/ZSM-5 powder at room temperature. The resulting solution was stirred at 80 °C for 24 h. The samples were filtered and dried by evaporation in air and then calcined at 550 °C for 4 h. The copper content of the $CuCe_xZr_{1-x}O_y/ZSM$ -5 catalysts was fixed at 3 wt. %, and the molar ratio of Cu/(Ce + Zr) was 1:1. These catalysts are denoted as $CuZr_1/ZSM$ -5, $CuCe_{0.25}Zr_{0.25}/ZSM$ -5, $CuCe_{0.5}Zr_{0.5}/ZSM$ -5. The copper,

cerium, and zirconium contents of each calcined catalyst were determined by atomic absorption spectroscopy (AAS) using a PerkinElmer AAnalyst 300 spectrometer, and the results are shown in Table **??**. To evaluate the catalytic activity, the catalysts were compressed by a pressure of 20 MPa, then granulated and screened to the size of a 20–40 mesh.

Samula	Element Composition Measured by AAS (wt. %)				
Sample	Cu	Ce	Zr	Si	
CuZr ₁ /ZSM-5	3.0	0	4.4	52.4	
CuCe _{0.25} Zr _{0.75} /ZSM-5	3.1	1.3	3.1	52.2	
CuCe _{0.5} Zr _{0.5} /ZSM-5	2.9	3.4	2.3	52.3	
CuCe _{0.75} Zr _{0.25} /ZSM-5	2.8	5.2	1.2	52.4	
CuCe ₁ /ZSM-5	3.1	6.5	0	52.2	

Table 3. Element compositions of all catalysts.

3.2. Characterization

Powder X-ray diffraction (XRD) was performed using a Rigaku D/MAC/max 2500 v/pc instrument with Cu K α radiation (40 kV, 200 mA, λ = 1.5418 Å) (Jananese Science, Tokyo, Japan). The scan was performed with a 2θ rate of 0.02° /min from 5–80°. Transmission electron microscopy (TEM) images of the catalysts were obtained using a Philips Tecnai G2 F20 microscope operating at 200 kV equipped with an Oxford-1NCA energy-dispersive X-ray detector (EDX) (FEI, Hillsboro, OR, USA). Prior to measurement, the catalysts were dispersed in pure ethanol through sonication, and then mounted on nickel grids with a carbon film. X-ray photoelectron spectroscopy (XPS) was recorded using a Perkin-Elmer PHI-1600 ESCA spectrometer with a Mg Kα X-ray source (Perkin Elmer, Wellesley, MA, USA). The binding energies (BEs) were calibrated by referencing the C 1s at 284.8 eV. A ChemBet Pulsar system was used to perform temperature-programmed reduction by hydrogen (H₂-TPR) (Quantachrome Instruments, Boynton Beach, FL, USA). For the H₂-TPR experiments, 100 mg of the catalyst was exposed to pure Ar at a flow rate of 30 mL/min for 1 h at 300 °C. Then, the same gas was used to cool the catalyst down to 50 °C. The H₂-TPR measurements were carried out by increasing the temperature to 700 °C at a heating rate of 10 °C/min in 5% hydrogen at a constant flow rate of 30 mL/min. The hydrogen consumption was monitored and quantified using a thermal conductivity detector.

3.3. Catalytic Activity

The activities of the catalysts were measured in a continuous flow apparatus at atmosphere pressure. Before each test run, the catalyst powder was first pressed into a wafer and sieved into 20–40 meshes, and then 0.5 g of catalyst was set into a fix-bed reactor made of a quartz tube with an internal diameter of 10 mm. The reaction was carried out in the temperature range 50–600 °C, and a K-type thermocouple was located inside the catalyst bed to monitor the reaction temperature. The feed gas was controlled by calibrated electronic mass flow controllers, and contained 1000 ppm NO, 1000 ppm NH₃, and 10% O₂ and N₂ as the balance gas. The space velocity was set at 15,000 h⁻¹. An online mass spectrometry (Dycor LC-D200) was used to monitor the effluent NO, NO₂, N₂O, and NH₃. From the concentration of the gases at steady state, the NO_x conversion was defined as:

NO_x conversion (%) =
$$\frac{[NO_x]_{in} - [NO_x]_{out}}{[NO_x]_{in}} \times 100, \ [NOx] = [NO] + [NO_2]$$

4. Conclusions

The present study has highlighted the effects of cerium and/or zirconium incorporation into Cu/ZSM-5 on the SCR activity in NO abatement. The CuCe_xZr_{1-x}O_y/ZSM-5 catalysts used in

this study present a more than 95% NO conversion rate in a wide temperature range (175–468 °C), which is an improvement relative to the Cu/ZSM-5 catalyst (209–406 °C) in terms of SCR activity. The zirconium doping improves copper dispersion, and the introduction of zirconium promotes surface copper enrichment and prevents copper crystallization, which favors catalytic NO reduction. Partial substitution of cerium and zirconium ions by copper ions increases the reactive lattice oxygen content and reducibility, and thus improves the low-temperature activity. Because of the existence of the Ce³⁺/Ce⁴⁺ redox couple in the cerium-containing catalysts, the SCR activities of cerium-rich catalysts are higher than those of zirconium-rich catalysts. Moreover, formation of N₂O is suppressed in the range of test temperatures due to the presence of zirconium in the catalysts.

Author Contributions: This study was conducted through contributions of all authors. C.S. and G.L. designed the study and wrote the manuscript. Y.L., C.F. and X.L. designed and performed the experiments and analyzed the data. Y.L. checked and corrected the manuscript.

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Article



A CeO₂/ZrO₂-TiO₂ Catalyst for the Selective Catalytic Reduction of NO_x with NH₃

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Abstract: In this study, $CeZr_{0.5}Ti_aO_x$ (with a = 0, 1, 2, 5, 10) catalysts were prepared by a stepwise precipitation approach for the selective catalytic reduction of NO_x with NH₃. When Ti was added, all of the Ce-Zr-Ti oxide catalysts showed much better catalytic performances than the $CeZr_{0.5}O_x$. Particularly, the $CeZr_{0.5}Ti_2O_x$ catalyst showed excellent activity for broad temperature range under high space velocity condition. Through the control of pH value and precipitation time during preparation, the function of the $CeZr_{0.5}Ti_2O_x$ catalyst could be controlled and the structure with highly dispersed CeO_2 (with redox functions) on the surface of ZrO_2 -TiO₂ (with acidic functions) could be obtained. Characterizations revealed that the superior catalytic performance of the catalyst is associated with its outstanding redox properties and adsorption/activation functions for the reactants.

Keywords: Ce-based catalyst; stepwise precipitation; selective catalytic reduction; diesel exhaust; nitrogen oxides abatement

1. Introduction

 NO_x (mainly NO and NO₂) in the atmosphere plays critical roles in the formation of severe air pollution problems, such as haze, acid rain, and photochemical smog. In the last few decades, great efforts have been devoted to the development of NO_x emission control technologies [1–3]. Selective catalytic reduction of NO_x with NH₃ (NH₃-SCR) has been widely applied for the removal of NO_x generated from stationary sources for many years, and it has also been used for the control of NO_x emission from diesel vehicles [2,4].

Catalysts play an important role in the development of NH₃-SCR technology [5,6]. Vanadium-based catalyst (especially V_2O_5 -WO₃/TiO₂), with excellent SO₂ resistance, is the most widely used NH₃-SCR catalyst for NO_x emission control from power plants, and it was also applied on diesel vehicles as the first generation of SCR catalyst [4]. However, this catalyst system still has some problems, including the toxicity of active V_2O_5 , narrow temperature window, and low thermal stability [2].

There has been strong interest in developing a vanadium-free catalyst that can be used on diesel vehicles [5–11]. Ce is a key component in three-way catalysts for emission control in automobiles for gasoline. CeO₂ provides an oxygen storage function through redox cycling between Ce³⁺ and

 Ce^{4+} . In recent years, Ce has also attracted great attention for applications as a support [12,13], promoter [14–18], or main active component [19–26] for NH₃-SCR catalysts.

Pure Ce oxide is not suitable for use as an NH₃-SCR catalyst [27,28]. When Zr oxide was introduced into Ce oxide, the thermal stability and the oxygen storage capacity of the oxide could be significantly improved. Therefore, Ce-Zr oxide was investigated for NH₃-SCR [12,13,29–34]. In the NH₃-SCR reaction, both redox functions and acidic functions of the catalyst are needed [4,35]. Therefore, a high dispersion of active sites and close coupling of redox with acid sites is the way to design a highly efficient NH₃-SCR catalyst.

In this study, starting from a preparation of Ce-Zr oxide by the co-preparation method, we developed a Ce-Zr-Ti oxide catalyst using a stepwise precipitation approach, under the theoretical guidance of the close combination of the Ce-Zr oxide with strong redox functions and Ti oxide with excellent acid properties [4,5]. This obtained catalyst showed superior catalytic performance for NH₃-SCR.

2. Results and Discussion

2.1. NH₃-SCR Activity

Figure 1A presents the NO_x conversion over the catalysts with different Ti contents under a relatively high gas hourly space velocity (GHSV) of 200,000 h⁻¹. The CeZr_{0.5}O_x just exhibited over 50% NO_x conversion in a narrow temperature range of 350–425 °C. When Ti was introduced, all of the Ce-Zr-Ti oxide catalysts exhibited much better activities. With the increase in Ti content, the low temperature firstly increased and then decreased. As a result, the CeZr_{0.5}Ti₂O_x catalyst presented the best activity in a low temperature range, together with a high NO_x conversion in a wide temperature range. On the other hand, the variation in high temperature activity with Ti content was contrary to that of low temperature activity, with the activity of CeZr_{0.5}Ti₂O_x slightly lower than those of the other Ce-Zr-Ti oxide catalysts in a high temperature range. In addition, adding Ti to the catalyst also enhanced the N₂ selectivity, and the Ce-Zr-Ti oxide catalysts all presented higher N₂ selectivity than CeZr_{0.5}O_x (Figure 1B).



Figure 1. (A) NO_x conversions and (B) N₂ selectivity over the CeZr_{0.5}O_x and Ce-Zr-Ti oxide catalysts. Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol.%, N₂ balance, and GHSV = 200,000 h⁻¹.

The influences of H₂O and space velocity on the NO_x conversion over CeZr_{0.5}Ti₂O_x were tested and the results are shown in Figure 2. The existence of 5% H₂O in the flow gas decreased the low temperature activity, but enhanced the high temperature activity. As a result, over 80% NO_x conversion could still be achieved from 250 to 450 °C. When the GHSV was decreased from 200,000 h⁻¹ to 100,000 h⁻¹, the activity of the catalyst at low temperatures was obviously improved.



Figure 2. NO_{*x*} conversion over $CeZr_{0.5}Ti_2O_x$ catalyst under different reaction conditions. Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol.%, [H₂O] = 5 vol.% (when used), N₂ balance, and GHSV = 100,000 or 200,000 h⁻¹.

2.2. Separated NO/NH₃ Oxidation

To analyze the effects of Ti on the catalyst, separated NO oxidation and NH₃ oxidation tests were carried out for the CeZr_{0.5}O_x and CeZr_{0.5}Ti₂O_x (Figure 3). The NO₂ production during NO oxidation over the CeZr_{0.5}Ti₂O_x was clearly higher than that over CeZr_{0.5}O_x at a low temperature. Since the presentation of NO₂ in the reaction gas could promote the SCR reaction at a low temperature by accelerating the fast SCR process (2 NH₃ + NO + NO₂ \rightarrow 2N₂ + 3H₂O), the enhanced low-temperature activity by the introduction of Ti should be associated with the promoted oxidation of NO to NO₂ over CeZr_{0.5}Ti₂O_x [10,35]. In addition, the introduction of Ti also promoted NH₃ oxidation over the catalyst at a high temperature. The NH₃-SCR reaction route at a high temperature mainly follows the Eley-Rideal mechanism, and the activation of NH₃ to form NH₂ species by oxidation plays the key role for the reaction with NO to form N₂ and H₂O, owing to NH₂ + NO(g) \rightarrow N₂ + H₂O. Therefore, promoted NH₃ oxidation would be beneficial for the improvement of high temperature activity.



Figure 3. (A) NO₂ productions during separate NO oxidation reaction and (B) NH₃ conversions during separate NH₃ oxidation reaction over the CeZr_{0.5}O_x and CeZr_{0.5}Ti₂O_x catalysts. Reaction conditions: (A) [NO] _ENREF_30= 500 ppm, (B) [NH₃] = 500 ppm, [O₂] = 5 vol.%, N₂ balance and GHSV = 200,000 h⁻¹.

2.3. XRD

The X-ray diffraction (XRD) results of the $CeZr_{0.5}O_x$ and Ce-Z-Ti oxide catalysts are presented in Figure 4. Both CeO_2 and ZrO_2 were detected in $CeZr_{0.5}O_x$. With the increase of Ti, the peaks for CeO_2 and ZrO_2 became more and more weak, and only anatase TiO₂ was observed for $CeZr_{0.5}Ti_{10}O_x$. Only weak peaks for CeO_2 with cubic fluorite structures (PDF# 43-1002) were observed in the $CeZr_{0.5}Ti_{2}O_x$, indicating that the introduction of Ti had induced the structural change of the $CeZr_{0.5}O_x$, and the crystallizations of Ce, Zr and Ti oxides in $CeZr_{0.5}Ti_{2}O_x$ were significantly inhibited. As a result, the $CeZr_{0.5}Ti_{2}O_x$ (165.1 m²/g) showed a higher Brunauer–Emmett–Teller (BET) surface area than $CeZr_{0.5}O_x$ (113.5 m²/g).



Figure 4. XRD patterns of the $CeZr_{0.5}O_x$ and Ce-Z-Ti oxide catalysts.

2.4. H_2 -TPR

The H₂ temperature-programmed reduction (H₂-TPR) profiles of $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$ are presented in Figure 5. The $CeZr_{0.5}O_x$ exhibited two peaks at 496 and 755 °C due to the surface and bulk reductions of CeO₂ (as detected by XRD), respectively [31,36–38]. During the test, coordinatively unsaturated surface oxygen anions are easily reduced by H₂ in the low temperature region, while the bulk oxygen species are reduced only after the transportation to the surface [39]. With the addition of Ti, a sharp H₂ consumption peak appeared at 567 °C, which indicates that another type of Ce species might be formed. Considering the XRD results, this sharp peak might be associated with the reduction of the highly dispersed Ce species from Ce⁴⁺ to Ce³⁺ [22,34]. In addition, the H₂ consumption of $CeZr_{0.5}Ti_2O_x$ was much higher than that of $CeZr_{0.5}O_x$ at a low temperature. The H₂-TPR results clearly indicated the enhancement of redox functions for $CeZr_{0.5}Ti_2O_x$.



Figure 5. H₂-TPR profiles of the $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$ catalysts.

Previous studies have indicated that the redox properties of NH₃-SCR catalyst play a dominant role in the low temperature activity [35,40,41]. Therefore, the enhanced redox function of $CeZr_{0.5}Ti_2O_x$ would beneficial for low temperature activity.

2.5. NO_x/NH₃-TPD

To investigate the NO_x and NH₃ adsorption/desorption properties of $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$, NO_x temperature-programmed desorption (NO_x-TPD) and NH₃ temperature-programmed desorption (NH₃-TPD) were performed for the catalysts (Figure 6).



Figure 6. (A) NO_x-TPD and (B) NH₃-TPD profiles of the CeZr_{0.5}O_x and CeZr_{0.5}Ti₂O_x catalysts.

The NO_x-TPD profiles are presented in Figure 6A. The first NO_x peak of $CeZr_{0.5}Ti_2O_x$ was at ca. 110 °C, mainly due to the desorption of physisorbed NO_x, while the other NO_x peak was at ca. 300 °C and was associated with the decomposition of chemsorbed NO_x species [42,43]. On the other hand, two weak peaks were observed for $CeZr_{0.5}O_x$ at ca. 270 °C and ca. 410 °C, respectively, which were due to the decomposition of different types of chemsorbed NO_x species. With the addition of Ti, the adsorbed NO_x on $CeZr_{0.5}Ti_2O_x$ was obviously more than that of $CeZr_{0.5}O_x$. Particularly, the desorbed NO₂ of $CeZr_{0.5}Ti_2O_x$ was much higher, owing to the enhanced low-temperature activity for NO oxidation (as shown by the separated NO oxidation results), which could facilitate the conversion of NO_x in NH₃-SCR.

Surface acidity plays a dominate role in the high-temperature SCR activity due to its effects on the adsorption and activation of NH₃ [35,41]. Previous studies have revealed that Ti species of NH₃-SCR catalysts mainly act as acid sites in the reaction for NH₃ adsorption [4]. Therefore, the adsorbed NH₃

of $CeZr_{0.5}Ti_2O_x$ was much more than that of $CeZr_{0.5}O_x$, which might be an important reason for the better NH₃-SCR activity of $CeZr_{0.5}Ti_2O_x$ at high temperatures.

2.6. XPS

The X-ray photoelectron spectroscopy (XPS) results for Ce 3d of the CeZr_{0.5}O_x and CeZr_{0.5}Ti₂O_x are shown in Figure 7. The sub-bands labeled with u'/v' and u⁰/v⁰ represent the 3d¹⁰4f¹ initial electronic state corresponding to Ce³⁺ and the 3d⁹4f² state of Ce³⁺, respectively [44]. The sub-bands labeled with u''' and v''' represent the 3d¹⁰4f⁰ state of Ce⁴⁺, and the sub-bands labeled with u, u'', v and v'' represent the 3d⁹4f¹ state corresponding to Ce⁴⁺ [44]. The presence of Ce³⁺ would induce a charge imbalance, which could lead to unsaturated chemical bonds and oxygen vacancies. The calculated Ce³⁺ ratio of CeZr_{0.5}Ti₂O_x (36.0%) was higher than that of CeZr_{0.5}O_x (33.8%), indicating that more surface oxygen vacancies presented in CeZr_{0.5}Ti₂O_x. In addition, the Ce³⁺ ratio of the catalyst could influence the redox ability and reactant adsorption and activation functions, and thereby contribute to NH₃-SCR performance.



Figure 7. XPS results of Ce 3d of the $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$ catalysts.

The surface oxygen vacancies of the catalysts might generate weakly-adsorbed oxygen species or additional chemisorbed oxygen on the surface of the catalyst [27,45]. The XPS results of O 1s of the CeZr_{0.5}O_x and CeZr_{0.5}Ti₂O_x are shown in Figure 8. The O 1s peak was fit into two sub-bands. The sub-bands at 531.2–531.5 eV and 529.1–529.6 eV were assigned to the surface adsorbed oxygen (O_{α}) , such as the $O_2^{2^-}$ and O^- belonging to defect-oxide or a hydroxyl-like group, and the lattice oxygen O^{2^-} (O_{β}), respectively [46]. The O_{α} ratios of the catalysts were calculated by $O_{\alpha}/(O_{\alpha} + O_{\beta})$, and the CeZr_{0.5}Ti₂O_x showed higher O_{α} ratio than CeZr_{0.5}O_x. The results confirmed that the addition of Ti indeed induced more surface-adsorbed oxygen, which would facilitate NO oxidation to NO₂ (as shown by the separated NO oxidation and NO_x-TPD results), and thus facilitates the conversion of NO by fast SCR effects.

2.7. Formation Process Analysis of the $CeZr_{0.5}Ti_2O_x$ Catalyst

Figure 9 shows the pH variations of the mixed solutions for the preparation of the $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$ catalysts. During the preparation of $CeZr_{0.5}O_x$, the initial pH value of the solution was 1.6. With the hydrolysis of urea, the pH increased gradually to be 7.6 after heating for 12 h. Due to the increase in pH, suspended particles began to appear in the solution in the second hour. The particles with the precipitation time of 2 h, 4 h, 6 h, and 12 h were collected and then calcined to be catalyst samples. The activity tests of these samples showed similar NO_x conversions with each other.



Figure 8. XPS results of O 1s of the $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$ catalysts.



Figure 9. The pH variation of the mixed solution during the preparation of the (A) $CeZr_{0.5}O_x$ and (B) $CeZr_{0.5}Ti_2O_x$ catalysts, and the NO_x conversions of the obtained samples at different precipitation time.

Due to the acidity induced by the added $Ti(SO_4)_2$, the initial pH value of the mixed solution during the preparation of $CeZr_{0.5}Ti_2O_x$ dropped to be 1.1. With the hydrolysis of urea, the pH increased gradually after heating, and some white particles generated in the first hour and suspended in the

solution. With the increase of time, the particles gradually turned yellow. The pH reached ca. 7.0 after 12 h of reaction. The particles with the precipitation times of 1 h, 4 h, 6 h, and 12 h were collected and then calcined to be catalyst samples. Interestingly, the activity test showed a remarkable enhancement of NO_x conversions for the four samples with the increase in precipitation time.

The surface metal atomic concentrations of the $CeZr_{0.5}Ti_2O_x$ samples with different precipitation times were analyzed using XPS, and the variations in Ce, Zr, and Ti concentrations with precipitation time are shown in Figure 10. For the 1-h precipitation sample, only Ti and Zr, without Ce, were detected. With the increase in precipitation time, surface Ce concentration increased gradually in the samples. At the same time, Ti and Zr concentrations gradually decreased with the increase in precipitation time. A TEM-EDS mapping image showed that Ce was highly dispersed in the CeZr_{0.5}Ti₂O_x catalyst (Figure 11).



Figure 10. Surface metal atomic concentrations of the $CeZr_{0.5}Ti_2O_x$ samples with different precipitation times.



Figure 11. TEM image (A) and the corresponding EDS mapping (B) for the Ce of the $CeZr_{0.5}Ti_{2}O_x$ catalyst.

Considering the variations in the solution pH value when preparing the $CeZr_{0.5}Ti_2O_x$, the formation process of the catalyst can be proposed as follows: The Ti and Zr species were first co-precipitated with the increase in solution pH. Then, the Ce species uniformly precipitated onto the precipitated Zr-Ti species with the further increase in pH. Finally, a $CeZr_{0.5}Ti_2O_x$ catalyst with a higher surface Ce concentration than Ti and Zr was obtained. Through control of the hydrolysis of urea, the variations in the solution pH can be controlled, and then we can control the precipitation process of

the catalyst, which is very important for the formation of highly-dispersed CeO₂ on ZrO_2 -TiO₂. Thus, the obtained catalyst can present excellent NH₃-SCR performance.

3. Experimental Section

3.1. Catalyst Preparation and Activity Test

The CeZr_{0.5}Ti_aO_x (a = Ti/Ce molar ratio = 0, 1, 2, 5, 10), with a Zr/Ce molar ratio fixed to be 0.5, was prepared using a precipitation method. Desired precursors of Ce(NO₃)₃·6H₂O (>99%, Sinopharm Chemical Reagent Co., Ltd., Shanghai, China), Zr(NO₃)₄·5H₂O (>99%, Sinopharm Chemical Reagent Co., Ltd., Shanghai, China) and Ti(SO₄)₂ (>98%, Sinopharm Chemical Reagent Co., Ltd., Shanghai, China) and Ti(SO₄)₂ (>98%, Sinopharm Chemical Reagent Co., Ltd., Shanghai, China) were dissolved together in distilled water, and urea (>99%, Sinopharm Chemical Reagent Co., Ltd., Shanghai, China) was added to the mixed solution as a slowly-releasing precipitator. Then, the solution was heated to 90 °C to facilitate the release of NH₃ and thereby raise the pH value gradually. The temperature of the mixed solution was held at 90 °C for 12 h under vigorous stirring (some samples with shorter precipitation times were also prepared). After that, the precipitated powders were collected via filtration, washed using distilled water, and dried for 12 h at 100 °C. Finally, the catalyst was obtained after calcination at 500 °C for 5 h.

The SCR activity of the catalysts (40–60 mesh) were tested in a fixed-bed quartz flow reactor. The reaction conditions were controlled as follows: 500 ppm NO, 500 ppm NH₃, 5 vol.% O₂, N₂ balance, and 400 mL/min total flow rate. Different gas hourly space velocities (GHSVs) were obtained by changing the volume of catalysts, i.e., 0.24 mL catalyst for a GHSV = 100,000 h⁻¹ and 0.12 mL catalyst for a GHSV = 200,000 h⁻¹. The concentrations of effluent N-containing gases (NO, NH₃, NO₂ and N₂O) were continuously measured by an online FTIR gas analyzer (Nicolet Antaris IGS analyzer, Thermo-Fisher Scientific, Waltham, MA, USA). NO_x conversion and N₂ selectivity were calculated using the following equations, respectively:

$$\begin{split} \text{NO}_{x} \text{ conversion} &= (1 - \frac{[\text{NO}]_{out} + [\text{NO}_{2}]_{out}}{[\text{NO}]_{in} + [\text{NO}_{2}]_{in}}) \times 100\% \\ \text{N}_{2} \text{ selectivity} &= (1 - \frac{2[\text{N}_{2}\text{O}]_{out}}{[\text{NO}_{x}]_{in} + [\text{NH}_{3}]_{in} - [\text{NO}_{x}]_{out} - [\text{NH}_{3}]_{out}}) \times 100\% \end{split}$$

3.2. Characterizations

X-ray diffraction (XRD) measurements were carried out on a computerized AXS D8 diffractometer (Bruker, GER), with Cu K α (λ = 0.15406 nm) radiation, from 20 to 80° at 8°/min.

Surface areas were tested using an ASAP 2020 (Micromeritics, Norcross, GA, USA) at -196 °C by N₂ adsorption/desorption and calculated using a BET equation in the 0.05–0.35 partial pressure range.

The X-ray photoelectron spectroscopy (XPS) results of Ce 3d and O 1s were measured on an ESCALAB 250Xi Scanning X-ray Microprobe (Thermo-Fisher Scientific, Waltham, MA, USA) using Al Ka radiation (1486.7 eV) and a C 1 s peak, with BE = 284.8 eV as the calibration standard.

The transmission electron microscopy (TEM) image and energy-dispersive X-ray spectroscopy (EDS) mapping of Ce were obtained using a JEM-2100F equipment (JEOL, Tokyo, Japan), combined with a specimen tilting beryllium holder for energy dispersive spectroscopy. The accelerating voltage was 200 kV.

The H₂ temperature-programmed reduction (H₂-TPR) was tested using an AutoChem_II_2920 chemisorption analyzer (Micromeritics, Norcross, GA, USA), and the temperature-programmed desorption of NH₃ and NO_x (NO_x-TPD and NH₃-TPD) were tested using the same reaction system as the activity tests. Experiment details can be found in Reference [42].

4. Conclusions

A series of Ce-Zr-Ti oxide catalysts were prepared using a stepwise precipitation approach for NH₃-SCR. CeZr_{0.5}O_x without Ti just showed a relatively low NO_x conversion. When Ti was introduced, Ce-Zr-Ti catalysts showed much better activities and N₂ selectivity. A CeZr_{0.5}Ti₂O_x catalyst, which contains moderate Ti amounts, showed the best performance, which is associated with its optimal ratios for the redox (CeO_x) and acidic (TiO₂) components.

 $CeZr_{0.5}O_x$ and $CeZr_{0.5}Ti_2O_x$ catalysts were characterized using various methods and the formation process during preparation was investigated. $CeZr_{0.5}Ti_2O_x$ catalyst showed superior redox properties (by H₂-TPR), good adsorption and NO_x/NH₃ activation functions (by NO_x-TPD and NH₃-TPD, respectively), and enhanced charge imbalance (by XPS).

During preparation, the Ti and Zr species were first co-precipitated with an increase in solution pH. Then, the Ce species uniformly precipitated onto the precipitated Zr-Ti species with the further increase in pH. As a result, $CeZr_{0.5}Ti_2O_x$ catalyst with a surface Ce concentration higher than those of Ti and Zr was obtained. This preparation process resulted in the formation of highly-dispersed CeO₂ on ZrO₂-TiO₂, and thus the catalyst can present excellent NH₃-SCR performance.

Author Contributions: W.S. and H.H. conceived the project; Y.G. and Y.Z. performed the experiments; W.S. and Z.L. carried out the data analysis; W.S. and Y.G. wrote the paper; H.H. supervised the study.

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Article



Gas-Phase Phosphorous Poisoning of a Pt/Ba/Al₂O₃ NO_x Storage Catalyst

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Abstract: The effect of phosphorous exposure on the NO_x storage capacity of a $Pt/Ba/Al_2O_3$ catalyst coated on a ceramic monolith substrate has been studied. The catalyst was exposed to phosphorous by evaporating phosphoric acid in presence of H₂O and O₂. The NO_x storage capacity was measured before and after the phosphorus exposure and a significant loss of the NO_x storage capacity was detected after phosphorous exposure. The phosphorous poisoned samples were characterized by X-ray photoelectron spectroscopy (XPS), environmental scanning electron microscopy (ESEM), N₂-physisorption and inductive coupled plasma atomic emission spectroscopy (ICP-AES). All characterization methods showed an axial distribution of phosphorous ranging from the inlet to the outlet of the coated monolith samples with a higher concentration at the inlet of the samples. Elemental analysis, using ICP-AES, confirmed this distribution of phosphorous on the catalyst surface. The specific surface area and pore volume were significantly lower at the inlet section of the monolith where the phosphorous concentration was higher, and higher at the outlet where the phosphorous concentration was lower. The results from the XPS and scanning electron microscopy (SEM)-energy dispersive X-ray (EDX) analyses showed higher accumulation of phosphorus towards the surface of the catalyst at the inlet of the monolith and the phosphorus was to a large extent present in the form of P_4O_{10} . However, in the middle section of the monolith, the XPS analysis revealed the presence of more metaphosphate (PO_3^-). Moreover, the SEM-EDX analysis showed that the phosphorous to higher extent had diffused into the washcoat and was less accumulated at the surface close to the outlet of the sample.

Keywords: LNT; NSR; NOx storage; phosphorous; deactivation; poisoning

1. Introduction

In diesel and lean burn gasoline engines, the engine operates with a large excess of air, and this results in an increased fuel economy, and thereby reduced emissions of CO_2 , which is a greenhouse gas. However, it is critical to remove NO_x , which can be done by different catalytic aftertreatment techniques. Emission standards have therefore been implemented in large parts of the world and the emission levels have been significantly reduced over the years [1].

Since the traditional Three-way catalyst (TWC), requires stoichiometric air:fuel ratio, the Lean NO_x Trap (LNT) was introduced as a catalytic concept for exhaust gas treatment in diesel vehicles [2–5]. In the LNT, usually barium is used as a storage component, and the nitrogen oxides during lean conditions are stored on barium as barium nitrates. Moreover, noble metals, such as platinum, play a key role to oxidize the NO to NO_2 , which facilities the NO_x storage [2,6–10]. Under short periods of time the engine is switched from operating under lean to rich conditions. Under rich

conditions, barium nitrates decompose and are reduced over the noble metal sites to mainly N_2 and H_2O . The reducing agents in the exhaust gas are hydrocarbons (HC), carbon monoxide (CO) and hydrogen (H₂), which react with the NO_x to produce CO₂, H_2O and N_2 . Nitrogen is the desired product from the reduction of stored nitrates. However, there are other possible bi-products from the reduction process such as ammonia (NH₃) and nitrous oxide (N₂O) [3].

A vehicle is expected to be in service for a long time, which requires the catalyst to be durable. Over time, the catalytic properties of the catalyst deteriorate. There are a few mechanisms that contribute to degradation of catalysts, such as sintering of catalytic particles as a result of high-temperature exposure [11]. The catalytic washcoat can also undergo mechanical tear and poisoning of active sites on the catalyst caused by chemisorption or reactions of catalytically active sites with poisons in the exhaust gas [12]. The poisons are often introduced to the system through the gasoline or diesel fuel, which is the case for SO₂, or through the lubricant of the engine. Lubricants, such as zinc dialkyldithiophosphate (ZDDP), are commonly used in vehicles and are a source for accumulation of both sulfates, zinc and phosphorus in the catalyst [13]. The deactivation effect of sulfur has extensively been studied over LNTs [12,14–16], and also how to regenerate the catalyst from sulfur [17].

There are only a few studies available in the open literature where phosphorus poisoning has been studied over noble metal catalysts used for emission control. Bunting et al. [13] studied a diesel oxidation catalyst (DOC) connected to an engine rig where the fuel was doped with ZDDP, which showed accumulation of foreign compounds such as phosphates, zinc and sulfates on the catalyst. Also, results from characterization of field-aged TWCs showed accumulation of elements found from ZDDP poisoning of catalysts [18–20]. Moreover, it has been shown that the distribution of phosphorus as it accumulates in the catalytic washcoat exhibits a gradient, with higher concentration at the inlet and lower along the washcoat length [19–23]. Compared to sulfur, which has a more even distribution over the washcoat and penetrates deep into the washcoat, phosphorus tends to be more concentrated close to the washcoat surface [13]. A study performed by Galisteo et al. [24], where a Pt/Ba/Al₂O₃ catalyst was impregnated with (NH₄)₂HPO₄, showed that the amount of phosphorus in the catalyst correlated with the loss of NO_x storage capacity.

However, there are, to our knowledge, no studies in the open literature that have examined phosphorus poisoning of LNT catalysts using vapour phase exposure that is closer to realistic poisoning conditions, which is the objective of the present work. In this study, we have examined a $Pt/Ba/Al_2O_3$ catalyst exposed to phosphorous in the gas phase using different phosphorous concentrations. The catalyst samples were studied in a flow reactor to determine changes in NO_x storage capacity, NO_x reduction activity and bi-product formation. The phosphorous-exposed samples were also characterized by X-ray photoelectron spectroscopy (XPS), nitrogen-physisorption, Inductive coupled plasma atomic emission spectroscopy (ICP-AES) and Environmental Scanning Electron Microscopy (ESEM) with energy dispersive X-ray (EDX) analysis.

2. Results and Discussion

2.1. Activity Measurements

Lean/rich cycling experiments were conducted in which the system was exposed to 400 vol.-ppm NO, 5 vol.-% CO₂, 5 vol.-% H₂O and 8 vol.-% O₂ during the lean phase. In this phase, NO is first oxidized over the platinum sites, and thereafter storage of NO_x species occurs as nitrates through a disproportionation mechanism, but also via nitrite formation at low temperature [5]. The NO oxidation capacity of the catalyst is clearly visible by the NO₂ formation seen in Figure 1. In the initial part of the lean phase, all NO_x is stored and after a certain amount of time, breakthrough of NO_x, due to saturation of the storage sites, is seen. Most NO_x is stored onto the barium sites in form of Ba(NO₃)₂, but some NO_x is also stored onto alumina sites [3]. At lower temperature, storage of NO_x occurs over

both the barium and alumina sites, but at higher temperature the barium sites dominate the NO_x storage [25], due to the lower stability of alumina nitrates.



Figure 1. Measured outlet NO, NO₂, N₂O and NH₃ concentrations for the Pt/Ba/Al₂O₃ sample "100 ppm, 34 h" before and after exposure to phosphorous (100 vol.-ppm H₃PO₄, 8 vol.-% O₂, 5 vol.-% H₂O at 200 °C) in a flow-reactor experiment at 400 °C. Lean phase; 400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and 8 vol.-% O₂. Rich phase; 400 vol.-ppm NO, 5 vol.-%H₂O, 5 vol.-% CO₂ and 1 vol.-% H₂.

As the system is switched from lean to rich phase, a NO_x peak is visible, which is related to release of stored NO_x in the catalytic material [26]. Directly after the switch from the lean to the rich (1 vol.-% H₂, 5 vol.-% CO₂ and 5 vol.-% H₂O) phase, minor formation of N₂O is seen and it is during this short period of time, that most of the N₂ formation is suggested to occur [27–29]. After the N₂O peak a large ammonia peak is observed, which is consistent with the work by Lindholm et al. [25]. Most of the formed NH₃, early in the rich phase, is consumed by a Selective Catalytic Reduction (SCR) reaction with the stored NO_x, according to Lietti et al. [27] and Lindholm et al. [6]. Hence, the NH₃ peak is delayed and will dominate more as the rich phase proceeds and this is also seen in Figure 1. Indeed, spatially resolved MS measurements performed by Partridge et al. [10] showed a clear axial consumption of ammonia.

The effect of phosphorus exposure was studied by exposing the Pt/Ba/Al₂O₃ catalyst to 100 vol.-ppm H₃PO₄, 8 vol.-% O₂ and 5 vol.-% H₂O at 200 °C for 34 h and thereafter repeating the same experiment, see Figure 1. In addition, a second catalyst was used, which was exposed to 50 vol.-ppm H₃PO₄ instead. These catalysts were studied under lean/rich cycling experiments, which were conducted both at 300 and 400 °C, using 400 vol.-ppm NO, 5 vol.-% CO₂, 8 vol.-% O₂, and 5 vol.-% H₂O for the lean conditions and 1 vol.-% H₂ or 1000 vol.-ppm C₃H₆, 400 vol.-ppm NO, 5 vol.-% CO₂ and 5 vol.-% H₂O during the rich conditions, and the resulting outlet NO_x concentrations are shown in Figure 2. The results show that exposure to phosphorus causes a significant loss of NO_x storage capacity for the Pt/Ba/Al₂O₃ catalyst. The loss of NO_x storage capacity is summarised in Table 1. It is clear that the loss in storage capacity is similar for both poisoning levels and, moreover, that the effect is more pronounced after exposure at 300 compared to 400 °C.

Our results are in line with those in the study by Galisteo et al. [24], where the effect of phosphorous on the NO_x storage capacity of Pt/Ba/Al₂O₃ was studied by impregnation of the catalyst with (NH₃)₃PO₄ dissolved in water. The authors also observed a decreased NO_x storage capacity after phosphorous exposure. At 300 °C, the NO_x reduction properties of C₃H₆ are poor compared to 400 °C (compare graph B and D in Figure 2), which previously has been observed by

Olsson et al. [30]. However, loss in NO_x storage capacity is still observed for this case after phosphorus exposure. For exposure at 400 °C, the phosphorus poisoning is similar for both H₂ and C₃H₆, which indicates that as long as the reductant can remove the NO_x, the phosphorus poisoning is related to poisoning of the reactions under lean conditions. For phosphorous exposure at 300 °C on the other hand, the reduction with propene is poor, which results in low NO_x storage and therefore the effect of phosphorus poisoning is not seen equally clear.



Figure 2. Measured outlet NO_x before and after exposing Pt/Ba/Al₂O₃ to 50 or 100 vol.-ppm H₃PO₄, 8 vol.-% O₂ and 5 vol.-% H₂O for 34 h. The lean phase (400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and 8 vol.-% O₂) starts at time = 0. The rich phase consist of 400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and H₂ or C₃H₆, where (**A**) 300 °C with 1 vol.-% H₂ in the rich phase, (**B**) 300 °C with 0.1 vol.-% C₃H₆ in the rich phase, (**C**) 400 °C with 1 vol.-% H₂ in the rich phase, and (**D**) 400 °C with 0.1 vol.-% C₃H₆ in the rich phase.

Condition	NO _x Storage Capacity (µmol)	Loss in NO_x Storage Capacity (µmol)
NO+H ₂ , 300 °C, fresh	92.4	-
NO+H ₂ , 300 °C, 50 ppm P	56.2	36.2
NO+H ₂ , 300 °C, fresh	88.8	-
NO+H ₂ , 300 °C, 100 ppm P	51.4	37.4
NO+H ₂ , 400 °C, fresh	85.9	-
NO+H ₂ , 400 °C, 50 ppm P	70.6	15.3
NO+H ₂ , 400 °C, fresh	70.3	-
NO+H ₂ , 400 °C, 100 ppm P	51.5	18.8

Table 1. NOx storage capacity for fresh and phosphorous exposed samples.

The resulting N₂O formation during these experiments is shown in Figure 3, and it is observed that the selectivity towards N₂O is higher at low temperatures, which is in accordance with the literature [31]. Most of the N₂O is formed in the first part of the rich period, but also a peak in the beginning of the lean phase is observed. Choi et al. [32] found that the second N₂O peak relates to reactions with NH₃. When shifting from rich to lean phase, the formation of NH₃ ceases due to the replacement of H₂ with O₂ in the feed. Therefore, it is plausible that the second N₂O peak relates to consumption of stored NH₃ to form N₂O.



Figure 3. Measured outlet N₂O before and after exposing Pt/Ba/Al₂O₃ to 50 or 100 vol.-ppm H₃PO₄, 8 vol.-% O₂ and 5 vol.-% H₂O for 34 h. The rich phase (400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and H₂ or C₃H₆) starts at time = 0. The lean phase consist of 400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and 8 vol.-% O₂, where (**A**) 300 °C with 1 vol.-% H₂ in the rich phase, (**B**) 300 °C with 0.1 vol.-% C₃H₆ in the rich phase, (**C**) 400 °C with 1 vol.-% H₂ in the rich phase, and (**D**) 400 °C with 0.1 vol.-% C₃H₆ in the rich phase.

In addition, more N₂O is formed when using propene as the reducing agent compared to hydrogen. The N₂O formation was previously studied by Olsson et al. [30] where a kinetic model for the reduction of NO_x by C_3H_6 was developed. The reason for the N_2O formation is the reaction between the hydrocarbon and NO_x to form N_2O [33]. Furthermore, in the beginning of the lean phase after propene reduction, there is an increase in the N_2O formation. This is due to reaction between NO_x and HC species, stored during the rich phase. In addition, the results clearly show that the formation of N₂O in the rich phase decreases after exposure to phosphoric acid. The decrease in N₂O formation occurs at both 300 and 400 $^{\circ}$ C for both H₂ and C₃H₆ as the reducing agent. When comparing the ratio of (N₂O formed)/(NO_x stored) at 300 $^{\circ}$ C with propene as reductant, the ratio clearly decreases after phosphorous exposure (0.015, 0.0084, 0.020 and 0.011 for first sample before and after exposure for 100 vol.-ppm P, and for second sample before and after exposure for 50 vol.-ppm P, respectively). Thus, it is clear that the N_2O formation not decreases only due that less NO_x is stored, but also due to that the selectivity changes because of the poisoning. De Abreu Goes et al. [11] studied thermal exposure of LNT catalysts, which resulted in sintering of the catalytic particles that in turn resulted in decreased formation of N₂O. Since the catalyst in the present study was not exposed to high temperature again after the degreening step, sintering as a cause for the loss of N₂O formation is not likely. Therefore, the reduced N₂O formation after phosphorous exposure is more likely to relate to poisoning of the platinum sites.

With H_2 as the reducing agent, the formation of NH_3 is higher compared to C_3H_6 as reductant, which can be seen in Figure 4. At 300 °C, the formation of NH_3 is even negligible for the C_3H_6 case. For both samples, the NH_3 formation is slightly higher after the exposure to phosphorus when H_2 is used as the reducing agent. The ammonia formed over the LNT depends on several reactions occurring simultaneously: (i) ammonia production from inlet NO_x ; (ii) ammonia production from stored NO_x ; and (iii) reaction of the formed ammonia with stored NO_x in an SCR process [6]. An increase in the

ammonia production has also been observed for field-aged, oven-aged and sulphur-poisoned LNT catalysts [11,34]. The reason for the increase in ammonia formation after phosphorus exposure could be related to the loss of NO_x storage capacity depicted in Figure 2. Since less NO_x is available in the form of nitrates during the lean phase, less ammonia can in turn react in the SCR reaction with the stored nitrates [6], and thereby the ammonia release is higher. However, less ammonia will also be produced from the stored NO_x , thus this means that the SCR reaction is more influenced by the phosphorous poisoning compared to the ammonia production. The increased ammonia production during rich conditions can be beneficial for the cases where an SCR catalyst is placed downstream of the LNT [8], in the so-called passive SCR technique.



Figure 4. Measured outlet NH₃ concentration before and after exposing Pt/Ba/Al₂O₃ to 50 or 100 vol.-ppm H₃PO₄, 8 vol.-% O₂ and 5 vol.-% H₂O for 34 h. The rich phase (400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and H₂ or C₃H₆) starts at time = 0. The lean phase consist of 400 vol.-ppm NO, 5 vol.-% H₂O, 5 vol.-% CO₂ and 8 vol.-% O₂, where (**A**) 300 °C with 1 vol.-% H₂ in the rich phase, (**B**) 300 °C with 0.1 vol.-% C₃H₆ in the rich phase, (**C**) 400 °C with 1 vol.-% H₂ in the rich phase, and (**D**) 400 °C with 0.1 vol.-% C₃H₆ in the rich phase.

The outlet N₂ concentration was not measured due to IR spectroscopy was used for the gas phase analysis. By assuming that all nitrogen leaving the catalyst are either in form of NO, NO₂, NH₃, N_2O or N_2 , the remaining N_2 can be determined. The quantities of NO_x , NO, NO_2 NH_3 and N_2O were estimated by integration of the data from the FTIR analysis. The results from these calculations are shown in Figure 5 and it is clear that the formed amount of N_2 is quite low. The reason for this is that we intentionally used long lean and rich cycles to come closer to saturation and also more complete regeneration. In a real application, the cycle times would be different, giving a high amount of N2. Moreover, the results clearly show a minor N2O formation in all conditions except with C_3H_6 as reducing agent at 300 °C. Hence, the main difference between the results from the different experimental conditions is the formation of NH_3 , N_2 and release of NO_x . The phosphorous-exposed samples tend to release higher amounts of NOx and NH3 at the cost of formation of the desired product; N2. According to the ICP-AES results shown in Table 2, the amount of accumulated phosphorous is similar for both H_3PO_4 concentrations. Thus the H_3PO_4 concentration does not seem to be such a critical factor for the poisoning at these levels. Therefore, the difference before and after phosphorous exposure between the two cases appears similar. For H₂ as the reducing agent, exposure to phosphorus has a more pronounced effect at 300 than 400 °C. The reason for this could be that since the ammonia

production is higher at lower temperature, and is to a large extent influenced by the phosphorous poisoning, it results in a more pronounced effect on the N₂ selectivity. However, for C_3H_6 as the reducing agent, there is no clear trend regarding the selectivities between the fresh and phosphorous exposed samples at 300 and 400 °C. The reason for this could be that the amounts of NH₃ and N₂O formed in these experiments are low, so most of the species are NO_x and N₂.



Figure 5. Balance over outlet nitrogen atoms over a lean/rich cycle before and after exposing $Pt/Ba/Al_2O_3$ to 50 and 100 vol.-ppm H_3PO_4 for 34 h, assuming that all nitrogen leaving the catalyst is either in form of NO_x, NH₃, N₂O or N₂. (**A**) 300 °C with 1 vol.-% H₂ in the rich phase, (**B**) 300 °C with 0.1 vol.-% C₃H₆ in the rich phase, (**C**) 400 °C with 1 vol.-% H₂ in the rich phase, (**D**) 400 °C with 0.1 vol.-% C₃H₆ in the rich phase.

2.2. Characterization of the Phosphorous Exposed Samples

One channel of the phosphorous-exposed monolith sample "50 ppm, 34 h" was cut out and the surface composition in the inlet, middle, and in the outlet section of the channel was analyzed by XPS. The overall spectra (see Figure 6A), show that the magnitude of the phosphorus peaks is highest from the surface of the inlet section and lowest from the outlet section of the phosphorous exposed sample. The P 2p binding energy region measured with higher resolution for the inlet and outlet sections for the phosphorous exposed sample are shown in Figure 6B. Due to different signal intensities from the two samples, the spectra are normalized to facilitate the interpretation of the peak positions when comparing different spectra with each other. The intensity of the P 2p peak from the middle section of the sample is significantly lower compared to the peak from the inlet section. In addition, the results show that the position of the Sample. Deconvolution of the P 2p peak for the inlet and middle section compared to the inlet section of the sample can be seen in Figure 7. For the inlet section of the sample the deconvolution shows a higher presence of P_4O_{10} with a binding energy ranging between 135.0 and 135.5 eV [35]. However, for the middle region of the sample, the deconvolution shows that metaphosphates, PO_3^- , dominate, which have a binding energy ranging between 134.0 and 134.5 eV [35]. Formation of

metaphosphates has previously been observed after exposing Cu/BEA [36] as well as Fe/BEA [37] for phosphoric acid in gas phase. In addition, formation of P_4O_{10} was also found on Fe/BEA [37], which is in line with our observations. Our results show that the phosphorus-containing surface species mainly are in the form of P_4O_{10} in the inlet section of the sample, where the phosphorus concentration and deposition is higher, while in the middle section more metaphosphates are seen.



Figure 6. (A) Normalized X-ray photoelectron spectroscopy (XPS) spectra from the inlet, middle and outlet section of the $Pt/Ba/Al_2O_3$ sample exposed to 50 vol.-ppm H_3PO_4 for 34 h. (B) Corresponding XPS spectra for the P 2p region from the inlet and middle section of the sample.



Figure 7. Deconvolution of the P 2p XPS spectra from the inlet (**A**) and middle (**B**) section of the Pt/Ba/Al₂O₃ sample exposed to 50 vol.-ppm H₃PO₄ for 34 h. The considered phosphorous species are P_4O_{10} , PO_3^- and PO_4^{3-} .

In order to achieve further insight into the effect of phosphorous exposure in the axial direction of the Pt/Ba/Al₂O₃ catalyst, the specific surface area and pore volume, and elemental composition of the inlet, middle and outlet sections of the samples exposed for 100 and 50 vol.-ppm phosphorous for 34 h were analysed together with a degreened sample. In Table 2, the phosphorus content from ICP-AES analysis is presented along with the specific surface area and pore volume. The results show that phosphorus has a gradient distribution, from the inlet to the outlet section of the two phosphorous-exposed samples, which is in line with other studies of both field-aged catalysts and experiments in engine rigs [13,19-23,38,39]. Moreover, it is found that the effect of the phosphorus concentration is minor, which is in line with the flow reactor experiments. Glisteo et al. [24] impregnated a Pt/Ba/Al₂O₃ catalyst with aqueous solutions of phosphorous of different concentrations. Their results showed that the specific surface area decreased more by higher phosphorous concentration accumulated on the catalyst, which is in line with our results. We find that the phosphorus content is the highest, and the surface area and pore volume have lowest values, for the inlet section of both catalysts exposed to phosphorous. Close to the outlet section, the phosphorus content is much lower for both samples, whereas the specific surface area and pore volume are higher and more similar to the degreened sample.

	P-content, (wt%)	S_{BET} (m ² /(g washcoat)) ¹	V_P (cm ³ /(g washcoat)) ¹
Fresh	-	134	0.46
50 ppm, 34 h, inlet	2.2	98	0.36
50 ppm, 34 h, middle	0.40	130	0.45
50 ppm, 34 h, outlet	0.11	134	0.45
100 ppm, 34 h, inlet	2.0	113	0.35
100 ppm, 34 h, middle	0.38	123	0.39
100 ppm, 34 h, outlet	0.07	132	0.39

Table 2. Phosphorous content, specific surface area and pore volume of the fresh $Pt/Ba/Al_2O_3$ sample and of the $Pt/Ba/Al_2O_3$ samples exposed to 50 and 100 vol.-ppm phosphorous for 34 h. The phosphorus contents are measured by inductive coupled plasma atomic emission spectroscopy (ICP-AES).

¹ Estimated based on washcoat amount.

The results from ESEM and EDX analyses acquired from two different positions; 2 mm from the front and 2 mm from the back of the monolith sample exposed for 50 vol.-ppm phosphorous for 34 h are presented in Figures 8 and 9, respectively. Besides phosphorus, other elements (Al and Ba) that are present in the washcoat are depicted in the images. Moreover, silicon, which is one of the major compounds of the cordierite substrate, is also shown. The analysis 2 mm from the front of the sample (Figure 8) indicates that phosphorus is located towards the surface of the washcoat, which previously has been shown for catalysts exposed to ZDDP doped fuel in engine rigs [13]. Closer to the outlet of the sample, less phosphorus is present and the distribution of the phosphorus is more uniform in the washcoat. Accumulation of phosphorus towards the front of the catalyst has been shown for field-aged catalysts [21]. The ESEM-EDX results in the present study can be related to the XPS results (Figure 7), which showed more P_4O_{10} in the inlet section of the sample, while more metaphosphates in the middle section of the phosphorus are observed with EDX, but it should be noted that phosphorus is still visible, which means that in a catalytic aftertreatment system it is possible that a catalyst placed downstream the NO_x storage catalyst also can be exposed to phosphorus.



Figure 8. Environmental scanning electron microscopy (ESEM) and energy dispersive X-ray (EDX) images from the cross-section 2 mm from the front of the $Pt/Ba/Al_2O_3$ sample exposed to 50 vol.-ppm H_3PO_4 for 34 h.


Figure 9. ESEM and EDX images from the cross-section 2 mm from the back of the $Pt/Ba/Al_2O_3$ sample exposed to 50 vol.-ppm H_3PO_4 for 34 h.

To summarize, the combined XPS, EDX, BET and ESEM-EDX data suggest that during phosphorus exposure in the vapour phase more phosphorus is deposited in the inlet of the monolith sample and this phosphorous is mainly located at the surface of the washcoat and contains a large fraction of P_4O_{10} . While the middle section of the monolith contains less phosphorus and has a higher fraction of metaphosphates. Moreover, closer to the outlet the phosphorus is more evenly distributed.

3. Materials and Methods

3.1. Catalyst Synthesis

The Pt/BaO/Al₂O₃ catalyst was synthesized using incipient wetness impregnation. To enhance the NO_x storage capacity of the chosen model catalyst, the impregnation of platinum on the support material was performed before incorporation of barium, which previously has shown beneficial to achieve high NO_x storage capacity of the catalyst [25]. Prior to the impregnation, the γ -alumina (Sigma-Aldrich) support was calcined for 2 h at 600 °C. Thereafter the γ -alumina support was impregnated with a platinum nitrate precursor (Heraeus platinum (II) nitrate; solution type K, batch: 12207), diluted with Milli-Q water (18 MΩ·cm) of the estimated volume targeting 2 wt.-% Pt. The impregnated support material was dried in air at 100 °C for 2 h and subsequently calcined in air at 550 °C for 2 h, starting from room temperature with a temperature increase of 5 °C/min. The calcined Pt/Al₂O₃ sample was then impregnated with a barium precursor (Sigma-Aldrich barium acetate, 99% A.C.S reagent, batch: 11415KA), dissolved in Milli-Q water in two steps also by incipient wetness impregnation, targeting a total of 16 wt.-% of BaO. The catalytic material was dried and calcined after each impregnation step under same conditions as for the impregnation of platinum.

3.2. Monolith Preparation

The catalytic material was coated on honeycomb-shaped ceramic monolith substrates (20 mm in length, 21 mm in diameter, a wall thickness of 0.106 mm and with 400 cpsi). To improve the adhesiveness, 10 wt.-% boehmite (Sasol disperal P2, product code: 538116) was mixed with the catalyst in a solution of 50 vol.-% Milli-Q water and 50 vol.-% ethanol. The monolith substrates were stepwise

dipped in the solution and dried at 90 °C until the target washcoat weight of 700 mg per monolith sample was reached. Thereafter, the catalysts were calcined at 550 °C for 2 h, starting from room temperature with a temperature increase of 5 °C/min.

3.3. Flow-Reactor Experiments

The flow-reactor experiments were carried out in a synthetic gas bench (SGB) reactor. The coated monolith samples were placed in a quartz tube, which was 750 mm in length and with an inner diameter of 22 mm. To prevent by-pass, quartz wool was wrapped around the monoliths. The gas flow and gas concentrations were regulated using mass flow controllers from Bronkhorst and water vapor was dosed using a Bronkhorst Controlled evaporation system (CEM) system. The quartz tube was placed in an electric heating coil and covered with insulation material, and the temperatures were measured by two thermocouples, one measuring and regulating the gas temperature before the catalyst and one measuring the catalyst temperature in a center channel of the monolith. An MKS Multigas 2030 FTIR spectrometer provided by MKS instruments, Andover, MA, USA was used to monitor the gas phase concentrations of NO, NO₂, N₂O, NH₃ and H₂O, and the total flow was 3500 mL/min, corresponding to 30,300 h⁻¹ space velocity, with argon as carrier gas for all measurements. All lines and connections prior and after the reactor were heated to 200 °C, to prevent condensation.

To avoid sintering of platinum particles under the experiments, the coated monoliths were first H₂-treated in 1 vol.-% H₂, 5 vol.-% CO₂ and 5 vol.-% H₂O for 20 min at 500 °C and thereafter degreened at 600 °C, according to the following procedure. First the catalyst was exposed to the rich phase (400 vol.-ppm NO, 5 vol.-% CO₂, 1 vol.-% H₂ and 5 vol.-% H₂O with argon as carrier gas) for 60 min, followed by exposure to the lean phase (400 vol.-ppm NO, 5 vol.-% CO₂, 8 vol.-% O₂, and 5 vol.-% H₂O in argon) for 15 min. Subsequently, the catalyst was reduced in 1 vol.-% H₂, 5 vol.-% CO₂ and 5 vol.-% H₂O in Ar for 20 min at 600 °C.

The NO_x storage and reduction performance of the catalysts was studied at 300 and 400 °C using lean/rich cycling. At each temperature, two cycle segments consisting of five cycles in each segment were conducted. Cycle number 4 for all sequences is shown in the Figures. The gas composition of the lean phase was 400 vol.-ppm NO, 5 vol.-% CO₂, 8 vol.-% O₂, and 5 vol.-% H₂O. In the first lean/rich phase cycle segment, hydrogen was used as the reductant (1 vol.-% H₂, 400 vol.-ppm NO, 5 vol.-% CO₂ and 5 vol.-% H₂O) and in the second cycle segments propene was used as the reducing agent (0.1 vol.-% C₃H₆, 400 vol.-ppm NO, 5 vol.-% CO₂ and 5 vol.-% H₂O). For all cycles, the duration of the lean and rich phase were 4 and 1 min, respectively. Prior to the experiment at each temperature, the catalyst was pre-treated with 1 vol.-% H₂, 5 vol.-% CO₂ and 5 vol.-% H₂O for 15 min at 400 °C. This pre-treatment was also conducted for the phosphorus-exposed catalysts.

3.4. Phosphorous Exposure

The NO_x storage capacity of the coated monoliths was measured in the flow reactor before and after the exposure to phosphorus, using the procedure described in the previous section. The same reactor as described above was used for the phosphorous exposure, however using a separate reactor tube. In order to reduce surfaces for the evaporated phosphoric acid to stick to, the monolith was not wrapped in quartz wool during the phosphorous exposure. Phosphoric acid was introduced to the gaseous flow by a syringe pump connected to a Teflon pipe disposing the acid inside the quarts tube. The composition of the gas phase during the phosphorous exposure was 50 or 100 vol.-ppm evaporated phosphoric acid, 5 vol.-% H₂O, 8 vol.-% O₂ in argon as carrier gas, and the temperature was 200 °C. Two different phosphorous exposure conditions were studied in this work; in the first case the catalyst was exposed to 50 vol.-ppm phosphoric acid for 34 h, and in the second case the sample was exposed to 100 vol.-ppm for phosphoric acid for 34 h. This corresponds to an exposure of 1.7 and 3.4 g P/l catalyst, respectively. Note that high amounts of phosphorus passed through the catalyst without adsorption. These samples are denoted "50 ppm, 34 h" and "100 ppm, 34 h", respectively. Note that new monolith samples were used for each poisoning condition.

3.5. XPS

To evaluate the oxidation state of phosphorus accumulated on the catalytic washcoat, an entire channel from the coated monolith was placed on a holder to measure the surface species of the washcoat. The use of a whole channel also enabled the identification of the axial position where measurements took place. The instrument used for the XPS analysis was a Perkin Elmer PHI 5000 ESCA system provided by PerkinElmer, Waltham, MA, USA equipped with an EDS elemental mapping system. The X-ray source for the XPS measurements was monochromatic Al K α radiation at 1486.6 eV. Correction for charging was performed by normalizing the spectra using the C 1s peak at 284.6 eV [40] as reference. Furthermore, since normalizing will cause the largest peak being equal to 1, all spectra were divided by 1.1 for better visualization. Deconvolution of the P 2p peak from each sample was performed by fitting a Gaussian function to the experimental data, with subtraction of a linear baseline under both peaks. The peak positions were optimized according to the same procedure for both samples to achieve the lowest standard deviation and found to be 134.25, 135.5 and 132.5 eV for PO₃⁻, P₄O₁₀ and PO₄³⁻, respectively.

3.6. N₂-Physisorption and ICP-AES

The specific surface area and pore volume of the samples were measured by N₂-physisorption isotherms at -195 °C and were collected using a TriStar 3000 gas adsorption analyzer. The inlet-, middle- and outlet section of the coated monoliths were crushed and grinded to powder form. Approximately 300 mg sample from each section was thermally dried at 90 °C in N₂ gas flow for 4 h and used for specific surface area and pore-volume measurements.

Powder samples from each section were used for elemental analysis by inductive coupled plasma atomic emission spectroscopy (ICP-AES), and the measurements were performed by ALS Scandinavia AB, Luleå, Sweden. The concentration of the measured phosphorus is reported as weight percentages.

3.7. ESEM

Environmental scanning electron microscopy (ESEM) was used to acquire an overview of the location of different elements over the catalytic washcoat. The instrument used to gain these images was a Quanta200 ESEM FEG from FEI, Hillsboro, OR, USA. An energy dispersive X-ray (EDX) system from Oxford Inca was equipped to the instrument to map the elements over scanned areas.

4. Conclusions

In the present study, a Pt/Ba/Al₂O₃ NO_x storage catalyst coated on a ceramic monolith substrate was exposed to two different phosphorus containing atmospheres for 34 h at 200 °C; 50 vol.-ppm H₃PO₄ + 8 vol.-% O₂ + 5 vol.-% H₂O and 100 vol.-ppm H₃PO₄ + 8 vol.-% O₂ + 5 vol.-% H₂O. The catalysts were characterized by N₂-physisorption, XPS, ESEM and ICP-AES, and the NO_x storage capacity was measured in flow reactor experiments.

The specific surface area and pore volume measurements, together with the elemental analysis by ICP-AES, showed a clear gradient distribution of phosphorous along the axial direction of the monolith. The phosphorus concentrations close to the inlet section of the sample were significantly higher compared to the middle and outlet section. The specific surface area and pore volume were found to be lower in the sections where the concentration of phosphorus was higher. At the outlet section, both the specific surface area and the pore volume were found to be comparable with those for the fresh sample, and the accumulated amount of phosphorus was only about 0.1 wt.-%, which can be compared with 2 wt.-% at the inlet section of the phosphorous concentrations, which indicates that the phosphorus concentration is not limiting the deposition of phosphorous on the catalyst, due to long exposure time.

Flow-reactor experiments showed that the NO_x storage capacity of Pt/BaO/Al₂O₃ decreased with presence of phosphorous species on the catalyst. The results also showed that the formation of N₂ decreased after the exposure to phosphorous, in favour of ammonia production. Furthermore, the formation of N₂O decreased after the phosphorus exposure. The increased ammonia production can be explained by the fact that the outlet ammonia concentration is dependent on three reactions over the catalyst: (i) the formation of ammonia from NO in the inlet gas; (ii) formation of ammonia from stored NO_x; and (iii) the consumption of ammonia through the selective catalytic reduction by the reaction with the stored nitrates. Since the amount of stored nitrates is lower for the phosphorus exposed sample, it means that less ammonia is produced in reaction (ii), but simultaneously less ammonia is consumed in reaction (iii). Since there is more ammonia in total, it means that the SCR reaction is more influenced by the phosphorus poisoning than the ammonia formation from the stored nitrates.

The XPS analysis showed that phosphorus exposure resulted in the formation of different species depending on the axial location of the washcoat. At the inlet section of the catalyst, where most of the phosphorus was accumulated, the dominating phosphorous species was found to be P_4O_{10} while further along the axial direction metaphosphates (PO_3^-) was the dominating species. From the ESEM and EDX analysis it could be concluded that most phosphorus, close to the inlet of the catalyst, was located at the washcoat surface. At the outlet, only a small amount of phosphorus could be detected by SEM-EDX, however, it was more diffused into the washcoat. To summarize, the results from XPS and SEM indicate that the accumulation of phosphorus differs between the front and the back of the monolith catalyst.

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Article "PdO vs. PtO"—The Influence of PGM Oxide Promotion of Co₃O₄ Spinel on Direct NO Decomposition Activity

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Abstract: Direct decomposition of NO into N₂ and O₂ (2NO \rightarrow N₂ + O₂) is recognized as the "ideal" reaction for NOx removal because it needs no reductant. It was reported that the spinel Co₃O₄ is one of the most active single-element oxide catalysts for NO decomposition at higher reaction temperatures, however, activity remains low below 650 °C. The present study aims to investigate new promoters for Co₃O₄, specifically PdO vs. PtO. Interestingly, the PdO promoter effect on Co₃O₄ was much greater than the PtO effect, yielding a 4 times higher activity for direct NO decomposition at 650 °C. Also, Co₃O₄ catalysts with the PdO promoter exhibit higher selectivity to N₂ compared to PtO/Co_3O_4 catalysts. Several characterization measurements, including X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), H_2 -temperature programmed reduction (H_2 -TPR), and in situ FT-IR, were performed to understand the effect of PdO vs. PtO on the properties of Co_3O_4 . Structural and surface analysis measurements show that impregnation of PdO on Co₃O₄ leads to a greater ease of reduction of the catalysts and an increased thermal stability of surface adsorbed NOx species, which contribute to promotion of direct NO decomposition activity. In contrast, rather than remaining solely as a surface species, PtO enters the Co_3O_4 structure, and it promotes neither redox properties nor NO adsorption properties of Co₃O₄, resulting in a diminished promotional effect compared to PdO.

Keywords: direct NO decomposition; PGM oxide promotion; PdO vs. PtO; in-situ FT-IR; NO adsorption properties; redox properties

1. Introduction

Nitrogen oxides (NOx) formed by combustion from fixed and mobile sources cause severe detrimental environmental problems, such as acid rain and photochemical smog [1–3]. Effectively controlling the emission of NOx is the topic of much research and has led to the introduction of many new catalyst technologies, such as three-way catalysts (TWC), NOx storage-reduction (NSR), and selective catalytic reduction (SCR) for NOx gas removal from mobile sources, and SCR and selective non-catalytic reduction (SNCR) for NOx gas removal from fixed sources [4–6]. Among various deNOx strategies, direct decomposition of NO (NO \rightarrow 1/2O₂ + 1/2N₂) has been considered to be the most desirable method because this reaction is thermodynamically favorable at low temperatures and does not need any reductants, such as NH₃, H₂, CO, or hydrocarbons. However, kinetic studies have indicated that the reaction needs to overcome a large activation energy (~335 kJ mol⁻¹) barrier [4–15]. Accordingly, there is an apparent need for a suitable catalyst to decompose NOx at a given temperature, and therefore, significant research has been undertaken towards development of active and stable catalysts.

Since the pioneering work of Jellinek on the catalytic decomposition of NO in 1906, much research has been reported on NO direct decomposition over several materials, including perovskites, rare earth

oxides, and Cu-zeolites [2–7]. Numerous metal oxides have also been examined as candidates for NO decomposition catalysts [16] and Co_3O_4 is often recognized as a significant component in many active catalysts at higher reaction temperatures [17–21]. However, Haneda et al. recently reported that NO decomposition takes place slowly, if at all, over pure Co_3O_4 at temperatures below 650 °C [18]. They reported that the presence of small amounts of alkali metals were essential to activate NO decomposition over Co_3O_4 oxide by enhancing NO adsorption [18–20]. This interesting effect of alkali metals, particularly Na, was also reported by Kung et al. [21], but dependence on alkali metals is not feasible for practical applications due to their volatile nature at temperatures above 600 °C.

Metal oxide supported platinum group metals (PGM metals) were also one of the earliest types of NO decomposition catalysts studied, and the results have been widely reported; mainly Au, Pt, Pd, and Ir at temperatures higher than 700 °C [22,23]. Suzuki et al. [24] synthesized a porous CaZrO₃/MgO/Pt composite and found that this catalyst could obtain a NO conversion rate of about 52% at 900 °C in the absence of O₂. Haneda et al. [25] found that the addition of Pt improved the direct NO decomposition performance of rare earth oxides. They [26,27] also compared the activity of [Pd(NH₃)₄] (NO₃)₂, Pd(NO₃)₂, Pd(CH₃COO)₂, and (NH₄)₂-[PdCl₄] as palladium precursors for NO decomposition in a Pd/Al₂O₃ catalyst at 700 °C, and the activity was found to decrease in the order of $Pd(NO_3)_2 > [Pd(NH_3)_4] (NO_3)_2 > Pd(CH_3COO)_2 >> (NH_4)_2[PdCl_4]. Almusaiteer et al. [28] reported and the second second$ that compared to Pd/Al₂O₃, the Pd/C (activated carbon) catalyst was found to be more beneficial for O₂ desorption, but both have similar activity. Oliveira et al. [29] investigated the catalytic performance of palladium and copper catalysts loaded in mordenite (MOR) and found that these catalysts were more active for NO decomposition than alumina supported catalysts. However, the reports on supported PtO catalysts for direct NO decomposition at temperatures below 700 °C are very limited in the literature. Similarly, metallic Pd catalysts always deactivate over time due to oxidation of Pd metal to PdO at temperatures below 650 °C [30].

To the best of our knowledge, Co_3O_4 supported PGM catalysts have never been explored for direct NO decomposition, likely due to the inactivity of the individual components at lower temperatures. However, the need for enhanced NO adsorption on Co_3O_4 suggests that the addition of PGM promotion can lead to increased low temperature activity. Hence, the present study aims to investigate the promotional effect of PdO vs. PtO on the Co_3O_4 for direct NO decomposition. The activity measurements show that the optimum PdO/ Co_3O_4 catalyst exhibits 4 times higher activity than PtO/ Co_3O_4 catalysts. Several characterization techniques such as X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), H₂-Temperature programmed reduction (H₂-TPR), and *in situ* FT-IR are employed to understand the influence of PdO and PtO on the structural and surface properties of Co_3O_4 .

2. Results and Discussion

2.1. Direct NO Decomposition Activity Measurements

To qualify as direct NO decomposition, the catalyst must decompose NO into just two products: N_2 and O_2 . The possibility for unwanted N_2O and NO_2 formation as side products cannot be neglected during the analysis of this reaction. Therefore, high NO conversion is desired, but not sufficient; it is also very important to maximize selectivity towards N_2 rather than N_2O or NO_2 . Considering all possible products, the reaction can be written as:

$$2 \text{ NO} \rightarrow \text{N}_2 + \text{O}_2$$

 $4 \text{ NO} \rightarrow 2\text{N}_2\text{O} + \text{O}_2$
NO + [O] \rightarrow NO₂ ([O]: catalyst lattice oxygen)

The selectivity to N₂ can be defined as:

$$N_2$$
 selectivity (%) = 2 × $[N_2]/(2 × [N_2] + 2[N_2O] + [NO_2])$

Also, when NO dissociates on the catalyst surface, it is possible for N_2 production to occur, but without simultaneous release of the stoichiometric amount of the O_2 product. This situation can occur quite frequently and is the topic of a previous study from some of us [30]. Rather than desorb as the O_2 product, oxygen atoms either remain strongly adsorbed on the surface or chemically react with catalyst material, oxidizing both surface and bulk, and changing the catalyst composition. In either case, the catalyst deactivates over time. Hence, it is very important to confirm that the catalyst releases both N_2 and O_2 as products.

Direct NO decomposition measurements were performed over the pure spinel oxide Co_3O_4 and over the PdO- and PtO-promoted Co₃O₄ catalysts, denoted PdO/Co₃O₄ and PtO/Co₃O₄, respectively. Activity was measured for 2 h at 400, 450, 550, and 650 °C. The direct NO decomposition activity to N_2 of PdO/Co₃O₄ and PtO/Co₃O₄ with varying PGM loading is presented in Figure 1 as a function of temperature and compared to the pure Co_3O_4 spinel. All numeric values of NO conversion and $N_{2,\ell}$ N₂O, NO₂ ppm product concentrations of the Co₃O₄, PdO/Co₃O₄, and PtO/Co₃O₄ catalysts at various reaction temperatures are presented in Tables S1 and S2. The raw NO conversion profiles (NO converted to all the products) of Co_3O_4 , $3PdO/Co_3O_4$, and $4PtO/Co_3O_4$ during the steady state direct NO decomposition obtained from FT-IR detector in the temperature region 400 to 650 °C are presented in Figure S1. High values for NO conversion at lower temperatures may seem counterintuitive, however, most of the NO conversion at low temperature is simply due to thermodynamically favorable NO oxidation to NO₂. For example, as shown in Table S1, the NO conversion values of Co₃O₄, 3 PdO/Co₃O₄, are 3 PtO/Co₃O₄ are 3.15, 2.92, and 7.66, respectively, giving the impression that 3 PtO/Co₃O₄ is the most active catalyst at 400 °C. However, for 3 PtO/Co₃O₄, the NO conversion specifically to N₂ is lower than that of 3 PdO/Co_3O_4 because most of the NO conversion is due to NO oxidation. Hence, for NO decomposition to N_2 , it is more instructive to consider NO converted into N_2 rather than total NO conversion. The activity values were calculated in this manner and are presented in Figure 1 in units of micromoles of NO converted to N₂ per gram per second.

NO and Ar partial pressure values obtained by mass spectrometry (MS) during the steady state measurements are presented in Figure S2 and compared with data obtained in the absence of the catalyst, which serves as a baseline. Inert Ar gas was introduced as a tracer to monitor for potential systematic variation in signal intensity during the experiment. As shown in Figure S2, no change in Ar signal intensity was observed during the experiment, however, the intensities of NO signal changed based on catalyst identity and reaction temperature. These measurements confirm the change in the NO signal is due to catalytic conversion of NO and not due to artifacts. Also, the MS signal for NO (m/z = 30) tracks with the conversion reported by the FT-IR measurements (Figure S1). Figures S3–S5 present the MS partial pressure values of the N2, N2O, and NO2 products, respectively, during direct NO decomposition and are also compared to the MS partial pressures obtained in the absence of the catalyst. As shown in Figure S3, the N2 and O2 signal intensities are higher compared to the background signals, which confirms the simultaneous release of N_2 and O_2 as expected for NO decomposition. Furthermore, the MS intensity of the N₂ signal qualitatively tracks with the N₂ concentrations calculated by nitrogen mass balance from the FTIR measurements (Tables S1 and S2), lending additional confidence in the activity results. Similarly, good correlation between the FT-IR detection of N_2O and NO_2 and the MS signals was observed (compare Tables S1 and S2 to Figures S3 and S4). Finally, the release of oxygen as a product (Figure S3) and stable NO conversion (Figures S1 and S2) during the steady state measurements suggest that the catalysts were not poisoned by the irreversible chemical adsorption of oxygen. Thus, the good correlation between the FT-IR data and MS signals suggests that the calculation of N_2 production from the FT-IR is reliable and can be used for the calculation of activity from NO conversion to N2. As shown in Figure 1b, NO decomposition activity to N_2 increases slightly with temperature up to 550 °C for the pure Co_3O_4 catalyst. Further increase in the temperature to 650 $^{\circ}$ C results in decreased activity. This result suggests that the Co₃O₄ spinel is not a good catalyst for NO decomposition at temperatures below 650 °C, as indicated by Hamada et al. [21]. The addition of PdO and PtO to the Co_3O_4 spinel improves the direct NOx decomposition activity of

 Co_3O_4 . Direct NOx activity increases with temperature for all palladium and platinum loadings, but unlike the pure Co_3O_4 , deactivation was not observed above 550 °C for the any of PdO or PtO catalysts.

The direct NO decomposition activity of PdO/Co₃O₄ and PtO/Co₃O₄ to N₂ at various temperatures is presented as a function of the weight percent loading of the respective PGMs in Figure 2. For PdO/Co₃O₄, activity increases with palladium loading up to 3 wt%, but further increase in the palladium loading leads to decreased activity (Figure 2a). For PtO/Co₃O₄, activity increases with PtO loading up to 4 wt%, however, the overall effect on activity is significantly diminished compared to PdO/Co₃O₄ (Figure 2b). The optimum loading of each PGM was found to be 3 wt% PdO/Co₃O₄ and 4wt% PtO/Co₃O₄. Interestingly, PdO-promoted catalysts exhibit higher activity than the PtO-promoted catalysts. At 650 °C, the optimum PdO/Co₃O₄ catalyst exhibits 4 times higher activity compared to the optimum PtO/Co₃O₄ catalyst.

To confirm the reaction is indeed direct NO decomposition to N₂ rather than the unwanted production of N₂O or NO₂, the selectivity to N₂ was calculated. The N₂ selectivity is presented as a function of PGM loading from 400–650 °C for PdO/Co₃O₄ and PtO/Co₃O₄ in Figure 3a,b, respectively. As expected, pure Co₃O₄ (0 wt% PGM loading) exhibited relatively low selectivity to N₂ (\leq 20%) at 400 and 450 °C. The N₂ selectivity increased to 80% at 550 °C and to 100% at 650 °C. Formation of N₂O was not observed and only N₂ and NO₂ products are detected during direct NO decomposition over the PtO- and PdO-promoted Co₃O₄ catalysts (Tables S1 and S2). Thus, the product distribution measurements suggest NO oxidation (NO₂ formation) is more favorable at lower reaction temperatures (\leq 450 °C), and at higher reaction temperatures, NO decomposition (N₂ formation) is predominant. Remarkably, the addition of 1 wt% PdO to the Co₃O₄ improves the N₂ selectivity from 1 to 40% at 450 °C and from 21 to 70% at 450 °C. The N₂ selectivity. When also considering the activity measurement in Figure 2, the N₂ selectivity measurements confirm 3 wt% PdO as the optimum loading on Co₃O₄ for direct NO decomposition.

Regarding N₂ selectivity as a function of PtO loading (Figure 3b), the addition of 1 wt% PtO yielded less improvement at 400 °C (1 to 16%) compared to the addition of 1 wt% PdO (1 to 40%). Furthermore, the N₂ selectivity decreases with increasing PtO loading, dropping to 9% for the 4wt% PtO/Co₃O₄ catalyst. Moreover, there is no improvement in the selectivity observed at reaction temperatures at and above 450 °C. Therefore, the direct NO decomposition measurements show that the addition of PdO to Co₃O₄ improves the decomposition activity and N₂ selectivity and 3wt% is the optimum Pd loading over Co₃O₄, whereas PtO loaded on Co₃O₄ leads to only slight improvement in the activity and almost no influence on the N₂ selectivity.



Figure 1. Direct NO decomposition activity to N₂ as a function of reaction temperature for (**a**) PdO- and (**b**) PtO-promoted Co₃O₄ catalysts with varying PGM loading. The pure Co₃O₄ support is included for comparison. (Gas hourly space velocity (GHSV) = 2100 h^{-1} , 1% NO/He).



Figure 2. Direct NO decomposition activity to N₂ as a function of Pd and Pt loading for (**a**) PdO- and (**b**) PtO-promoted Co₃O₄ catalysts at varying temperature. The pure Co₃O₄ support is included for comparison. (GHSV = $2100 h^{-1}$, 1% NO/He).



Figure 3. N₂ selectivity as a function of PdO and PtO loading for (**a**) PdO- and (**b**) PtO-promoted Co_3O_4 catalysts at varying temperature. The pure Co_3O_4 support is included for comparison. (GHSV = $2100 h^{-1}$, 1% NO/He).

2.2. Catalyst Characterization

2.2.1. Structural and Textural Properties

These catalysts have been evaluated using several characterization techniques, like XRD, XPS, H₂-TPR, BET surface area, and *in-situ* FT-IR during NO adsorption, to understand the influence of PdO and PtO on the structural and surface properties of Co_3O_4 and to explain the greater promoter effect of PdO on Co_3O_4 compared to PtO during NO decomposition. The palladium and platinum loadings of the studied catalysts prepared by impregnation were verified with XRF spectrometry. For the nominal 1.0, 2.0, 3.0, and 4.0 wt% of PdO, the experimental values were 0.83, 1.94, 2.80, and 4.15, respectively (Table 1). The experimental values for PtO doped Co_3O_4 catalysts were 0.93, 2.12, 3.23, and 4.02, respectively. Differences may be due to surface heterogeneity or incomplete precursor dispersion during the impregnation procedure or the inherent uncertainty related to the employed XRF method, which did not utilize a standard material to aid the data analysis.

The BET surface area values of PdO/Co₃O₄ and PtO/Co₃O₄ catalysts and the pure Co₃O₄ are presented in Table 1. The pure Co₃O₄ catalyst exhibits a BET surface area of 36 m²/g. Little change in the surface area is observed after impregnating Co₃O₄ with 1, 2, and 3 wt% Pd. However, increasing Pd loading from 3 to 4 wt% on to Co₃O₄ lead to a decrease in the surface area from 33 to 26 m²/g. These values suggest that PdO dispersed very well on the surface of Co₃O₄ until 3wt% and further

increase in the loading to 4wt% likely leads to surface agglomeration, which can block access to the active surface. The BET surface area measurements are corroborated by the activity measurements, which showed that the activity of Co_3O_4 increases with increasing palladium doping only until 3 wt%. Further increase in the Pd loading to 4 wt% lead to a decrease in the activity. Little change in surface area was observed for the Co_3O_4 with platinum impregnation (Table 1), as only a slight decrease in the surface area was observed at the highest loading. As suggested above, the blockage of the active surface may be the cause of the decrease in the activity for the 4 wt% PdO/ Co_3O_4 catalyst. It is hypothesized that the formation of PdO crystallites is responsible for this behavior, and this hypothesis will be investigated below.

Catalyst Loading -	PGM Loading (wt%) *		BET Surface Area (m ² /g)		
	Pd wt%	Pt wt%	PdO/Co ₃ O ₄	PtO/Co ₃ O ₄	
0	-	-	36	36	
1	0.83	0.93	36	35	
2	1.94	2.12	35	39	
3	2.8	3.23	33	34	
4	4.15	4.02	26	33	

Table 1. Co/M (M = Pd, Pt) and BET surface area values of PdO/Co₃O₄ and PtO/Co₃O₄ catalysts.

* As measured by XRF.

The X-ray diffraction (XRD) patterns of the fresh PdO/Co₃O₄ and PtO/Co₃O₄ catalysts are shown in Figure 4a,b, with the pattern of the pure Co_3O_4 for reference. The X-ray diffraction lines characteristic of the cubic cobalt spinel structure were indexed within the Fd3m space group (JCPDS card no. 01-080-1533) in the case of the pure Co_3O_4 catalyst [31]. As shown in Figure 4a, all the PdO/Co_3O_4 catalysts exhibit peaks due to Co_3O_4 . The diffractograms provide evidence that the spinel structure was preserved after Pd impregnation, revealing no observable structural changes compared to the pure Co₃O₄ carrier. Diffraction peaks related to Pd or PdO were not detected on samples up to a nominal Pd loading of 3 wt%. For the 4 wt% Pd sample, a low-intensity diffraction peak indicative of tetragonal PdO (0 0 2) (JCPDS card no. 75-584) was visible at 2θ of 33.6° [32]. These measurements suggest that PdO is well-dispersed on the Co_3O_4 support up to 3 wt%, and above this loading, crystalline PdO forms on Co₃O₄ surface. As stated above, the BET surface decreases from 33 to 26 m^2/g with increasing Pd loading from 3 wt% to 4 wt%. XRD measurements confirm that the decrease in the surface area is due to the formation of crystalline PdO on the surface of Co_3O_4 and blocking of the active surface. The X-ray diffraction patterns of PtO/Co₃O₄ catalysts are shown in Figure 4b and only peaks due to Co_3O_4 are present at all Pt loadings, suggesting the absence of bulk metallic Pt or PtO with long-range order. However, unlike PdO/Co_3O_4 , the XRD peaks of the PtO/Co_3O_4 samples were shifted to higher values relative to the pure Co_3O_4 spinel for all Pt loadings. The shift in the peak position to higher values indicates that Pt is likely incorporating into the Co_3O_4 spinel structure in contrast to PdO/Co_3O_4 , where the absence of the peak shift indicates that Pd remained dispersed on the Co₃O₄ surface.

Figure 5a,b display X-ray diffraction patterns of the spent PdO/Co₃O₄ and PtO/Co₃O₄ catalysts after direct NO decomposition. As shown in Figure 5a, the spent Co₃O₄ exhibits only peaks due to spinel structure. There are no peaks due to either CoO or metallic Co, suggesting the Co₃O₄ spinel is structurally stable during direct NO decomposition. In addition to the Co₃O₄ spinel peaks, the spent PdO/Co₃O₄ catalysts with 1, 2, and 3 wt% Pd loading also exhibit peaks at 20 values of 40.3, 46.79, and 68.4°. These peaks are due to the (1 1 1), (2 0 0), and (2 2 0) facets of Pd metal (JCPDS no: 46-1043). The XRD measurements of the spent catalysts show that the dispersed PdO reduced to Pd metal during direct NO decomposition. Similar to the XRD pattern for the fresh 4 wt% PdO/Co₃O₄, the spent XRD pattern also exhibits peaks due to PdO along with the Pd metal and Co₃O₄ peak, which suggests that the crystalline PdO remains even after direct NO decomposition. The X-ray diffraction patterns of the spent PtO/Co₃O₄ catalysts after direct NO decomposition are displayed in Figure 5b. The spent

 PtO/Co_3O_4 catalysts exhibit peaks due only to Co_3O_4 after direct NO decomposition. In contrast to the metallic phase observed in the spent PdO/Co_3O_4 catalysts, there are no peaks due to metallic Pt observed in the spent PtO/Co_3O_4 catalysts.

The XRD measurements show that in the case of PdO/Co₃O₄ catalysts, the PdO reduced to metallic Pd during direct NO decomposition and promotes the activity of Co₃O₄ catalysts. On the other hand, no metallic Pt formation occurred in the PtO/Co₃O₄ catalysts, leading to a greatly diminished promoter effect compared to the PdO/Co₃O₄ catalysts. Also, the NO decomposition measurements show that the catalytic activity decreases with increasing Pd loading from 3 wt% to 4 wt%. The formation of crystalline PdO and decrease in the surface area (blocking of the active surface) explains the lower activity of 4 wt% sample compared to 3 wt% sample. Hence, XRD and BET surface area measurements corroborate with the activity measurements. The spent PtO/Co₃O₄ catalysts also exhibit a shift in the peak positions to higher 2 θ values compared to the pure Co₃O₄ catalyst, which suggests the incorporation of Pt into the Co₃O₄ spinel structure even after direct NO decomposition.



Figure 4. X-ray diffraction patterns of fresh (**a**) PdO- and (**b**) PtO-promoted Co_3O_4 catalysts. The pattern for the fresh pure Co_3O_4 support is included for reference.



Figure 5. X-ray diffraction patterns of spent (a) PdO- and (b) PtO-promoted Co_3O_4 catalysts. The pattern for the spent pure Co_3O_4 support is included for reference.

2.2.2. Redox Properties

The influence of PdO and PtO on the redox properties of Co_3O_4 are investigated using H₂-temperature programmed reduction (H₂-TPR) measurements. The H₂-TPR profiles of PdO/Co₃O₄ and PtO/Co₃O₄ catalysts are presented in Figure 6a,b, along with that of the pure Co_3O_4 for comparison. Several authors reported that the reduction behavior of Co_3O_4 is strongly dependent on the preparation method, catalyst composition, and dispersion on a support [33,34]. The reduction behavior of Co_3O_4 was widely accepted as a stepwise process, including the reduction of Co^{3+} to Co^{2+} and Co^{2+} to metallic Co. There are three well-defined reduction peaks in the TPR profile of Co_3O_4 (Figure 6). The peak at 235 °C is attributed to the reduction of surface oxygen species. The other two peaks are for the stepwise reduction of Co_3O_4 to metallic cobalt. According to the literature, the second reduction peak centered at 275 °C is due to the reduction of Co_3O_4 to CoO, and the third peak at the region of 305 °C is due to the reduction of CoO to metallic cobalt [33,34].

$$Co_{3}O_{4} + H_{2} \rightarrow 3CoO + H_{2}O$$
$$CoO + H_{2} \rightarrow Co + H_{2}O$$

The addition of 1 wt% PdO to the Co_3O_4 leads to a drastic change in the redox profile of Co_3O_4 (Figure 6). No peaks were observed in the 250–310 °C temperature region. Both PdO and Co₃O₄ were reduced at much lower temperature and all reduction events completed below 150 °C. These measurements show that PdO promotes the reduction of Co₃O₄. Two reduction peaks were observed in the 1 wt% PdO/Co₃O₄ H₂-TPR profile at 79 and 104 $^{\circ}$ C. The first reduction peak at 79 $^{\circ}\text{C}$ is due to the reduction of PdO to metallic Pd and the second reduction peak is due to the reduction of Co₃O₄. The H₂-TPR profiles for 1, 2, and 3 wt% PdO/Co₃O₄ were all very similar (Figure 6). The promotion of Co_3O_4 reduction by Pd observed in H₂-TPR is possibly ascribed to hydrogen spillover and the synergistic effect between Pd species and Co₃O₄. The synergistic effect can weaken the Co-O bond. Chen et al. [35] also reported a similar promotional effect for PdO impregnated on Co₃O₄ catalysts with different morphologies, and the synergistic effect between Pd and Co existed, regardless of Co₃O₄ morphology. In the present study, the intensity of the first reduction peak increases with increasing PdO loading from 1 to 3 wt%, and the increase is accompanied by a slight shift in the peak temperature from 79 to 85 °C. This may be due to the increase in the PdO loading on the Co₃O₄ surface. The reduction profile of 4 wt% PdO/Co₃O₄ is slightly different from the PdO promoted catalysts of lower loading. Along with the peaks due to Co₃O₄ and PdO, a small additional peak is observed at 220 °C. Given the identification of crystalline PdO in the XRD pattern of the 4 wt% PdO/Co_3O_4 , it is reasonable to assign this peak to reduction of crystalline PdO.

The H₂-TPR profiles of PtO/Co₃O₄ catalysts are presented in Figure 6b. Two types of reduction features were observed in the case of PtO promoted Co₃O₄ catalysts, one from 130 to 190 °C and another from 200 to 325 °C. The first feature corresponds to reduction of PtO to metallic Pt and the second is reduction of Co₃O₄. Unlike PdO/Co₃O₄, little to no shift in the Co₃O₄ reduction temperature of PtO/Co₃O₄ catalysts was observed relative to the pure Co₃O₄. The reduction of Co³⁺ to Co²⁺ occurred in the 260–275 °C temperature region for Co₃O₄ and PtO/Co₃O₄ catalysts, irrespective of PtO loading, and the reduction of PtO occurred separately at a distinctly lower temperature. Yang et al. [36] observed similar behavior for Pt promoted Co₃O₄/Al₂O₃ catalysts, wherein both PtO and Co₃O₄ reduced separately in distinct temperature regions. Even though the Co₃O₄ reduction shifted to slightly lower temperatures at higher Pt loadings in their study, a synergistic effect, similar to Pd and Co in PdO/Co₃O₄ is consistent with the smaller promotional effect of Pt on direct NO decomposition activity compared to Pd promotion. Conversely, the decreased reduction temperature of Co₃O₄ observed in H₂-TPR measurements of PdO/Co₃O₄ illustrates how Pd can promote direct NO decomposition by enhancing the reducibility of the catalyst.



Figure 6. H₂- Temperature programmed reduction profiles of fresh (**a**) PdO- and (**b**) PtO-promoted Co_3O_4 catalysts. The pure Co_3O_4 profile is shown for comparison.

2.2.3. Surface Properties

The X-ray photoelectron spectroscopy (XPS) was used to investigate the surface elemental compositions, metal oxidation states, and adsorbed oxygen species of the as-prepared and spent samples. The O1s XPS spectra of fresh PdO- and PtO-promoted Co_3O_4 catalysts are presented in Figure 7, with that of the pure Co_3O_4 for comparison. The pure Co_3O_4 exhibits two peaks in the O1s spectra. The large peak at lower binding energy (BE = 530.2–530.7 eV) is attributed to the surface lattice oxygen in Co_3O_4 (denoted as O_{lat}) [37]. The shoulder at higher BE (532.0–532.7 eV) is associated with oxygen atoms present as surface adsorbed oxygen or surface hydroxyl groups or defect oxide (denoted as O_{ad}). The PdO- and PtO-promoted Co_3O_4 samples also exhibit two peaks in their O1s spectra due to the O_{lat} and O_{ad} species., however, little difference in the peak energies is observed. This is may be due to lower loadings of promoters.



Figure 7. O1s XPS spectra of fresh (**a**) PdO- and (**b**) PtO-promoted Co₃O₄ catalysts. Co₃O₄ spectrum is shown for reference.

The fitted Co2*p* XPS spectra of the fresh PdO- and PtO-promoted Co₃O₄ catalysts are presented in Figure 8, with that of the pure Co₃O₄ for comparison. In the pure Co₃O₄ XPS spectrum, the main peak in the BE range of 780.7–782.2 eV is assigned to Co2*p*_{3/2}, and the shoulder at 795.9–797.9 eV is attributed to Co2*p*_{1/2}. Pure Co₃O₄ exhibits peaks due to both Co³⁺ and Co²⁺ and their satellites. The main Co2*p*_{3/2} feature can be further resolved into two components, with BE values centered at 778.7–780.4 eV and 779.8–781.6 eV, and corresponding to Co³⁺ and Co²⁺, respectively [38]. Furthermore, the presence of the satellite peaks also confirms the presence of Co²⁺ in the catalysts. As expected for samples containing Co₃O₄ spinel, all catalysts exhibited peaks and satellites due to both Co³⁺ and Co²⁺, irrespective of Pd or Pt promoter loading. Also, no significant change in the position of the peaks was observed upon impregnation of Co₃O₄ with PdO or PtO. The spent catalysts also exhibit peaks due to the Co³⁺ and Co²⁺ ions, irrespective of promoter identity or loading. The Co2*p* XPS measurements show that the Co₃O₄ spinel is very stable during direct NO decomposition, which agrees with XRD measurements.



Figure 8. Co_2p XPS profiles of fresh (a) PdO- and (b) PtO-promoted Co_3O_4 . The pure Co_3O_4 spectra is presented for reference.

The Pd3*d* XPS spectra of the fresh and spent 2, 3, and 4 wt% PdO/Co₃O₄ are presented in Figure 9a,b. In general, Pd may exist as Pd⁰ (335.1–335.4 eV [39]), Pd²⁺ (336.8.1–337.2 eV or 336.3–336.8 eV [40–44]), Pd⁴⁺ (337.8–339.3 eV), or a combination thereof. All the PdO/Co₃O₄ catalysts exhibit peaks due to the Pd²⁺ and Pd⁴⁺ after calcination at all PdO loadings. However, XRD measurements show no peaks corresponding to PdO or PdO₂ up to 3 wt% loading, which indicates that the PdO present on Co₃O₄ is in amorphous form and dispersed very well on the surface. In agreement with the above XRD analysis of the spent samples (see Figure 6a), XPS indicates PdO/Co₃O₄ catalysts exhibit peaks due to PdO and metallic Pd after direct NOx decomposition. These results suggest that some of the PdO reduced to metallic Pd during direct NOx decomposition, which corroborates the evidence from H₂-TPR and XRD of the promotional effect of Pd on the activity of Co₃O₄ spinel catalysts. The intensity of the metallic Pd increases with increasing PdO loading from 2 to 3 wt%, however, the intensity of the metallic Pd peak decreases drastically with further PdO loading from 3 to 4 wt%. This is due to the formation of the separate bulk PdO phase in the spent 4 wt% PdO/Co₃O₄ sample, which is clearly observed in the spent 4 wt% PdO/Co₃O₄ Pd3*d* spectrum. The formation of a separate PdO phase leads to less reduction of PdO to metallic Pd during direct NOx decomposition and is the likely cause of the lower activity compared to the 3 wt% PdO/Co₃O₄ catalyst. These results agree with the conclusions made based on the spent XRD patterns (see Figure 5a), further strengthening the evidence that 3 wt% Pd is the optimum loading for promoting activity.



Figure 9. Pd3d XPS spectra of (a) fresh and (b) spent PdO-promoted Co₃O₄ catalysts.

Figure 10a,b present the Pt4*f* XPS spectra of the fresh and spent 2, 3, and 4 wt% PdO/Co₃O₄. All fresh and spent PtO/Co₃O₄ catalysts only exhibit peaks due to Pt²⁺ (72.3, 74.1 eV) at all Pt loadings [45]. There are no observed peaks due to either Pt⁴⁺ or metallic Pt⁰ in contrast to the PdO/Co_3O_4 catalysts, wherein the PGM underwent significant changes in oxidation state with exposure to reaction conditions. Overall, XPS measurements show that the support Co_3O_4 is very stable during the promoter impregnation, as well as during the direct NO decomposition. PdO reduced to metallic Pd during the direct NO decomposition and improves the direct NO decomposition activity of Co_3O_4 . On the other hand, PtO stays in an oxidized state (no metallic Pt formation) during the direct NO decomposition and exhibits less promotional effects compared to PdO.

2.2.4. NO Adsorption Properties

The adsorption of NO and formation of surface intermediates is essential to establishing activity for direct NO decomposition. In situ FT-IR spectroscopy was performed on pure Co₃O₄ and the PdOand PtO-promoted Co₃O₄ catalysts to understand the interaction of NO with the catalyst during adsorption. In situ FT-IR measurements were collected during NO adsorption over the catalysts at 300 °C. Before NO adsorption, all the catalysts were pretreated at 350 °C in the presence of 10% O_2 in a helium balance and cooled to 300 °C in the presence of helium. All the spectra collected were normalized with respect to the gas phase NO peak at 1874 cm⁻¹. The in situ FTIR spectra of Co₃O₄, PdO- and PtO-promoted Co₃O₄ catalysts during NO adsorption are presented in Figure 11a,b. As shown in Figure 11a, little to no NO adsorption occurs over the pure Co_3O_4 spinel oxide at 300 °C, as no clear peaks were present relative to the noise level. Interestingly, impregnating PdO over Co_3O_4 leads to a formation of chelating surface nitrate intermediates (1577 and 1254 cm⁻¹) during the NO adsorption [46]. The intermediate formation was observed for all the catalysts irrespective of loading and the intensity of the peak increases with PdO loading. On the other hand, PtO-promoted Co₃O₄ catalysts do not produce spectroscopically relevant amounts of intermediates during the NO adsorption and exhibit spectra similar to the pristine Co_3O_4 catalyst. The catalysts in the current study exhibit activity at temperatures \geq 400 °C, however, at these temperatures, no spectroscopically relevant surface NOx species were observed by in situ FTIR (not shown). This observation indicates that neither the surface chelating nitrate nor any other surface NOx species is the most abundant reactive intermediate in the direct NO decomposition mechanism. The in situ FTIR results at 300 °C are, therefore, interpreted as a probe of the affinity of the NO reactant molecule to interact with the catalyst surface. In this interpretation, it is concluded that the presence of PdO improves the affinity of the catalyst to interact with NO compared to PtO or the pure Co_3O_4 support.

The direct NO decomposition measurements show that PdO promotes direct NO decomposition activity of Co_3O_4 much better compared to the PtO. The Co_3O_4 is a normal spinel with AB_2O_4 formula, where A (T_d) sites are occupied by Co^{2+} ions and B (Oh) sites are occupied by Co^{3+} ions. According to the general mechanism proposed by Haneda et al. [18], initially NO adsorbs on the surface and decomposes into N and O. The oxygen atoms adsorb on Co^{2+} ions and are oxidized to Co^{3+} . Then, Co^{3+} ions reduce back to Co^{2+} upon release of the oxygen as a product. Hence, NO adsorption and oxygen release (redox) properties are very important for direct NO decomposition. The *in-situ* FT-IR results reveal that PdO increases the affinity of the catalyst to form surface NOx species compared to PtO or a pure Co_3O_4 support. The H₂-TPR studies in our work show that the Co_3O_4 reduction temperature is significantly decreased by the presence of dispersed PdO, thus suggesting a more facile reduction of Co^{3+} to Co^{2+} to release of cobalt reduction explains the better direct NO decomposition activity of PdO catalysts compared to PtO catalysts.



Figure 10. Pt4f XPS profiles of (a) fresh and (b) spent PtO-promoted Co₃O₄ catalysts.



Figure 11. *In-situ* FT-IR spectra of Co_3O_4 , PdO (a), and PtO (b) promoted Co_3O_4 catalysts during NO adsorption at 300 °C.

3. Materials and Methods

3.1. Catalyst Synthesis

Palladium and platinum promoted Co_3O_4 catalysts were synthesized using the wet impregnation method. Commercial Co_3O_4 was purchased from Sigma-Aldrich (St. Louis, MO, USA) (99.5% trace metal basis) and used as received without any further modification for the synthesis. In a typical synthesis procedure, 5 g of commercial Co_3O_4 were mixed with 50 mL of water. Then, the required quantity of palladium nitrate hydrate (Sigma-Aldrich), or tetraamine platinum (II) nitrate (Sigma-Aldric, 99.995% trace metal basis) was dissolved separately in deionized water and combined with the Co_3O_4 suspension. The mixture was heated to 80 °C with continuous stirring. The obtained powder was then dried in an oven at 120 °C for 12 h under air. Finally, the catalyst was calcined

at 400 °C for 4 h with a 1 °C/min ramp. Different loadings of palladium and platinum on Co_3O_4 (nominally 1 to 4 wt%) catalysts were prepared by varying the amount of palladium nitrate or platinum nitrate. For reference, the commercial Co_3O_4 support was also calcined at 400 °C for 4 h.

3.2. Catalyst Characterization

X-ray diffraction: X-ray powder diffraction (XRD) patterns were obtained using a Rigaku SmartLab X-ray diffractometer (Rigaku, The Woodlands, TX, USA) using Cu K α radiation (1.5405 A). A glass holder was used to support the sample. The scanning range was from 10 to 80° (20) with a step size of 0.02° and a step time of 1 s. The XRD phases present in the samples were identified with the help of ICDD-JCPDS [31] data files.

BET Surface Area Measurements: The surface area of the PdO and PtO promoted Co₃O₄ materials were measured using a Micromeritics 3Flex surface characterization instrument (Micromeritics, Atlanta, GA, USA). N₂ physisorption isotherms was conducted at -196 °C, and the surface area was measured by the BET method. Prior to the analyses, the samples were outgassed at 300 °C under vacuum (5 × 10⁻³ Torr) for 3 h.

X-ray Fluorescence Measurements: XRF was collected using a Rigaku ZSX, primus II X-ray spectrometer (Rigaku, The Woodlands, TX, USA). Impurities in the crystals were gained by X-ray fluorescence in operation of spectrometer in standard fewer modes with coverage of a full element. The amount of any elements and oxides particles was detected by the XRF experiment.

 H_2 -Temperature Programmed Reduction (H_2 -TPR) Measurements: The redox properties of the PtO/Co₃O₄ and PdO/Co₃O₄ catalysts were studied using H_2 temperature programmed reduction (H_2 -TPR) experiments. H_2 -TPR experiments were performed using a Micromeritics 3Flex surface characterization instrument (Micromeritics, Atlanta, GA, USA) equipped with a thermal conductivity detector. Before the experiment, the catalysts were preheated to 300 °C in the presence of 20% O₂/He (30 mL/min). After the pretreatment, the temperature was decreased to 50 °C. The H_2 -TPR measurements were performed by heating the catalyst from 50 to 600 °C in the presence of 10% H_2 /Ar (30 mL/min).

X-ray Photo Electron Spectroscopy: The XPS measurements were performed using a PHI 5000 Versa Probe II X-ray photoelectron spectrometer (Physical Electronics, East Chanhassen, MN, USA) using an Al K α source. Charging of the catalyst samples was corrected by setting the binding energy of the adventitious carbon (C1s) to 284.6 eV [47]. The XPS analysis was performed at ambient temperature and at pressures typically on the order of 10⁷ Torr. Prior to the analysis, the samples were outgassed under vacuum for 30 min.

In Situ FTIR Spectroscopy Measurements during NO Adsorption: The NO adsorption properties were measured using in situ Fourier transform infrared (FTIR) spectroscopy. The Harrick High Temperature Cell with environmental (gas flow) and temperature control was used for in situ diffuse-reflectance FTIR spectroscopy. Spectra were recorded using a Thermo Scientific Nicolet 8700 Research FT-IR Spectrometer (Thermo Scientific Fidher, Waltham, MA USA) equipped with a liquid N₂ cooled MCT detector. Spectra were obtained with a resolution of 2 cm⁻¹ and by averaging 64 scans. In situ diffuse-reflectance FTIR spectra were collected during NO adsorption at 300 °C. Prior to NO adsorption, the sample was first pretreated at 350 °C in 30 mL/min of 10% O₂/He. The background spectrum (64 scans) was of the catalyst after cooling to 300 °C in 30 mL/min UHP He. Adsorption of NO was achieved by flowing 30 mL/min of 10,000 ppm NO over the catalyst for 25 min. Adsorption of NO proceeded for 25 min, while spectra were obtained every minute using a series collection. To compare peak intensities among different catalyst samples, the adsorption spectra were normalized to the NO gas phase peak at ~1876 cm⁻¹.

3.3. Direct NO Decomposition Measurements

Direct NO decomposition measurements were performed in a fixed bed flow reactor. A gas mixture of ~1% NO in helium balance was used with a gas hourly space velocity of ~2100 h^{-1} and in

the temperature region of 450–650 °C. Before the reaction, catalysts were pretreated at 500 °C in the presence of 20% O₂/He. After pretreatment, the bed temperature was decreased to 400 °C and direct NO decomposition measurements were collected. The measurements were performed at 400, 450, 550, and 650 °C, with 2 h of steady state at each temperature. The NO, N₂O, and NO₂ concentrations were analyzed with a FTIR detector (CAI 600 SC FTIR California Analytical Instruments, Inc., Orange County, CA, USA). The N₂ concentration was calculated by mass balance of the total nitrogen species. The raw NO conversion (NO converted to all the products) during the steady state measurements are presented in Figure S1 of the supporting information and activity to N₂ was reported in Figure 1 of the manuscript. The steady state direct NO decomposition measurements were also performed in a reactor, which was equipped with the mass spectroscopy (MKs, Cirrus 2). The changes in the NO, N₂, O₂, N₂O, and NO₂ signal intensities were monitored during the reaction (Figure S2–S5). The inert Ar gas was introduced as a tracer to monitor for potential systematic variation in signal intensity during the experiment (Figure S2).

4. Conclusions

The direct NO decomposition measurements show that PdO promotes the activity of Co_3O_4 and is 4 times more active compared to PtO at 650 °C. Also, the activity increases with increasing PdO loading until 3 wt% and further increase in the loading leads to a decrease in the activity. On the other hand, only a slight increase in the activity was observed with increasing PtO loading up to 4 wt%. Surface area measurements indicated that both PdO and PtO have little to no influence on the surface area of Co₃O₄, except for a decrease in surface area for 4 wt% PdO/Co₃O₄. The X-ray diffraction measurements show that Pt incorporated into the Co₃O₄ structure during the synthesis and PdO stays mostly on the surface. The diffraction measurements also suggested that PdO is in an amorphous form up to 3 wt% over Co₃O₄ surface and crystalline PdO forms at 4 wt% loading, whereas PtO mostly stays as amorphous from or incorporated into Co_3O_4 structure until 4 wt%. Due to the synergistic effect between Pd species and Co₃O₄, an improvement in the redox properties of Co₃O₄ was observed in the case of PdO/Co₃O₄ catalysts. Conversely, PtO do not have any influence on the redox properties of Co₃O₄. The X-ray photo electron spectroscopic measurements reveal that PdO reduced to Pd metal during the direct NO decomposition reaction and Pt was in 2+ oxidation state before and after the direct NO decomposition reaction. In situ NO adsorption measurements show that PdO improve the NO adsorption properties of Co₃O₄ by forming the nitrate ion intermediates, whereas PtO/Co₃O₄ do not form any intermediates during the NO adsorption at 300 °C. Overall, PdO ease the redox properties of Co₃O₄ and forms surface adsorbed species during NO adsorption and improves the direct NO decomposition activity of Co_3O_4 . On the other hand, PtO do not have any influence on the redox or NO adsorption properties of Co₃O₄ and exhibits lesser promotional effects compared to PdO. For PdO/Co₃O₄ catalysts, the PdO remains in amorphous form until 3PdO/Co₃O₄ and improves the activity of Co₃O₄ with loading. However, further increase in the loading to 4 wt% leads to formation of crystalline PdO, which reduces separately during H₂-TPR and inhibits the PdO reduction to metallic Pd during direct NO decomposition and exhibits lesser activity compared to 3 wt% PdO/Co₃O₄.

Supporting information: The NO conversion, and N₂, N₂O, and NO₂ ppm values of various PdO/Co₃O₄ and PtO/Co₃O₄ catalysts are presented in Tables S1 and S2. The total NO conversion profiles of the Co₃O₄, 3PdO/Co₃O₄, and 4PtO/CoO calculated from the FT-IR detector during the steady state direct NO decomposition measurements are presented in Figure S2. NO, Ar, N₂, O₂, N₂O, and NO₂ partial pressure values (obtained from mass spectroscopy) of Co₃O₄, 3 wt% PdO/Co₃O₄, and 3 wt% PtO/Co₃O₄ during steady state direct NO decomposition in the 400 to 650 °C temperature region are presented in Figure S3 and S4.

Supplementary Materials: The following are available online at http://www.mdpi.com/2073-4344/9/1/62/s1. Table S1: NO conversion, N₂, N₂O, NO₂ ppm values of the Co_3O_4 and various PdO/ Co_3O_4 catalysts in the temperature region 400–650 °C. Table S2: NO conversion, N₂, N₂O, NO₂ ppm values of the Co_3O_4 and various PtO/ Co_3O_4 catalysts in the temperature region 400–650 °C. Figure S1: Steady state NO conversion values of the Co₃O₄ and various PtO/ Co_3O_4 catalysts in the temperature region 400–650 °C.

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the Co₃O₄, 3PdO/Co₃O₄, and 4PtO/Co₃O₄ catalysts during the direct NO decomposition in the temperature region 400 to 650 °C. Figure S2: NO and Ar M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄, and (c) 3PtO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C. Figure S3: N₂ and O₂ M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C. Figure S3: N₂ and O₂ M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄, and (c) 3PtO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C. Figure S4: N₂O M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄, and (c) 3PtO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C. Figure S4: N₂O M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄, and (c) 3PtO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C. Figure S5: NO₂ M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄, and (c) 3PtO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C. Figure S5: NO₂ M.S. partial pressures of the (a) Co₃O₄, (b) 3PdO/Co₃O₄, and (c) 3PtO/Co₃O₄ catalysts during the steady state direct NO decomposition in the temperature region 400 to 650 °C.

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Article Investigation of Various Pd Species in Pd/BEA for Cold Start Application

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Abstract: A series of Pd/BEA catalysts with various Pd loadings were synthesized. Two active Pd²⁺ species, Z^- -Pd²⁺- Z^- and Z^- -Pd(OH)⁺, on exchanged sites of zeolites, were identified by in situ FTIR using CO and NH₃ respectively. Higher NO_x storage capacity of Z^- -Pd²⁺- Z^- was demonstrated compared with that of Z^- -Pd(OH)⁺, which was caused by the different resistance to H₂O. Besides, lower Pd loading led to a sharp decline of Z^- -Pd(OH)⁺, which was attributed to the 'exchange preference' for Z^- -Pd²⁺- Z^- in BEA. Based on this research, the atom utilization of Pd can be improved by decreasing Pd loading.

Keywords: Pd/BEA; Cold start; Pd species; NOx abatement

1. Introduction

The exhaust regulation on NO_x emissions is getting more stringent [1]. Presently, NH₃ selective catalytic reduction (SCR) [2] and NO_x storage reduction (NSR) [3] are widely used for NO_x removal. However, standard NO_x aftertreatment technologies fail to function efficiently at low temperatures, which results in a large proportion of the tailpipe NO_x emission [4]. Meanwhile, high efficiency internal combustion engines require new and/or improved technologies which specifically address their low exhaust temperatures. In response to difficulties of low temperature emissions control, numerous efforts are underway to develop catalysts that light-off at temperatures below 150 °C. Passive NO_x adsorbers (PNAs) could play a critical role in enabling high efficiency advanced combustion systems.

Recently, Pd/zeolite serving as passive NO_x adsorbers (PNAs) was first proposed by Chen et al. [4]. This catalyst is emerging as effective passive NO_x adsorbent technology because of its NO_x storage/release capabilities, resistance to sulfur poisoning and hydrothermal deactivation [4–8]. Due to these excellent characteristics, Pd/zeolite has attracted great interest recently and has been further optimized [6–18].

Several recent studies show that isolated Pd ions in Pd/zeolite are the main active sites for NO_x trapping [6,7,11,13,16]. It is reported that there are nine skeletal T sites with various chemical environments in BEA [19]. So, isolated Pd ions located in various positions of zeolite framework are likely to be formed. As reported by Gao et al. and Giordanino et al. [20–22], two kinds of isolated Cu species (Cu²⁺ and [Cu(OH)]⁺) are identified on various framework positions of Cu/SSZ-13, which indicates that species of isolated cations may be influenced by their locations in zeolites. They also mentioned that Cu species were significantly affected by Cu loading. So, various isolated Pd species may co-exist in Pd/BEA, and Pd loading may be capable of influencing the content of them. Actually, two kinds of active isolated Pd ions (bare Pd²⁺ and Pd(OH)⁺) have been observed by Zheng et al. [10]

and Khivantsev et al. [12]. However, they did not point out the difference in adsorption between these two species.

Therefore, these two kinds of isolated Pd ions were further studied in this research. A series of in situ FTIR experiments in CO and NH_3 were carried out, and a semiquantitative method was adopted to distinguish these two species. Based on this method, the NO_x storage capacity of each isolated Pd species was compared. Further, by increasing the proportion of isolated Pd species with higher NO_x storage capacities, the atom utilization of Pd can be improved.

2. Results

2.1. Ex-Situ FTIR

Ex-situ FTIR spectra of each sample are exhibited in Figure 1. Compared with 0-Pd (Figure 1a), an extra vibration at 926 cm⁻¹ is observed on the FTIR spectrum of 1-Pd. This peak is attributed to the vibration distortion of the skeletal T–O–T bond due to the strong interaction of Pd ions [4]. This is an obvious piece of evidence for the existence of Pd ions on the exchange sites of zeolites. Peaks at 926 cm⁻¹ appear on spectra of 0.2-Pd, 0.5-Pd and 2-Pd, too (Figure 1b). So, the existence of Pd ions on exchange sites in these samples is confirmed.



Figure 1. (a) Ex-situ FTIR spectra of 0-Pd, 1-Pd, Na-1-Pd; (b) Ex-situ FTIR spectra of 2-Pd, 1-Pd, 0.5-Pd, 0.2-Pd (Temperature: 200 °C; Flow: N₂, 1 L/min).

Compared with 1-Pd (Figure 1a), the peak at 926 cm⁻¹ disappears in the spectrum of Na-1-Pd, which indicates the elimination of isolated Pd ions on exchange sites. This is firm evidence of which isolated Pd ions in Na-1-Pd have been completely exchanged by Na⁺ ions through the titration process. So, Pd loaded on Na-1-Pd mainly exists in form of Pd oxidations.

2.2. Na⁺ Titration

Na⁺ titration was adopted to measure the content of isolated Pd ions in each sample [23], and the proportion of isolated Pd ions in total Pd loading (marked as Isolated Pd/Pd loading in Table 1) was

further calculated. Considering that Pd is sensitive to Cl [24,25], the titration process was carried out in a NaNO₃ solution. Besides, this measurement is believed to have no negative effect on zeolite structures since the titration process was carried out in mild conditions and no calcination was done. As shown in Table 1, lower Pd loading leads to less isolated Pd^{2+} content, whereas the proportion of isolated Pd ions in total Pd loading becomes larger. This phenomenon indicates that the increase of Pd loading leads to the formation of more Pd oxidations (PdOx), which is consistent with Jaeha Lee et al.'s study [11].

Catalyst	0-Pd	0.2-Pd	0.5-Pd	1-Pd	2-Pd
Isolated Pd (wt %)	0.00	0.22	0.53	0.87	0.91
Isolated Pd/Pd loading (%)	0.00	95.7	96.4	77.7	42.7

Table 1. The content of Pd ions on exchange sites of each sample.

2.3. CO In Situ FTIR

Pd species were further probed by CO using in situ FTIR, and the result is displayed in Figure 2. Peaks below 2100 cm^{-1} are attributed to CO signals on metallic Pd (Pd⁰) formed via CO reduction [10,26], among which Pd⁰–CO (atop) are found at 2098 cm⁻¹, 2076 cm⁻¹ and Pd₂⁰–CO (bridging) are found at 1951 cm⁻¹. Besides, the peak at 2117 cm⁻¹ is attributed to the C-O vibration on Pd⁺ [10,27]. As reported by Vu et al. [9], Pd⁺ is formed by the CO reduction of ion-exchanged Pd species [9]. CO signals on isolated Pd²⁺ species are observed above 2100 cm⁻¹. The peak at 2152 cm⁻¹ with a shoulder peak at 2137 cm⁻¹ is attributed to the C-O vibration on isolated Pd²⁺ bonded with the hydroxy of zeolites (marked as Z⁻-Pd²⁺-Z⁻) [10]. The peak at 2179 cm⁻¹ is attributed to the vibration of C-O adsorbed by another kind of isolated Pd²⁺ [10], which was first determined as Z⁻-Pd(OH)⁺ by Okumura et al. [28].



Figure 2. CO in situ FTIR spectra of 2-Pd, 1-Pd, 0.5-Pd, 0.2-Pd (Temperature: 80 $^{\circ}$ C; Flow: 1000 ppm CO, balanced with N₂, 500 mL/min).

In short, two kinds of isolated Pd^{2+} (Z^- - Pd^{2+} - Z^- and Z^- - $Pd(OH)^+$) were identified in 0.2-Pd, 0.5-Pd, 1-Pd and 2-Pd. Note that all spectra in Figure 2 were obtained when the steady state had been achieved, and corresponding peaks of Z^- - Pd^{2+} - Z^- and Z^- - $Pd(OH)^+$ were still observed. So, these two

isolated Pd²⁺ species cannot be completely reduced by CO at this temperature, which indicated that the reduction reaction is a reversible one. As shown in Figure S1, the existence of Z⁻-Pd²⁺-Z⁻ in 1-Pd-80 is also demonstrated by CO whereas no Z⁻-Pd(OH)⁺ is observed. Since the complete reduction of Z⁻-Pd(OH)⁺ cannot be achieved by CO, there is only one isolated Pd²⁺ species, Z⁻-Pd²⁺-Z⁻, in 1-Pd-80 (see Figure S1).

2.4. Catalyst Evaluation

Profiles of the NO adsorption stage are shown in Figure 3. NO_x storage capacities are calculated by the integration of negative peaks on these profiles, and the dead volume has been subtracted. The result is displayed in Table 2. Note that samples with the same Pd loading as much as 1 wt % (see Figure S2) exhibit entirely different NO_x storage capacities (53.3 µmol/gcat and 9.9 µmol/gcat respectively), which indicates that only part of Pd loaded on samples is efficient. As shown in Figure 3, NO_x storage capacities of 0-Pd and Na-1-Pd are very low. So, Brønsted hydroxyl group and PdOx are inefficient active centers for NO_x storage, whereas they do trap NO_x in this condition as reported [10,27]. Since the NO_x storage capacity of 1-Pd is much higher, isolated Pd^{2+} ions are likely to be the main active sites for NO trapping.



Figure 3. NO_x adsorption profiles (Temperature: 80 °C; Flow: NO_x, CO, H₂O, O₂, CO₂, balanced with N₂, 1 L/min).

Table 2. NO_x storage capacities in NO_x , CO, H_2O , O_2 and CO_2 .

Catalyst	0-Pd	0.2-Pd	0.5-Pd	1-Pd	2-Pd
NO _x storage capacity (µmol/gcat)	3.7	21.6	47.5	53.3	50.1

2.5. NH₃ In Situ FTIR

Acidity over samples is probed by NH₃ with in situ FTIR, and the result is exhibited in Figure 4. The peak at 1463cm⁻¹ observed in both spectra of 1-Pd and 0-Pd (Figure 4a) is attributed to the vibration of NH⁴⁺ in Brønsted hydroxyl groups (NH⁴⁺-B) [29]. In the spectrum of 1-Pd, two additional peaks at 1625 cm⁻¹ and 1313cm⁻¹ corresponding to the vibration of NH₃ on Lewis acid (NH₃-L) [30,31] is observed. Nevertheless, these two peaks do not exist in Na-1-Pd in which isolated Pd²⁺ ions (Z⁻-Pd²⁺-Z⁻ and Z⁻-Pd(OH)⁺) are replaced by Na⁺ completely. So, it is reasonable to believe that Lewis acid is generated from Z⁻-Pd²⁺-Z⁻ and Z⁻-Pd(OH)⁺ species.



Figure 4. (a) NH₃ in situ FTIR spectra of 0-Pd, 1-Pd, Na-1-Pd; (b) NH₃ in situ FTIR spectra of 2-Pd, 1-Pd, 0.5-Pd, 0.2-Pd (Temperature: 80 °C; Flow: 500 ppm NH₃, balanced with N₂, 500 mL/min).

As shown in Figure S3, there is only one peak at 1625cm⁻¹ corresponding to the vibration of NH₃-L observed in the spectrum of 1-Pd-80. Since there is only one kind of isolated Pd²⁺, Z⁻-Pd²⁺-Z⁻, in this sample as discussed above, the peak at 1625cm⁻¹ should be assigned to the vibration of NH₃-L originated from Z⁻-Pd²⁺-Z⁻. In this case, the peak at 1313cm⁻¹ should be attributed to the vibration of NH₃-L originating from Z⁻-Pd(OH)⁺.

Since Z⁻-Pd²⁺-Z⁻ and Z⁻-Pd(OH)⁺ are capable of being probed by NH₃, the ratio of the height of the peaks at 1625 cm⁻¹ (P_Z⁻-_{Pd}²⁺-_Z⁻) and 1313 cm⁻¹ (P_Z⁻-_{Pd(OH)}⁺) in Figure 4b can be defined as the relative content between these two isolated Pd²⁺ species. Considering that extinction coefficients for NH₃ molecules adsorbed on Z⁻-Pd²⁺-Z⁻ and Z⁻-Pd(OH)⁺ are both constants, they have no effect on the tendency of peak height ratios. So, the extinction coefficient is not considered in this part [32–34]. The result is shown in Figure 5a. It is worth nothing that P_Z^{-} - $P_d(OH)^{+}/P_Z^{-}$ - P_d^{2+} - Z^{-} rises in parallel with the increase of isolated Pd²⁺, which means that the higher content of isolated Pd²⁺ leads to a much more obvious increase of Z⁻-Pd(OH)⁺ than that of Z⁻-Pd²⁺-Z⁻. So, isolated Pd²⁺ in 0.2-Pd mainly exists in the form of Z⁻-Pd²⁺-Z⁻, whereas a large amount of Z⁻-Pd(OH)⁺ ions are formed in 2-Pd.



Figure 5. (a) tendency between isolated Pd^{2+} and $P_Z^{-}_{Pd(OH)}^{+}/P_Z^{-}_{Pd}^{2+}_{-Z}^{-}$; (b) tendency between $P_Z^{-}_{Pd(OH)}^{+}/P_Z^{-}_{Pd}^{2+}_{-Z}^{-}_{-Z}$ and NO_x/Pd^{2+} .

Mole ratios of NO_x adsorbed to isolated Pd²⁺ for each sample (marked as NO/Pd²⁺) are calculated via data in Tables 1 and 2. Figure 5b is plotted by $P_Z^--_{Pd(OH)}^+/P_Z^--_{Pd}^{2+}-_Z^-$ on the horizontal axis and NO_x/Pd²⁺ on the vertical. As reported, NO_x is believed to be trapped in the form of Pd⁹⁺-NO [12,15], so the maximum NO_x/Pd²⁺ ratio is 1 in theory. However, the downward trend

between $P_Z^- P_{d(OH)}^+ / P_Z^- P_d^{2+} Z^-$ and NO_x / Pd^{2+} means that higher $P_Z^- P_{d(OH)}^+ / P_Z^- P_d^{2+} Z^-$ leads to lower NO_x storage capacity of isolated Pd^{2+} , which indicates that the NO_x storage capacity of $Z^- Pd(OH)^+$ is obviously lower than that of $Z^- Pd^{2+} Z^-$ in this condition.

A further discussion is given below so as to give a reasonable explanation from the perspective of the structure-function relationship.

In BEA zeolite, nine skeletal T sites with various chemical environments are determined [19]. Among these T sites, T5 and T6 are demonstrated to have the lowest Al substitution energy [35]. Thus, Al substitution on these two sites (marked as Al_{T5} and Al_{T6}) is preferential, and the amount of Al_{T5} and Al_{T6} is larger than that of Al atoms on the other T sites in BEA. Besides, exchange sites formed on these two Al atoms are more active due to the lowest deprotonation energy [36] of them. Meanwhile, Al_{T5} and Al_{T6} are located in the meta-position of the same five-membered ring [37], and the co-ion-exchange of protons on these two exchange sites formed on Al_{T5} and Al_{T6} is feasible. These characteristics of BEA can explain the 'exchange preference' for $Z^-Pd^{2+}-Z^-$ formed in Pd/BEA with lower content of isolated Pd ions to some extent. Besides, with the increase of Pd loading, more exchange sites are taken up. Limited by the amount of protons which are capable of being exchanged, $Z^-Pd(OH)^+$ ions, which take up less exchange sites than $Z^-Pd^{2+}-Z^-$, are preferred to be formed. So, higher content of isolated Pd²⁺ leads to much more obvious increase of $Z^-Pd(OH)^+$ than that of $Z^-Pd^{2+}-Z^-$.

As discussed above, the maximum NO_x/Pd^{2+} ratio is 1 in theory. As reported by Khivantsev et al. [12], H₂O molecules can occupy NO_x storage sites due to strong hydration of isolated Pd²⁺ species. Since isolated Pd²⁺ is the main active site for NO_x storage, the NO_x/Pd^{2+} ratio of Pd/zeolite-based PNA materials will significantly smaller than 1 in the presence of H₂O. However, as shown in Figure 5b, NO_x/Pd^{2+} of 0.2-Pd is as much as 1, which indicates that isolated Pd²⁺ in this sample is not occupied by H₂O. As discussed above, isolated Pd²⁺ in 0.2-Pd mainly exists in the form of Z⁻-Pd²⁺-Z⁻. Thus Z⁻-Pd²⁺-Z⁻ may be insensitive to H₂O in this condition. Furthermore, Z⁻-Pd(OH)⁺ probably tends to hydrate due to the hydrogen bond interaction between H₂O and hydroxy. Accordingly, the difference in NO_x storage capacities of Z⁻-Pd(OH)⁺ and Z⁻-Pd²⁺-Z⁻ is probably caused by the different resistance to H₂O in this condition.

The atom utilization of Pd can be represented by mole ratios of NO_x adsorbed to total Pd loading. According to Figure 6, the Pd utilization of 0.2-Pd is as much as 100% whereas that of 2-Pd is only 25%. So, it is obvious that lower Pd loading benefits to the increase of Pd utilization. Besides, Khivantsev et al. [13] have reported that the Pd utilization of the 1 wt% Pd/SSZ-13 (Si/Al ratio = 6) sample is 100%, which is much higher than that of 1-Pd (Si/Al ratio = 16) using BEA as a support. Accordingly, zeolite structure and Si/Al ratio can also influence the Pd utilization. In this part, only the influence of Pd loading will be discussed.

As discussed above, lower Pd loading leads to lower content of isolated Pd²⁺, whereas the proportion of isolated Pd ions in total Pd loading becomes larger. On the on hand, a larger proportion of isolated Pd ions in total Pd loading is beneficial for the improvement of Pd utilization, since isolated Pd ions are identified as the main active center for NO_x storage. On the other hand, less isolated Pd²⁺ results in a preference for the formation of Z⁻-Pd²⁺-Z⁻ which exhibits higher NO_x storage capacity. Thus, the atom utilization of Pd can be improved by decreasing Pd loading.

Nevertheless, lower Pd loading also leads to less NO_x storage capacities of unit mass of Pd/BEA. Note that the coating amount of catalysts is limited in order to obtain acceptable back pressure. Accordingly, to determine optimum Pd loading for low temperature NO_x adsorption, the NO_x storage capacity and the atom utilization of Pd should be both considered. As shown in Figure 6, 0.2-Pd and 0.5-Pd exhibits much higher Pd atom utilization than that of the other two samples. Meanwhile, Pd atom utilization of 0.5-Pd is 92%, which is only slightly lower than that of 0.2-Pd. However, the NO_x storage capacity of 0.5-Pd is larger than twice of that of 0.2-Pd. So, 0.5wt% should be determined as the optimum Pd loading for Pd/BEA (Si/Al ratio = 16) served as PNA material.



Figure 6. Tendency between Pd loading and the atom utilization of Pd.

3. Materials and Methods

3.1. Catalyst Preparation

H-BEA zeolite (Si/Al = 16.2) was supplied by Novel Chemistry. Pd/BEA samples with various Pd loading were prepared by incipient wetness impregnation, and Pd(NO₃)₂ solution (15.47wt% Pd, Heraeus Materials Technology Shanghai Ltd., Shanghai, China) was used. Then, the powders obtained were dried under ambient temperature followed by a 4-h calcination at 550 °C in air. Finally, samples were stabilized at 750 °C for 12 h in air with 10% H₂O. Pd loading of each sample was detected by inductively coupled plasma (ICP) analysis (USA Agilent 5100 ICP-OES, Santa Clara, CA, USA). Si/Al ratios were measured by X-ray fluorescence (XRF, ThermoFisher PERFORM'X, Waltham, MA, USA) analysis. Detailed information is exhibited in Table 3. In the following text, samples are abbreviated as x-Pd, where "x" represents the Pd loading of corresponding samples. Besides, the sample treated by Na⁺ titration was named as Na-1-Pd.

Table 3. Pd loading of samples.						
Catalyst	0-Pd	0.2-Pd	0.5-Pd	1-Pd	2-Pd	Na-1-Pd
ICP Pd (wt %)	0	0.23	0.55	1.12	2.13	0.25

3.2. Catalyst Characterization

Na⁺ titration was used to quantify isolated Pd ions as reported by Ogura et al. [23]. Each sample was mixed with NaNO₃ (Purity above 99.0%, Tianjin Jiangtian Chemical Technology Co., Ltd., Tianjin, China) solution (0.1M) at 80 °C. The solution was stirred for 4 h followed by suction filtration. At last, the powders obtained were washed by deionized water and dried at 100 °C for 6 h. The whole process above was repeated three times. By analyzing the change of Pd loading in each sample after the titration process, the content of isolated Pd can be measured.

NO_x storage capacities were measured by standard catalyst evaluation tests carried out in a plug flow reactor system. 0.25 g sample (60–80 mesh) mixed with 0.75 g quartz (60–80 mesh) was loaded in a quartz reactor with a thermocouple. The sample was oxidized in the flow of 10% O_2/N_2 for 30 min at 500 °C and was cooled to 80 °C in N₂. Then the flow (200 ppm NO_x, 200 ppm CO, 5% CO₂, 5%H₂O, 10% O_2 , balanced with N₂) mixed in the bypass in advance was introduced into the reactor. The NO_x

adsorption stage continued for 3 min. After that, the sample was heated to 500 °C with a ramping rate of 10 °C/min in the flow of N₂. Gas concentrations were measured by an online MKS MultiGas 2030 FTIR gas analyzer in the whole process. Besides, a space velocity of 28,800 h^{-1} was adopted.

Ex-situ FTIR was carried out on a Nicolet iS10 FTIR spectrometer equipped with a liquid N_2 cooled mercury cadmium telluride (MCT) detector to identify the existence of Pd ions. The sample (20 mg) was pressed into a self-supporting wafer with a diameter of 13 mm and was inserted into a cell sealed with ZnSe windows connected with a gas manifold. KBr was used to obtain background spectra. All samples, as well as KBr, were oxidized in the flow of 10% O_2/N_2 for 30 min at 500 °C in advance. Spectra were obtained at 200 °C in N_2 .

In situ FTIR was carried out on a Nicolet iS10 FTIR spectrometer, too. Test temperatures and feed compositions varied according to the needs of different experiments, and detailed information will be reported below.

4. Conclusions

In this work, two isolated Pd^{2+} ions, $Z^--Pd^{2+}-Z^-$ and $Z^--Pd(OH)^+$, on exchange sites of zeolites are confirmed as the main active sites for NO trapping in cold-start applications. Lower Pd loading leads to a lower content of isolated Pd^{2+} whereas the proportion of isolated Pd ions in total Pd loading becomes larger. Lower content of isolated Pd^{2+} further leads to a sharp decline of $Z^--Pd(OH)^+$, which is attributed to the 'exchange preference' for $Z^--Pd^{2+}-Z^-$ in BEA. Besides, the higher NO_x storage capacity of $Z^--Pd^{2+}-Z^-$ is demonstrated compared with that of $Z^--Pd(OH)^+$, which is caused by the different resistance to H₂O. In conclusion, the atom utilization of Pd can be improved by using lower Pd loading. 0.5wt% should be determined as the optimum Pd loading for Pd/BEA (Si/Al ratio = 16) serving as PNA material among all samples.

Supplementary Materials: The following are available online at http://www.mdpi.com/2073-4344/9/3/247/s1, Table S1: Detailed information of 1-Pd-80, Figure S1: CO in situ FTIR spectra of 1-Pd-80, Figure S2: NO_x adsorption profiles of 1-Pd and 1-Pd-80, Figure S3: NH₃ in situ FTIR spectra of 1-Pd-80.

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Article Optimization of Ammonia Oxidation Using Response Surface Methodology

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Abstract: In this paper, the design of experiments and response surface methodology were proposed to study ammonia oxidation process. The following independent variables were selected: the reactor's load, the temperature of reaction and the number of catalytic gauzes, whereas ammonia oxidation efficiency and N₂O concentration in nitrous gases were assumed as dependent variables (response). Based on the achieved results, statistically significant mathematical models were developed which describe the effect of independent variables on the analysed responses. In case of ammonia oxidation efficiency, its achieved value depends on the reactor's load and the number of catalytic gauzes, whereas the temperature in the studied range (870–910 °C) has no effect on this dependent variable. The concentration of nitrous oxide in nitrous gases depends on all three parameters. The developed models were used for the multi-criteria optimization with the application of desirability function. Sets of parameters were achieved for which optimization assumptions were met: maximization of ammonia oxidation efficiency and minimization of the N₂O amount being formed in the reaction.

Keywords: ammonia oxidation; response surface methodology; desirability function; Box-Behnken design

1. Introduction

Nitric acid is mainly used for producing nitrogen fertilizers: ammonium nitrate (AN) and calcium ammonium nitrate (CAN) which constitute 75–80% of its entire production. The remaining amount of nitric acid is used in other industrial applications for example as a nitration agent for the production of explosives and other semi-organic products (aliphatic nitro compounds and aromatic nitro compounds) for the production of adipic acid, for metallurgy (etching steel) [1].

The industrial production of nitric acid is based on Ostwald process [1] which involves three basic stages: the catalytic oxidation of ammonia to nitrogen oxide (NO) with the use of oxygen from air, oxidation of nitrogen oxide (NO) to nitrogen dioxide (NO₂) and absorption of nitrogen oxides in water with the formation of HNO₃.

Ammonia consumption depends on the selectivity of the applied ammonia oxidation catalyst and on the process conditions. Among numerous catalysts [2–8], packages of gauzes made of noble metal alloys such as platinum and rhodium are most commonly applied in industrial practice [7–10]. Properly selected catalyst package allows to obtain ammonia conversion to main product (NO) in the range of 90–98% depending mainly on oxidation pressure [1,7,8]. Oxidation pressure has an inversely proportional effect on ammonia oxidation efficiency. In order to alleviate this effect, the temperature of reaction should be higher. However, this leads to the increased platinum losses and as a consequence, shortens the lifetime of the catalytic gauzes. For example, platinum losses are six times higher after increasing the temperature of reaction from 820 to 920 °C [1,2]. Therefore, both these aspects should be taken into account to determine the temperature of reaction.
The application of medium pressure in the oxidation unit (0.35–0.55 MPa) and high pressure in the absorption unit (0.8–1.5 MPa) is optimal for specific ammonia consumption and efficient energy use. Therefore, modern nitric acid plants are dual-pressure ones. The average pressure in the oxidation unit is a kind of trade-off between the capacity that is possible to achieve per 1 m² of catalytic gauzes, oxidation efficiency, number of gauzes in package, lifetime of gauzes and noble metals losses during exploitation [1,2].

In the context of global warming and climate changes, a very important issue related to ammonia oxidation process is the amount of the by-product formed that is nitrous oxide (N_2O). In Kyoto Protocol, N_2O was qualified as a greenhouse gas with a very high global warming potential, about 300 times higher than CO_2 [11]. At room temperature, N_2O is a colourless, non-flammable gas with a delicate pleasant smell and sweet taste [12]. Since it was isolated at the end of 17th century and because of its pain-relieving and anaesthetic properties, it has been widely applied in dentistry and surgery. Currently, due to some concerns, there is an ongoing discussion on its safe use which has the effect of decreasing the N_2O application in medicine [12,13]. At the same time, the increasing trend of its use for recreational purposes is observed. Inhaling the 'laughing gas' causes euphoria and hallucinations [13].

Microbial nitrification and de-nitrification in land and aqueous eco-systems are the natural sources of N_2O in environment. The anthropogenic sources are cultivated soils fertilized intensely with nitrogen fertilizers and industrial processes such as burning fossil fuels and biomass as well as the production of adipic acid and nitric acid with the last one being regarded as the biggest source of N_2O in the chemical industry [14,15]. Nitrous oxide formed in nitric acid plant does not undergo any conversions and it is released to atmosphere. Currently, the emission of this gas is monitored and industry is obliged to reduce it. Pursuant to BAT requirements, concentration of this gas in outlet gases cannot exceed 20–300 ppm depending on the type of nitric acid plants [16,17]. However, due to the battle against climate change and global warming, further restrictions in emission limits can be expected.

There are several methods of limiting N₂O emissions from nitric acid plants [14]. Generally, they can be classified as primary and secondary methods. Primary methods involve preventing the formation of N₂O during ammonia oxidation. They include modification of catalytic gauzes (so-called low-emissions systems) and parameters optimization of ammonia oxidation process. Secondary methods involve the removal of N₂O. At the temperature over 800 °C, thermal decomposition of N₂O occurs but the efficient decomposition requires ensuring adequately long residence time at high temperatures [14,15].

The achievement of low level of N_2O emissions requires the application of the catalytic methods such as high temperature N_2O decomposition from nitrous gases, low- or middle temperature N_2O decomposition or reduction from tail gases. High temperature method is more common. In some cases, the combination of primary method (application of modified catalytic gauzes packages and/or optimization of ammonia oxidation parameters) and high temperature N_2O decomposition ensures meeting the emission standards.

Optimization of production process requires extensive knowledge and understanding the effect of particular parameters on the process. Until recently, the most commonly applied approach of researchers to study simple and complex processes was 'one-factor-at-a-time' (*OFAT*), which is time consuming and ineffective method for processes with multiple complicated dependencies between parameters. Over the last years, mathematical and statistical methods for design of experiments and parameters optimization have been applied more frequently [18]. Because of its usability, this method is applied for the design, improvement and optimization of production processes and products [19–21]. It is a widely applied method in research of various processes [22–27] and approx. 50% of all applications is in medicine, engineering, biochemistry, physics and computer science [28]. In this method, reaction kinetics equations and process mechanism are not taken into account and they are regarded as a 'black-box' [29] (Figure 1).



Figure 1. "Black-box" model of the research issue in design of experiments methodology.

The choice of experiment plan depends mainly on the issue which is the subject matter of investigations as well as on objectives which are set. The most commonly applied experiments plans include: full or fractional factorial, Plackett-Burman, central composite, Box-Behnken and Taguchi designs.

As a result of modelling of the data obtained, empirical equations with statistically significant importance are received which describe the effect of process variables (independent variables) on the process result (response variable).

Desirability function (*DF*) can be applied in search for optimal operational parameters. The method proposed by Derringer and Suich [18] involves the construction response surface model and then finding the values of independent variables which ensure the most desirable value. The objective of the presented studies was the analysis of the impact of reactor's operational parameters on ammonia oxidation reaction. To the best of our knowledge, the approach presented here to describe ammonia oxidation process is published for the first time.

2. Results and Discussion

2.1. Design of Experiments

Ammonia oxidation reaction depends on a few process variables. In this study, the effect of the reactor's load (X_1), the temperature of nitrous gas specifying the temperature of reaction (X_2) and the number of catalytic gauzes (X_3) on ammonia oxidation reaction was investigated. The oxidation efficiency of NH₃ to NO (R_1) and N₂O concentration in nitrous gases (R_2) were selected as measures for ammonia oxidation reaction. The matrix of 15 experiments including particular levels of coded variables and achieved values of response variables R_1 and R_2 are presented in Table 1. In the regarded experimental area of independent variables, ammonia oxidation efficiency ranged from 91.4% to 96.4%, whereas N₂O concentration in nitrous gases ranged from 1011 to 1762 ppm.

Standard	Run	X_1	<i>X</i> ₂	X_3	R_1	R_2
Order	Order	Reactor's Load, kg NH ₃ /(m ² h)	Temperature, °C	No. of Gauzes, pcs	Ammonia Oxidation Efficiency, %	N ₂ O Concentration in Nitrous Gases, ppm
3	1	-1	1	0	96.1	1011
15	2	0	0	0	96.1	1279
7	3	-1	0	1	96.2	1238
6	4	1	0	-1	91.4	1620
9	5	0	-1	$^{-1}$	92.0	1762
5	6	-1	0	-1	93.7	1348
14	7	0	0	0	96.2	1265
12	8	0	1	1	96.4	1074
2	9	1	-1	0	95.6	1457
8	10	1	0	1	96.2	1312
1	11	$^{-1}$	-1	0	96.3	1423
4	12	1	1	0	95.8	1114
13	13	0	0	0	96.2	1271
11	14	0	-1	1	96.0	1506
10	15	0	1	$^{-1}$	92.7	1207

Table 1. The Box-Behnken design matrix and experimental data.

2.2. Model Fitting

The first task was to find out which equation would allow to obtain the best correlation between independent variables and responses. Analysis of Variance (ANOVA) was carried out for most frequently applied equations: linear, two-factor interaction (2FI), quadratic and cubic. Table 2 includes the summary statistics of both responses for different mathematical equations.

Response Variable: R ₁ —Ammonia Oxidation Efficiency							
Source	Std. Dev.	R^2	Adjusted R ²	Predicted R^2	PRESS		
Linear	1.08	0.6969	0.6142	0.4268	24.38		
2FI	1.20	0.7294	0.5265	-0.1215	47.70		
Quadratic	0.2790	0.9908	0.9744	0.8557	6.14	Suggested	
Cubic	0.0577	0.9998	0.9989		*	Aliased	
	Resp	onse variat	ole: R_2 —N ₂ O co	ncentration in n	itrous gases		
Source	Std. Dev.	R^2	Adjusted R ²	Predicted R^2	PRESS		
Linear	87.82	0.8524	0.8121	0.6993	1.728×10^5		
2FI	93.58	0.8781	0.7867	0.4155	3.359×10^5		
Quadratic	49.78	0.9784	0.9396	0.6574	1.969×10^{5}	Suggested	
Cubic	7.02	0.9998	0.9988		*	Aliased	

Table 2. Model summary statistics for response variables R_1 and R_2 .

* - case(s) with leverage of 1.0000; PRESS statistic not defined.

Based on the achieved results, it was found that the experimental data is described best with quadratic and cubic equations. For both responses, high values of R^2 and adjusted R^2 were achieved. The number of conducted experiments caused that the cubic model was aliased. It means that the experimental matrix contains an insufficient number of experimental points for independent estimation of all effects for these models. Therefore, quadratic equation was selected for further analysis.

The statistical significance of these equations and their particular terms was specified based on Analysis of Variance (ANOVA). Results of this analysis are presented in Tables 3 and 4, for response variables R_1 and R_2 respectively. Large *F*-value indicates that most changes of independent variable can be explained with the developed regression equation. The correlated probability *p*-value is used to estimate whether *F*-value is large enough to show statistical significance.

Source	Sum of Squares	df	Mean Square	F-Value	<i>p</i> -Value	Significance		
Model	42.14	9	4.68	60.16	0.0001	highly significant		
X_1	1.36	1	1.36	17.49	0.0086	significant		
X2	0.1512	1	0.1512	1.94	0.2221	not significant		
X_3	28.13	1	28.13	361.35	< 0.0001	highly significant		
X_1X_2	0.0400	1	0.0400	0.5139	0.5055	not significant		
X_1X_3	1.32	1	1.32	16.99	0.0092	significant		
$X_{2}X_{3}$	0.0225	1	0.0225	0.2891	0.6139	not significant		
X_{1}^{2}	0.0126	1	0.0126	0.1614	0.7044	not significant		
X_2^2	0.0926	1	0.0926	1.19	0.3252	not significant		
X_3^2	11.09	1	11.09	142.53	< 0.0001	highly significant		
Residual	0.3892	5	0.0778					
Lack of Fit	0.3825	3	0.1275	38.25	0.0256	significant		
Pure Error	0.0067	2	0.0033			Ŭ.		
Corrected total SS	42.53	14						
<i>p</i> < 0.000	p < 0.0001—highly significant, $0.0001 —significant, p > 0.05—not significant$							

Table 3. ANOVA results for response variable *R*₁.

The probability *p*-value for the achieved model of variable R_1 is 0.0001. It means that the model is statistically significant but some terms of equation are statistically not significant. Coefficients: R^2 , adjusted R^2 and predicted R^2 are very high: 0.9908, 0.9744 and 0.8577, respectively. There is also high compliance between coefficients: predicted R^2 and adjusted R^2 (difference <0.2). The achievement of statistically significant value lack of fit (0.0256) is the incompliance of this model as this parameter should be statistically not significant.

Table 4. /	ANOVA	results	for	response	variable R ₂
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Source	Sum of Squares	df	Mean Square	F-Value	<i>p</i> -Value	Significance
Model	5.623×10^5	9	62,480.28	25.21	0.0012	highly significant
X_1	29,161.12	1	29,161.12	11.77	0.0186	significant
X_2	3.793×10^{5}	1	3.793×10^{5}	153.05	< 0.0001	highly significant
X_3	81,406.13	1	81,406.13	32.85	0.0023	significant
X_1X_2	1190.25	1	1190.25	0.4803	0.5192	not significant
X_1X_3	9801.00	1	9801.00	3.95	0.1034	not significant
X_2X_3	3782.25	1	3782.25	1.53	0.2716	not significant
X_{1}^{2}	732.33	1	732.33	0.2955	0.6101	not significant
X_{2}^{2}	148.10	1	148.10	0.0598	0.8166	not significant
X_{3}^{2}	54,881.26	1	54,881.26	22.14	0.0053	significant
Residual	12,391.92	5	2478.38			Ŭ.
Lack of Fit	12,293.25	3	4097.75	83.06	0.0119	significant
Pure Error	98.67	2	49.33			Ŭ.
Corrected total SS	5.74710^5	14				
p < 0.0001—highly significant, $0.0001 —significant, p > 0.05—not significant$						

In case of response variable R_2 , the probability *p*-value (0.0012) indicates that the assumed quadratic equation is statistically significant but some of its terms are statistically not significant. High coefficients R^2 , adjusted R^2 and predicted R^2 are also achieved for the second response variable and they are: 0.9784, 0.9396 and 0.6574, respectively. However, the difference between predicted R^2 and adjusted R^2 is larger than the recommended one (>0.2). This may demonstrate a large block effect or problems with model or data. This model is also characteristic of statistically significant parameter lack of fit (p = 0.0119).

At a further stage of analysis, statistically not significant terms of initial equation were eliminated from the analysis. The reduction was made using step-by-step method (from the most insignificant term). For both these response variables, only statistically significant terms were left and higher R^2 ,

adjusted R^2 and predicted R^2 coefficients were achieved. Results of Analysis of Variance (ANOVA) are presented in Tables 5 and 6.

Source	Sum of Squares	df	Mean Square	F-Value	<i>p</i> -Value	Significance
Model	41.83	4	10.46	148.66	< 0.0001	highly significant
X1	1.36	1	1.36	19.35	0.0013	significant
X3	28.13	1	28.13	399.85	< 0.0001	highly significant
X_1X_3	1.32	1	1.32	18.80	0.0015	significant
X_{3}^{2}	11.02	1	11.02	156.63	< 0.0001	highly significant
Residual	0.7034	10	0.0703			
Lack of Fit	0.6967	8	0.0871	26.13	0.0374	significant
Pure Error	0.0067	2	0.0033			, i i i i i i i i i i i i i i i i i i i
Corrected total SS	42.53	14				
R^2	0.9835					
Adjusted R^2	0.9768					
Predicted R ²	0.9550					

Table 5. ANOVA results for reduced model of the response variable *R*₁.

The obtained mathematical model for response R_1 is highly significant (*p*-value < 0.0001). The dependence on linear terms X_1 , X_3 , interaction X_1X_3 and quadratic term X_3^2 are significant. High determination coefficients are obtained 1 ($R^2 = 0.9835$, adjusted $R^2 = 0.9768$, predicted $R^2 = 0.9550$).

The final model is presented in Equation (1).

$$R = 96.04 - 0.4125X_1 + 1.88X_3 + 0.575X_1X_3 - 1.72X_3^2 \tag{1}$$

Source	Sum of Squares	df	Mean Square	F-Value	<i>p</i> -Value	Significance
Model	5.467×10^5	4	1.367×10^5	48.81	< 0.0001	highly significant
X_1	29,161.12	1	29,161.12	10.41	0.0091	significant
X_2	3.793×10^{5}	1	$3.793 imes 10^5$	135.47	< 0.0001	highly significant
X_3	81,406.13	1	81,406.13	29.07	0.0003	significant
X_3^2	56,826.52	1	56,826.52	20.30	0.0011	significant
Residual	28,000.13	10	2800.01			Ť
Lack of Fit	27,901.46	8	3487.68	70.70	0.0140	significant
Pure Error	98.67	2	49.33			Ť
Corrected total SS	5.747×10^{5}	14				
R ²	0.9513					
Adjusted R ²	0.9318					
Predicted R ²	0.8743					

Table 6. ANOVA results for reduced model of the response variable R_2 .

The obtained mathematical model for response R_2 is highly significant (*p*-value < 0.0001). The dependence on linear terms X_1 , X_2 , X_3 and quadratic term X_3^2 are significant. High determination coefficients ($R^2 = 0.9513$, adjusted $R^2 = 0.9318$, predicted $R^2 = 0.8743$) were achieved for the model. The final model is presented in Equation (2).

$$R = 1260 + 60.37X_1 - 217.75X_2 - 100.88X_3 + 123.37X_3^2$$
(2)

2.3. Model Diagnostics

Before the process optimization, the model diagnostics for both equations was performed because of occurrence of statistically significant Lack of fit parameter. Results of model diagnostics: a normal probability of the residuals, residuals analysis and actual data versus predicted values plots were analysed.

Figure 2 presents model diagnostics for response variable R_1 , whereas Figure 3 presents model diagnostics for response variable R_2 . Normal plot of studentised residuals should be approximately a straight line, whereas studentised residuals versus predicted response values and versus run should be a random scatter. Points in plots of real response values with reference to predicted response values line up accurately along the axis at the angle of 45° .



Figure 2. Model diagnostics for response variable R_1 (**a**) A normal probability plot of the residuals; (**b**) Residuals versus predicted value of R_1 ; (**c**) Residuals versus run number; (**d**) Predicted versus actual value of response variable R_1 . Points on graphs correspond to results of particular experiments and colour points correspond to the value in accordance with the scale.

These diagnostics show that despite the fact that *Lack of fit* parameter is statistically significant, experimental and predicted points for both equations correlate well with each other.



Figure 3. Model diagnostics for response variable R_2 (**a**) A normal probability plot of the residuals; (**b**) Residuals versus predicted value of R_2 ; (**c**) Residuals versus run number; (**d**) Predicted versus actual value of response variable R_2 . Points on graphs correspond to results of particular experiments and colour points correspond to the value in accordance with the scale.

2.4. The Effect of Independent Variables

In case of response variable R_1 (ammonia oxidation efficiency), the mathematical model shows a strong linear effect of the reactor's load (X_1) and the number of catalytic gauzes (X_3) and interaction between these two variables (X_1X_3) and the quadratic term number of catalytic gauzes (X_3^2) on the achieved response variable. The temperature of reaction in the studied range does not affect the ammonia oxidation efficiency. The effect of X_1 and X_3 on response R_1 were shown as contour plot (Figure 4). According to the presented plot of variable of R_1 , a small number of catalytic gauzes causes lower ammonia oxidation efficiency for the entire range of studies reactor's load. The increase in the number of catalytic gauzes to $X_3 = 0$ causes increase in oxidation efficiency within the entire range of studied reactor's load.

Studies related to dependency of N_2O concentration in nitrous gases on operating parameters are relatively new research issue. Therefore, there is a lack of scientific reports dedicated to systematic studies in this field. In case of the N_2O concentration in nitrous gases, the achieved mathematical model demonstrates a significant effect of the selected process variables (X_1 , X_2 , X_3) and the quadratic term number of catalytic gauzes (X_3^2) on the achieved response variable. This effect is illustrated in Figures 5–7. The analysis of Equation (2) and Figures 5–7 indicates that the temperature of reaction has the biggest quantitative effect on N₂O concentration in nitrous gases. From the comparison of plots (Figure 5a–c) it can be concluded that despite the presence of statistically significant terms of equations derived from variable X_3 , plots of contour line corresponding to levels 0 and 1 are similar. Only for $X_3 = -1$, higher values of R_2 are achieved. Profiles of response variable R_2 presented in Figures 6a–c and 7a–c confirm the effect of the number of catalytic gauzes. Both these figures show that the number of catalytic gauzes has little effect on the amount of N₂O being formed. For the level of $X_3 = 0$ –0.4 (10–12 gauzes), the local optimum is observed. For this number of gauzes, increasing the reactor's load ($X_1 = 1$) at the fixed reaction temperature (Figure 6a–c) and decrease in N₂O concentration in nitrous gases.



Figure 4. Interaction effect between the reactor's load (X_1) and the number of catalytic gauzes (X_3) on ammonia oxidation efficiency (R_1) by contour plot.

2.5. Multi-Response Desirability Optimization

The major optimization task is to find the number of catalytic gauzes and the permissible reactor's load ensuring the maximization of ammonia oxidation efficiency (R_1) and the minimization of N₂O concentration in nitrous gases. Results of experiments discussed in Section 2.4. indicate that statistically, the temperature has no significant effect on ammonia oxidation efficiency but on the other hand, the amount of N₂O formed is reversely proportional to the temperature of reaction. For desirability function, it was assumed that independent variables are in the variability range. Assumptions for the optimization are presented in Table 7.

 Table 7. Assumptions for the optimization of the ammonia oxidation process using desirability function.

 Variables symbol identification according to the Table 1.

Name	Goal	Lower Limit	Upper Limit	Lower Weight *	Upper Weight *	Importance **
X_1	in range	-1	1	1	1	3
X2	in range	-1	1	1	1	3
X_3	in range	-1	1	1	1	3
R_1	maximize	91.4	96.4	1	1	3
R_2	minimize	1011	1762	1	1	3

* Weight: 1—linear change of values in the range from 0 to 1; ** Importance: 5—high, 3—medium, 1—low.

For such optimization assumptions, the area of detailed set of parameters was achieved. It confirms that optimization assumptions are met. Desirability functions for three temperature levels are presented as contour plot in Figure 8.



Figure 5. Interaction effect between the reactor's load (X_1) and the temperature (X_2) at fixed number of catalytic gauzes (X_3) on N₂O concentration in nitrous gases (R_2) by contour plot. (**a**) $X_3 = -1$; (**b**) $X_3 = 0$; (**c**) $X_3 = 1$.



Figure 6. Interaction effect between the reactor's load (X_1) and the number of catalytic gauzes (X_3) at fixed temperature (X_2) on N₂O concentration in nitrous gases (R_2) by contour plot. (a) $X_2 = -1$; (b) $X_2 = 0$; (c) $X_2 = 1$.



Figure 7. Interaction effect between the temperature (X_2) and the number of catalytic gauzes (X_3) at fixed reactor's load (X_1) on N₂O concentration in nitrous gases (R_2) by contour plot. (**a**) $X_1 = -1$; (**b**) $X_1 = 0$; (**c**) $X_1 = 1$.



Figure 8. Desirability function plots. Effect of the reactor's load (X_1) and the number of catalytic gauzes (X_3) at three levels of temperature (**a**) $X_2 = -1$; (**b**) $X_2 = 0$; (**c**) $X_2 = 1$.

High values of desirability function (DF > 0.9) at 910 °C are described with dependency according to which for the load of 456 kg NH₃/(m²h), the sufficient number of catalytic gauzes is 8. However, for the maximum load studied, 10 catalytic gauzes should be applied. At the temperature of 910 °C and when all optimization criteria are met, the expected value of N₂O concentration ranges from

1000 ppm to 1100 ppm (Figure 8a). Lowering the reaction temperature to 890 °C means that desirability function DF > 0.8 is within the region where the minimum catalyst gauzes is 9 for the loading not higher than 480 kg NH₃/(m²h) and 12 gauzes for load of 645 kg NH₃/(m²h) (Figure 8b). At this temperature, the expected concentration of N₂O in nitrous gases ranges from 1180 to 1200 ppm. At the lowest temperature (within the studied range) of 870 °C, the highest value of desirability function is 0.69. For 12 gauzes and the load of 456 kg/(m²h), the expected concentration of N₂O in nitrous gases is 1400 ppm.

Taking into account the amount of the primary emissions of N₂O (the environmental aspect), it is favourable to conduct the reaction at the temperature of 910 °C. However, this leads to the increased platinum losses. Platinum losses at 910 °C are higher by approx. 25% as compared to losses at 890 °C and by 45% as compared to losses at 870 °C [1,2]. Lowering the reaction temperature to 890 °C with maintaining the optimal range of other parameters causes the increase of N₂O concentration in nitrous gases by 100–200 ppm.

The assumption of other values of 'weight' and 'importance' for particular variables leads to obtain other profiles of desirability function. Under industrial conditions, the assumed value of 'weight' and 'importance' should take into account the process economics with regard to platinum losses.

2.6. Validation

Validation of the developed optimization model should be carried out under conditions specified as optimal. Optimization results indicate a wide set of parameters for which desirability function achieved high values. Therefore, in order to carry out additional measurements, the point with the independent variables value of: 1, 1, 1 was selected. This point is in the range of high desirability function value. In Table 8 levels of independent variables, results of validation experiment and predicted mean values of response variables with standard deviation are presented.

Independent Variable	Reactor's Load	Temperature	Number of Gauzes
Level	1	1	1
Response variable	Experimental Data	Predicted Mean Value	Std Dev
R_1	96.2	96.3625 ± 0.4672	0.265216
R_2	1120	1125.12 ± 83.37	52.9151

 Table 8.
 The assumed levels of independent variables in validation studies and predicted responses values.

For the assumed independent variables, the values of predicted mean with 95% two-sided confidence intervals met by 99% of population were estimated based on the achieved mathematical model. High conformity of results expected according to mathematical models with the obtained measurements results was achieved.

3. Materials and Methods

3.1. Materials

Standard knitted gauzes made of platinum alloy with the addition of 10% wt. Rh made of 0.076 mm wire and specific weight of 600 g/m^2 were used for ammonia oxidation studies. The catchment gauzes which are most commonly used in industrial process were not used in studies. The prefiltered compressed air and gaseous ammonia were used as raw materials for ammonia oxidation reaction.

3.2. Experimental Procedure

Ammonia oxidation studies were performed in a pilot plant equipped with a flow reactor (inner diameter: 100 mm). The Pt-Rh catalytic gauzes were installed inside the basket. After initiating the reaction, the stable ammonia and air ratio was maintained in reaction mixture amounting to approx. 10.9% vol. The air–ammonia mixture temperature was controlled in such a manner as to obtain the temperature of nitrous gas as assumed in the experiment plan. Air–ammonia mixture temperature was variable in the range of 135–195 °C. The flow of the air-ammonia mixture was also controlled in order to obtain the assumed reactor's load. All experiments were conducted under the pressure of 0.5 MPa. The range of temperature of reaction at which studies were carried out is similar to that applied in industrial practice. The reactor's load was selected in such a manner as to ensure that the gas flow through the catalytic package in the range applied for medium-pressure industrial reactor namely 1–3 m/s. The scheme of the pilot plant is presented in Figure 9. For measuring ammonia oxidation efficiency, samples of ammonia-air mixture were taken at the inlet to the reactor and samples of nitrous gases were taken.



Figure 9. Scheme of the pilot plant. Symbols: C1—absorption column, C2—bleaching column, E1–E6—heat exchangers, M1—air–ammonia mixer, R1—ammonia oxidation reactor, R2—selective catalytic reduction reactor.

3.3. Analytical Methods

The ammonia oxidation efficiency was calculated based on concentrations of ammonia in the air-ammonia mixture at the inlet and concentrations of NO in nitrous gases at the outlet of the reactor. Both analysis were determined according to a titration method. Ammonia from air-ammonia mixture samples was absorbed in water with the formation of ammonia-water solution which was then titrated with sulphuric acid. The nitrous gases samples were absorbed in 3% water solution of hydrogen peroxide. After ensuring the sufficient period of time, NO oxidized completely to NO₂ and then, it reacted with water to HNO₃. The formed HNO₃ was titrated with the sodium hydroxide solution in the presence of an indicator.

Ammonia oxidation efficiency (R_1) was calculated according to the following formula:

$$R_1 = \left(\frac{C_2}{C_1}\right) \cdot 100\% \tag{3}$$

where: C_1 —ammonia concentration in ammonia-air mixture, w w/w; C_2 —concentration of oxidized ammonia, w w/w.

The result of each measurement is an average value, calculated from 7 independent samplings. The difference in the extreme individual values were not greater than $\pm 0.3\%$ in comparison to the average one.

N₂O concentration in nitrous gases (R_2) was determined by gas chromatography using a Unicam 610 system with a discharge ionization detector. Gaseous samples were collected in the vacuum flasks containing 3% water solution of hydrogen peroxide. After the absorption of nitrous gases and water vapor condensation, exhaust gas from the flasks was injected to the gas chromatograph through 1 mL sample loop. The result of each measurement was an average value calculated from 3 independent samplings. The difference in the extreme individual values was not greater than \pm 35 ppm in comparison to the average one.

3.4. Statistical Methods

The experimental procedure was carried out according to Box–Behnken design matrix. The reactor's load (X_1), the temperature of nitrous gas specifying the temperature of reaction (X_2) and the number of catalytic gauzes (X_2) were selected as independent variables. Ammonia oxidation efficiency (R_1) and N₂O concentration in nitrous gas (R_2) were specified as response variables. Each level of independent variables were coded according to the Equation (4).

$$X_i = \frac{x_i - x_0}{\Delta x_i} \tag{4}$$

where, X_i is the dimensionless, coded level of independent variable (-1, 0 or 1), x_i is the actual value of the independent variable, x_0 is the value of the independent variable at the centre point, Δx_i is the step change in x_i .

Ranges and levels of independent variables are presented in Table 9.

	X_1	X_2	X_3	
Independent Variable:	Reactor's Load, kg NH ₃ /(m ² h)	Temperature of Reaction, °C	Number of Catalytic Gauzes, pcs	
Coded variable level:				
Low level (-1)	456	870	5	
Mid-level (0)	582	890	10	
High level (+1)	708	910	15	

Table 9. Coded and uncoded levels of independent variables used in experiments.

The total number of the experiments (*N*) was calculated using the Equation (5).

$$N = 2k(k-1) + c_0 \tag{5}$$

where, *k* is the number of independent variables, c_0 —number of the replicates run of the centre point (in our research $c_0 = 3$). For three independent variables, the total number of experiments assumed in the plan was 15.

The experiments were conducted in a randomized order to avoid the influence of uncontrolled variables on the dependent responses.

A mathematical relationship between the independent variables and response variables was determined by fitting the experimental data with second-order polynomial Equation (6).

$$R_i = b_0 + \sum_{i=1}^3 b_i X_i + \sum_{i=1}^3 \sum_{j=1, i < j}^3 b_{ij} X_i X_j + \sum_{i=1}^3 b_{ii} X_i^2$$
(6)

where R_i is the estimate response variable, b_0 , b_i , b_{ii} , b_{ij} are regression coefficients fitted from the experimental data, X_i , X_i are coded independent variables, listed in Table 9.

The significance of the model equation, individual parameters, were evaluated through ANOVA with the confidence interval (CI) of 95%. A simultaneous optimization of several dependent variables requires the application of multi-criteria methodology. In this case, the desirability function (*DF*) was used. The particular desirability functions are combined using the geometric mean which allows to achieve overall desirability function [16], according to Equation (7).

$$DF = \left((d_1)^{w_1} \times (d_2)^{w_2} \times \dots \times (d_n)^{w_n} \right)^{1/\sum w_i}$$
(7)

where *n* is the number of responses, d_i is an individual response desirability, w_i is a response 'weight'.

The adjustment of the shape of particular desirability function can be performed by assigning the specified 'weight.' Setting a different 'importance' for each objective with respect to the remaining objectives is also possible. For these studies, identical 'weight' for all the independent variables and response variables was assumed. Desirability function assigns values from 0 to 1 where 1 means meeting all the optimization criteria. It is not always necessary to search for the solution aiming at achievement of the highest value of desirability function but it is vital to search for the set of parameters which would meet the optimization objectives to the particular extent (e.g., DF > 0.75). The statistical software used to experimental design and analysis was Design Expert 11.0.6.0 version (Stat-Ease, Inc., Minneapolis, MN, USA).

4. Conclusions

The conducted studies allowed us to develop statistically significant mathematical models describing the course of variables of ammonia oxidation efficiency and N_2O concentration in nitrous gases depending on three selected independent variables.

The design of the experiment allowed the reduction of the costs of studies and to achieve a number of results accurate for modelling. It was found that, within the studied range of variability, the temperature of reaction has no significant effect statistically on the achieved ammonia oxidation efficiency, whereas it has the effect on the amount of N_2O formed in the side reaction (primary emission of N_2O).

The developed models were used to optimize the process. As a result of this optimization, the set of the independent variables was developed for which optimization assumptions are met, which are expressed as a high value of desirability functions. It is possible to specify the optimum number of gauzes with the determined reactor's load for the studied package of catalytic gauzes.

In validation experiments, the developed model of desirability function achieved the high conformity of experimental values with the expected ones.

The presented methodology can be used to minimize the primary N_2O emission at high ammonia oxidation efficiency. It can be applied for optimization of operating parameters of ammonia oxidation reactor with two types of catalysts: catalytic gauzes and catalyst for high temperature of N_2O decomposition. As a result, it is possible to obtain the set of independent variables ensuring low N_2O emission and to meet the binding environmental regulations.

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Tuning the Catalytic Properties of Copper-Promoted Nanoceria via a Hydrothermal Method

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Abstract: Copper-cerium mixed oxide catalysts have gained ground over the years in the field of heterogeneous catalysis and especially in CO oxidation reaction due to their remarkable performance. In this study, a series of highly active, atomically dispersed copper-ceria nanocatalysts were synthesized via appropriate tuning of a novel hydrothermal method. Various physicochemical techniques including electron paramagnetic resonance (EPR) spectroscopy, X-ray diffraction (XRD), N₂ adsorption, scanning electron microscopy (SEM), Raman spectroscopy, and ultraviolet-visible diffuse reflectance spectroscopy (UV-Vis DRS) were employed in the characterization of the synthesized materials, while all the catalysts were evaluated in the CO oxidation reaction. Moreover, discussion of the employed mechanism during hydrothermal route was provided. The observed catalytic activity in CO oxidation reaction was strongly dependent on the nanostructured morphology, oxygen vacancy concentration, and nature of atomically dispersed Cu²⁺ clusters.

Keywords: copper-ceria catalysts; hydrothermal method; CO oxidation; copper clusters; nanoceria

1. Introduction

Carbon monoxide (CO) is a harmful, toxic gas that is present in many industrial processes. Due to its negative impact for both humans and the environment, the catalytic oxidation of CO into CO_2 has always been a research topic of great interest [1–3]. Moreover, the catalytic oxidation of CO is an important reaction in the technological fields of fuel cells [4–6], gas sensors [7], and CO_2 lasers [8].

Cerium oxide or ceria (CeO₂) has been thoroughly studied as a catalyst or support in CO oxidation reaction due to its defective structure enriched with oxygen vacancies and high oxygen storage capacity (OSC) resulting from the interaction between Ce³⁺ and Ce⁴⁺ [9–12]. In the case of ceria synthesized in a nanosized form, more remarkable functions can be obtained due to the nanosize effects. For this reason, research has focused on the understanding of the properties of nanoceria as well as improving its OSC, surface to volume ratio, and redox properties [13–16]. Computational studies have shown that the catalytic activity of nanoceria is strongly associated with the exposed surface plane. Sayle et al. [17] predicted that the (110) and (100) surfaces are catalytically more active for CO oxidation than the (111) surface, due to more oxygen vacancies located in the former. According to Conesa [18], the formation of oxygen vacancies on the (110) surfaces requires less energy than the (111) surface. It is noteworthy that the exposition of the reactive surface plane is dependent on the morphology of the material at the nanoscale. Zhou et al. [19] have shown that CeO₂ nanorods, which exposed the (110)

and (100) planes achieved higher catalytic activity for CO oxidation than nanoparticles exposing the (111) planes. Wu et al. [20] studied the morphology dependence of CO oxidation over ceria nanocrystals. They discovered that the activity for CO oxidation of those CeO_2 nanostructures follows the order: Rods > cubes > octahedra, whereas the activity of different planes follows the order: (110) > (100) > (111). These results were also confirmed by Tana et al. [21]. In order to prepare various shapes of nanoceria, a number of methods have been applied, such as sol-gel [22], precipitation [23], hydrothermal or solvothermal methods [24–27], and electrochemical deposition [28]. Among these methods, the hydrothermal method has attracted great interest because a desired morphology can be obtained via appropriate control of the hydrothermal parameters such as reaction time, temperature, and concentration [29–31].

Despite the attractive physicochemical properties of ceria, poor catalytic activity of pure ceria [32] can by highly promoted via doping with a series of metal ions, in order to change its surface chemistry and promote the active oxygen content [33]. It is well known that the reduction behavior of ceria can be rapidly altered by the addition of a minimal amount of Au, Pd, and Pt precious metals and/or transition metals [34–38]. While the activity of the catalysts is improved by the addition of precious metals, their high cost prohibits their application. Numerous reports have indicated that the activity of ceria in oxidation reactions is enhanced by transition metals like copper. The copper-ceria system presents a cost-effective material with unique catalytic properties, comparable to noble metal catalysts, in many catalytic reactions and especially in the CO oxidation and in the preferential oxidation of CO in excess of hydrogen [39–44]. The superiority of CuCeO_x catalytic system has been attributed to a synergistic effect. Reports on the mechanism of CO oxidation reaction over these catalysts have demonstrated the significance of both copper and ceria species in the adsorption of CO and CO₂ production, as the former takes place in the copper-ceria interface [45]. Particularly, the main reasons that trigger the highly catalytic performance of these Cu-Ce catalysts are the large amount of well-dispersed copper species in the ceria support, the creation of oxygen vacancies due to incorporation of Cu²⁺ ions into the ceria structure and the presence of high concentration of active lattice oxygen [45-47]. Several Cu²⁺ entities (e.g., amorphous clusters, isolated ions, dimers, and discrete crystallites) have been detected, which can take part in the catalytic mechanism, displaying high levels of activity [5,39,48–50]. In order to form these entities, conventional preparation methods have been used such as the deposition of Cu²⁻ species onto pre-synthesized ceria support or the coprecipitation of Cu and Ce precursors. The obtained materials are calcined at high temperatures, which enable the dispersion of copper species and the chemical bonding between the Cu and Ce components. For example, Harrison et al. [39] prepared CuO/CeO₂ catalysts via coprecipitation and impregnation routes, and tested in CO oxidation. After thermal pretreatment of materials at 400 °C, the copper content on the surface of ceria consisted of amorphous clusters of Cu^{2+} ions, which presented high catalytic activity for CO oxidation. In several cases, the high temperatures required for the activation of copper component (>600 °C for 4 h) may lead to particle sintering and phase segregation, which facilitates the formation of tenorite particles of CuO, which are inactive for CO oxidation [51].

In the present work, copper clusters were atomically dispersed in ceria nanostructures via a novel hydrothermal route, yielding highly active catalysts in CO oxidation reaction. Tuning of the physicochemical and catalytic properties was studied by varying the hydrothermal parameters (temperature and concentration). A set of analytical techniques such as electron paramagnetic resonance (EPR) spectroscopy, X-ray diffraction (XRD), N₂ adsorption-desorption, scanning electron microscopy (SEM), Raman spectroscopy, and ultraviolet-visible diffuse reflectance spectroscopy (UV-Vis DRS) was used to assess the physicochemical characteristics of the materials and correlate with the catalytic performance.

2. Results

2.1. EPR Measurements

Figure 1 illustrates the EPR spectra of the hydrothermally prepared ceria-based catalysts, using various concentrations of NaOH. The concentrations of the detected species are shown in Table 1. In general, the EPR spectra present three characteristic signals; A signal at ca. 1580 Gauss (g = 4) is attributed to Ce³⁺ ions in high spin (S = 5/2) state, corresponding to reduced Ce³⁺ ceria centers located inside the lattice of ceria particles [15,52]. The sharp signal at around 3400 Gauss (visualized better in the right column spectra of Figure 1) is ascribed to [Ce³⁺-O⁻-Ce⁴⁺] (S = 1/2) units localized on the surface of the ceria particles [5,53,54]. According to the geometrical disposition of the two types of Ce³⁺ centers—lattice and surface—it is considered that surface Ce³⁺ ions can be correlated with high catalytic activity, i.e., since they may result to the formation of oxygen vacancies [55–57]. Regarding the Cu²⁺ signals, the EPR signals of the sample treated at 180 °C show a strong signal at 2600–3300 Gauss, which is attributed to Cu²⁺ (S = 1/2, I = 3/2) ions. The line shape of this EPR spectrum indicates that the copper atoms show dipolar Cu-Cu interactions, i.e., the Cu²⁺ is clustered within 10 ± 3 Angstroms from each other [58]. The g-and A values of Cu²⁺, (g₁ = 2.035, g_{1/2} = 2.305, A₁ = 15, A_{1/2}=150) suggest that the Cu²⁺ ions are located in the octahedral sites in ceria with a tetragonal distortion [39,49,55,56,59–61].

Table 1. Concentrations of the species as detected by EPR spectroscopy.

Samples	Bulk Ce ³⁺ (μ M/gr)	Surface Ce ³⁺ (µM/gr)	Cu ²⁺ Species (wt %)
Ce-120-0	0.43	0.1	N.D.
CuCe-120-0.05	0.50	1.3	0.20 (isolated ions)
Ce-120-0.1	0.40	0.4	N.D.
CuCe-120-0.5	0.50	4.0	0.002 (isolated ions)
CuCe-120-1	0.29	4.0	0.02 (isolated ions)
Ce-120-5	0.50	4.0	N.D.
CuCe-150-0	1.00	1.3	0.15 (isolated ions)
CuCe-150-0.05	0.20	1.1	0.3 (>95% clusters)
CuCe-150-0.1	0.40	0.8	3.9 (>95% clusters)
Ce-150-0.5	0.50	3.0	N.D.
CuCe-150-1	0.27	4.0	0.002 (isolated ions)
Ce-150-5	0.50	4.0	N.D.
CuCe-180-0	0.02	< 0.05	2.1 (>95% clusters)
CuCe-180-0.05	N.D. ¹	< 0.05	2.8 (>95% clusters)
CuCe-180-0.1	0.40	0.1	1.4 (>95% clusters)
CuCe-180-0.5	0.50	3.0	0.02 (isolated ions)
CuCe-180-1	0.25	3.0	0.01 (isolated ions)
Ce-180-5	0.40	4.0	N.D.

¹ Not-detected.



Figure 1. Cont.



Figure 1. Cont.



Figure 1. EPR spectra of the catalysts prepared with 0–5 M NaOH. Magnification of the region in which the bulk Ce^{3+} ions were detected, is shown in the inset figure; the figures on the right side present an enlarged region where the Cu^{2+} species and the surface Ce^{3+} ions were detected.

Quantitate data on the Ce³⁺ and Cu²⁺, estimated from the EPR spectra, are listed in Table 1. Copper loading was also confirmed via XRF measurements (within an experimental error of $\pm 10\%$; not shown here), in line with EPR results. Concerning the surface Ce³⁺ ions, there is a clear trend vs. the NaOH concentration: Their concentrations increases when concentrated solutions of sodium hydroxide were employed in the hydrothermal route. With regard to the copper entities, the combination of low concentrations of NaOH and high hydrothermal temperatures caused the formation of a high amount of copper species (especially Cu²⁺ clusters). On the other hand, the high concentrations of NaOH halted the dispersion of copper species in the ceria phase, due to the high basicity of NaOH resulting in Cu(OH)_n clusters' formation in the CeO₂ phase.

2.2. XRD Measurements

The XRD diffractograms of all the samples are shown in Figure S1. Noticeably, all peaks can be indexed to (111), (200), (220), (331), (222), (400), (331), (420), and (422) planes corresponding to the pure cubic phase [space group: Fm-3m, JCPDS: 00-043-1002, $\alpha = 0.54113$ nm] of CeO₂ [30]. No diffraction peaks of crystalline copper species can be observed, due to the presence of highly dispersed amorphous copper species or/and the low copper loading, which can be hardly detected at XRD [62,63]. However, as it was discussed in EPR section, the copper species might be atomically dispersed in the ceria matrix. The calculated average crystallite size and the lattice parameter of the ceria-based materials

are presented in Table 2. It has to be noted that there is no clear relationship between Cu content (see Table 1) and crystallite size or lattice parameter of the catalysts.

Samples	d ₁₁₁ (nm)	α (nm)
Ce-120-0	8.6	0.54078
CuCe-120-0.05	6.9	0.54228
Ce-120-0.1	10.0	0.54136
CuCe-120-0.5	13.2	0.54118
CuCe-120-1	13.8	0.54123
Ce-120-5	8.2	0.54185
CuCe-150-0	7.2	0.53966
CuCe-150-0.05	7.5	0.54147
CuCe-150-0.1	19.0	0.54211
Ce-150-0.5	9.5	0.54164
CuCe-150-1	15.2	0.54127
Ce-150-5	25.3	0.54133
CuCe-180-0	37.7	0.54168
CuCe-180-0.05	29.2	0.54168
CuCe-180-0.1	18.2	0.54142
CuCe-180-0.5	11.3	0.54134
CuCe-180-1	13.9	0.54138
Ce-180-5	23.5	0.54159

Table 2. Average crystallite size and lattice parameter of the catalysts based on the crystal plane of (111) of CeO₂.

The dependence of the crystallite size from the hydrothermal parameters is depicted in Figure 2. It can be seen in Figure 2a that, for concentrations of NaOH ≤ 0.05 M, small-size crystallites were formed at temperatures of 120–150 °C, while at 180 °C, a dramatic increase of their size can be observed. A more rapid increase of crystallites size was obtained in the case of 0.1 M NaOH. For instance, the sample Ce-120-0.1 presents an average crystallite size at 10 nm, while at 150 °C, the average size is ca. 20 nm. In the range of 0.5–1 M NaOH, the crystallites size varies from ~10 to 15 nm following an increase in the hydrothermal temperature. Finally, for 5 M NaOH, the combination of high concentration and high temperatures caused the formation of large crystallites. In Figure 2b, at 120 and 150 °C, a fluctuation of the crystallite size is observed for concentrations ≤ 0.5 M. Further increase of the concentration caused the rise of the crystallites size.



Figure 2. Dependence of the crystallites size of the catalysts with respect to the hydrothermal parameters: (a) Based on the temperature of the hydrothermal treatment; and (b) based on the concentration of NaOH.

Concerning the lattice parameter, the vast majority of the samples present higher values of lattice parameter than the pure ceria, suggesting the lattice expansion for the obtained materials. According to several studies, the lattice expansion is closely correlated with the presence of Ce^{3+} ions in the crystal lattice because the radius of Ce^{3+} ions (0.114 nm) is higher than the radius of Ce^{4+} ions (0.097 nm), inducing the lattice expansion [27,64–66]. On the other hand, only two samples (Ce-120-0 and CuCe-150-0) presented a smaller value of lattice parameter than the pure ceria. According to Pan et al. [30], the hydroxyl groups may stabilize the smaller nanoparticles resulting in the smaller value of lattice parameter.

2.3. N₂ Adsorption Measurements

The N₂ adsorption/desorption isotherms and the pore size distribution for the ceria-based materials are illustrated in Figure 3. Additionally, the specific surface areas (SSA), the pore volume and the pore size of all samples are shown in Table 3. For hydrothermal solutions of NaOH \leq 0.1 M, the materials present type II isotherms with type H2 hysteresis loops, independently of the hydrothermal temperature. Materials that present type H2 hysteresis loops are often disordered without well-defined pore distribution. However, for NaOH concentrations \geq 0.5 M, it can observed that the type of isotherm becomes type IV, which is characteristic of mesoporous materials [67] with type H2 hysteresis loops. It is worth mentioning that a different type of hysteresis loops is revealed for Ce-150-5 and Ce-180-5 samples. Specifically, the type of hysteresis loop is type H1, which is associated with well-defined cylindrical pores.

Samples	SSA (m ² /g)	Pore Volume (cm ³ /g)	Pore Size (nm)
Ce-120-0	36.3	0.0860	10.34
CuCe-120-0.05	67.0	0.0920	5.33
Ce-120-0.1	22.2	0.0420	10.02
CuCe-120-0.5	70.6	0.1286	5.17
CuCe-120-1	78.8	0.1447	5.63
Ce-120-5	137.1	0.2294	5.55
CuCe-150-0	63.7	0.1070	5.51
CuCe-150-0.05	73.3	0.1050	4.80
CuCe-150-0.1	40.4	0.0320	5.23
Ce-150-0.5	86.2	0.1462	4.97
CuCe-150-1	58.6	0.1390	5.88
Ce-150-5	37.3	0.2114	18.06
CuCe-180-0	9.2	0.0450	15.30
CuCe-180-0.05	14.8	0.0530	12.30
CuCe-180-0.1	27.8	0.0730	10.00
CuCe-180-0.5	76.5	0.1437	5.50
CuCe-180-1	59.4	0.1343	6.50
Ce-180-5	35.0	0.1952	16.80

Table 3. Specific Surface Area, Pore Volume, and Pore Size of the catalysts.

The pore size distribution diagrams (Figure 3) denote that these ceria-based catalysts have not got a well-defined pore size. Indeed, the combination of hydrothermal parameters affected this distribution. Pores ranging in the mesoporous region were formed in the case of NaOH concentration \geq 0.5 M, while different distributions can be observed with NaOH concentrations \leq 0.1 M. For example, the CuCe-150-0 catalyst mainly presents mesopores, whereas the CuCe-180-0 sample mainly consists of macroporous.

The effect of the hydrothermal parameters on the specific surface area (SSA) of the obtained materials is illustrated in Figure 4. It can be seen in Figure 4a that an increase of the hydrothermal temperature from 120 to 150 °C resulted in higher SSA when the NaOH concentration was ≤ 0.5 M. Further increase of the temperature lowered the surface area of the catalysts. On the other hand, for higher NaOH concentration (≥ 1 M), the highest surface area was obtained at 120 °C. The highlight of

this trend was the SSA of the Ce-120-5 catalyst (137.1 m² g⁻¹). A general trend depicted in Figure 4b, suggests that for elevated concentration of NaOH, higher surface areas can be obtained for the catalysts prepared hydrothermally at 120 °C. A similar trend can be seen at 150 and 180 °C, however a maximum of SSA corresponds to NaOH concentration of 0.5 M. Therefore, it can be concluded that the combination of high concentrations and high hydrothermal temperatures favors the formation of catalysts with poor surface area.



Figure 3. N₂ adsorption/desorption isotherms and pore size distribution curves of CeO₂ and Cu/CeO₂ catalysts: (**a**,**c**,**e**) N₂ adsorption/desorption isotherms; and (**b**,**d**,**f**) pore size distribution curves.



Figure 4. Dependence of the specific surface area of the catalysts with respect to the hydrothermal parameters: (a) Based on the temperature of the hydrothermal treatment; and (b) based on the concentration of NaOH.

2.4. SEM Measurements

Figure 5 illustrates representatives SEM images of the materials which were hydrothermally synthesized at 120 °C. A morphology of bulk rods with various aggregates onto their surface was obtained for the Ce-120-0 sample. In the presence of NaOH, the morphology changed and spheres with a diameter of 4-8 µm were formed for the CuCe-120-0.05 sample. Spherical aggregates with a size of few hundred nanometers were formed onto this material surface. Further increases of the NaOH concentration resulted in rods with various lengths (2–10 µm) for the Ce-120-0.1 sample. No particles or aggregates can be observed on the surface of these rods, a fact that confirms the high crystallinity of the rods. These rods disappeared at elevated concentrations of NaOH, and very big aggregates with non-defined morphology were formed. In the case of hydrothermal synthesis at 150 °C, a spherical morphology dominates (Figure 6), while at high NaOH concentrations (≥ 0.5 M), bulky aggregates with particles without well-defined geometry, were formed. Figure 7 illustrates representative SEM images of the materials, which were treated hydrothermally at 180 °C. The CuCe-180-0 sample maintained the spherical morphology. The spheres composed of particles without well-defined geometry with size of ca. 70 nm. Adding low amounts of NaOH (0.05 M) resulted in a mixed morphology with rods of few micrometers, spheres and to a smaller extent sheets. All these structures contained irregular particles with an average size of a few dozen nanometers. The CuCe-180-0.1 sample appears to possess a spherical morphology, while similar structures with the lower hydrothermal temperatures were obtained at elevated NaOH concentrations (≥ 0.5 M).

2.5. Raman Measurements

Raman spectra of all the catalysts are shown in Figure 8. Ceria presents a fluorite structure with only one allowed Raman mode, which has an F_{2g} symmetry and can be viewed to the symmetrical stretching mode of oxygen ions around Ce⁴⁺ ions [68–70]. For bulk ceria, this band appears at 465 cm⁻¹. However, a shift to lower frequencies (Table 4), together with a non-linear linewidth, can be viewed for all the samples. According to Spanier et al. [71], a large number of factors can contribute to the changes in the Raman peak position and linewidth of the 465 cm⁻¹ peak, including phonon confinement, broadening associated with size distribution, defects, strain, and variations in phonon relaxation as a function of particle size. Apart from this main band, several other bands can be clearly distinguished in the corresponding spectra. The band at ca. 265 cm⁻¹ is attributed to the tetrahedral displacement of oxygen from the ideal fluorite lattice [25,69,72]. The band at ca. 600 cm⁻¹ is assigned to the defect-induced mode (D), related to the presence of lattice defects, mostly oxygen vacancies [5,42,68,73]. Noticeably,

the samples synthesized at 150 °C with low concentration of NaOH (CuCe-150-0, CuCe-150-0.05, and CuCe-150-0.1) exhibit one additional band at ca. 830 cm⁻¹. According to Choi et al. [72], this band is associated with peroxo-like oxygen species adsorbed on the oxygen vacancies, in close relation with reduced ceria species. No separated copper phase (CuO) can be confirmed via the appearance of the corresponding Raman peaks at ca. 295 cm⁻¹ and ca. 350 cm⁻¹. Therefore, both the XRD and Raman results clearly indicate that Cu species onto ceria are highly dispersed.



Figure 5. Catalysts prepared hydrothermally at 120 °C: (a) Ce-120-0, (b) CuCe-120-0.05, (c) Ce-120-0.1, (d) CuCe-120-0.5, (e) CuCe-120-1, and (f) Ce-120-5.



Figure 6. Catalysts prepared hydrothermally at 150 °C: (a) CuCe-150-0, (b) CuCe-150-0.05, (c) CuCe-150-0.1, (d) Ce-150-0.5, (e) CuCe-150-1, and (f) Ce-150-5.

The position and the FWHM for the F_{2g} band, and the relative intensity ratio of I_D/I_{F2g} for the ceria-based catalysts are shown in Table 4. In general, the ratio of I_D/I_{F2g} , where I_{F2g} and I_D correspond to the maximum intensity of F_{2g} and D bands, respectively, can roughly reflect the amount of lattice defects (oxygen vacancies) in the obtained materials [72,74,75]. With appropriate tuning of the employed hydrothermal parameters, materials with high concentration of oxygen vacancies are obtained. For instance, the Ce-120-5 sample presents the highest value of this ratio, i.e., 0.106, suggesting a material with high perspectives in catalytic CO oxidation reaction. According to previous studies, the FWHM of the F_{2g} band is influenced to a great extent by the crystallite size of ceria and the concentration of oxygen vacancies [75–77]. However, there is no specific trend in this work and mainly the high FWHM may be closely correlated with the high concentration of oxygen vacancies, since more factors might also play a role in these features.



Figure 7. Catalysts prepared hydrothermally at 180 °C: (**a**) CuCe-180-0, (**b**) CuCe-180-0.05, (**c**) CuCe-180-0.1, (**d**) CuCe-180-0.5, (**e**) CuCe-180-1, and (**f**) Ce-180-5.



Figure 8. Raman spectra of the catalysts synthesized hydrothermally (a) 120, (b) 150, and (c) 180 °C.

Table 4. F_{2g} peak position, full width at half maximum (FWHM) for the F_{2g} band, relative intensity ratio of I_D/I_{F2g} , and energy band gap (E_g) of the catalysts.

Samples	F_{2g} (cm ⁻¹)	FWHM (cm $^{-1}$)	I _D /I _{F2g}	Eg (eV)
Ce-120-0	463.0	14.8	0.027	3.05
CuCe-120-0.05	462.3	22.8	0.046	3.03
Ce-120-0.1	463.1	17.7	0.023	3.13
CuCe-120-0.5	463.0	13.3	0.035	3.18
CuCe-120-1	463.0	12.5	0.060	3.24
Ce-120-5	462.3	36.5	0.106	3.12
CuCe-150-0	463.0	24.2	0.031	3.12
CuCe-150-0.05	462.3	26.7	0.050	3.06
CuCe-150-0.1	460.7	36.0	0.046	3.07
Ce-150-0.5	463.8	17.6	0.032	3.20
CuCe-150-1	464.6	12.7	0.014	3.28
Ce-150-5	464.6	17.0	0.046	3.29
CuCe-180-0	463.7	22.8	0.047	3.11
CuCe-180-0.05	463.0	23.2	0.080	3.00
CuCe-180-0.1	461.4	28.1	0.035	3.21
CuCe-180-0.5	463.7	15.8	0.038	3.20
CuCe-180-1	464.5	13.1	0.020	3.25
Ce-180-5	463.8	14.8	0.034	3.29

2.6. UV-Vis DRS Measurements

Figure S2 presents the UV-Vis DRS spectra of the materials. Each spectrum indicates three different bands, which according to the literature correspond to different electronic transitions between cerium and oxygen ions. The band at ca. 360 nm is ascribed to the $O^{2-} \rightarrow Ce^{4+}$ interband-transfer transition. The bands at ca. 270 and ca. 230 nm are attributed to the $O^{2-} \rightarrow Ce^{4+}$ and $O^{2-} \rightarrow Ce^{3+}$ charge-transition, respectively [25,78]. Interestingly, the CuCe-180-0, CuCe-180-0.05, CuCe-180-0.1, and CuCe-150-0.1 samples present one extra wide band at 670 nm. It is important to mention that there is a lot of controversy about this band. According to Rakai et al. [79], this band is correlated with a surface redox couple of cerium ions (Ce³⁺/Ce⁴⁺). A similar band was found from Bensalem et al. [80] and Binet et al. [81]. However, there is a number of studies which ascribe this band to d-d transitions of Cu²⁺ in an octahedral environment [2,82,83]. It also has to be noted that, for the CuCe-150-0.1, CuCe-180-0.1, CuCe-180-0.5, and CuCe-180-0.1, catalysts a shoulder appears in the spectra at ca. 400–500 nm. This feature represents the charge transfer between "Support — Oxygen—Active Phase", suggesting strong interactions between the copper species and ceria [43,44].

The energy band gap (Eg) of the catalysts, calculated via the Tauc plots method, is summarized in Table 4. Taking into account the reference value for the bulk CeO₂ as a direct band gap semiconductor (Eg = 3.19 eV), the obtained values of the ceria-based materials, especially the ones synthesized with low concentrations of NaOH (≤ 0.1 M), are smaller than the energy band gap of bulk ceria. On the other hand, the catalysts prepared with higher concentrations of NaOH (≥ 0.5 M) resulted in a higher value of Eg. The former values indicate the presence of defects and especially oxygen vacancies [26,84], while the latter values of Eg are closely related to the quantum confinement effect [77,85].

2.7. Formation Mechanism

Over the years, the utilization of organic additives such as polyvinylpyrrolidone (PVP) [86], cetyltrimethylammonium bromide (CTAB) [87] and oleic acid [88] has been established in the hydrothermal method in order to obtain well-defined particles and morphologies. However, due to adsorption effects on the surface of particles these additives lead to quenching because of their high-energy vibration [89]. Compared with the above surfactants, citric acid (CA) presents weaker morphology control ability. Therefore, when the citric acid is used as a chelating agent, products with various morphologies and sizes can be obtained [90,91].

According to Levien [92], citric acid separates into different ionic species:

$$H_{3}cit (citric acid) \rightarrow H^{+} + H_{2}cit^{-}, (pK_{\alpha} = 3.2),$$
(1)

$$H_2 cit^- \to H^+ + Hcit^{2-}, (pK_n = 4.9),$$
 (2)

$$\text{Hcit}^{2-} \to \text{H}^+ + \text{cit}^{3-}, \ (\text{pK}_{\alpha} = 6.4),$$
 (3)

The concentrations of these ionic species are closely depended on the pH value. In an acid solution (ca. pH = 2) the citric acid is not effectively dissociated to citric ions. Further increase of pH at values of 4, 6, and 8, the main citric species are the H₂cit⁻, Hcit²⁻ and cit³⁻, respectively. The citric ions can form complexes with the cerium ions, which are depended on the pH of solution [93].

The pH values from each step of synthesis route are illustrated in Table 5. It can be observed that the low concentrations of NaOH (≤ 0.1 M) did not increase the pH of the final solution (acid solution). As a result, the citric acid was not effectively dissociated, as there were H₃cit and H₂cit⁻ species in the final solution. Concerning the partial dissociation of citric acid, it is not possible to form several and stable complexes with the cerium ions under ambient conditions. On the other hand, the high concentrations of NaOH (≥ 0.5 M) caused the formation of an alkaline solution. Under these conditions, the citric acid has been completely dissociated and can form several and stable complexes with the cerium ions.

Samples	pH (Ce ³⁺) ²	pH (CA) ³	pH (Ce ³⁺ + CA) ⁴	pH (NaOH) ⁵	pH Total ⁶
0 M ¹	3.75	1.56	0.95	0	1.68
0.05 M	3.75	1.56	0.95	12.76	2.10
0.1 M	3.75	1.56	0.95	13.03	2.52
0.5 M	3.75	1.56	0.95	13.35	13.11
1 M	3.75	1.56	0.95	13.51	13.46
5 M	3.75	1.56	0.95	13.91	13.86

Table 5. pH values from the different part of the catalyst's preparation.

¹ The sample 0M is referred to the sample that no addition of NaOH is occurred and so on; ² pH of the solution that contains Ce^{3+} ions; ³ pH of the solution that contains citric ions; ⁴ pH during the mixing of the previous solutions; ⁵ pH of the NaOH solution; and ⁶ pH of the final solution before the hydrothermal treatment.

It is believed that the increase of the temperature (from room temperature to the desired hydrothermal temperature) ensures the complete complexation of citric ions with the cerium ions [93–95], while the formed complex is polymerized and becomes stable in the solution. As the hydrothermal treatment proceeds the elevated pressure and temperature triggers the appearance of two events:

- The weakening of the complex stability and precipitation of the latter as hydroxide into the solution.
- Depending on the hydrothermal parameters, the leaching and migration of copper species from the copper ring to the solution where their deposition in the ceria phase is taking place.

The final catalyst is obtained after the steps of filtration, drying, and calcination.

Concerning the various morphologies which are illustrated in Figures 5–7, it is proposed that the successful combination of hydrothermal parameters (temperature and concentration of NaOH) is the main reason behind these morphologies. It has been mentioned in Section 2.4 that the combination of low concentrations (≤ 0.1 M) and hydrothermal temperatures caused the formation spheres and/or rods. Given that the pH of the solution at these specific concentration was acidic, Ostwald Ripening seems to be the most dominant particle's formation mechanism [96–98]. Once the particles have been formed, a process of self-organization starts to happen, which results in a spherical and/or rod-like morphology depending on the hydrothermal temperatures resulted in the formation of bulky aggregates consisting of particles with a spherical and/or irregular geometry. At these conditions, the pH of the solution was alkaline and so the proposed formation mechanism of particles appears to be the oriented attachment [87,99]. Once the formation of particles has been completed, self-organization process initiates, but simultaneously the excess of OH⁻ groups and the hydrothermal temperatures cause the formation of bulky aggregates.

2.8. CO Oxidation Catalytic Studies

The conversion of CO to CO₂ as a function of the temperature of the reaction over the ceria-based catalysts is shown in Figure 9. As a general trend, it can be commented that the catalysts prepared with low concentrations of NaOH (\leq 0.1 M) and contained Cu²⁺ species onto their surface, presented better catalytic behavior than the catalysts which were hydrothermally treated with high concentrations of NaOH (\geq 0.5 M). Noticeably, the latter catalysts illustrated a significant amount of surface Ce³⁺ ions (see Table 1), which is usually correlated with high catalytic activity. However, this is not the case in this work and other factors control the catalytic activity, as will be discussed in the next section.

An indicator of the catalytic activity behavior are the temperatures where 50% and 90% CO conversion is achieved (T_{50} and T_{90} , respectively) (Figure 9). For the samples treated hydrothermally at 120 °C, it can be seen that the most active catalysts are the CuCe-120-0.05 and the Ce-120-0, presenting T_{50} equal to 194 °C and 224 °C, respectively, and T_{90} equal to 272 °C and 263 °C, respectively. Additionally, the above samples achieved full removal of CO. The less active sample was the Ce-120-0.1, which revealed a T_{50} at 306 °C and T_{90} at 351 °C. One possible explanation for the poor catalytic

activity of this sample is its morphology, which presented rods with high crystallinity. Increasing the concentration of NaOH (≥ 0.5 M), the catalytic activity of the samples was improved, but it cannot be comparable with the activities of the CuCe-120-0.05 and Ce-120-0. The morphology of the samples prepared with high concentrations of NaOH (see Figure 5) seems to influence the catalytic activity in a negative way. Moreover, although the sample Ce-120-5 illustrated the highest SSA among all catalysts (see Table 3), this is not related to the achievement of high catalytic activity.



Figure 9. (a,c,e) CO conversion diagrams of the catalysts which were treated hydrothermally at 120, 150 and 180 °C, respectively; and (b,d,f) Bar graphs which are shown the temperatures where 50% and 90% conversion of CO to CO_2 is achieved.

The catalytic activity of the samples synthesized at 150 °C, was dramatically improved, especially for NaOH concentrations ≤ 0.1 M, reaching in some cases 100% of CO conversion. For instance, the CuCe-150-0 sample presents a T₅₀ at 257 °C and the sample CuCe-150-0.1 shows T₅₀ at 165 °C. It should be mentioned that the former sample contained Cu²⁺ isolated ions dispersed in ceria, as confirmed via EPR spectroscopy, while the latter sample presented Cu²⁺ clusters onto ceria surface. Higher reactivity of Cu²⁺ clusters than Cu²⁺ isolated ions, was also reported by Harrison et al. [39]. Higher concentrations of NaOH (≥ 0.5 M) resulted in catalytic materials with poor activity.

Similar trends can be also depicted in the catalytic behavior of the materials synthesized with a hydrothermal temperature of 180 °C. Overall, the highest catalytic activity was achieved over the CuCe-180-0.05 sample which present in Figure 9, T_{50} and T_{90} at 132 °C and 180 °C, respectively. The rod-like morphology, the presence of high-content Cu²⁺ clusters (see Table 1) and high concentration of oxygen vacancies (see Table 4) seems to be the crucial reasons behind this extraordinary catalytic activity. It is noteworthy that the CuCe-180-0.05 as both CuCe-180-0 and CuCe-180-0.1 illustrated 100% conversion of CO. On the other hand, the samples that were prepared with high concentrations (\geq 0.5 M) presented similar catalytic behavior with the samples hydrothermally treated at 120 °C and 150 °C.

3. Discussion

Correlation of the Physicochemical Characteristics with the Catalytic Activity

Taking into consideration both the physicochemical characterization and catalytic evaluation, it can be proposed that the high catalytic activity is controlled by the successful combination of specific materials characteristics such as the morphology, oxygen vacancies, and type of copper entities. The specific surface area seems to be the feature with less impact in order to achieve high activity. More specifically:

- The most active samples presented spherical and/or rod-like morphology (samples with concentrations of NaOH \leq 0.1 M), which was attributed to the effective combination of hydrothermal parameters. Exception to this trend was the sample Ce-120-0.1, which illustrated rods with high crystallinity. The bulky aggregates which, were revealed by the samples with high concentrations (\geq 0.5 M), seemed to negatively affect the catalytic activity.
- While the samples with high NaOH concentrations (≥0.5 M) showed high SSA, they could
 not compete the activities of the samples prepared with low concentrations of NaOH (≤0.1 M).
 In general, the high SSA is a desirable parameter for enhanced catalytic activity, but in our work,
 it is not a determining factor, as the observed catalytic performance is a combination of the copper
 content, copper species nature, morphology, oxygen vacancies and Cu-Ce interactions.
- Two different copper species were detected in ceria-based materials, which promoted the catalytic activity. This effect was more pronounced in the case of Cu²⁺ clusters as compared with the effect caused by the presence of Cu²⁺ isolated ions. The former species were favored under high hydrothermal temperatures and low NaOH concentrations.
- The presence of lattice defects and especially oxygen vacancies, which were certified by the spectroscopic techniques, affected the catalytic behavior, but not following a conventional trend. As shown via EPR spectroscopy, the catalysts, which were prepared with high concentrations of NaOH (≥0.5 M), presented a high concentration of Ce³⁺ ions (see Table 1), they did not result in a high catalytic activity. On the other hand, the samples with low concentrations of NaOH (≤0.1 M), although possessing small concentrations of Ce³⁺ ions, they exhibited better catalytic behavior. Strong copper-ceria concentrations were confirmed via DRS analysis for several samples prepared at 150 and 180 °C with low concentrations of NaOH. Conclcuding the above remarks, it can be commented that the formation of a high-defective structure is not a crucial factor for high catalytic activity, but the formation of a structure that will have a suitable concentration of oxygen vacancies. Similar results have been illustrated by Lin et al. [99]. Pulse calorimetry
measurements showed that the high concentration of oxygen vacancies promoted the buildup of adsorbed carbonates that prohibit the adsorption and activation of CO, but not O_2 . According to several studies, oxygen vacancies tend to form oxygen vacancy clusters [100,101]. Wang et al. [102] showed that oxygen vacancy clusters with suitable size and distribution are responsible for high activity. Thus, it is believed that the various morphologies, which were obtained in this study, can present oxygen vacancies with different size and distribution.

4. Materials and Methods

4.1. Catalysts Preparation

All the chemicals used in this work were of analytical reagent grade. Cerium(III) nitrate hexahydrate Ce(NO₃)₃ · 6H₂O (purity 99.99%, Sigma-Aldrich) and a pure metallic copper ring were used as precursors for the preparation of ceria and copper-promoted ceria nanomaterials. Moreover, citric acid monohydrate C₆H₈O₇ · H₂O (purity 99.5–101.0%, Ing. Petr Švec PENTA, Prague, Czech Republic) and sodium hydroxide NaOH (purity \geq 98.0%, Ing. Petr Švec PENTA, Prague, Czech Republic) were also employed in the synthesis procedure as a chelating agent and precipitating agent, respectively.

At first, 3.7834 gr of Ce(NO₃)₃ · 6H₂O and 1.8314 gr of C₆H₈O₇ H₂O were dissolved under continuous stirring into 15 mL of triple distilled (3D) water, respectively. The molar ratio of citric acid to metal nitrate was adjusted according to the stoichiometry of the reaction (citric acid/Ce = 1/1). When the dissolution of the compounds was completed, the two aqueous solutions were mixed under continuous stirring, while at the same time, a 150 mL aqueous solution of NaOH (C_{NaOH} = 0–5 M) was prepared. Both solutions were mixed in a Teflon beaker, and this mixture was kept under stirring for 20 min. The Teflon beaker was placed in a lab-made stainless-steel autoclave (chamber volume of 200 mL), where the copper ring was placed over the beaker (not immersed into the solution). The autoclave was sealed tightly and heated at various temperatures (120, 150, and 180 °C) for 24 h. After the hydrothermal treatment, the autoclave was opened, excess water was decanted, and the precipitates were filtered, washed several times with triple distilled water (until pH = 7), and dried under vacuum at 70 °C overnight. Finally, the obtained powders were calcined at 400 °C for 2 h (heating ramp = 2 °C min⁻¹).

In order to facilitate the presentation of results, the following encoding of catalysts is used: Ce-T-M and CuCe-T-M, where T represents the temperature (°C) of the hydrothermal reaction and M the molarity of NaOH (M). For example, the Ce-120-5 catalyst was synthesized hydrothermally at 120 °C, using 5 M NaOH.

4.2. Catalysts Characterization

An X-ray powder diffractometer (Bruker D8 Advance, Bruker, Birmingham, UK) employing Cu K_a radiation (λ = 0.15418 nm) at 40 kV and 40 mA was used to analyze the crystalline structure of the catalysts.

The specific surface area (SSA), the pore volume and the pore size distribution of the materials were determined from the adsorption and desorption isotherms of nitrogen at -196 °C using a TriStar 3000 Micromeritics instrument (Norcross, GA, U.S.A). Prior to the measurements, the samples were outgassed at 150 °C for 1 h, under N₂ flow.

The morphology of the obtained materials was observed with scanning electron microscopy (SEM, Leo Supra 35VP (Carl Zeiss SMT AG Company, Oberkochen, Germany).

Raman spectra were accumulated with the 441.6 nm laser line as the excitation source emerging from a He–Cd laser (Kimon). The scattered light was analyzed by the Lab-Ram HR800 (Jobin-Yvon, Horiba, Montpellier, France) micro Raman spectometer at a spectral resolution of about 2.0 cm⁻¹. A microscope objective with magnification $50 \times$ was used to focus the light onto a spot of ~3 µm in

diameter. Low laser intensities were used (\sim 0.37 mW on the sample) to avoid spectral changes due to heat-induced effects. The Raman shift was calibrated using the 520 cm⁻¹ Raman band of crystalline Si.

The diffuse reflectance spectra of the obtained materials were recorded in the range 200–800 nm at room temperature using a UV-vis spectrophotometer (Varian Cary 3; Varian Inc. Palo Alto, CA, USA) equipped with an integration sphere. The DR spectra were collected on calcined samples with PTFE disks. The powder samples were mounted in a quartz cell, which provided a sample thickness >3 mm to guarantee the "infinite" sample thickness.

Ceria-based nanomaterials were also characterized by Electron Paramagnetic Resonance (EPR) spectroscopy employing a Bruker ER200D spectrometer (Billerica, MA, USA) equipped with an Agilent 5310A frequency counter (Agilent Technologies, Santa Clara, CA, USA). EPR spectra were recorded at 77 K in suprasil-quarz tubes (3 mm inner diameter; Willmad Glass). 10 mg of nano-powders were introduced into the sample EPR tube followed by outgassing at 300 K for 10 min under 10-4 bar vacuum. The spectrometer was running under home-made software based on LabView. Numerical simulation of experimental EPR spectra was performed with EasySpin 5.2.21 software (The MathWorks Inc., Natick, MA, USA) [103]. Quantification of Ce3+-O-Ce4+ (S = 1/2) was performed using 2,2-diphenyl-1-picrylhydrazyl (DPPH) [104] as a spin standard (Sigma Aldrich). Quantification of Ce³⁺ (S = 5/2) was performed using FeIII (S = 5/2)-EDTA complex [105,106], while the quantification of the Cu²⁺ centers was done using a Cu(NO₃)₂ standard. Copper content was also determined via WDXRF analysis (Wavelength Dispersive X-Ray Fluorescence; ZSX PRIMUS II, RIGAKU, Austin, TX, USA).

4.3. Catalytic Studies

Activity measurements for the catalytic oxidation of CO were conducted in a conventional fixed-bed reactor (described in detail elsewhere [107]) at atmospheric pressure, in the temperature range of 30–400 °C. The catalyst was in the form of powder with a mass of 120 mg and the total flow rate of the reaction mixture was 25 cm³ min⁻¹, yielding a contact time of W/F = 0.288 g s cm⁻³, where W is the weight of catalyst and F the total flow rate of the reactant gas. The reaction feed stream contained 1 vol.% CO, 20 vol.% O₂ and He as balance. Product and reactant analysis was carried out by a gas chromatograph (Shimadzu GC-14B) equipped with a thermal conductivity detector. The CO conversion calculation was based on the CO₂ formation or CO consumption:

$$CO \text{ conversion } (\%) = \frac{[CO_2]_{OUT}}{[CO]_{IN}} \times 100\%, \tag{4}$$

5. Conclusions

In this study, highly active ceria and copper-promoted ceria catalysts were synthesized via a novel hydrothermal method and evaluated in CO oxidation reaction. The physicochemical characteristics, and as a consequence, the catalytic properties were controlled via appropriate combination of hydrothermal parameters (temperature and concentration of the precipitating agent). EPR spectroscopy demonstrated the presence of different copper species (isolated ions and amorphous clusters), dispersed in ceria nanostructrures (mainly nanorods and nanospheres, depending on the pH of the hydrothermal solution). Elevated hydrothermal temperatures and NaOH concentrations favored the formation of silated copper ions, which resulted to be less active, as compared with copper clusters. Moreover, the catalysts prepared with high concentrations of NaOH (\geq 0.5 M) presented a significant amount of surface Ce³⁺ ions, while Raman spectroscopy and UV-Vis DRS measurements revealed the presence of lattice defects and especially oxygen vacancies. Overall, ceria-based catalysts prepared at elevated temperatures with low concentrations of NaOH (\leq 0.1 M) were more active in CO oxidation, and this behavior can be mainly related with the obtained morphology and the nature of oxygen vacancies and dispersed copper species, and to a lesser extent, with the specific surface area and the concentration of defects.

Supplementary Materials: The following are available online at http://www.mdpi.com/2073-4344/9/2/138/s1, Figure S1. XRD diffractograms of CeO₂ and Cu/CeO₂ catalysts: (a) Catalysts prepared hydrothermally at 120 °C; (b) Catalysts prepared hydrothermally at 150 °C; (c) Catalysts prepared hydrothermally at 180 °C, Figure S2. UV-Vis DRS spectra of CeO₂ and Cu/CeO₂ catalysts: (a) Catalysts prepared hydrothermally at 120 °C; (b) Catalysts prepared hydrothermally at 150 °C; (c) Catalysts prepared hydrothermally at 120 °C; (b) Catalysts prepared hydrothermally at 150 °C; (c) Catalysts prepared hydrothermally at 120 °C; (b) Catalysts prepared hydrothermally at 150 °C; (c) Catalysts prepared hydrothermally at 120 °C; (b) Catalysts prepared hydrothermally at 150 °C; (c) Catalysts prepared hydrothermally at 180 °C.

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Catalytic Properties of Double Substituted Lanthanum Cobaltite Nanostructured Coatings Prepared by Reactive Magnetron Sputtering

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Abstract: Lanthanum perovskites are promising candidates to replace platinum group metal (PGM), especially regarding catalytic oxidation reactions. We have prepared thin catalytic coatings of Sr and Ag doped lanthanum perovskite by using the cathodic co-sputtering magnetron method in reactive condition. Such development of catalytic films may optimize the surface/bulk ratio to save raw materials, since a porous coating can combine a large exchange surface with the gas phase with an extremely low loading. The sputtering deposition process was optimized to generate crystallized and thin perovskites films on alumina substrates. We found that high Ag contents has a strong impact on the morphology of the coatings. High Ag loadings favor the growth of covering films with a porous wire-like morphology showing a good catalytic activity for CO oxidation. The most active composition displays similar catalytic performances than those of a Pt film. In addition, this porous coating is also efficient for CO and NO oxidation in a simulated Diesel exhaust gas mixture, demonstrating the promising catalytic properties of such nanostructured thin sputtered perovskite films.

Keywords: perovskite; catalytic coating; CO oxidation; cathodic sputtering method

1. Introduction

The family of perovskite oxides is known for its catalytic properties, hydrothermal stability, high recyclability and low cost compared with Platinum Group Metal (PGM) [1–3]. The general formula of perovskites is ABO₃ where the larger size A-cation presents a 12-coordination number and the B-cation coordinates with 6 neighboring atoms [4]. Partial substitution of A and/or B atoms with other elements showing redox properties may enhance the catalytic activity due to the generation of structural defects such as anionic or cationic vacancies and/or modification of the oxidation state of B cations to maintain the electro-neutrality [5]. Lanthanum perovskites are promising candidates to replace noble metals (Pt, Pd, etc.) [6,7], especially regarding catalytic oxidation reactions. Lanthanum cobaltite (LaCoO₃) is one of the most promising catalysts for the oxidation of gaseous pollutants such as carbon monoxide, unburnt hydrocarbons and nitrogen oxide [7,8]. Lanthanum cobaltites are used in many others fields due to their magnetic properties as well as their mixed ionic and electronic conductivity [9–12]. Most of these applications require the implementation of thin films that act



as electrodes for fuel cells [13], thermoelectric processes [14], sensors [15], and magnetoresistance devices [16]. These electric, magnetic and electrocatalytic properties depend on the composition but also on the microstructure and morphology of coatings, which strongly depend on the preparation method [17,18]. Several techniques are reported for the synthesis of perovskite thin films such as the ground-frost [7], casting in band [18], chemical vapor deposition (CVD) [17], spray painting [19], sol-gel [7], atomic layer deposition (ALD) [20], laser pulsed deposition (PLD) [21], and physical vapor deposition (PVD) [22,23]. Regarding catalytic applications, the development of submicrometric catalytic films may optimize the surface/bulk ratio to save raw materials. Indeed, a thin catalytic coating can combine a large exchange surface with the gas phase with an extremely low loading. This is particularly suitable for air cleaning since gaseous pollutants only lick the surface of the catalyst. In addition, thin catalytic coatings could be deposited on hot substrates such as collector walls in thermal engine exhausts for removing pollutants or of radiators for improving indoor air quality. Few studies on the development of thin perovskite coatings for catalytic applications are reported in the literature [24–26]. The challenge lies in preparing adherent, thermally stable and pure crystallized perovskite films without any secondary parasite phase [18], showing appropriate compositions and nanostructures for catalysis [7,8,16,18]. The porosity and the specific surface areas have to be optimized to counter-balance the low quantity of materials involved in submicrometric catalytic coatings. In the present study, we used the cathodic co-sputtering magnetron method in reactive condition to prepare submicrometric nanostructured coatings of perovskites. This technique allows the deposition of pure and adherent coatings of complex oxides with a controlled and reproducible manner, while respecting the environment [27,28]. We have chosen to prepare Sr and Ag-doped lanthanum cobaltite as this perovskite composition is one of the most active for CO oxidation [8]. The partial substitution of La^{3+} cations by Sr^{2+} ones improves the thermal stability and then the specific surface area of pure $LaCoO_3$ [29] and also enhances the surface concentration of oxygen vacancies involved in the oxidation catalytic mechanism [30]. The partial substitution of La^{3+} by Ag⁺ can also increase the catalytic properties of lanthanum cobaltites for oxidation reactions due to the formation of oxygen vacancies [31] and the stabilisation of Ag nanoparticles on the oxide surface [6,32]. The different parameters (intensities of current applied on the metallic targets, total pressure in the chamber, oxygen partial pressure, etc.) of the reactive magnetron sputtering preparation method were tuned to achieve pure and adherent coatings of LaCoO₃ on alumina dense membranes. After a calcination at 500 °C, crystallized and dense cubic perovskite films of around 1.5 µm thick were achieved with the targeted La/Co stoichiometry. Different compositions of La_{1-x-y}Sr_xAg_yCoO_{3- α} (x = 0.13–0.28, y = 0.14–0.48) doped perovskites were sputtered on alumina disks. We found that the incorporation of high contents of Ag can strongly modify the morphology of the coatings, increasing their porosity. The catalytic performance of the perovskite catalytic coatings for CO oxidation was found to be improved by the double substitution of Sr and Ag. The most active composition, La_{0.40}Sr_{0.1}Ag_{0.48}Co_{0.93}O₃, displays similar catalytic performances than those of a Pt film, for CO oxidation. In addition, this porous coating is also active for CO and NO oxidation in a simulated Diesel exhaust gas mixture, demonstrating the promising catalytic properties of such nanostructured thin sputtered perovskite films.

2. Results

2.1. Preparation and Characterization of LaCoO₃ Catalytic Coatings

Layers of LaCoO₃ were synthetized on alumina substrates by using the reactive magnetron sputtering preparation method. Depositions have been performed in a reactive mode, mixing oxygen and argon in the chamber. To achieve the suitable La/Co ratio, the discharge current applied to the Co target was adjusted while maintaining a constant current of 1 A on the La target (Table 1). As expected, the La/Co ratio, estimated by Energy Dispersive Spectroscopy (EDS), decreases with the current dissipated on the Co target (Figure 1). We determined that a current intensity of 0.3 A is required to reach the targeted La/Co ratio. The diffractogram recorded on this as-deposited coating

(Figure 2) shows that the layer is amorphous as no peak is detected. The difference between the radii of La³⁺ (136 pm) and Co³⁺ cations (72 pm) suggests an amorphous as-deposited coating due to steric effects as predicted by the confusion principle [33–35]. Then, a calcination step was performed at different temperatures (from 100 °C to 500 °C) for 2 h in air. X-Ray Diffraction (XRD) patterns (Figure 2), obtained under a Bragg-Brentano configuration ($\theta/2\theta$), evidence that the oxide coating crystallizes from 500 °C as a cubic perovskite phase (JCPDS 01-075-0279). This temperature is approximately 100 °C lower than those reported by H. Seim et al. [20] and H.J. Hwang et al. [36] on LaCoO₃ films prepared by PLD and sol-gel methods, respectively. This demonstrates that the reactive magnetron sputtering technique is efficient for preparing crystallized lanthanum cobaltite films at rather low temperatures.

Target	Intensity (A)	Pulse Frequency (kHz)	Dead Time (Toff μs)	Ar Flow Rate (sccm) *	O ₂ Flow Rate (sccm) *	Total Pressure (Pa)	Draw Distance (mm)	Sputtering Time (h)
La Co	1 0.2 to 0.4	50	5	50	20	1.5	45	3

Table 1. Sputtering parameters used for LaCoO₃ coating.



* sccm = Standard Cubic Centimeter per Minutes.

Figure 1. Evolution of the atomic La/Co ratio measured by Energy Dispersive Spectroscopy (EDS) as a function of the current dissipated on the Co target ($I_{La} = 1 A$). The total pressure is 1.5 Pa.

This calcination step at 500 °C was performed for all coatings (Figure 3). The thin film with the highest La/Co ratio is not crystallized and only presents small peaks of La₂O₃ (JCDPS 00-005-0602), probably coming from a high excess of La. On the opposite, for larger atomic ratios, coatings are crystallised with various space groups of the perovskite-type structure. Indeed, the film is cubic (Pm3m space group) for a ratio of 0.87, exhibits a rhomboedric symmetry (R-3c space group) for a ratio of 0.56 and again becomes cubic for a ratio of 0.3.



Figure 2. X-Ray diffractograms of LaCoO₃ sputtered films prepared with a current dissipated on the Co target of 0.3 A (La/Co = 0.87) after different post-calcination treatments during 2 h in air from 100 °C to 500 °C.



Figure 3. XRD patterns as a function of the atomic composition ratio after calcination treatment at 500 °C for 2 h under air.

The morphology of the LaCoO₃ coating with the suitable ratio (La/Co = 0.87) was observed by SEM. Figure 4 shows SEM images of the top view and the brittle cross section of the sample before and after the calcination step at 500 °C. The as-deposited coating (Figure 4a,c) covers the surface of the alumina substrate and follows its morphology. This film is quite dense, as shown in the cross-section image (Figure 4a) and adherent with a vitreous appearance characteristic of an amorphous material, in agreement with XRD (Figure 2) [37]. After calcination (Figure 4b), some cracks can be observed especially in the cross section (Figure 4d). The crystallized film is not sticking well with the alumina support and remains quite dense. The cracks were probably formed under stresses during the crystallization or the thermal treatment. In this latter case, the mismatch of thermal expansion coefficients between the perovskite film and the alumina substrate could be at the origin of the delamination ($\alpha_{LaCoO3} \approx 20 \times 10^{-6} \text{ °C}^{-1}$ [38] and $\alpha_{Al2O3} \approx 7 \times 10^{-6} \text{ °C}^{-1}$ [39]) of the coating. The thickness of the annealed coating is around 1.5 µm (Figure 4c,d), resulting in a 500 nm.h⁻¹ deposition rate. This result is in agreement with the thickness measured by tactile profilometry.



Figure 4. SEM images of the surface and the brittle cross section of the LaCoO₃ sample synthesized with $I_{Co} = 0.3$ A and $I_{La} = 1$ A: (**a**,**c**) as deposited coating and (**b**,**d**) after calcination treatment at 500 °C for 2 h under air.

2.2. Preparation and Characterization of $La_{1-x-y}Sr_xAg_yCoO_{3-\alpha}$ (LSACO) Catalytic Coatings

The experimental parameters used for the synthesis of double substituted cobaltite (denoted as LSACO) coatings on alumina supports are reported in Table 2. The protocol is similar to that developed for the synthesis of pure lanthanum cobaltite (Ar flow rate = 50 sccm, O₂ flow rate = 20 sccm, total pressure = 1.5 Pa, draw distance = 45 mm) except the sputtering time which was extended to 4 h to achieve thicker films. The control of the substitution rates of Sr and Ag was achieved by tuning the intensity of the currents dissipated on the metallic targets (Table 2). We assumed that Sr²⁺ and also Ag⁺ cations will partially substitute La³⁺ cations in A sites of the perovskite. This hypothesis is based on the similarity of the ionic radius of La³⁺ (136 pm) and Ag⁺ (128 pm) [40]. The Ag content in the perovskites was gradually enhanced by increasing the current applied to the Ag target (LSACO-1 to 4). This was counter-balanced by a progressive decay of the current dissipated to the La target. The Co content has been slightly increased in LSACO-5 compared with LSACO-4, while maintaining constant the Ag, La, and Sr contents (Table 2)

Table 2. Sputtering pa	arameters used for	LSACO coatings.
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Coating	I _{La} (A) Pulse (kHz)/Toff (μs)	I _{Sr} (A) Pulse (kHz)/Toff (µs)	I _{Ag} (A) Pulse (kHz)/Toff (µs)	I _{Co} (A) Pulse (kHz)/Toff (μs)
LSACO-1	1.2 50/5	1.2 350/1.4	0.008 0	0.35 50/5
LSACO-2	0.8 50/5	1.1 350/1.4	0.009 0	0.35 50/5
LSACO-3	0.75 50/5	1.1 350/1.4	0.01 0	0.35 50/5
LSACO-4	0.6 50/5	1.1 350/1.4	0.013 0	0.35 50/5
LSACO-5	0.6 50/5	1.1 350/1.4	0.013 0	0.4 50/5

The chemical composition of the as-prepared LSACO coatings was estimated by EDS analysis. Table 3 displays the variations of the atomic percent of each element, without considering the oxygen level which is tricky to quantity by using EDS. The Sr and Ag contents inversely vary in LSACO-1, 2 and 3 while La and Co loadings are fairly stable. Let us note that the chemical composition of LSACO-2 and LSACO-3 is fairly similar with 11 at% of Ag and around 10 and 30 at% of Sr and La, respectively. The high current applied to the Ag target during the preparation of LSACO-4 and LSACO-5 leads to a significant increase in the Ag concentration and a concomitant drop of the La content. In LSACO-4 and LSACO-5, the Ag atomic percent reaches around 25% while the La level is approximately 20%. Surprisingly, the Co concentration decreases in LSACO-5 despite the largest applied current to the Co target. Except LSACO-5, the atomic ratio between cations located in A sites (theoretically La³⁺, Sr²⁺ and Ag⁺) and Co³⁺ ones located in B sites is close to the target of 1. The morphology of the as-deposited coatings (Figure 5) was found to drastically change with the chemical composition. The LSACO-1 coating only contains few perovskite clusters (Figure 5a, white spots at the top right) dispersed on the surface of the micrometric alumina grains (Figure 5, dark grey), despite of the 4 h deposition time. For LSACO-2 and LSACO-3, the number of perovskite clusters significantly increases on the surface of the alumina grains (white spots in Figure 5b,c). Interestingly, larger starfish shaped perovskites islands are growing on the substrate defects (cavity of alumina substrate or grain boundaries) showing arms like filaments. The morphology of the surface of LSACO-4 and LSACO-5 coatings (Figure 5d,e) are quite different. The alumina substrate is now fairly fully covered by a porous film with a wire-like morphology (Figure 5d,e). These wires have grown parallel to each other, leading void between each other, reaching a length of the order of 1 µm and a diameter of around 100 nm. Therefore, high Ag contents seem to strongly enhance the porosity and then the coverage of the films.

Table 3. Chemical composition determined by EDS of the as-prepared $La_{1-x-y}Sr_xAg_yCoO_{3-\alpha}$ (LSCAO) perovskite coatings.

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Figure 5. SEM images of the surface morphology of as-deposited LSACO coatings on alumina substrate: (a) LSACO-1, (b) LSACO-2, (c) LSACO-3, (d) LSACO-4, and (e) LSACO-5.

These series of LSACO coatings were annealed at 500 °C for 2 h in air. This calcination step was sufficient to crystallize all the films as a cubic perovskite phase (JCPDS 01-075-0279) whatever the composition. On the other hand, an additional XRD pattern at 44.3° on the diffractogramm of LSACO-4 and LSACO-5, corresponding to (2 0 0) planes of fcc metallic silver, proves the presence of metallic Ag, out of the perovskite structure. These results demonstrate the limited solubility of Ag in the perovskite in agreement with previous studies [32]. For high contents of Ag, reaching 25 at. %, i.e., around 10 at% in the overall oxide including the oxygen atoms, part of Ag is not incorporated into the perovskite structure. No XRD peaks corresponding to Ag° were observed for LSACO-1, LSACO-2, and LSACO-3 (Figure 6), but the low content of Ag, below 5 at%, makes difficult their detection. Therefore, we cannot exclude the presence of metallic Ag on the surface of these samples.



Figure 6. XRD patterns after the calcination treatment at 500 °C for 2 h under air of the different LSACO coatings.

Figure 7 displays the surface morphology of the calcinated films. The alumina substrates are cracked probably due to the thermal treatment. The LSACO-1 coating only contains isolated perovskites clusters on the surface of alumina grains. As before calcination, the morphology of LSACO-2 and LSACO-3 coatings are similar with perovskite islands mainly located in the cavity and interstices of the alumina substrate. Filaments that were present around of the perovskite clusters have disappeared. Coatings containing high Ag loadings (LSACO-4 and LSACO-5) are the only ones able to fully cover the alumina substrate with a porous wire-like morphology. These nanowires are less ordered than before calcination and interlock, then decreasing the porosity of the film. Ag particles (Figure 7e) can be observed on the surface of LSACO-5 (white particles), confirming that a part of Ag was not incorporated into the perovkiste structure.



Figure 7. SEM images of the surface morphology of calcined LSACO coatings on alumina substrate: (a) LSACO-1, (b) LSACO-2, (c) LSACO-3, (d) LSACO-4, and (e) LSACO-5.

2.3. Catalytic Performances of the Perovskite Coatings

2.3.1. Catalytic Performances for CO Oxidation

The catalytic activity of LSACO-2, LSACO-4, and LSACO-5 coatings, containing respectively 11, 25 and 28 at% of Ag, was measured for CO oxidation during heating ramps up to 400 °C. We have selected LSACO-4 and LSACO-5 as these films exhibit a porous and wire-like morphology. We have also tested the catalytic performances for CO oxidation of pure lanthanum cobaltite coatings. These samples have shown no activity below 500 °C in good agreement with their dense morphology. The catalyst LSACO-2 was also tested for comparison as a non-porous and non-covering representative layer. In addition, we have also tested, for comparison, a Pt coating (3 μ m thick, 5 mg Pt/cm²) prepared by spray-painting and annealed at 500 °C. Figure 8 displays the Light-off (LO) curves of the different catalytic coatings for CO oxidation. The results show that the catalytic performances increase with the silver content in the film. Values of T20, temperature at 20% conversion, significantly decrease from 280 to 215 °C when increasing the Ag content. Furthermore, the catalytic performances of LSACO-5 are rather close to those of the Pt coating up to 220 °C. This underlines the promising catalytic activity of the sputtered Ag-doped perovskite coating.

Two successive LO up to 400 °C have been performed on LSACO-4 (Figure 9a). The value of T20 increases from 215 to 242 °C during the second LO. This indicates that a modification of the coating morphology during the first LO, most probably due to the Ag particles sintering. Nevertheless, the onset temperature slightly decreases from 175 to 150 °C (Figure 9a). Similar catalytic experiments have been carried out on LSACO-5 (Figure 9b) without any significant modification of the catalytic performances between the two successive LO. This indicates a better stability of the morphology of LSACO-5 in the presence of the reactive mixture up to 400°C compared with LSACO-4. XPS measurements performed after catalytic tests evidence a La surface seggregation with a concommitant drop of the Co concentration. The two catalytic coatings show a similar Ag surface atomic concentration, i.e., around 10 at%, which is far lower from the loading estimated by EDS before the catalytic tests (Table 3). The Ag3d XPS peaks (Figure 10) show a binding energy of Ag3d5/2 at 367.7 eV, attributed to metallic silver. The catalytic properties during the second LO of the two catalytic perovskite coatings are fairly similar, in good agreement with an equivalent Ag surface concentration (Table 4). This confirms the direct link between the surface Ag concentration and the catalytic activity for CO oxidation.



Figure 8. Light-off curves for CO oxidation recorded on the sputtered perovskite coatings and on Pt film. Reactive mixture: $CO/O_2 = 3000$ ppm/3%. Overall flow = 3.6 L/h.



Figure 9. Two successive LO recorded on (a) LSACO-4 and (b) LSACO-5. Reactive mixture: $CO/O_2 = 3000 \text{ ppm/3\%}$. Overall flow = 3.6 L/h.



Figure 10. Ag3d XPS spectrum of LSAC-5.

Sample	La at%	Sr at%	Ag at%	Co at%
LSACO-4	47	11	12	30
LSACO-5	51	13	10	26

Table 4. Surface composition determined by XPS of LSACO-4 and LSACO-5 after catalytic tests.

2.3.2. Catalytic Performances in a Model Lean Diesel Exhaust Gas

The catalytic performances of LSACO-4 were explored in a model lean diesel exhaust gas mixture containing 8% O₂, 950 ppm CO, 270 ppm NO, 1000 ppm C₃H₈ and 10% H₂O. The LO was recorded up to 375 °C to investigate the low temperature activity. Figure 11 shows the ability of the catalytic coating to oxidize CO and NO into CO₂ and NO₂, respectively. Below 350 °C, we did not observed any propane oxidation. However, the catalytic coating shows a remarkable activity for NO oxidation from around 225 °C. The NO conversion achieves 30% at 375 °C. CO conversion starts from 250 °C but seems to reach a maximum at only 18% from 350 °C. These results show that, despite a very low mass of catalyst and a complex mixture including the inhibiting effect of H₂O, the perovkite catalytic coating can be active at low temperature for low temperature NO and CO oxidation.



Figure 11. Variation of CO and NO conversion as a function of temperature on LSACO-4. Reactive mixture: CO/O₂/NO/C₃H₈/H₂O: 950 ppm/8%/270 ppm/1000 ppm/10%. Overall flow: 3.6 L/h.

3. Materials and Methods

3.1. Preparation of the Catalytic Coatings

LaCoO₃ perovskite coatings were synthesized by magnetron sputtering from two metallic targets of La (kurt J. Lekser, purity 99.9%, Hastings, England) and Co (kurt J. Lekser, purity 99.9%). The Alcatel SCM 604 reactor, described elsewhere [33], was a 90-l sputtering chamber where a pressure of 10^{-4} Pa was maintained by a primary pump assisted by a molecular turbo pump. The La metallic target (thickness = 3 mm; diameter = 50 mm) and the Co one (thickness = 1 mm; diameter = 50 mm) were attached to two magnetrons distant from 120 mm and connected to a pulsed direct current generator (PINACLE+ purchased by advance energy). The distance between the two targets and the substrate holder (DT-S) can be independently modified. The draw distance between the targets and the alumina

substrate was adjusted to 45 mm (Table 1). The targets are powered with pulsed currents to avoid any electric micro-arcs that can damage the quality of the films [41]. To avoid any accumulation of positive charge (Ar+) on the targets, the current is periodically stopped for very short dead time of 5 μ S. Gases were introduced with mass-flow controllers (Brooks, 5850 SLA, Hatfield, PA, USA) and the total working pressure was measured with a MKS Baratron gauge. Films were deposited on dense alumina pellets (Keral 99, diameter = 16 mm, thickness = 0.60 mm purchased by Kerafol Gmbh, (Koppe Platz 1, Eschenbach i. d. Opf). The synthesis of doped perovskites with Sr and Ag was performed with the addition of two magnetrons into the sputtering chamber and metallic targets of Sr and Ag (thickness = 3 mm diameter = 50 mm, kurt J. Lekser, purity 99.9%). The Sr and Ag targets were powered by a two pulsed direct current generator (PINACLE+ and MDX500 both purchased by advance energy, Metzingen, Germany).

3.2. Characterizations of the Catalytic Coatings

The crystal structure of the coatings was determined thanks to a BRUKER D8 focus X-Ray diffractometer (Co K α 1+ α 2 radiations, in Bragg Brentano configuration and equipped with a LynxEye linear detector (Bruker, Billerica, MA, USA). XRD patterns were collected under air during 10 min in the (20–80°) scattering angle range by steps of 0.019°. The surface and the morphology of films were observed with a scanning electron microscope (JEOL JSM 7800F, Akishima, Japan) equipped with an EDS detector allowing the estimation of the chemical composition of samples. The observation of brittle cross section by SEM led also to the determination of the thickness of films. Coating thickness was also determined by the step method with an Altysurf profilometer produced by Altimet society (Marin, France) equipped with tungsten micro force probe inductive allowing an accuracy of about 20 nm. Before each measurement, the calibration of the experimental device was realized with a reference sample number 787569 accredited by CETIM organization.

XPS spectra were recorded for each catalysts on an AXIS Ultra DLD from Kratos Analytical (Manchester, UK) using a monochromatized Al X-ray source (hv = 1486.6 eV) between 0 and 1200 eV with a pass energy of 40 eV. Sample were pretreated at 200 °C in He before measurements to clean the surface. Peaks were referenced using C1S peak of carbon (BE = 284.6 eV).

3.3. Catalytic Activity Measurements

A quartz reactor was operated under continuous flowing conditions at atmospheric pressure. The samples were placed on a fritted quartz, 18 mm in diameter, with the catalytic coating side facing the fritted quartz [42]. Gases were mixed by using mass-flow controllers (Brooks) to generate the different reactive mixtures. H₂O vapor was introduced using an atmospheric pressure Pyrex saturator heated at 46 °C. Reactants and products were analyzed using a gas micro-chromatograph (SRA 3000 equipped with two TCD detectors, a molecular sieve and a Porapak Q column for O₂, CO, C₃H₆, and CO₂ analysis) and a CO₂ Infra-Red analyzer (Horiba VA 3000, Horiba Europe Gmbh, Leichlingen, Germany). NO and NO₂ concentrations were measured with IR and UV online analyzers (EMERSON NGA2000).

Reactants were Air Liquide certified standards of NO in He (8000 ppm), C_3H_8 in He (8005 ppm), CO in He (1%), O_2 (99.999%), which could be further diluted in (99.999%). The carbon and nitrogen balance closure was found to be within 2%.

4. Conclusions

Different compositions of $La_{1-x-y}Sr_xAg_yCoO_{3-\alpha}$ (x = 0.13-0.28, y = 0.14-0.48) doped perovskites were synthetized as thin coatings deposited on alumina disks with the cathodic co-sputtering magnetron method in reactive conditions. The control of different parameters during the sputtering process can tune the morphology of the catalytic films. In particular, we found that the incorporation of high Ag loadings can generate covering films with a porous wire-like morphology showing good catalytic activity for CO oxidation. The most active composition, $La_{0.40}Sr_{0.1}Ag_{0.48}Co_{0.93}O_3$, displays similar catalytic performances than those of a Pt film. In addition, this porous coating is also efficient for CO and NO oxidation in a simulated Diesel exhaust gas mixture, demonstrating the promising catalytic properties of such nanostructured thin sputtered perovskite films.

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Article Effect of Preparation Method of Co-Ce Catalysts on CH₄ Combustion

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Abstract: Transition metal oxides have recently attracted considerable attention as candidate catalysts for the complete oxidation of methane, the main component of the natural gas, used in various industrial processes or as a fuel in turbines and vehicles. A series of novel Co-Ce mixed oxide catalysts were synthesized as an effort to enhance synergistic effects that could improve their redox behavior, oxygen storage ability and, thus, their activity in methane oxidation. The effect of synthesis method (hydrothermal or precipitation) and Co loading (0, 2, 5, and 15 wt.%) on the catalytic efficiency and stability of the derived materials was investigated. Use of hydrothermal synthesis results in the most efficient Co/CeO₂ catalysts, a fact related with their improved physicochemical properties, as compared with the materials prepared via precipitation. In particular, a CeO₂ support of smaller crystallite size and larger surface area seems to enhance the reducibility of the Co₃O₄/CeO₂ materials, as evidenced by the blue shift of the corresponding reduction peaks (H₂-TPR, H₂-Temperature Programmed Reduction). The limited methane oxidation activity over pure CeO₂ samples is significantly enhanced by Co incorporation and further improved by higher Co loadings. The optimum performance was observed over a 15 wt% Co/CeO₂ catalyst, which also presented sufficient tolerance to water presence.

Keywords: Co₃O₄; CeO₂; complete CH₄ oxidation; hydrothermal synthesis; precipitation

1. Introduction

Today, CH₄, N₂O, and CO₂ emissions represent approximately 98% of the total greenhouse gas (GHG) inventory worldwide, a percentage expected to increase even further in the 21st century. Methane is the main compound of natural gas (NG) and is today used in various industrial processes and also as energy source in gas turbines or natural gas fuelled vehicles. Heavy-duty (HD) natural gas–fuelled fleet represents today less than 2% of the total fleet, however this percentage is expected to grow to represent as much as 50% in the near future. Moreover, methane is considered a substantially more potent greenhouse gas (GHG) than CO₂ and, hence, leakage of only a small percentage of methane from the supply chain could alter the net climate benefits of NG, which is lately suggested as an alternative fuel for heavy-duty (HD) transportation. Methane can be also a result of incomplete combustion processes, similarly to CO, while both of them are well-known environmentally detrimental pollutants and, thus, proper catalytic oxidation converters are needed for their abatement [1,2]. Noble metals and transition metals are the most commonly used components of such catalysts, investigated for methane and/or CO complete oxidation. Both these groups of catalysts have been extensively studied in order to develop efficient catalytic systems for combustion of methane, recognized as the lowest reactivity molecule among alkanes. All up to date research

efforts have disclosed the superiority of the noble metals (especially Pt, Pd, and their mixtures) on oxide supports [3–5]. However, there are many factors defining the effectiveness of these supported catalysts, including the nature and properties of the support, the metal loading, the size, shape, and electronic state of metal nanoparticles and their interaction/synergy with the support, their preparation method and, finally, any pretreatment conditions possibly applied for their improvement [6]. However, in spite of their excellent activity, their applications at a larger scale have been limited because of their prohibitively high cost, their shortage in worldwide reserves, and their prompt deactivation during methane oxidation, mainly due to sintering of supported noble metal nanoparticles [7]. On the other hand, transition metal oxides have recently attracted considerable attention as alternative catalytic materials, among which cobalt oxide is considered as the most promising catalyst for methane combustion. Most crucial property of the early transition metal oxides is the fact that they can generate oxygen vacancies, suggested as the necessary sites that can activate molecular oxygen for oxidation applications [5]. Co_3O_4 nanosheets are reported to show better catalytic activity than Co_3O_4 nanobelts and nanocubes for CH₄ oxidation, in spite of their low specific surface area. Co₃O₄ catalysts with various structures (nanoparticles, two-dimensional and three-dimensional structures) were also investigated for methane catalytic combustion, exhibiting that enhanced catalytic performance of methane of the 2D-Co₃O₄ and 3D-Co₃O₄ catalysts is related with their noticeable reducibility and existence of abundant active Co³⁺ species [8,9]. Highly porous Co₃O₄ nanorods, with narrow pore-size distribution and a high surface area were synthesized by simple hydrothermal methods and similarly exhibited an enhanced catalytic performance for CH₄ oxidation, especially at high GHSV (Gas Hourly Space Velocity) conditions [10]. Co₃O₄ nanomaterials were also synthesized by hydrothermal method and tested for low temperature CO oxidation and methane combustion in another report. In that case, addition of ethylene glycol and its varying concentration significantly influenced the size and shape of the Co_3O_4 oxides (resulting in nanosheets or nanospheres). During catalyst evaluation, it was evidenced that the Co_3O_4 nanosheets exhibited more active oxygen species than the Co_3O_4 nanospheres, a fact accounting for their higher activity in the oxidative reactions [11].

In addition to cobalt, several ceria-based, low-cost catalysts have recently attracted much attention as a reliever of costly, unstable noble metal catalysts, since they exhibit good performance in oxidation reactions of carbon monoxide and lower hydrocarbons. Once more, the synthesis method is very crucial; for example nanocrystalline CeO₂, as small as 5 nm, were prepared by the precipitation method, using hydrogen peroxide as an oxidizer and is reported to present even 100 °C lower methane oxidation temperature, as compared with microcrystalline CeO₂. Nabih et al. has also reported the synthesis of mesoporous ceria nanoparticles (NPs) by combining the sol-gel (SG) process with the inverse mini-emulsion technique. These materials were of high specific surface area and exhibited an even better catalytic performance than the reference samples, synthesized by precipitation [12,13]. The addition of MO_x (M: di- or tri-valent transition metal ion) in CeO₂ is interestingly reported to promote the catalytic activity for the oxidation reaction, a fact possibly related with improved redox properties and high oxygen storage capacity, induced by the synergistic effect between CeO₂ and MO_x [14–17]. A nanocomposite catalyst Co₃O₄/CeO₂ is very recently reported to exhibit high activity in complete oxidation of CH₄, suggesting a synergistic effect of CeO₂ nanorods and the supported Co₃O₄ nanoparticles [7].

In addition to oxidation performance, catalyst tolerance to deactivation, caused by poisoning and thermal aging, is also a problem in all natural gas applications. Poisoning is either due to adsorption of impurities, e.g., S, P, Zn, Ca, and Mg, present in the exhaust gases, on the catalytic active sites or their reaction with the active sites causing the formation of non-active compounds. Sulfur in the NG vehicles exhaust gas actually originates from odorants and lubricating oils and less from gas itself. Sulfur is considered as the most noxious component readily deactivating Pd catalysts, recognized as the most efficient for methane oxidation. Only small amounts of SO₂ can significantly inhibit catalyst activity by blocking the active noble metal sites by sulfur compounds [18–21]. Moreover, water is always present in the process of the catalytic combustion of methane, either as gas humidity

or as a combustion product. Water presence may affect the reaction kinetics, resulting in a serious inhibition of the oxidation reactions [14]. Thus, activity of Pd/Al₂O₃ is reported as severely inhibited by water at temperatures below 450 °C, a fact attributed to the formation of surface hydroxyl groups, which blocks active catalyst sites, while suppressed oxygen exchange between support and PdO via the formed hydroxyl groups is another recent explanation. Based on experiments and kinetic modelling two are the suggested routes of the water inhibition effect; rapid adsorption of water species on the active sites and slow build-up of hydroxyl species. Formation of these surface hydroxyls is enhanced by high concentration of oxygen and high concentration of water vapor [22]. In general, while various mechanisms have been proposed in the literature to explain the deactivation observed due to water presence, all studies agree on a strong effect of the support. Thus, tuning opportunely the metal-support interaction and the oxygen exchange capability of the support is a suggested way to deal with water inhibition problem over Pd/CeO₂ catalysts, otherwise very highly active for methane oxidation in dry conditions [23]. In a recent study, Pt addition was suggested to improve the performance of Pd-modified Mn-hexaaluminate catalyst in the high-temperature oxidation of methane, especially in SO₂ and water presence, correlating the improved water and sulfur resistance with the presence of particles of PtPd alloy [24]. In a similar study, presence of water vapor is reported to significantly inhibit the catalytic performance of CeO_2 -MO_x (M = Cu, Mn, Fe, Co, and Ni) mixed oxide catalysts for CH₄ combustion, while CeO₂-NiO samples showed the best durability for CH₄ wet combustion (T50: 528 °C, when 20% water vapor was added in the reaction stream) among all materials tested [14]. Interestingly, water presence in the reaction feed is even reported to promote methane conversion over a NiO catalyst, prepared from thermal decomposition of the corresponding metal nitrate. As activation of oxygen plays a crucial role in methane oxidation, in the surface kinetics controlled region, the observed high catalytic efficiency of NiO is attributed to its larger capability for oxygen adsorption, even at 500 °C. The promoting role of water could be due to the fact that H_2O molecules could modify the NiO surface and, thus, promote further activation of O₂ and/or CH₄ on the surface [25].

In a series of recent studies, we have investigated the impact of synthesis parameters on the solid state properties of CeO₂ materials. In detail, we followed four different, time- and cost-effective, preparation methods, i.e., thermal decomposition, precipitation, and hydrothermal method of low and high NaOH concentration, employing in all cases Ce(NO₃)₃·6H₂O as the cerium precursor. A thorough characterization of all samples was carried out to gain insight into the impact of synthesis route on the textural, structural, morphological, and redox properties of the derived materials. The results revealed the superiority of the hydrothermal method towards the development of ceria nanoparticles of high specific surface area (>90 m² g^{-1}), well-defined geometry (nanorods) and improved redox properties. Employing CO oxidation as a probe reaction evidenced a direct quantitative correlation between the catalytic activity and the abundance of easily reduced, loosely bound oxygen species. More specifically, the rod-like morphology of the hydrothermally synthesized CeO₂ nanoparticles, with well-defined (100) and (110) reactive planes, favored the enhanced reducibility and lattice oxygen mobility, rendering this material appropriate as catalyst or supporting carrier [26–28]. The current study aims in the preparation of novel Co-Ce catalytic materials targeting in the complete combustion of CH₄. Co incorporation on an optimum selected CeO₂ support aims to promote the catalytic activity for the oxidation reactions, via an attempted improvement of the redox property and high oxygen storage capacity, deliberately induced by the synergistic effect between CeO_2 and CoO_x . We purposely explored the effect of both the synthesis method of the CeO₂ support (following either the hydrothermal or precipitation technique) and the metal loading (0, 2, 5, and 15 wt% Co) on the catalytic performance and durability for complete methane oxidation. Deactivation studies were also performed examining tolerance of the optimum selected catalysts to thermal aging and presence of water vapor in the reaction feed (wet CH₄ combustion).

2. Results and Discussion

2.1. Catalyst Characterization

2.1.1. Textural and Structural Characterization

The physicochemical properties of catalytic materials are presented in Table 1. As expected [26], the synthesis method of CeO_2 greatly affects the specific surface area, the pore volume and the average pore diameter of the carrier. In addition, the mean particle size of CeO₂ presents important differences between the two synthesis methods. Applying the hydrothermal method (CeO₂-H), the formed ceria presents higher surface area (110.9 m^2/g) and smaller CeO₂ crystallites (11.8 nm) than the precipitated ceria (CeO₂-P), where the growing up of the ceria crystallites (18.6 nm) resulted in decreased surface area (49.5 m^2/g). The effect of synthesis method on the crystallite size is probably related with slower rates of crystal migration and crystallites grow up, when following the hydrothermal synthesis method. Incorporation of cobalt, on both CeO₂ carriers, as well as the increase of its metal loading, affects the texture of the CeO₂ support, decreasing the specific surface area of the derived, CeO₂-supported, Co catalysts. On the contrary, in the case of pure Co₃O₄, the different synthesis method does not appear to significantly affect the specific surface area. However, applying the hydrothermal method (Co₃O₄-H), the formed Co oxide presents lower pore volume and average pore diameter and similarly smaller Co₃O₄ crystallites than Co₃O₄-P (33.3 nm as compared to 42.2 nm, respectively). Lower crystallite sizes (16.1–16.8 nm) of Co₃O₄ nanomaterials, synthesized by hydrothermal method are reported in the recent literature, probably related with lower calcination temperature, thus limiting aggregation of Co_3O_4 crystallites in the sample [11]. It is noteworthy that much smaller Co_3O_4 crystallites are also formed on the CeO₂-H rather than the CeO₂-P support (16.0 as compared to 76.6 nm, respectively), when incorporating 15 wt% Co on either support via wet impregnation, a fact indicating synergetic effects between the CeO₂-H support and the incorporated Co oxide.

Catalwat	C = (19/)	S _{BET} (m ² /g)	Pore Volume (cm ³ /g)	Average Pore	Mean Particle Size (nm) ¹	
Catalyst	C0 (W1 /6)			Diameter (nm)	CeO ₂	Co ₃ O ₄
CeO ₂ -P	-	49.5	0.06	5.2	18.6	-
CeO ₂ -H	-	110.9	0.55	19.9	11.8	-
2Co/CeO ₂ -P	2.1	43.3	0.06	5.8	16.3	-
2Co/CeO ₂ -H	2.3	100.3	0.49	19.5	8.1	-
5Co/CeO ₂ -P	5.1	39.6	0.05	5.3	20.5	-
5Co/CeO ₂ -H	5.1	92.1	0.23	9.5	9.4	17.8
15Co/CeO ₂ -P	14.6	38.8	0.09	9.3	16.2	76.6
15Co/CeO ₂ -H	15.4	78.9	0.14	7.3	6.4	16.0
Co ₃ O ₄ -P	-	15.5	0.14	47.6	-	42.2
Co ₃ O ₄ -H	-	12.8	0.04	11.4	-	33.3

Table 1.	Physicochemical	properties of a	catalytic mate	erials.

 1 Measured by XRD diffractograms using the Scherrer equation, at (20) 28.6°, 47.5° for CeO_2 and 31.3°, 36.9° for Co_3O_4.

Representative XRD (X-Ray Diffraction) patterns of bare CeO₂, Co₃O₄ and *x*Co/CeO₂ prepared with either the precipitation or the hydrothermal method, are shown in Figure 1a,b respectively. As expected, both samples of CeO₂ supports (CeO₂-P and CeO₂-H) exhibit characteristic peaks predominately located at 2θ : 28.6°, 33.1°, 47.5°, 56.36°, 59.1°, 69.4°, 76.9°, and 79.2° and have face centered cubic structure with the lattice parameters a = b = c = 5.411 Å and $\alpha = \beta = \gamma = 90^{\circ}$ [29], which are also preserved in the Co-supported catalysts.

Bare Co oxides (Co₃O₄-P and Co₃O₄-H) show diffraction peaks at 2θ : 19.0°, 31.3°, 36.9°, 38.5°, 44.8°, 55.9°, 59.4°, and 65.2°, which are attributed to the cobalt oxide Co₃O₄ and suggest that the cobalt precursor led to the formation of a face centered cubic unit cell of Co₃O₄ (space group *Fd*3*m*) with a

spinel type structure after calcination [9,11]. At low Co loadings (2 wt.%), no diffraction peak related to Co_3O_4 was observed, while increasing Co loading (to 5 wt.% and 15 wt.%) leads to the development of low intensity Co_3O_4 peaks (2 θ : 31.3°, 36.9°, and 44.8°) for Co/CeO₂-P (Figure 1a) and Co/CeO₂-H (Figure 1b) catalytic materials, respectively.



Figure 1. XRD patterns of CeO_2 , Co_3O_4 and xCo/CeO_2 prepared following the (**a**) precipitation method; and (**b**) the hydrothermal method. (*x*: 2, 5 or 15 wt% Co loading).

2.1.2. Morphological Characterization

The HR-TEM (High Resolution-Transmission Electron Microscopy) images further confirm the exposed planes of the two 15 wt.%. Co catalysts, supported on differently synthesized CeO₂ carriers. In both cases, two different face-centered cubic phases are identified, on taken electron diffraction patterns: a face-centered cubic phase of CeO₂ with a = 0.541 nm, and a second face-centered cubic phase of Co₃O₄, with a = 0.808 nm. However, the morphology of the Co₃O₄/CeO₂ catalysts is further explored with HR-TEM. In detail, low magnification images in Figure 2 (Figure 2a,b correspond to CeO₂-P and CeO₂-H, respectively) show an overview of the two catalytic samples. A very well defined, crystalline hexagonal formation (nanocrystalline structures of CeO₂ are forming plates) is observed in the first case, while in the second case the CeO₂ nanocrystals show a nanorod structure, an expected morphology related with the hydrothermal synthesis applied during ceria preparation [26]. Moreover, in both HR-TEM micrographs showing a very well defined hexagonal formation of CeO₂-P (Figure 2c), and a nanorod structure of CeO₂ the (Figure 2d) respectively, the characteristic d-spacings of the (111) planes are evidenced for both CeO₂ supports, in agreement with the XRD patterns (characteristic peak, predominately located at 2 θ : 28.6°) [7].

As shown in Figure 3, both CeO₂ nanocrystals and Co₃O₄ nanoparticles are well crystallized, as evidenced by the clear lattice fringes in these HR-TEM micrographs. The coexistence of the Co₃O₄ and CeO₂ cubic phases is evidenced by fast Fourier transformation analysis (FFT). However, Co₃O₄ nanoparticles are supported on bigger CeO₂ nanocrystals with a size of about 25–30 nm (slightly larger than estimated via XRD patterns) over the Co/CeO₂-P sample (Figure 3a). The experimental lattice spacings of 0.285 nm and 0.243 nm, as well as the lattice spacings of 0.312 nm, can be assigned to (220) planes and (311) planes of the Co₃O₄ cubic phase, and to the (111) planes of the CeO₂ respectively. On the contrary, HR-TEM image of the Co/CeO₂-H sample (Figure 3b) shows a much smaller CeO₂ nanorod (<20 nm), a trend in agreement with the XRD characterization. A smaller Co₃O₄ nanoparticle with a size of about 15 nm, seems to be attached to these nanorod support. In this case, the experimental lattice spacings of 0.285 nm and 0.202 nm, as well as the lattice spacings of 0.312 nm, can be defined to (220) planes and (400) planes of the Co₃O₄ cubic phase, and to the (111) planes of the CeO₂, respectively [7,9].



Figure 2. Low magnification images (overview) of the CeO₂ specimens, synthesized using the precipitation, CeO₂-P (**a**) or the hydrothermal CeO₂-H (**b**) method. The corresponding HR-TEM micrographs of the CeO₂-P sample (**c**), and the CeO₂-H sample (**d**), respectively.



Figure 3. HR-TEM micrographs of the CeO₂-P (**a**) and CeO₂-H (**b**) specimens. As insets, FFT images taken at the white marked area are present for the two different cases, CeO₂-P (**a**) and CeO₂-H (**b**) samples, respectively.

2.1.3. H₂-TPR and O₂-TPD Studies

The reducibility of the catalysts was also investigated with H₂-TPR experiments and the corresponding reduction profiles are presented in Figure 4. Bare CeO₂-P and CeO₂-H (Figure 4a,b insets, respectively) exhibit two peaks; one at 500–515 °C and one at 780–790 °C. The former is attributed to the reduction of surface oxygen of ceria, while the second one to the reduction of the bulk oxygen of CeO₂ [29,30]. Regarding pure Co oxides synthesized by different methods (Co₃O₄-P and Co₃O₄-H in Figure 4a,b insets, respectively), a broad double peak is observed in both cases, which is

associated with the reduction of Co_3O_4 to CoO (T: 330–340 °C) and CoO to Co^0 (T: 400–460 °C) [30,31]. It is worth to mention that for both materials prepared with hydrothermal method (Co_3O_4 -H and CeO_2 -H) a blue shift is observed; the reduction peaks are shifted to lower temperatures, which indicates enhanced reducibility of both oxide materials under study.

Concerning the H₂-TPR profiles of xCo/CeO₂ catalysts, all samples present three reduction peaks, which are also shifted to lower temperatures, in comparison with the corresponding bare CeO₂ and Co₃O₄ oxides (Figure 4a,b), an observation assigned to the interaction of Co with ceria. The first two peaks are related to the reduction of Co₃O₄, while the third one is attributed to the reduction of bulk oxygen of ceria support. In addition, the corresponding H₂ consumption tends to increase as expected, as Co loading increases from 2 to 5, and finally to 15 wt.%.



Figure 4. TPR profiles of bare CeO_2 , Co_3O_4 (insets) and xCo/CeO_2 catalysts prepared with: (a) Precipitation and (b) Hydrothermal Method. (*x*: 2, 5m or 15 wt% Co loading).

As suggested in several studies, during methane combustion the lattice oxygen in the oxide surface contributes to the oxidation of CH_x , thus leading to the formation of an oxygen vacancy, available for oxygen adsorption from the gas-phase. Both the adsorbed oxygen, as well as the surface lattice oxygen, are considered crucial for oxidation reactions, as they contribute to the adsorption/activation of oxygen and, thus, the desired oxidation of intermediates on the catalyst surface [11,32,33]. Oxygen-temperature programmed desorption (O₂-TPD) experiments have evidenced that oxygen molecules were released from Co_3O_4 nanomaterials, suggesting that these oxygen molecules were adsorbed on the surface oxygen vacancies of the oxides [32]. These oxygen vacancies facilitate oxygen mobility of the oxide, forming the active oxygen species and, thus, enhancing the reaction activity with reductive molecules like H_2 and CH_4 [8,34]. Tuning the shape of the oxide catalyst, e.g. nanorods, nanocubes and nanosheets, is suggested to improve availability and abundance of these active oxygen species [26,35]. Especially, synthesis of Co_3O_4 catalysts of specific shape in order to provide more active oxygen species in the catalyst surface is reported in the recent literature [11].

Figure 5 shows the profiles of O₂ temperature-programmed desorption (O₂-TPD) of bare CeO₂ (prepared either following the precipitation or the hydrothermal method) and the corresponding 15 wt.% Co/CeO₂ supported catalysts. The bare ceria supports exhibit one peak at 418 °C, related with the adsorbed atomic oxygen evolved from the bulk of CeO₂ support [36]. The incorporation of 15 wt.% Co to CeO₂ carriers shifted the peak to lower temperatures, suggesting that O₂ desorption is facilitated, especially for the 15Co/CeO₂-H catalyst (observed shift from 418 to 376 °C) [30]. In addition, both Co-based catalysts exhibit two extra peaks at a higher temperature area (500–650 °C), which are related with the oxygen desorption from the surface of the Co oxides, while the peak observed at T > 700 °C corresponds to O₂ desorption from the bulk of Co₃O₄ [11].



Figure 5. O_2 -TPD profiles of the two bare CeO₂ supports (prepared with either the precipitation or the hydrothermal method) and the corresponding 15 wt.% Co/CeO₂ supported catalysts.

2.2. Catalytic Activity of Pure CeO₂, Pure Co₃O₄ and Co₃O₄/CeO₂ Oxides for Methane Oxidation

The catalytic performances of all materials (pure CeO₂, pure Co₃O₄ and Co₃O₄/CeO₂ composite catalysts) in complete oxidation of methane were investigated using a fixed-bed reactor and are presented in Figure 6a. Below 500°C, CH₄ conversion was <10% for pure CeO₂ supports, while the CeO₂ nanorods (CeO₂-H) exhibit a slightly improved performance, finally reaching 41% conversion at 600 °C (as compared to 31% conversion achieved with the CeO₂-P sample, respectively). Co incorporation via deposition of Co₃O₄ nanoparticles on both CeO₂ supports largely increased their catalytic activity for CH₄ combustion, which is further enhanced as Co loading increases from 2 to 5 and, finally, 15 wt.%. However, an improved catalytic performance is always observed on the CeO₂-H (ceria nanorods) supported materials. Co incorporation on CeO₂ nanorods achieves 10% CH₄ conversion

from 420 °C and finally reaches up to ~90% CH₄ conversion at 600 °C (15Co/CeO₂-H sample). It is worth to mention that on higher Co loadings (5 and 15 wt.%) a significant improvement (>15% higher conversion) is always observed, when using the CeO₂ nanorods support. This is probably related with the optimized physicochemical properties of this CeO₂-H support: smaller CeO₂ cystallites, combined with higher surface area and possibly some synergetic effects between the deposited Co₃O₄ and the CeO₂ carrier. These catalysts exhibit similar performance ($T_{50\%}$ = 520 °C) with materials reported in the recent literature [7], evaluated however at milder conditions (lower GHSV:18,000 h⁻¹ and different feedstock ~6.67% CH₄ and ~33.3% O₂). Efficiency at high GHSV feeds is very important and still remains a significant challenge, as reported in the literature [10].



Figure 6. Catalytic performance in complete CH_4 oxidation on pure CeO_2 , Co_3O_4 and nanocomposite Co/CeO_2 catalysts. (a) CH_4 conversion vs. temperature Flow rate: 900 cm³/min; GHSV~40,000 h⁻¹; (b) Effect of WHSV on CH_4 conversion for the optimum selected 15Co/CeO₂-H catalyst. Feed: 0.5% vol. CH_4 and 10% vol. O_2 , balanced with He.

In order to further exhibit superiority of our optimum material, $15Co/CeO_2$ -H was further evaluated differentiating the reaction conditions used; varying GHSV between 10,000, 20,000, and 40,000 h⁻¹ (Figure 6b). As expected, $15Co/CeO_2$ -H catalyst exhibits a significantly improved performance at lower GHSV reaching $T_{50\%} = 440$ °C, which is obviously much better when compared with similar reported catalysts (Co_3O_4/CeO_2 exhibiting $T_{50\%} = 475$ °C at WHSV of 9000 mL g⁻¹ h⁻¹) [7]. Moreover, the much lower temperature for 50% conversion of CH₄ ($T_{50\%}$) on Co_3O_4/CeO_2 -H, as compared with the bare CeO₂-H support ($T_{50\%} = 440$ °C as compared with $T_{50\%} = 575$ °C, respectively, at low GHSV: 10,000 h⁻¹ conditions), once more supports the significant role of the loaded Co_3O_4 catalysts is very interesting, as these materials reached complete (100%) oxidation of CH₄ from 520 °C independently of their synthesis method and thus deserve further investigation.

2.3. Catalyst Stability Tests

Thermal stability of a catalyst used in the CH₄ catalytic combustion is very important because of its expected operation at high temperatures. In general, thermal aging may cause crystal growth and loss of catalytic surface. Thus, catalyst stability over time of the optimum selected material (15Co/CeO₂-H) was tested under the flow of mixture of 0.5 vol% CH₄ and 10 vol% O₂, balanced with He. As shown in Figure 7, the conversion at 600 °C is rather stable (only a small decrease from 88 to 81% is observed) after 10 hours on stream, under a GHSV of 40,000 h⁻¹. This is probably attributed to the thermal stability of the particular oxides (CeO₂ and Co₃O₄), which are expected to deteriorate at higher operating windows (700–750 °C). Preservation of finely dispersed Co₃O₄/CeO₂ species after aging at 750 °C is reported to retain high methane oxidation activity and tolerance to thermal ageing, suggesting once more the importance of the preparation method to achieve intimate contact between the cobalt–ceria species leading to good oxidation activity [17].



Figure 7. Response of CH₄ combustion to long-term stability tests over the optimum selected 15Co/CeO₂-H catalyst; Flow rate: 900 cm³/min; GHSV~40,000 h⁻¹.; Reaction temperature 600 °C; Feed: 0.5% vol. CH₄ and 10% vol. O₂, balanced with He.

The presence of ~10 vol.% of water in the NG exhausts still remains one of the main challenges for the required catalytic system to control the emissions of NG vehicles. Thus, in order to further test the water vapor effect on the catalytic performance of our optimum catalyst ($15Co/CeO_2$ -H) for methane oxidation, 5% and 10% water vapor was added in the process of CH₄ combustion and the corresponding catalytic performance is shown in Figure 8 in comparison with the case using the dry reaction gas. The amount of water added significantly exceeds the stoichiometric amount produced during combustion, under the conditions used in this work (0.5% CH₄, 10% O₂). Increasing the water vapor content in the reaction stream, slightly descends its catalytic performance, as $T_{50\%}$ is increased from 520 to 530 °C, when water vapor is added, while increase of water content (from 5 to 10 wt.%) does not induce any further inhibition effect. Thus, water presence does not lead to a significant activity fall (observed increase in $T_{50\%}$ is less than 20 °C), while water effect is even more limited at higher temperatures of methane combustion. On the contrary, effect of water presence is most evident on the initiation temperature of the CH₄ combustion, increasing T10 value from 420 to 450 $^{\circ}C$ [14]. Removing water from the reaction feed completely restored the initial activity of the catalyst, without the need to subject the catalyst to any regeneration/thermal treatment. The restored catalyst efficiency of the used catalyst suggests that the limited water inhibition effect is probably related with adsorption of water species on the active sites at the low temperature range. There is no evidence of hydroxyl species' build-up or sintering of Co species, which could lead to a more severe/permanent deactivation, in spite the high concentration of oxygen (10 wt%) and high concentration of water vapor (10 wt%) used, expected to facilitate formation of surface hydroxyls [22]. The steam reforming reaction of methane might be involved, during the CH₄ catalytic combustion in the presence of water vapor, in which CH₄ is partially oxidized to CO and H₂. However, analysis of the effluent gas hardly detected any CO, confirming that CH₄ is totally oxidized to CO₂ and H₂O over 15Co/CeO₂-H, even in the presence of water vapor (wet CH₄ catalytic combustion).



Figure 8. CH₄ catalytic combustion over 15Co/CeO₂-H (fresh and used) with different contents of water vapor addition (0, 5 and 10 wt%). Flow rate: 900 cm³/min; GHSV~40,000 h⁻¹.

3. Materials and Methods

The CeO₂ carrier was prepared with two different methods; the precipitation method (CeO₂-P) and the hydrothermal method (CeO₂-H). Cerium nitrate hexahydrate (Ce(NO₃)₃·6H₂O, 99% purity, supplied by Fluka, Steinheim, Germany) was used as a precursor salt for both synthesis methods. For the final production of 25 g of CeO₂-P we initially prepared an aqueous solution (0.5 M) of cerium nitrate dissolving ~63 g of Ce(NO₃)₃·6H₂O in 290 mL of double distilled water. Then a precipitating agent (25 vol% NH₃ solution) was added at room temperature to the continuously stirred solution of the cerium nitrate, until the pH of the solution reached the value of 10. Then the pH remained stable at this value for additional 3 h. Following precipitation, the resulting precipitate was filtered, dried overnight at 110 °C and then calcined at 500 °C for 5 h under air flow. For the preparation of CeO₂-H, 0.023 mol of cerium nitrate was dissolved in 46 mL of double distilled water. Then, 125 mL of a 1 M NaOH solution was added rapidly under vigorous stirring. The mixed solution was placed in a 1 L Teflon bottle, which was sealed and placed in an oven at 110 °C for 5 h. The as-obtained material was washed several times with double distilled water and ethanol, and then dried and calcined, following the same thermal treatment with the precipitated one.

The incorporation of cobalt over the CeO₂ supports was conducted by applying the incipient wetness impregnation method. A series of Co catalysts supported on CeO₂-P and CeO₂-H, were prepared in three different metal loadings (2, 5 and 15 wt%). Incorporation of the Co active phase was accomplished by using Co(NO₃)₂·6H₂O (99% purity, supplied by Merck, Darmstadt, Germany). Aqueous solution of the metal salt was impregnated on CeO₂ supports, in successive stages, with intermediate drying in an oven at 100 °C for 30 min. The derived samples were dried at 110 °C overnight and finally calcined under air flow at 500 °C for 5 h. The as prepared composites are herein labeled as *x*Co/CeO₂-P or H, where *x* is the cobalt loading.

For comparison reasons, bare Co_3O_4 was prepared by precipitation (Co_3O_4 -P) and hydrothermal (Co_3O_4 -H) method. $Co(NO_3)_2 \cdot 6H_2O$ was used as a precursor salt and the same precipitation method was followed as for the synthesis of CeO_2 -P. For the preparation of Co_3O_4 -H, 0.0515 mol of cobalt nitrate was dissolved in 51 mL of double distilled water and 125 mL of a 1 M NaOH solution was added rapidly under vigorous stirring. Then, the next steps (aging, washing, and thermal treatment) were the same with the hydrothermal synthesis of ceria.

The physicochemical characteristics of as-synthesized materials were evaluated by various complimentary techniques. The total Co loading (wt%) of the final catalysts was determined by the Inductively Coupled Plasma Atomic Emission Spectroscopy (ICP-AES -on a Perkin-Elmer Optima 4300DV apparatus, Waltham, MA, USA). The textural characteristics of the as prepared catalysts were determined by the N₂ adsorption-desorption isotherms at -196 °C (Nova 2200e Quantachrome flow apparatus, Boynton Beach, FL, USA). Specific surface areas (m^2g^{-1}) were obtained according to the Brunauer-Emmett-Teller (BET) method at relative pressures in the 0.05–0.30 range. The specific pore volume (cm^3g^{-1}) was calculated based on the highest relative pressure, whereas the average pore size diameter (dp, nm) was determined by the Barrett-Joyner-Halenda (BJH) method. Prior to measurements the samples were degassed at 250 °C under vacuum. The crystalline structure of the catalysts was determined by powder X-ray diffraction (XRD) on a Siemens D 500 diffractometer (Bruker, Karlsruhe, Germany) operated at 40 kV and 30 mA with Cu K α radiation (λ = 0.154 nm). Diffractograms were recorded in the 5–80° 2θ range and at a scanning rate of 0.02° s⁻¹. The Scherrer equation was employed to determine the primary particle size of a given crystal phase based on the most intense diffraction peaks of CeO₂ (28.6° , 47.5°) and Co₃O₄ (31.3° , 36.9°). The size, morphology and structure of the two 15 wt% Co-supported catalysts (either on CeO₂-H or CeO₂-P supports) were also investigated by transmission electron microscopy using a (JEOL JEM 2011 TEM, Zaventem, Belgium) operating at 200 kV with an atomic resolution of 0.194nm. The HR-TEM micrographs were analyzed using Digital Micrograph v2 software (Gatan, Inc., München, Germany).

Hydrogen temperature-programmed reduction (H₂-TPR) experiments were performed in a different bench-scale unit, also available in our facilities for TPX studies, loading a fixed bed tubular reactor with 0.1 g of sample, while the reactor exit was coupled with a mass spectrometer (MS). A flow (50 cm³/min) of 5 vol% H₂/He gas mixtures was used, while the temperature of the solid catalyst was increased to 800°C at the rate of 10 °C/min. The effluent gas from the reactor was analysed using the MS detector (m/z = 2 was used for H₂). Prior to reduction, the as prepared sample was loaded to the reactor and pre-treated at 300°C for 1h under He flow. The catalyst sample was then cooled in He gas flow to room temperature, before recording the H₂-TPR trace. Oxygen temperature-programmed desorption was performed (O₂-TPD) in the same unit. The temperature was increased to 400 °C (heating rate 5 °C/min) under 20 vol% O₂/N₂ (50 cm³/min), for 30 min and were cooled to room temperature to adsorb oxygen for 30 min. After that, the samples were purged under He flow for 1 h, in order to purge physically adsorbed oxygen. The desorption step was conducted under He flow (50 cm³/min) up to 750 °C at a rate of 10 °C/min and the signal of O₂ (m/z = 32) were detected by the mass spectrometer.

The catalytic combustion of CH_4 was performed in a fixed-bed reactor loaded with 0.6 g of catalyst. The total gas flow rate was 900cm³/min, corresponding to a gas hourly space velocity (GHSV) of ~40,000 h⁻¹. The composition of the feed was 0.5 vol% CH_4 and 10 vol% O_2 , balanced with He.

The effect of GHSV was also investigated at 40,000, 20,000, and 10,000 h⁻¹. The CH₄ conversion was monitored in the 300–600 °C range (in a decreasing temperature mode) and was held constant at each temperature for ~20 min prior taking measurements. The composition of the effluent gas from the reactor was analysed using gas chromatography (GC-FID, Agilent 7890, Walborn, Germany). Stability of optimum catalyst's performance (time-on-stream) was tested at 600 °C for 10 h (under 0.5 vol% CH₄, 10 vol% O₂ in He flow). In addition, the effect of H₂O content (0, 5, and 10% vol. H₂O in He) was studied on CH₄ oxidation efficiency.

4. Conclusions

Various novel Co-Ce catalysts were prepared following different preparation techniques for the synthesis of ceria support (precipitation or hydrothermal method) and incorporating different loadings of Co on either support. The synthesis route results in ceria carriers of different physicochemical properties, further modified after Co incorporation. Hydrothermal synthesis leads to an improved CeO₂ support, with smaller crystallite size, larger surface area and enhanced reducibility. The ceria support and, thus, the synthetic procedure is important in terms of dispersing the active Co component (smaller Co₃O₄ crystallites formed on the CeO₂-H support) and exhibiting higher oxygen mobility $(O_2$ desorption peak of larger area, additionally appearing at lower temperature), as deduced by comparing the 15Co/CeO₂-H and 15Co/CeO₂-P catalytic samples. In summary, a nanocomposite catalyst Co₃O₄/CeO₂-H exhibits higher activity in complete oxidation of CH₄, than Co₃O₄/CeO₂-P and pure CeO₂ materials. The higher dispersion of the deposited Co species and the enhanced reducibility of Co-Ce catalysts advocate synergistic effects of CeO2 nanorods and the supported Co3O4 nanoparticles. Moreover, the addition of water vapor in the reaction stream only slightly inhibited the catalytic activity of the optimum selected 15Co/CeO₂-H catalyst, a fact mainly observed at lower temperature ranges. Thus, the corresponding material is very promising for methane oxidation purposes, as it additionally presented good water resistance properties in wet CH₄ combustion.

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Article Electrochemical Promotion of Nanostructured Palladium Catalyst for Complete Methane Oxidation

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Abstract: Electrochemical promotion of catalysis (EPOC) was investigated for methane complete oxidation over palladium nano-structured catalysts deposited on yttria-stabilized zirconia (YSZ) solid electrolyte. The catalytic rate was evaluated at different temperatures (400, 425 and 450 °C), reactant ratios and polarization values. The electrophobic behavior of the catalyst, i.e., reaction rate increase upon anodic polarization was observed for all temperatures and gas compositions with an apparent Faradaic efficiency as high as 3000 (a current application as low as 1 μ A) and maximum rate enhancement ratio up to 2.7. Temperature increase resulted in higher enhancement ratios under closed-circuit conditions. Electrochemical promotion experiments showed persistent behavior, where the catalyst remained in the promoted state upon current or potential interruption for a long period of time. An increase in the polarization time resulted in a longer-lasting persistent promotion (p-EPOC) and required more time for the reaction rate to reach its initial open-circuit value. This was attributed to continuous promotion by the stored oxygen in palladium oxide, which was formed during the anodic polarization in agreement with p-EPOC mechanism reported earlier.

Keywords: electrochemical promotion; NEMCA; palladium; ionic promoter; nanoparticles; yttria-stabilized zirconia

1. Introduction

Natural Gas Vehicles (NGVs) have gained considerable attention in the last decade due to much lower greenhouse gas emissions and lower price of methane compared to diesel or gasoline. Not only CH_4 is abundant in natural gas form, but methane can also be produced using anaerobic digestion technologies of bio-derived sources [1–4]. Despite lower emissions of NGVs they often suffer from incomplete methane combustion. Because methane is also 23 times more potent in warming the atmosphere than carbon dioxide its complete conversion to CO_2 is paramount. Therefore, the development of efficient low temperature catalysts for deep oxidation of methane (CH_4) has recently attracted significant attention [5–10].

Palladium-based catalysts are considered the most efficient for methane activation in excess of oxygen and their activity depends on temperature, methane/oxygen ratio, and catalyst surface oxidation state and composition [5]. The nature of the active surface sites PdO_x vs. Pd was a subject of several studies [6,7]. It was shown that chemisorbed oxygen on Pd metal is poorly active, whereas Pd oxidation with an optimum of 3–4 monolayers forms an active PdO catalyst. Chemisorption of a first layer of oxygen is fast; however, partial bulk oxidation is relatively slow [7].

The innovative field of catalysis that could boost complete methane oxidation reaction over Pd is electrochemical promotion of catalysis (EPOC) or non-Faradaic electrochemical modification of catalytic activity (NEMCA). This general, well-established phenomenon in catalysis aims at controlling in-situ both the activity and the selectivity of a catalyst through application of electric stimuli [8–11].

EPOC is observed with solid electrolyte materials that serve as catalyst support. Ions contained in these electrolytes (O^{2-} , H^+ , Na^+ OH^- , etc.) are electrochemically pumped to the catalyst surface, where they act as promoting species leading to modification of catalyst electronic properties and as a result its catalytic activity and selectivity. More precisely, applying an anodic polarization results in the strengthening of electron-donor adsorbates, e.g., chemisorbed methane, and weakening of the binding strength of electron-accepting adsorbates, e.g., dissociatively chemisorbed oxygen [12]. The resulting electrochemical activation magnitude is much higher than that predicted by Faraday's law. [13,14]. The increase in the catalytic rate, $\Delta r \pmod{O/s}$ divided by the electrochemical rate, I/nF(I is current, F is Faraday's constant and *n* is number of electrons, 2 for O^{2-}) is denoted as the apparent Faradaic efficiency, Λ , and the process is considered non-Faradaic when $|\Lambda|$ is greater than 1. Another parameter used to quantify EPOC is the rate enhancement ratio, ρ , which is the ratio between the promoted closed-circuit catalytic rate, *r* and the unpromoted open-circuit catalytic rate, r_0 .

The electrochemical promotion of complete methane oxidation was investigated on palladium catalysts prepared using various methods as summarized in Table 1 [15–23]. Electrochemical promotion of Pd thick film catalyst electrode prepared using wet impregnation was investigated in the temperature range of 470–600 °C [18]. The rate enhancement of 40% and an apparent Faradaic efficiency of 1.85 were found in this work. The addition of CeO₂ layer between the YSZ solid-electrolyte and Pd film catalyst increased the open-circuit catalytic rate but decreased the apparent Faradaic efficiency due to the higher electric resistance [18]. Another study on a Pd film catalyst prepared using commercial organometallic paste showed higher enhancement ratio ($\rho = 5.6$) and Faradaic efficiency ($\Lambda = 579$) at 560 °C; however, instability of the catalyst over time and rapid deactivation within 900 min of experiment was also noted [20]. Furthermore, the effect of CeO_2 layer was studied in [24]. It was shown that presence of ceria increased catalytic activity of Pd due to the formation of an active PdO phase, which was stabilized by CeO_2 acting as a continuous source of oxygen, similarly to the oxygen migration from the YSZ electrolyte under EPOC. In another study, an addition of porous YSZ layer between Pd film-catalyst and the dense YSZ solid-electrolyte resulted in a high open-circuit catalytic rate. The authors reported an enhancement ratio of 1.2 and apparent Faradaic efficiency, Λ , of 17 under fuel-rich conditions [23]. EPOC of sputtered Pd catalyst-electrode was compared to impregnated Pd for methane complete oxidation. The sputtered catalysts showed slightly higher Λ of 12 but similar enhancement ratio ($\rho = 1.6$) [21]. The effect of metal loading and catalyst thickness on EPOC of Pd was studies on Pd catalyst prepared by physical vapor deposition (PVD). It was found that metal loading and catalyst thickness have significant effect on the open-circuit rate and electrochemical promotion, where the thinner films resulted in the highest reaction rates per gram of catalyst at 500 °C [17]. A scaling-up of the system was attempted by electroless deposition of Pd in the channels of a YSZ monolith honeycomb; however, decrease in the conversion of methane occurred under positive and negative polarization [19].

Therefore, from previous EPOC studies on Pd for complete CH_4 oxidation, it is clear that the catalyst preparation method has a strong influence on Pd morphology, structure, oxidation state and as a result, on its catalytic activity, degree of promotion and stability under open and closed circuit conditions. Furthermore, from the practical point of view it is essential to work with low loadings of noble-metal catalysts that exhibit high dispersion and large active surface area In the present work, we report electrochemical promotion of nano-structured, highly dispersed Pd catalyst prepared by polyol reduction method for CH_4 complete oxidation in the temperature range of 400 and 450 °C and various gas compositions.

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Catalyst Synthesis Method	Loading	I or U Applied	T/°C	P _{CH4} /kPa	P _{O2} /kPa	Total Flow/ccm	Rate Change/mol O.s ⁻¹ 10 ⁻⁸	σ	V	Authors, Year & Ref.
Paste coating	n/a	+300 μA	400	2.75	1.55	n/a	4.52 to 20.5	4.5	103	^a Giannikos et al. (1998) [15]
Paste coating	n/a	+1 V	400	2.6	1.9	n/a	0.295 to 20	68	153	^a Frantzis et al. (2000) [16]
PVD	24 µg	+100 μA	500	2	10	166	7.3 to 20.4	2.8	258	Roche et al. (2008) [17]
Thermal decomposition	$1.1 \text{ mg/cm}^2$	+10 mA	600	0.4	1	150	0.47 to 0.68	2.6	7	^{a,b} Jimenez-Borja et al. (2009) [18]
Electroless deposition	5 mg total	+100 μA	400	2	10	166	136.4 to 135.2	0.99	-23	Roche et al. (2010) [19]
Paste coating	$7 \text{ mg/cm}^2$	+25 μA	560	0.4	1.2	150	1.6 to 9.1	5.6	579	^a Jimenez-Borja et al. (2011) [20]
Sputtered	$0.4 \text{ mg/cm}^2$	+1 mA	350	1.3	4.5	200	11.4 to 18.2	1.6	12	Matei et al. (2012) [21]
Impregnation	$0.4 \text{ mg/cm}^2$	+300 µA	350	1.3	4.5	200	22 to 26	1.18	25	Jimenez-Borja et al. (2012) [22]
Impregnation	$0.4 \text{ mg/cm}^2$	+5 mA	400	1.4	2.8	200	135 to 158	1.2	17	Matei et al. (2013) [23]
Poly ol method	$0.3 \mathrm{mg/cm^2}$	+0.5 V	425	2	4	100	6 to 16	2.66	383	^a This work
		^a Contir	nuous incre	ease in closed	-circuit rate	e post EPOC; ^t	Electrolysis effect.			

# 2. Results and Discussion

Transmission electron microscopy (TEM) was used to determine the palladium morphology and particle size (Figure 1a,b). The resulting Pd particles are spherical in shape with a diameter of approximately 5 nm, that are coalesced together in larger aggregates of roughly 50 nm in size. Figure 1c,d shows SEM images of as-prepared Pd/YSZ catalysts and the same catalyst after catalytic measurements under open circuit and EPOC conditions. It can be seen that as-prepared catalyst-electrode forms a highly dispersed, non-continuous layer on YSZ surface that consists of fine grains and pores. After the reaction, the "spent" catalyst shows much larger catalyst islands indicating a change in the morphology due to the catalyst agglomeration that takes place during the reaction. The resulting energy dispersive X-ray spectroscopy (EDS) spectrum (Figure 2a) of as-prepared Pd shows that the only element present is Pd. The x-ray diffraction pattern (XRD) contains several diffraction peaks of Pd (111), (200) and (220) corresponding to face-centered cubic (fcc) structure typical for bulk palladium metal. The crystallite size of Pd found from Pd (111) was 8 nm.



**Figure 1.** (**a**,**b**) TEM images of stand-alone Pd NPs at 5 nm and 50 nm scale; (**c**,**d**) SEM images of as-prepared (**c**) and post-experiment (**d**) Pd catalyst deposited on YSZ solid electrolyte.



Figure 2. (a) EDS spectrum and (b) XRD pattern of Pd catalyst.

The open-circuit catalytic rate of methane oxidation was tested at two different temperatures and at various gas compositions (Figure 3). It can be seen that for both temperatures the rate increased as a function of oxygen partial pressure. The increase was more significant at 4 kPa of methane where oxygen-to-methane partial pressures ratios were lower. The rate increase in fuel-rich condition is indicative of a Langmuir-Hinshelwood mechanism [15] where methane is able to competitively adsorb on palladium as seen at 450 °C under 4 kPa of CH₄, while the quasi-stable value at higher ratio of oxygen (at 2 kPa of CH₄) can be explained by an Eley-Rideal mechanism as CH₄ reacts on the oxygen covered surface [25].



Figure 3. Effect of oxygen-to-methane ratio on catalytic rate at 350 and 450 °C.

Figure 4 shows the transient rate response to an application of positive current (20  $\mu$ A) between the Pd working electrode and the counter electrode at 450 °C. Under open-circuit conditions (t < 0.5 h), the catalytic rate was at 8.35 × 10⁻⁸ mol/s. When a constant current was imposed, the reaction rate gradually increased due to pumping of O²⁻ promoters to the catalyst surface. After 2 h of polarization, the closed-circuit rate increase was 180% higher than its corresponding open-circuit value. In addition, the non-Faradaic behavior resulted in an apparent Faradaic efficiency,  $\Lambda$ , of 610 denoting that the back-spillover of O²⁻ at a I/2F rate gave a 610 times increase in the overall oxidation rate [26]. It should be noted that slight increase in the catalyst-working electrode potential (U_{WR}) was observed upon positive polarization (from 0.39 to 0.42 V). Furthermore, the slow reaction rate increase did not reach a steady-state value even after 2 h of applied polarization and took over 1 h to reach the open circuit value observed before polarization. According to the mechanism of EPOC the reaction rate is due to the supply of oxygen ionic species from YSZ and the formation of an effective double layer at the surface of the catalyst that changes the work function of Pd leading to weakening of the chemisorbed oxygen bond strength thus facilitating the  $-O_2C$  desorption from the catalyst surface.

A similar behavior was observed under potentiostatic conditions ( $U_{WR} = 0.25$  V) and a stoichiometric flow of reactants at 425 °C (Figure 5). The catalytic rate slowly increased in value until it reached 7.0 × 10⁻⁸ mol s⁻¹ after 2 h. As in the galvanostatic conditions (Figure 4) the closed-circuit reaction rate was continuously increasing with time without reaching a steady-state. After 2 h, the rate enhancement ratio of 1.31 and, an apparent Faradaic efficiency of 1107 were obtained. The high  $\Lambda$  value is due to the low current that passed through the cell, which was sufficient to promote Pd catalyst.



**Figure 4.** Transient rate response of Pd nanoparticles to a current step change. o.c.: open-circuit. Conditions: T = 450 °C, 2 kPa of CH₄ and 4 kPa of O₂. Flow rate: 100 ccm.

The continuous increase of the catalytic rate in Figures 4 and 5 can be explained by the continuous oxidation of palladium catalyst to  $PdO_x$ , which makes the catalyst more active for methane oxidation. In Figure 5, this continuous increase in catalytic rate occurred simultaneously with a decrease in current at a constant applied potential value indicating Pd oxide formation. Upon current or potential interruption the rate slowly returned to its initial state, because oxygen species stored in  $PdO_x$  continued acting as sacrificial promoters for methane oxidation reaction.



Figure 5. Transient rate response of Pd nanoparticles to potential step changes. o.c.: open-circuit. Conditions: T = 425 °C, 2 kPa of CH₄ and 4 kPa of O₂. Flow rate: 100 ccm.

To confirm  $PdO_x$  formation and oxygen storage effect on p-EPOC, the catalyst was polarized for a different duration 3, 6 and 10 h. As seen in Figure 6, the closed-circuit catalytic rate was continuously increasing with time under constant current (40  $\mu$ A for 3 h) and potential (0.5 V for 6 and 10 h) application. It can be seen that even after ten hours of polarization, the closed-circuit rate kept on rising without reaching a steady-state. At the same time, the longer polarization time resulted in longer decrease of the open-circuit rate after polarization was stopped. A proportional relationship (as shown in the inset Figure 6) was found between the duration of EPOC and the time to reach the open-circuit rate  $r_{o.c.}$ . The slope of this relationship was 0.5. This persistent electrochemical promotion (p-EPOC) is due to the stored oxygen ions in PdO_x that act as sacrificial promoters when the electrical circuit is open [27].



Figure 6. Cont.



**Figure 6.** (a) Transient rate response of Pd at different duration of EPOC and (b) the potential/current read at potentiostatic or galvanostatic application. Conditions:  $T = 425 \degree C$ , 2 kPa of CH₄ and 4 kPa of O₂. Flow rate: 100 ccm.

This indicates that during the positive polarization two parallel processes take place: i. Migration of  $O^{\delta-}$  promoters to the gas exposed catalyst surface and ii  $PdO_x$  formation at the three-phase boundary (tpb) according to the electrochemical reaction:

$$Pd + xO^{2-} \to PdO_x + xe^- \tag{1}$$

Current decrease and potential increase upon positive potentiostatic (Figures 5 and 6) and glavanostatic polarization (Figure 4), respectively, confirms the formation of  $PdO_X$ . Palladium oxide has lower conductivity than Pd metal, therefore current that flows through the solid-state cell or potential difference of the working catalyst-electrode ( $U_{WR}$ ) are the clear indication of an electrochemical oxide formation [28].

Figure 7 shows a transient rate response at a constant applied potential 0.5 V for 24 h. The reaction rate continuously increased for up to 20 h followed by 2 h of a steady-state rate and then a slight rate decrease. The continuous rate increase indicates constant catalyst activation due to the growth of  $PdO_x$ , whereas somewhat rate decrease after 22 h of polarization at 425 °C may be linked with morphology change observed for the "spent" catalysts (Figure 1d). The open-circuit rate took more than 6 h to return to its initial value showing a persistent promotional effect.



Figure 7. Long period transient rate response of Pd to a potentiostatic step change. o.c.: open-circuit. Conditions: T = 425 °C, 2 kPa of CH₄ and 4 kPa of O₂. Flow rate: 100 ccm.

Figure 8 shows the catalytic rate and catalyst potential change upon application of constant current as low as 1  $\mu$ A. This resulted in the continuous catalytic rate increase for 1, 2 and 4 kPa of O₂, accompanied by catalyst potential (U_{WR}) increase, confirming palladium oxide formation. The corresponding  $\Lambda$  values are summarized in Figure 9.



**Figure 8.** (a)Transient rate response of Pd catalyst at different O₂ partial pressure under 1  $\mu$ A galvanostatic application and (b) the corresponding catalyst potential response. Conditions: T = 450 °C, 2 kPa of CH₄. Flow rate: 100 ccm.

In Figure 9, the effect of partial pressure on closed-circuit reaction rate was tested at different galvanostatic conditions. The highest increase in catalytic rate was found at slightly fuel-rich conditions resulting in a  $\rho$  value of 1.3; the higher rate can be explained by the advantaged adsorption of gaseous methane over oxygen. At this condition, gaseous methane can be expected to directly adsorb onto Pd, resulting in a competition between oxygen and methane adsorption following a Langmuir-Hinshelwood mechanism. In addition, the desorption of oxygen from the surface becomes facilitated at lower oxygen partial pressure in the atmosphere as the overall chemical potential of oxygen is reduced [29]. At a partial pressure ratio higher than the stoichiometric ratio, it is perceived that the catalytic rate increase is slightly lower. The slight decrease in the enhancement is due to the

competing adsorption of oxygen on the surface of Pd, putting slight mass-transfer limitations on the chemisorption of  $CH_4$ .



Figure 9. Current effect on catalytic rate in function of methane/oxygen ratio. T = 450 °C. Flow rate: 100 ccm.

Similarly, the effect of temperature on closed-circuit reaction rate was tested at different galvanostatic conditions. Figure 10 shows that there is an increase in catalytic rate as a function of temperature at all applied positive current values for 400, 425 and 450 °C. It can be noticed that upon application of small current of 1  $\mu$ A, a significant rate increase was detected, resulting in a logarithmic-shape relationship of rate increase versus applied current. In addition, a highest value of ~2400 was found for the apparent Faradaic efficiency, constructing that a very minimal current was able to result in a change of the Pd surface oxidation state and hence the adsorption strength of methane reactant [15,16,18,20].



Figure 10. Current effect on catalytic rate in function of temperature: 400, 425 and 450 °C.  $P_{CH4}$ = 2 kPa and  $P_{O2}$  = 4 kPa. Flow rate: 100 ccm.

Table 1 compares EPOC of methane oxidation on Pd/YSZ found in this work to previous studies carried out on Pd catalyst-electrode deposited on YSZ solid-electrolyte. The table depicts where our results fall in comparison with previous experiments. It should be noted that the metal loading used in this work is the lowest in the temperature range of interest (T  $\leq$  450 °C), which is important for cold-start emission application. In agreement with previous work, Pd nanostructured catalyst synthesized by polyol method shows electrophobic type of EPOC, where only positive polarization promotes the reaction. Applied polarization led to the supply of O^{$\delta$ -} promoters from YSZ electrolyte to the catalyst surface, resulting in the formation of a more active phase of PdO, on the surface first and in the bulk gradually.

Ionic oxygen migration to the surface altered the adsorption properties of the catalyst surface, resulting in the weakening of gaseous oxygen adsorption and strengthening that of electron-donor methane. The alteration of the catalytic oxidation state was similar under both potentiostatic and galvanostatic application, which have resulted in a continuous increase in catalytic rate and a very slow open circuit rate decrease when the circuit was interrupted. This was explained by the formation of PdO during polarization and oxygen storage in its bulk, which was continuously providing the promoting oxygen species to the surface post-polarization according to p-EPOC mechanism [27].

# 3. Experimental

# 3.1. Synthesis of Pd Nanoparticles

Mono-metallic Pd NPs were synthesized using 0.133 g of palladium chloride (Fisher Scientific[®], Canada) precursor salts dissolved in 25 mL of ethylene glycol and 0.8 M NaOH. The mixture was heated up to 160 °C and kept under stirring conditions for 3 h. The final colloidal solution was cooled down and washed repeatedly with ethanol.

# 3.2. Catalyst Characterization

X-ray diffraction (XRD) was performed on the fresh Pd sample using Rigaku Ultima IV multipurpose diffractometer (Rigaku, The Woodlands, TX, USA). The diffractometer was equipped with an X'Celerator detector with monochromatic CuK $\alpha$  radiation ( $\lambda$  = 1.5418 Å) at 40 kV and 44 mA with a divergence slit of 2/3 degree, a scan speed of 0.03 °/s and a scan step of 0.02 degrees between 30 and 80° 20.

The transmission electron microscopy (TEM) micrographs were obtained using JEOL JEM 2100F FETEM (JEOL, Peabody, MA, USA) operating with a field emission gun at an acceleration voltage of 200 kV. SEM micrographs were recorded using PhenomTM SEM (Nanoscience Instruments, Virginia, USA). Additional elemental analysis was performed using the energy dispersive X-ray spectroscopy (EDS) attachment.

#### 3.3. Electrochemical Cell and Reactor

The solid electrolyte is a 19 mm diameter and 1 mm thickness disk of 8 mol %  $Y_2O_3$ -stabilized  $ZrO_2$  (YSZ) (TOSOH[®], Grove city, OH, USA) fabricated following the procedure reported earlier [30]. Inert gold reference and counter electrodes were deposited on one side of the disk by applying thin gold paste coating (Gwent Group, Pontypool, UK) of 0.2 and 1 cm² surface areas, respectively. This was followed by annealing in air at 500 °C. The catalyst-working electrode was deposited on the other side of the solid electrolyte disk (1 cm² surface area) opposing to the counter electrode. To this end, mono-metallic Pd were dispersed in isopropanol and 10  $\mu$ L of a suspension were deposited at a time with intermediate drying at room temperature. The resulting total metal loading was 0.5 mg of Pd on YSZ. Catalytic measurements were carried out at atmospheric pressure in the single-chamber capsule reactor reported earlier [31,32]. The working electrode side of the electrolyte was pressed against a gold mesh (1 cm²) that served as a current collector, while the counter and reference electrodes were

pressed directly against gold wires [33]. Two type K thermocouples (Omega[®], Quebec, Canada) were placed in vicinity of the electrochemical cell, one for temperature control and one for data acquisition.

# 3.4. Catalytic and Electrochemical Measurements

The reaction gases were CH₄ (Linde, 99.99%), O₂ (Linde, 99.99%), and He (Linde, 99.997%) as a carrier gas. The total flow rate was constant at 100 mL min⁻¹. Gas composition was varied using MKS, 1259 C and 1261-C series flow meters and detected using non-dispersive infrared (NDIR) CO₂ gas analyzer (Horiba, VA-3000, Burlington, Canada). Constant electric current or potential were applied using a potentiostat-galvanostat (Arbin Instruments[®], MSTAT, College Station, TX, USA) connected to the electrodes of solid-electrolyte electrochemical cell.

# 4. Conclusions

Electrochemical promotion of Pd nanostructured catalyst was investigated for the methane oxidation reaction in the 400–450 °C temperature range. The promotion of the catalytic rate of Pd NPs was achieved under anodic polarization. Upon various potentiostatic and galvanostatic tests, non-Faradaic enhancement was achieved, most notably at 450 °C, under the application of 1  $\mu$ A, where  $\Lambda$  was equal to ~3000, higher than any previous EPOC study on Pd. Continuous increase in the reaction rate was found under EPOC conditions, due to the Pd oxide formation in the vicinity of the tpb. A proportional relationship was found between the duration of polarization, a persistent promotion (p-EPOC) was observed due to the promotion of the reaction by the stored oxygen, which was accumulated during positive polarization. Overall, our work revealed interesting behavior of Pd synthesized by polyol method, providing further insight into the application of electrochemical promotion for complete methane oxidation with highly dispersed Pd.

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Article

# Synthesis of Sulfur-Resistant TiO₂-CeO₂ Composite and Its Catalytic Performance in the Oxidation of a Soluble Organic Fraction from Diesel Exhaust

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**Abstract:** Sulfur poisoning is one of the most important factors deteriorating the purification efficiency of diesel exhaust after-treatment system, thus improving the sulfur resistibility of catalysts is imperative. Herein, ceria oxygen storage material was introduced into a sulfur-resistant titania by a co-precipitation method, and the sulfur resistibility and catalytic activity of prepared TiO₂-CeO₂ composite in the oxidation of diesel soluble organic fraction (SOF) were studied. Catalytic performance testing results show that the CeO₂ modification significantly improves the catalytic SOF purification efficiency of TiO₂-CeO₂ catalyst. SO₂ uptake and energy-dispersive X-ray (EDX) results suggest that the ceria doping does not debase the excellent sulfur resistibility of bare TiO₂, the prepared TiO₂-CeO₂ catalyst exhibits obviously better sulfur resistibility than the CeO₂ and commercial CeO₂-ZrO₂-Al₂O₃. X-ray powder diffraction (XRD) and Raman spectra indicate that cerium ions can enter into the TiO₂ lattice and not form complete CeO₂ crystals. X-ray photoelectron spectroscopy (XPS), H₂-temperature programmed reduction (H₂-TPR) and oxygen storage capacity (OSC) testing results imply that the addition of CeO₂ in TiO₂-CeO₂.

**Keywords:** cerium-doped titania; sulfur-tolerant materials; organic compounds purification; diesel oxidation catalyst; vehicle exhaust

# 1. Introduction

Diesel engines have been widely used in passenger cars and vans, due to excellent fuel efficiency and durability. However, diesel exhaust gases, such as carbon monoxide (CO), unburned hydrocarbons (HC_x), nitrogen oxides (NO_x), particulate matter (PM) and soluble organic fraction (SOF), are considered major sources of air pollutants [1–3]. Among these hazardous pollutants, SOF are the heavy liquid hydrocarbons (C > 16 [4,5], aromatics and oxygenated compounds [6]) adsorbed on soot, which mainly come from unburned fuel and lube oils [7,8]. The content of SOF is known to vary with engine operating conditions and can reach about 5–60% of the whole mass of the particulate matter [6,9–11]. Due to the fact that diesel SOF contains types of polycyclic aromatic hydrocarbons (PAHs) [12,13] which are recognized as strong carcinogens [14,15], purifying the SOF from diesel exhaust is an important and essential work.

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Diesel oxidation catalyst (DOC) was employed to accelerate the oxidation and purification of diesel exhaust gases of CO,  $HC_x$  and SOF. In recent decades,  $CeO_2$ - $ZrO_2$ ,  $Al_2O_3$  and their mixed oxide-based catalysts were widely used as commercial DOC and displayed excellent catalytic CO and  $HC_x$  oxidation activity. Focusing on the purification of SOF,  $CeO_2$ -based oxygen storage materials (OSM) have been greatly impressed by considerable researchers due to the superior catalytic activity on hydrocarbons and SOF oxidation [16–18]. Meanwhile, controlled synthesized of nanostructured  $CeO_2$ -based catalysts and their catalytic performance in diesel soot oxidation are lucubrated [19,20]. However,  $CeO_2$ -based catalysts are easily poisoned by SO_2 [21–24]. And SO_2 is a subsistent in the diesel exhaust, since sulfur is present in almost all commercial diesel fuels [25–27], sulfur in fuels would be oxidized to SO_2 and then emitted from diesel engines [28,29]. Furthermore, sulfur poisoning resulting from sulfur species accumulation is more destructive, since even using ultra-low sulfur diesel (ULSD), cumulative exposure of a catalyst over its lifetime in a heavy-duty diesel may amount to kilograms of sulfur [30]. A large amount of sulfur species accumulation inevitably results in the blocking of channels of monolithic catalyst, and hence the strike of diesel exhaust after-treatment system. Thus, it is important and realistic for a diesel oxidation catalyst to improve the sulfur resistibility.

 $TiO_2$  is known as an effective sulfur-resistant material [31–33], and our previous works [34–36] also prove that  $TiO_2$ -based diesel oxidation catalysts display excellent sulfur resistibility. However,  $TiO_2$  is not active enough for catalytic diesel SOF combustion. Reports about the sulfur resistance catalyst for diesel SOF oxidation are still scarce. Considering all of this, in this work,  $CeO_2$  was introduced in  $TiO_2$  by a co-precipitation method, and its effects on sulfur resistibility and catalytic activity for diesel SOF combustion were investigated.

# 2. Results and Discussion

#### 2.1. Sulfur Resistibility

The sulfur resistibility values of catalysts were measured by sulfur uptake testing. As shown in Figure 1, under SO₂ exposure, the weight of all samples increased with time; final weight increments tended to be flat after 1–3 h SO₂ exposure, except for CeO₂. After 4 h, chosen as representative exposure time of simulative 160,000 km vehicle aged catalyst, the final weight increments of TiO₂, TiO₂-CeO₂ and CeO₂ catalysts are 1.63 wt. %, 2.01 wt. % and 4.72 wt. %, respectively. The normalized sulfur uptake results are calculated by supposing 1 g of sample as standard, and the results are listed in Table 1. The normalized sulfur uptake values of TiO₂, TiO₂-CeO₂ and CeO₂ catalysts are 166 µg/m², 170 µg/m², and 891 µg/m², respectively. From the results it can be seen that the sulfur species accumulation is severe on CeO₂ catalysts (TiO₂ and TiO₂-CeO₂) present obviously better sulfur resistibility than the CeO₂ catalyst. Since the non-sulfating material of TiO₂ displays low SO₂ adsorption and hence relieves sulfate generation and exhibits superior sulfur resistibility [32,37]. Furthermore, the introduction of moderate amounts of ceria in TiO₂ has essentially no effect on the naturally excellent sulfur resistibility of TiO₂.

Table 1. Sulfur accumulation and normalized sulfur	uptake of the TiO ₂	, TiO ₂ -CeO ₂ and	CeO ₂ samples.
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Samples	Surface Area ^a (m ² /g)	Sulfur Content	(wt. %)	- Normalized Sulfur Uptake ^c (ug/m ² )	
Samples	Surface Area (III /g)	Sulfur Uptake	EDX ^b	Normanzeu Surrur Optake (µg/m)	
TiO ₂	98	1.63	1.02	166	
TiO ₂ -CeO ₂	118	2.01	1.04	170	
CeO ₂	53	4.72	4.26	891	

^a Surface area was calculated by BET method from the N₂ adsorption-desorption results; ^b EDX results were obtained by detecting the simulative 160,000 km vehicle aged samples; ^c The normalized sulfur uptake = sulfur uptake/(100 × surface area) [37,38].



Figure 1. Sulfur uptake of the TiO₂, TiO₂-CeO₂ and CeO₂ catalysts.

Additionally, the accumulation amounts of sulfur species on simulative 160,000 km vehicle aged TiO₂, TiO₂-CeO₂ and CeO₂ catalysts tested by EDX (Table 1) are 1.02 wt. %, 1.04 wt. % and 4.26 wt. %, respectively, which shows the same trend with the sulfur uptake results. It is clear that the prepared TiO₂ and TiO₂-CeO₂ present significantly less sulfur species accumulation and better sulfur resistibility than the CeO₂ catalyst under the long-term exposure of diesel exhaust ambiences.

#### 2.2. Catalytic Performance

Figure 2 shows the thermogravimetry-differential thermal analysis-differential thermogravimetry (TG-DTA-DTG) curves of the bulk lube without catalysts and the lube impregnated on catalysts; all of the DTA curves are the positive peaks indicating exothermic peaks, and all of the originally DTG negative peaks are inverted into positive peaks for a better readability of the graph. The combustion of lube without catalyst under air flow is shown in Figure 2a; the weight loss of bulk lube is tersely distinguished into two stages; about 90% of lube is deflagrated at 220-350 °C, and then the rest of 10% lubricating oil is consumed tardily after 350 °C till 500 °C, which implies that the commercial lube contains a fraction of hydrocarbons hard to pyrolyze (may come from the lubricant additive). The onset combustion temperature of  $T_{10\%}$  (the temperature at which 10% of the initial lube is converted) is about 264 °C, the lube combustion fastest temperature of  $T_m$  (the temperature of weightlessness fastest point in DTG curves) is about 324  $^{\circ}$ C, and the final reaction temperature of  $T_f$  (the temperature of lube is completely converted) is about 507 °C. As shown in DTA curves of Figure 2a, a sharp and large exothermic peak is seen at about 325 °C which result from the rapid pyrolysis of bulk lube. Due to the decomposition of lube being an exothermic reaction, once ignition occurs, the heat continually increases and accumulates, and hence, most of lube is removed rapidly. The multiple peaks at 400–500 °C imply that the commercial lube contains multifarious hydrocarbons (lubricant additives) that are hard to decompose. Figure 2b plots the lubricant oxidation on CeO₂ catalyst. About 93% of the lube is rapidly oxidized between 140 °C and 280 °C, and the rest of 7% of lube is fully burnt at 280–340 °C. The onset combustion temperature of  $T_{10\%}$  is about 162 °C, the fastest weightlessness temperature of  $T_m$  is about 186 °C and the final reaction temperature of  $T_f$  is about 322 °C. Figure 2c shows the decomposition of lubricant with TiO₂, the TG-DTG curves can be divided into four stages with different decomposition rates. About 16% of lube is decomposed at 205–266 °C, another 28% of the lube is burnt at 266–324 °C, about 49% of the lubricant is consumed at 324–396 °C and the rest of 7% of uninflammable lube is ignited between 396 °C and 420 °C. The onset combustion temperature of  $T_{10\%}$  is about 249 °C, the final reaction temperature of  $T_f$  is about 420 °C, and three obviously fast weightlessness peaks of lube combustion are observed at about 252, 289 and 363 °C. For the TiO₂-CeO₂ catalyzed lube

combustion (Figure 2d), about 97% of the lubricating oil is rapidly combusted between 180 °C and 330 °C, and the rest of about 3% of lube is burnt out at 334–362 °C. The  $T_{10\%}$ ,  $T_m$  and  $T_f$  is about 212, 261 and 362 °C, respectively.



**Figure 2.** Simultaneous TG-DTA-DTG curves for simulating the catalytic performance for the combustion of SOF on (**a**) without catalyst; (**b**) CeO₂; (**c**) TiO₂ and (**d**) TiO₂-CeO₂ catalysts.

Due to the fact that the weight loss of lube can be ascribed either to evaporation or to combustion, the activity of prepared catalyst is also identified by the integrated area of the exotherm. The normalized DTA peak areas are described in units of  $\mu V^{\circ}C/(mg lube)/(mg sample)$ , and then different catalysts can be directly compared on a common basis [39,40]. The larger the value of normalized DTA exotherm peak area, the greater the fraction of lubricant combusted verses evaporated, and the better the catalytic performance [39-41]. The oxidation activity data of catalysts for lube combustion are listed in Table 2. CeO₂ catalyst exhibits an outstanding catalytic lube oxidation activity, which is consistent with our previous reports [40,41]. Although the introduction of TiO₂ slightly lowers the combustion temperature of lube, bare  $TiO_2$  is not active enough for catalytic SOF oxidation. About 60% of lube is burnt at 350 °C in the lube/TiO₂ sample, while the value of lube without catalyst sample is about 90%. This is because the lube oxidation is an exothermic reaction; once ignition occurs, the heat continually increases and accumulates. Thus, the burn of lube without catalyst (containing more lube oil) is more violently than the lube/TiO2. The TiO2-CeO2 catalyst obviously lowers the onset temperature of lube combustion and considerably promote the removing of lube resulting from combustion. Compared to TiO₂, the prepared TiO₂-CeO₂ catalyst presents obviously lower SOF removal temperature and larger exothermal peak area, which indicates that the TiO2-CeO2 catalyst presents better catalytic lube combustion activity.

Samples	$T_{10\%}$ (°C)	$T_m$ (°C)	$T_f$ (°C)	Exothermal Peak Area (μV·°C/mg Lube)	Exothermal Peak Area (µV·°C/(mg Lube)/(mg Sample))
lube	264	324	507	283.9	-
lube/CeO2	162	186	322	1068.1	119.5
lube/TiO ₂	249	_ a	420	901.6	102.6
lube/TiO ₂ -CeO ₂	212	261	362	1029.5	115.5

Table 2. Catalytic performances for the combustion of SOF over the prepared catalysts.

^a Three obviously fast weightlessness peaks are observed over the lube/TiO₂ sample.

#### 2.3. Catalyst Characterization

# 2.3.1. XRD and Raman Spectra

The XRD patterns of prepared catalysts are shown in Figure 3; both TiO₂ and TiO₂-CeO₂ display only characteristic peaks which refer to the typical anatase structure of TiO₂. The peaks of TiO₂ are sharper than those of TiO₂-CeO₂, which indicates that the addition of CeO₂ impedes the crystal growth and sintering and lower crystallinity of the TiO₂-CeO₂ composite materials. In the case of TiO₂-CeO₂, the typical reflections of CeO₂ crystals at 28.7°, 33.2°, 47.7°, 56.6° and 77.1° are not observed, and the positions of typical anatase structure of TiO₂ shift obviously to smaller angles, which suggest that a complete CeO₂ crystal is not formed and Ce ions (Ce⁴⁺ radius: 0.087 nm) possibly enter into the TiO₂ (Ti⁴⁺ radius: 0.06 nm) lattice and resulting in an expansion of TiO₂ unit cell, the unit cell volume of anatase tetragonal cell of TiO₂ is 134.95 Å³, for TiO₂-CeO₂, the value enlarges to 135.15 Å³. Thus, it can be inferred that Ce ions entered into the TiO₂ unit cell, and this is a possible reason why the addition of ceria into TiO₂ has no effect on the naturally excellent sulfur resistibility of TiO₂.



Figure 3. XRD patterns of the TiO₂ and TiO₂-CeO₂ catalysts.

To further confirm the above conjecture, Raman spectra were employed. As can be seen in Figure 4, CeO₂ presents a strong peak at about 464 cm⁻¹, which can be associated with the cubic CeO₂ [42]. TiO₂ and TiO₂-CeO₂ catalysts show five visible peaks at 145, 196, 397, 517 and 639 cm⁻¹, which are the A1g + 2B1g + 3Eg Raman-active modes of TiO₂ anatase phase (the peak at 517 cm⁻¹ is complex of A_{1g} and B_{1g}) [43], for the TiO₂-CeO₂ catalyst, the characteristic Raman peak of CeO₂ at 464 cm⁻¹ is not observed. This result further confirmed that a CeO₂ crystal is not formed in the TiO₂-CeO₂ catalyst, which is consistent with the XRD result.



Figure 4. Raman spectra of the CeO₂, TiO₂ and TiO₂-CeO₂ catalysts.

# 2.3.2. Nitrogen Adsorption-Desorption

The nitrogen adsorption–desorption isotherms and pore size dispersion of  $TiO_2$  and  $TiO_2$ -CeO₂ are shown in Figure 5; both  $TiO_2$  and  $TiO_2$ -CeO₂ are mesoporous materials and show distinct H3 and H4 complex hysteresis loops indicating slit pore features; compared to  $TiO_2$ , the prepared  $TiO_2$ -CeO₂ displays obviously larger pore size; the textual features are listed in Table 3.



Figure 5. Nitrogen adsorption-desorption isotherms and pore size dispersion (inset) of the  $TiO_2$  and  $TiO_2$ -CeO₂ catalysts.

Table 3. The texture properties of TiO₂ and TiO₂-CeO₂ catalysts.

Samples	Surface Area (m ² /g)	Pore Volume (cm ³ /g)	Average Pore Diameter (nm)
TiO ₂	98	0.22	7.2
TiO ₂ -CeO ₂	118	0.26	8.6

The surface areas of TiO₂ and TiO₂-CeO₂ catalysts are about 98 m²/g and 118 m²/g, respectively, which indicates that the introduction of CeO₂ slightly increases the surface area of TiO₂. The pore volume and average pore size of TiO₂ are about 0.22 cm³/g and 7.2 nm, respectively; the pore volume and average pore diameter of TiO₂-CeO₂ catalyst are 0.26 cm³/g and 8.6 nm, respectively. It can be seen that the addition of CeO₂ to TiO₂ increases its surface area, pore volume and average pore size; this is possibly due to the addition of CeO₂ which impedes crystal growth and sintering and lowers crystallinity of the TiO₂-CeO₂ composite materials, and this speculation is consistent with the XRD results (Figure 3). The larger surface area, pore volume and pore size can be advantages to the contacting, transmitting and diffusion of the lube molecules on the catalyst, and hence, be beneficial to the catalytic SOF oxidation activity.

# 2.3.3. XPS

Figure 6 shows the XPS spectra of O 1s region of the CeO₂, TiO₂ and TiO₂-CeO₂ catalysts, all samples show two XPS peaks, the peak at about 530.1 eV can be assigned to lattice oxygen, and the peak with a binding energy of 532.1 eV is characteristic of surface adsorbed oxygen [37,44]. Due to the superior oxygen storage capacity, the peak of surface adsorbed oxygen of CeO₂ is obviously stronger than the lattice oxygen; the surface adsorbed oxygen ratio is about 55%. For the TiO₂ catalyst, the surface adsorbed oxygen ratio is about 38%, and the addition of CeO₂ obviously increases the surface adsorbed oxygen ratio, where the value of TiO₂-CeO₂ is about 42%. Usually, surface adsorbed oxygen is more reactive than lattice oxygen is more active than the lattice oxygen in the catalytic SOF oxidation reaction [40]. Thus, the addition of CeO₂ in TiO₂ enhances the amounts of surface adsorbed oxygen of TiO₂-CeO₂, which can be responsible for the improved catalytic activity of SOF oxidation.



Figure 6. XPS (O 1s) spectra of the CeO₂, TiO₂ and TiO₂-CeO₂ catalysts.

Additionally, the XPS (Ti2p) spectra of TiO₂ and TiO₂-CeO₂ catalysts are both located at 458.5 eV  $(2p_{3/2})$  and 464.2 eV  $(2p_{1/2})$ , which are characteristic of TiO₂ species. Compared to bare CeO₂, the cerium peaks of TiO₂-CeO₂ are very weak and indiscernible, which indicates that the surface concentration of Ce in TiO₂-CeO₂ catalyst is very low; this phenomenon further confirms that Ce dopants are not gathering on the surface and are possibly entering into TiO₂ lattice, which is consistent with the XRD results (Figure 3).

# 2.3.4. H2-TPR

The redox property of a catalyst is closely related to the catalytic performance. The redox behavior of the prepared catalysts is described by hydrogen temperature-programmed reduction ( $H_2$ -TPR) and shown in Figure 7.



Figure 7. H₂-TPR profiles of the CeO₂, TiO₂ and TiO₂-CeO₂ catalysts.

The TPR peak of individual TiO₂ shows a weak peak at about 500–700 °C which may be ascribed to the reduction of TiO₂, and this phenomenon has been observed by other researchers [47,48]. The CeO₂ obviously shows a two-step reduction, the multiple low-temperature reduction peak at about 300–600 °C can be assigned to the reduction of surface oxygen, and the peak after 700 °C is ascribed to the reduction of CeO₂ bulk oxygen [47]. For the TiO₂-CeO₂, the reduction peak at 300–700 °C can be ascribed to the reduction of TiO₂ and surface oxygen of CeO₂; interestingly, the reduction peak of bulk oxygen of CeO₂ after 700 °C is poorly identified, which indicates that the TiO₂-CeO₂ catalyst exhibits lots of surface oxygen, and this result is consistent with XPS result. Consequently, the prepared TiO₂-CeO₂ catalyst with lots of surface oxygen species presents superior catalytic activity, due to the surface oxygen is highly reactive [45,46].

# 2.3.5. OSC

The CeO₂-based catalyst presents excellent oxygen storage capacity (OSC) and exhibits superior catalytic SOF oxidation activity [16,41]. The OSC of prepared catalysts were tested by an oxygen pulse injection technique and the results are listed in Table 4.

Samples	OSC (µmol O ₂ /g Sample)	Normalized OSC (µmol O ₂ /g CeO ₂ )
TiO ₂	2.9	-
TiO ₂ -CeO ₂ ^a	101	524
CeO ₂	73	73
$Ce_{0.35}Zr_{0.60}Nd_{0.05}O_2$	269 [49]	638

Table 4. The oxygen storage capacity (OSC) of TiO₂, TiO₂-CeO₂ and CeO₂-based catalysts.

^a The prepared TiO₂-CeO₂ catalyst in this work is Ce_{0.1}Ti_{0.9}O₂.

It can be seen that the OSC of TiO₂ is about 2.9  $\mu$ mol/g, which is very slight and may be within the measurement uncertainties. While the addition of CeO₂ significantly increases the OSC of TiO₂-CeO₂ catalyst, the OSC is about 101 ( $\mu$ mol O₂)/(g sample), and the normalized OSC of TiO₂-CeO₂ catalyst is

about 524 ( $\mu$ mol O₂)/(g CeO₂). For the pure CeO₂ sample, the OSC is about 73 ( $\mu$ mol O₂)/(g sample). It should be mentioned that the static oxygen storage capacity testing for OSC of pure CeO₂ is an undervalued result, due to the oxygen mobility of bare CeO₂ being very low. Thus, a CeO₂-based oxygen storage materials in our previous work [49] is used as reference, the normalized OSC of Ce_{0.35}Zr_{0.60}Nd_{0.05}O₂ is about 638 ( $\mu$ mol O₂)/(g CeO₂). It can be seen that the normalized OSC of prepared TiO₂-CeO₂ catalyst is almost as good as the CeO₂-based oxygen storage materials. Based on the results, it can be suggested that although Ce ions enter into the TiO₂ lattice, the OSC of CeO₂ in TiO₂ is not degraded, the TiO₂-CeO₂ still exhibits good OSC. The good OSC of TiO₂-CeO₂ is one of the reasons that TiO₂-CeO₂ catalyst presents excellent catalytic SOF oxidation activity, this is consistent with the catalytic performance results (Figure 2 and Table 2) and the reports [16,41].

# 3. Materials and Methods

# 3.1. Catalyst Preparation

The TiO₂-CeO₂ catalyst was prepared by a co-precipitation method. Desired TiOSO₄·2H₂O and Ce(NO₃)₃·9H₂O mixture solutions with the molar ratio of Ti:Ce = 9:1, which was the optimal ratio to expose the single TiO₂ crystal structure in our previous studies [35,36], were slowly added to NH₃·H₂O solutions under vigorous stirring. And then the precipitate was filtered and washed many times, after dried at 120 °C overnight and calcined for 3 h at 500 °C under airflow, the TiO₂-CeO₂ catalyst powder was obtained. The TiO₂ and CeO₂ catalysts were prepared by the same method.

The simulative 160,000 km vehicle aged sample was obtained by following reference [50] and our previous works [35,37]. Fresh catalyst was placed in a reactor and aged at 670 °C for 15 h and then at 250 °C for 15 h in the gases mixture at flow rate of 800 mL/min: 600 ppm C₃H₆, 1500 ppm CO, 200 ppm NO, 50 ppm SO₂, 5% O₂, 4% CO₂, 8% vapor and N₂ balance.

# 3.2. Catalyst Evaluation

The catalytic activity for SOF combustion of prepared catalysts were tested using TG-DTA method [16,39]. Due to the fact that the diesel SOF is comprised primarily of lube with a small amount of unburned fuel [51], commercial lubricating oil was often used to simulate the diesel SOF catalytic combustion [16,39].

For this test, the prepared powder catalyst was dried overnight at 120 °C to remove the effects of surface adsorbed water, and then impregnated with 5.0 wt. % commercial lube (Shell Helix HX7 5W-40, Shell Petrochemicals Company Limited, Jiaxing, China), the slurry of lube/catalyst mixture was stirred and mixed till a homogeneous state was obtained. About 10 mg of the lube/catalyst mixture powder was placed in the sample pan of TG-DTA unit (HCT-2, Beijing Henven Instruments, Beijing, China) and dried at 120 °C for 1 h to eliminate the effects of adsorbed substances (water, volatile matters etc.), and then heated to 550 °C with a temperature rate of 5 °C/min under airflow at 30 mL/min. The TG-DTA curves were recorded to determine the catalytic performance for SOF combustion of the prepared catalysts.

The test on lube without catalyst was carried out as a reference. About 10 mg of the lube was placed in the sample pan of TG-DTA unit and dried at 120 °C for 1 h, and then heated to 550 °C with a temperature rate of 5 °C/min under airflow at 30 mL/min.

# 3.3. Catalyst Characterization

Sulfur uptake was tested on a thermogravimetric analyzer (TGA) HCT-2 (Henven Instruments, Beijing, China). Consulting the references [38,52] and our previous works [37,53], about 15 mg of catalyst was placed in the sample crucible and pretreated under a 35 mL/min of N₂ flow for 5 h at 300 °C, and then the gas mixture of 43 mL/min SO₂(0.05 vol. %)-N₂ and 31 mL/min O₂(15.0 vol. %)-N₂ was introduced at 300 °C for 4 h, the weight increase as a function of time was recorded by the TGA.

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The sulfur cumulant of catalysts after mimicking 160,000 km vehicle aging were analyzed by an energy-dispersive X-ray (EDX) spectroscopy (IE-250, Oxford Instruments, Oxford, UK).

X-ray powder diffraction (XRD) patterns were collected on a powder X-ray diffractometer (DX-1000, Dandong Fangyuan Instrument Ltd., Dandong, China) employing Cu K $\alpha$  radiation ( $\lambda$  = 0.1542 nm).

Raman spectra were recorded by a LabRAM HR Laser Raman spectrometer (HORIBA Jobin Yvon Inc., Paris, France) with an excitation wavelength of 532 nm.

 $\rm N_2$  adsorption-description isotherms were measured on a QUADRASORB SI automated surface area and pore size analyzer (Quantachrome Instruments Ltd., Boynton Beach, FL, USA). The specific surface area and pore size were calculated by the Brunauer-Emmett-Teller (BET) method and Barret-Joyner-Halenda (BJH) method, respectively. Before adsorption measurements, the samples were degassed at 300  $^\circ C$  for 3 h under vacuum.

X-ray photoelectron spectroscopy (XPS) data were collected on a Kratos XSAM 800 spectrometer (Kratos Analytic Inc., Manchester, UK) with Al K $\alpha$  radiation. The binding energy shifts of the samples were calibrated by fixing the C1s binding energy (BE 284.8 eV).

 $H_2$ -temperature programmed reduction ( $H_2$ -TPR) experiments were performed in a quartz tubular reactor. Samples were pretreated at 450 °C for 1 h under the  $N_2$  flow (35 mL/min) and then cooled to room temperature; after that, the samples were heated from room temperature to 800 °C with a heating rate of 10 °C/min under the flow of  $H_2$  (5.0 vol. %)- $N_2$  mixture. The hydrogen consumption as a function of reduction temperature was recorded by a thermal conductivity detector (TCD) cell.

The oxygen storage capacity (OSC) of the samples was measured by a pulse injection technique [54]. The sample was firstly reduced in a H₂ flow (30 mL/min) at 550 °C for 1 h; after cooling to 200 °C, an oxygen pulse was injected every 5 min to obtain a breakthrough curve, from which the OSC was calculated.

# 4. Conclusions

From the aforementioned results, it can be concluded that moderate amounts of ceria dopants in titania can obviously enhance the catalytic SOF oxidation activity of TiO₂-CeO₂ catalyst. Meanwhile, the prepared TiO₂-CeO₂ catalyst can maintain the naturally excellent sulfur resistibility of titania; the sulfur resistibility of TiO₂-CeO₂ is as well as the bare TiO₂. The prepared TiO₂-CeO₂ catalyst significantly enhances the sulfur tolerance of conventional CeO₂-based SOF oxidation catalysts and displays a good catalytic SOF oxidation activity. The TiO₂-CeO₂ catalyst exhibits a typical phase of anatase, and the cerium ions can enter into the TiO₂ unit cell, impede the crystal growth and sintering and lower crystallinity of the TiO₂-CeO₂ catalyst. Moreover, the addition of CeO₂ in TiO₂ can significantly enhance the surface oxygen concentration and oxygen storage capacity of TiO₂-CeO₂ is almost as good as the CeO₂-based oxygen storage materials. The improvement of textual features and surface oxygen concentration of TiO₂-CeO₂ catalyst are the main reasons for the enhancement of catalytic SOF oxidation activity.

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# Low-Temperature Electrocatalytic Conversion of CO₂ to Liquid Fuels: Effect of the Cu Particle Size

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Abstract: A novel gas-phase electrocatalytic system based on a low-temperature proton exchange membrane (Sterion) was developed for the gas-phase electrocatalytic conversion of CO₂ to liquid fuels. This system achieved gas-phase electrocatalytic reduction of CO₂ at low temperatures (below 90 °C) over a Cu cathode by using water electrolysis-derived protons generated in-situ on an IrO₂ anode. Three Cu-based cathodes with varying metal particle sizes were prepared by supporting this metal on an activated carbon at three loadings (50, 20, and 10 wt %; 50% Cu-AC, 20% Cu-AC, and 10% Cu-AC, respectively). The cathodes were characterized by N₂ adsorption–desorption, temperature-programmed reduction (TPR), and X-ray diffraction (XRD) and their performance towards the electrocatalytic conversion of CO₂ was subsequently studied. The membrane electrode assembly (MEA) containing the cathode with the largest Cu particle size (50% Cu-AC, 40 nm) showed the highest CO₂ electrocatalytic activity per mole of Cu, with methyl formate being the main product. This higher electrocatalytic activity was attributed to the lower Cu–CO bonding strength over large Cu particles. Different product distributions were obtained over 20% Cu-AC and 10% Cu-AC, with acetaldehyde and methanol being the main reaction products, respectively. The CO₂ consumption rate increased with the applied current and reaction temperature.

**Keywords:** CO₂ electroreduction; CO₂ valorization; Cu catalyst; particle size; PEM; acetaldehyde production; methanol production

# 1. Introduction

Fossil fuels and biomass are the most common feedstocks for the production of liquid fuels. Since the burning of these feedstocks results in  $CO_2$  emissions into the atmosphere, it is necessary to develop strategies for the upgrading of this gas into useful products. One of these approaches involves the recycling of  $CO_2$  into sustainable hydrocarbon fuels [1] by different methods such as catalytic processes (e.g., hydrogenation to alkanes, alkenes or other oxygenated, or reforming with hydrocarbons) [2], biological processes [3], microwave and plasma systems [4–6], and photocatalytic and electrocatalytic routes [7]. Among these methods, the electrochemical reduction of  $CO_2$  is highly interesting since it allows to directly transform  $CO_2$  to syngas and light hydrocarbons with electricity, which may be obtained from renewable energy sources [1,8,9]. This approach is advantageous in that electrochemical cell reactors are typically compact, modular, and easy to scale-up. While electrolysis of  $CO_2$  and/or  $H_2O$  can be carried out in solid oxide cells (SOCs) [10], the high temperatures required for these systems to operate (above 600 °C) usually result in catalyst sintering and stability losses. In addition, high-temperature  $CO_2$ - $H_2O$  co-electrolysis produces syngas as the only product, and

further conversion steps (e.g., Fischer–Tropsch synthesis) are required to produce hydrocarbon fuels. Alternatively, low-temperature electrolyzers containing protonic exchange membranes (PEM) have been proposed to directly transform CO₂ into hydrocarbons and oxygenates [11–13]. Despite the overall single-pass conversions being typically low in these reactors, the unreacted CO₂ can be easily separated from the liquid fuels and recycled again to the reactor. The mild working conditions of these systems (typically below 90 °C and atmospheric pressure) facilitate the utilization of renewable energies such as solar heating and electrical energy for driving the electrochemical process. Moreover, unlike conventional catalytic hydrogenation of CO₂, PEM-based electrolyzers do not require external hydrogen since CO₂ directly reacts with the protons produced in-situ by water electrolysis [7]. These advantages have motivated researchers to investigate on gas-phase low-temperature CO₂ electroreduction [9,13–16], opening the way for incorporating renewable energies into the value chain of chemical industries. Professor Centi has developed most of these works with Pt, Fe, and Cu catalyst supported on a variety of carbonaceous materials such as carbon black, carbon nanofibers, and graphene, among others.

In this work, we carried out the electroreduction of  $CO_2$  in the gas phase at low temperature over Cu cathodes. We performed a systematic study with three different Cu-based cathodic catalysts supported on a high-surface-area activated carbon. The main objective of the work was to study the influence of the Cu particle size on the electrocatalytic activity of the system. With this aim, three different membrane electrode assemblies (MEAs) were fabricated with three Cu cathodic catalysts having different metal loadings and particle sizes. These MEAs were characterized and subsequently used for the electrocatalytic conversion of  $CO_2$  into synthetic fuels.

# 2. Results and Discussion

# 2.1. Characterization of The Cu Powder Catalysts and The Cu Electrodes

The different Cu powders and Cu cathodic-catalysts were characterized by  $N_2$  adsorption–desorption, atomic absorption spectroscopy (AAS), temperature-programmed reduction (TPR), and X-ray diffraction (XRD). Table 1 shows the physicochemical properties of the activated carbon support (AC) and the three Cu cathodic-catalysts.

Sample	Powder Metal Loading/wt %	Electrode Metal Weight/mg∙cm ⁻²	Surface Area/m²⋅g ⁻¹	Total Pore Volume/cm ³ ·g ^{−1}	TPR-T _{max} /°C	Mean Particle Size from XRD/nm
AC	-	-	866	0.293	-	-
50% Cu-AC	55	0.22	773	0.186	171	40
20% Cu-AC	19	0.18	797	0.260	185	14
10% Cu-AC	12	0.16	817	0.274	193	12

Table 1. Physicochemical properties of the support material, the catalysts, and the fresh electrodes.

The main difference between the three Cu cathodic catalysts is the metal loading. Thus, atomic absorption spectroscopy revealed metal loadings of 55, 19, and 12 wt % corresponding to electrode metal weights of 0.22, 0.18, and 0.16 mg·cm⁻² for the 50% Cu-AC, 20% Cu-AC, and 10% Cu-AC cathodic-catalysts, respectively. The electrocatalytic rates discussed below were normalized to the corresponding metal weight of the electrode.

The surface area and total pore volume of the different Cu cathodic-catalysts were determined by N₂ adsorption–desorption (Table 1). The activated carbon support showed high Langmuir areas and total pore volumes (866 m²·g⁻¹ and 0.293 cm³·g⁻¹, respectively), as previously reported in the literature [17,18]. As shown in Table 1, metal addition resulted in an important decrease of both the surface area and the total pore volume. As expected, the surface area and the total pore volume decreased with the metal loading, probably as a result of partial pore blockage by the metal particles [18,19]. Combined IUPAC (International Union of Pure and Applied Chemistry) types I and  $IV N_2$  adsorption–desorption isotherms (not shown) were obtained in all cases, revealing the presence of a microporous structure.

The TPR profiles of 50% Cu-AC, 20% Cu-AC, and 10% Cu-AC and the activated carbon support are shown in Figure 1. These TPR profiles result from the following sequential Cu reduction  $Cu^{2+} \rightarrow Cu^{+} \rightarrow Cu^{0}$  [20]. The first two reduction peaks at 171–193 °C and 223–276 °C (depending on the catalyst) can be attributed to the reduction of the more dispersed Cu particles and the reduction of CuO (II), respectively. The third reduction peak that appeared at 316–380 °C can be assigned to the reduction of Cu₂O (I). Finally, the peaks at higher temperatures are typically associated with the gasification of activated carbon and the reduction of surface oxygenated groups on the activated carbon support [21,22]. The temperature of the most intense consumption peak (T_{max}) is given in Table 1. The TPR profiles revealed that the interaction between the metal phase and the support varied depending on the metal particle size. Thus, stronger interactions were obtained for those catalysts having smaller Cu particles since T_{max} decreased with the Cu particle size (Table 1) [20,23,24]. Thus, the reducibility of the catalysts followed the sequence: 50% Cu-AC > 20% Cu-AC > 10% Cu-AC. Furthermore, given the TPR profiles, 400 °C was selected as a suitable reduction temperature for ensuring complete metal activation while maintaining the surface properties of the support.



Figure 1. Temperature-programmed reduction (TPR) profiles of the fresh catalysts and the activated carbon support.

Figure 2 shows the XRD patterns of the powder catalyst 50% Cu-AC before and after the reduction treatment at 400 °C. No significant differences were appreciated between the three Cu-AC powder catalysts. As shown in Figure 2a, diffraction peaks corresponding to metallic copper (Cu) and copper oxide (Cu₂O and CuO) were observed before the reduction treatment. The main diffraction Cu peaks were (111), (200), (220), and (331) observed at 43.3, 50.4, 74.1, and 89.8°, respectively. These peaks are associated with a metallic Cu phase with face-centered cubic (FCC) crystalline structure (JCPDS, 85–1326). Diffraction peaks corresponding to Cu₂O (JCPDS, 78–2076) and CuO (JCPDS, 80–1917) were also observed, in line with the TPR results (Figure 1). However, these peaks were lower in intensity as compared to those of metallic copper. Figure 2b shows the XRD patterns of the catalysts after the reduction treatment. As shown by the XRD patterns, only peaks corresponding to metallic Cu were observed at 20 = 43.3, 50.4, 74.1, and 90° for the (111), (200), (220), and (331) planes. These results revealed that Cu was completely reduced at 400 °C, as anticipated by the TPR profiles. Additionally, the metal precursor was completely calcined, since no peaks corresponding to the metal precursor

were observed in the XRD patterns. The mean Cu particle sizes of the different catalysts were estimated using the Scherrer equation, and the results are summarized in Table 1. As expected, the mean Cu particle size increased significantly with the metal loading. The Cu particle sizes obtained herein (10–40 nm) were similar to those measured for similar electrocatalytic systems prepared by direct impregnation with Cu precursor solutions [16,17,25].



**Figure 2.** X-ray diffraction (XRD) patterns of 50% Cu-AC supported on carbon paper substrates: (a) before reduction at 400  $^{\circ}$ C and (b) after reduction at 400  $^{\circ}$ C.

# 2.2. CO₂ Conversion Electrocatalytic Experiments

Figure 3 shows the production rates of different compounds as a function of the time on stream during the electroreduction of  $CO_2$  at a constant applied current of -20 mA and 90 °C. Note that no products were obtained under open circuit conditions (OCC, no current application). A constant current of -20 mA was subsequently applied at 90 °C for approximately 350 min under the same

reaction atmosphere. This polarization was maintained until products were obtained at a steady state rate. During this current application step, hydrogen (not shown in Figure 3, reaction (1)) and different products such as methanol, acetaldehyde, methyl formate, acetone, and n-propanol were obtained by reactions (2)–(6), respectively:

$$2\mathrm{H}^{+} + 2\mathrm{e}^{-} \to \mathrm{H}_{2} \tag{1}$$

$$CO_2 + 6H^+ + 6e^- \rightarrow CH_3OH + H_2O \tag{2}$$

$$2CO_2 + 10H^+ + 10e^- \rightarrow C_2H_4O + 3H_2O$$
 (3)

$$2CO_2 + 8H^+ + 8e^- \rightarrow C_2H_4O_2 + 2H_2O$$
 (4)

$$3CO_2 + 16H^+ + 16e^- \rightarrow C_3H_6O + 5H_2O$$
 (5)

$$3CO_2 + 18H^+ + 18e^- \rightarrow C_3H_8O + 5H_2O$$
 (6)

Most of these products have been previously obtained during the electrocatalytic conversion of  $CO_2$  on Fe, Co, Pt, and Cu supported on carbon nanotubes [15,16,26–28] and Cu supported on different carbonaceous supports (activated carbon, carbon nanofibers, and graphite) [17] at similar temperatures. These products reached maximum production rates after ca. 300 min on stream and decreased upon OCC. A slow dynamic behavior was observed (the steady state was reached after 4–5 h of reaction), and this can be attributed to the high residence times used herein. At this point it should be mentioned that low Faradaic efficiencies to hydrocarbon products (below 10%) were obtained in all the experiments. This is due to the high kinetics of the hydrogen evolution reaction (reaction (1)) vs. hydrocarbon formation reactions (reactions (2)–(6)) due to the low surface concentration of  $CO_2$ .

The configuration used herein is advantageous in that it allows the direct supply of H⁺ (more reactive than H₂) to the cathodic side of the cell. This configuration allows lower temperatures (around 90 °C) as compared to catalytic CO₂ hydrogenation processes (above 250 °C) [21,29–31].

Finally, in all experiments, the cathodic side of the cell was purged with N₂ (30 NmL·min⁻¹) and returned to OCC to remove all the products for subsequent reaction experiments. Sample 50% Cu-AC showed the highest intrinsic electrocatalytic activity among the three cathodic-catalyst studied herein. This higher electrocatalytic activity can be associated with the higher Cu particle size of 50% Cu-AC. Thus, large Cu particles have been previously reported to favor the formation of reaction products by reducing the strength of the metal–CO interaction [32]. In the electroreduction of CO₂, CO₂ is adsorbed on active centers and subsequently converted into CO and O₂. Since this adsorbed CO is more reactive than  $CO_2$ , it reacts with protons to generate the different products observed. This reaction is favored on large Cu particles. Since small Cu particles adsorb CO more strongly than large particles, the subsequent reaction of CO is hindered on small Cu particles, leading to CO and H₂ as the main products [32].

With regard to the composition of the reaction products, methyl formate was the main reaction product over 50% Cu-AC, followed by acetaldehyde and methanol. Acetaldehyde and methanol were the main reaction products obtained over samples 20% Cu-AC and 10% Cu-AC, respectively. In line with our results, acetaldehyde and methanol were previously obtained as main products over Cu-AC catalysts [17]. In the case of catalysts 50% Cu-AC, the high Cu particle size favored the production of methyl formate by dehydrogenation of methanol Reaction (7), as previously reported [33].

$$2CH_3OH \rightarrow HCOOCH_3 + 2H_2$$
 (7)


**Figure 3.** Time-on-stream variation of the rate of production for the different products at a constant current of -20 mA over the cathodic-catalysts on carbon paper substrates: (a) 50% Cu-AC, (b) 20% Cu-AC, and (c) 10% Cu-AC. Conditions: temperature = 90 °C, F_(CO_2), cathode = 1.65 NmL·min⁻¹ y F_(H_2O), anode = 6 NmL·min⁻¹.

Figure 4a,b summarize the effect of the applied current and the reaction temperature on the steady state  $CO_2$  consumption rate (after 350 min of polarization), respectively. The reaction rates were normalized by the amount of Cu deposited on each cathodic-catalyst. In line with the previous experiments, sample 50% Cu-AC showed larger electrocatalytic activities than samples 20% Cu-AC and 10% Cu-AC for all the reaction conditions studied. As expected, the consumption rate of  $CO_2$  increased with the applied current, most likely because of an increase in the electrochemical supply of H⁺. In line with previous studies [15], an increase in the reaction temperature also resulted in higher electrocatalytic activities for all the catalysts tested. While the kinetics of the electrochemical reactions can be improved by increasing the temperature [34], the protonic membrane prevented us from testing the system above 90 °C since its stability and conductivity under appropriate humidity conditions is not ensured at these conditions.



**Figure 4.** (a) Effect of the current at 90 °C and (b) temperature at I = -20 mA on the steady state CO₂ consumption rate for the cathodic-catalysts 50% Cu-AC, 20% Cu-AC, and 10% Cu-AC. Conditions:  $F_{-(CO_2)}$ , cathode = 1.65 NmL·min⁻¹ y F_(H_2O), anode = 6 NmL·min⁻¹.

Finally, we calculated the energy consumption for the three different cathodic-catalysts evaluated. The three different MEAs were compared at 90 °C and -20 mA. The overall energy consumption for CO₂ conversion (kW·h·mol⁻¹ CO₂) and the energy consumption for the production of methanol (kW·h·mol⁻¹ CH₃OH), acetaldehyde (kW·h·mol⁻¹ CH₃CHO), and methyl formate (kW·h·mol⁻¹ HCO₂CH₃) were calculated. As indicated above, the cathodic-catalyst with the highest metal loading, 50% Cu-AC, showed the highest electrocatalytic activity. This catalyst showed the lowest energy

consumption for the conversion of CO₂ (119.01 kW·h·mol⁻¹) among the catalysts tested. In addition, the MEA containing 50% Cu-AC consumed less energy per kg of methanol, acetaldehyde, and methyl formate (1549, 1054, and 444 kW·h·mol⁻¹, respectively) than the other two.

## 3. Materials and Methods

Cu catalysts supported on activated carbon were used as cathodes for the electrochemical reduction of  $CO_2$ , while Ir (IV) oxide (IrO₂) supported on carbon was used as a cell anode.

A commercial high surface area activated carbon (Sigma Aldrich, St. Louis, MO, USA) was used as a support. Metal particles were impregnated on activated carbon under vacuum at room conditions using an ethanolic precursor solution of  $Cu(NO_3)_3 \cdot 3H_2O$  (Panreac, Barcelona, Spain) in a rotary evaporator. The catalysts were dried at 120 °C overnight and then calcined for 2 h at 350 °C in an N₂ atmosphere. A final reduction step under H₂ at 400 °C for 2 h was performed with a heating rate of 5 °C·min⁻¹. Three different catalysts were prepared with total Cu loadings of 50, 20, and 10 wt % (50% Cu-AC, 20% Cu-AC, and 10% Cu-AC, respectively).

The catalyst inks were prepared by mixing appropriate amounts of the different catalysts;  $IrO_2$  commercial catalyst powders (Alfa Aesar, Haverhill, MA, USA, 99% for the anode, and Cu-activated-carbon powder for the cathode) with a Nafion solution (5 wt %, Aldrich Chemistry, St. Louis, MO, USA, Nafion[®] 117 solution) in isopropanol (Sigma Aldrich) containing a Nafion/solvent volume ratio of 0.04.  $IrO_2$  was selected as an anode as in conventional PEM water electrolyzers [17]. The different inks were deposited on carbon paper (Fuel Cell Earth, Woburn, MA, USA) substrates with a geometric surface area for both electrodes of 12.56 cm² (circular electrode) at 65 °C. A commercial proton conducting Sterion[®] membrane (Hydrogen Works) was used as the electrolyte. Prior to use, the Sterion[®] membrane was successively pre-treated at 100 °C for 2 h in H₂O₂ and then into H₂SO₄ (to promote activation), and finally in deionized water to wash it. Then, the membrane electrode assembly (MEA) was prepared by sandwiching the membrane between the electrodes. Finally, the whole system was hot-pressed under heating conditions at 120 °C, and a final pressure of 1 metric ton was applied for 3 min.

The Cu metal loading on the powder catalysts was measured by Atomic Absorption Spectroscopy (AAS) using a SPECTRA 220FS analyzer. The sample (ca. 0.5 g) was dissolved in a mixture containing 2 mL of HCl, 3 mL of HF, and 2 mL of H₂O₂ followed by microwave digestion at 250 °C. The surface area and volume porosity of the different materials were measured using a Micromeritics ASAP 2010. N₂ was used as the sorbate at -193 °C, and the microporosity was measured by the Howath–Kawazoe (HK) method. Prior to the analyses, the samples were outgassed at 180 °C under vacuum (5 × 10⁻³ Torr) for 12 h. Temperature Programmed Reduction (TPR) experiments were conducted on a commercial Micromeritics AutoChem 2950 HP unit provided with a thermal conductivity (TCD) detector. The samples (ca. 0.15 g) were loaded into a U-shaped reactor and ramped from room temperature up to 900 °C (10 °C·min⁻¹) under a reducing H₂/Ar gas mixture of 17.5% v/v (60 cm³·min⁻¹). X-Ray diffraction analysis (XRD) were conducted on the Cu-AC powder catalysts prior and after reduction with a Philips PW-1710 instrument, using Ni-filtered Cu K $\alpha$  radiation ( $\lambda = 1.5404$  Å). The samples were analyzed at a rate of 0.02°·step⁻¹ over a 2 $\theta$  range of 20–80° (scan time 2 s·step⁻¹) and the obtained spectra were compared with the corresponding JCPDS-ICDD references.

 $CO_2$  electro-reduction experiments were performed in an electrochemical cell reactor working at room pressure conditions [17]. Water was introduced into the anode side using a saturator. The water content in the anodic feed stream (25% H₂O/N₂) was controlled by the heating the saturator at 65 °C. All the pipelines downstream from the saturator were heated to avoid condensation. On the anode side, water electro-oxidation was carried out on IrO₂ to generate protons across the Sterion[®] membrane. The water stream was also used to hydrate the Sterion[®] membrane and to keep its ionic conductivity under working conditions [16]. The cathodic side of the cell operates under a gas flow of pure CO₂ (Praxair Inc. certified standards 99.999% purity, Danbury, CT, USA). Both gas flow rates (N₂ for the anode and CO₂ for the cathode) were controlled by mass flowmeters (Brooks, Seattle, WA, USA).

The electrocatalytic experiments were carried out at atmospheric pressure with an overall gas flow rate of 0.5 NmL·min⁻¹ of CO₂ for the cathodic stream and 6 NmL·min⁻¹ for the anodic stream working at 80 and 90 °C. The reactant and products outer stream from the cathodic chamber were analyzed by using a double channel gas chromatograph (Bruker 450-GC) equipped with Hayesep Q-Molsieve 13X consecutive columns and flame ionization detectors (FIDs). Hydrogen, methanol, acetaldehyde, methyl formate, acetone, and n-propanol were the reaction products detected. The carbon atom balance closed within a 10% error. A potentiostat/galvanostat (Voltalab 21, Radiometer Analytical, Lyon, France) was used for the different applied polarizations, and the two electrodes were connected to the potentiostat by using gold wires.

## 4. Conclusions

Three different Cu-based cathodic-catalysts with Cu loadings of 50, 20, and 10 wt % were synthesized by impregnation, characterized, and tested in the electrocatalytic conversion of CO₂.

50% Cu-AC showed the highest CO₂ electrocatalytic activity among the catalysts tested under all the explored reaction conditions. These results could be explained by the higher Cu particle size of this material. Considering that CO is an important intermediate in the process, large Cu particles are believed to favor electroreduction by weakening the metal–CO interaction.

Methyl formate was the main reaction product for 50% Cu-AC, while acetaldehyde and methanol were the main products for 20% Cu-AC and 10% Cu-AC, respectively. This fact can be attributed to the higher particle size of Cu that favored the production of methyl formate via dehydrogenation of methanol.

The  $CO_2$  consumption rate increased with the applied current and the reaction temperature due to an enhancement of the kinetic of the electrocatalytic reactions.

Finally, 50% Cu-AC showed the highest electrocatalytic activity and the lowest energy consumption values for the conversion of  $CO_2$  (119 kW·h·mol⁻¹). In addition, the MEA containing 50% of Cu consumed less energy per kg of methanol, acetaldehyde, and methyl formate than the other two MEAs containing less amount of Cu.

Author Contributions: J.L.V. conceived the project, N.G.-G. performed the experiments and J.C.S.-R. and A.d.L.-C. wrote the manuscript.

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# CuO Nanoparticles Supported on TiO₂ with High Efficiency for CO₂ Electrochemical Reduction to Ethanol

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**Abstract:** Non-noble metal oxides consisting of CuO and TiO₂ (CuO/TiO₂ catalyst) for CO₂ reduction were fabricated using a simple hydrothermal method. The designed catalysts of CuO could be in situ reduced to a metallic Cu-forming Cu/TiO₂ catalyst, which could efficiently catalyze CO₂ reduction to multi-carbon oxygenates (ethanol, acetone, and n-propanol) with a maximum overall faradaic efficiency of 47.4% at a potential of -0.85 V vs. reversible hydrogen electrode (RHE) in 0.5 M KHCO₃ solution. The catalytic activity for CO₂ electroreduction strongly depends on the CuO contents of the catalysts as-prepared, resulting in different electrochemistry surface areas. The significantly improved CO₂ catalytic activity of CuO/TiO₂ might be due to the strong CO₂ adsorption ability.

Keywords: electrochemical reduction; CO2; CuO; TiO2; ethanol

## 1. Introduction

Electrochemical reduction of  $CO_2$  ( $CO_2ER$ ) is a promising and feasible method with which to sustainably transform this waste stream into value-added low-carbon fuels [1–5], which has the following advantages: (1) An electrochemical system can be operated under moderate reaction conditions [5,6]; (2) The reaction process involves highly complex multiple protons and electrons transfer steps, which lead to wide product distribution, such as methanol, ethanol, acetone, and so on [7,8]; (3) The electricity could be supplied by clean and sustainable energy, such as solar, wind, and hydropower [9,10]; and (4) This reaction requires minimal chemical intake and is convenient for large-scale applications [5,8]. However,  $CO_2ER$  is facing severe challenges, including poor faradaic efficiency (FE), high overpotential, and low selectivity [11,12], which urges us to design new catalysts to address the above efficiency and selectivity issues.

Over the past few decades, numerous trials have been made to explore catalysts with distinguished performance for CO₂ER. Until now, multifarious catalysts including metals [5,11–13], metal oxides [14–16], and metal complexes [17,18] have been reported. Among these catalysts, Cu, as a relatively low-cost and earth abundant metal, has a unique capacity to produce hydrocarbons through a multiple protons and electrons transfer pathway; however, the traditional Cu catalysts show high overpotential [16] and low selectivity for diversiform products [5,12,19–21]. Substantial efforts have been pursued to enhance energetic efficiency for CO₂ER through altering surface structures, morphologies, and the nature compositions. Recently, Cu-based catalysts have been reported to possess enhanced FEs for CO₂ER; for instance, Takanabe et al. [22] used Cu-Sn alloy for the efficient and selective reduction of CO₂ to CO over a wide potential range. Yu and her team [23] reported that Cu nanoparticle (NP) interspersed MoS₂ nanoflowers facilitate CO₂ER to hydrocarbon, such as

 $CH_4$  and  $C_2H_4$  with high FE at low overpotentials. Sun et al. [24] introduced a Core/Shell Cu/SnO₂ structure that shows high selectivity to generate CO with FE reaching 93% at -0.7 V (vs. the reversible hydrogen electrode (RHE)).

Normally, TiO₂, as a semiconductor material, is one of the most widely used photocatalysts and electrocatalysts for CO₂ reduction because of its nontoxicity, low cost, and high chemical stability [25,26]. Furthermore, TiO₂ has been reported to act as a redox electron carrier to facilitate a variety of reduction reactions, including CO₂ conversion [27,28], and to assist in CO₂ adsorption [29,30]; thus, it may stabilize the CO₂ER intermediate and reduce overpotential. Some early works could verify this point in the CO₂ER, such as Ag/TiO₂ [31], Cu/TiO₂ [32], and Cu/TiO₂/N-graphene [33].

CuO/TiO₂ has been previously prepared using different synthetic methods and is used for various applications, such as photodriven reduction of CO₂ [34] and hydrogen production reaction [35]. Nevertheless, to the author's knowledge, CuO/TiO₂ as electrocatalysts for CO₂ER have been rarely investigated. In previous works, we have demonstrated that CuO with various morphologies can be highly efficient electrocatalysts for CO₂ reduction to ethanol in simple aqueous medium, but at high overpotential (-1.7 V vs. the saturated calomel electrode (SCE)) [36]. Inspired by previous studies, in this work, we systematically and carefully synthesized a series of nano-sized CuO/TiO₂ catalysts for CO₂ER. The prepared CuO/TiO₂ catalysts were optimized by varying the amount of loaded CuO NPs. CuO/TiO₂ with an intended CuO content of 60%, as an efficient electrocatalyst, exhibits the most outstanding activity (achieved total FE of 47.4% at the potential of -0.85 V vs. RHE) for CO₂ER in 0.5 M KHCO₃ solution at room temperature among all as-prepared CuO/TiO₂ catalysts.

## 2. Results and Discussion

#### 2.1. Catalyst Characterizations

A simple and mild hydrothermal synthesis method was used to prepare six different weight ratios of CuO/TiO₂ catalysts, which are defined as CuO/TiO₂-1, CuO/TiO₂-2, CuO/TiO₂-3, CuO/TiO₂-4, CuO/TiO₂-5, and CuO/TiO₂-6, corresponding to the intended CuO content of 5 wt %, 10 wt %, 20 wt %, 40 wt %, 60 wt %, and 80 wt %, respectively. X-ray diffractometer (XRD) patterns shown in Figure 1 indicated that these catalysts consist of both CuO and TiO₂, which were represented by solid lines and dashed lines, respectively. It can be clearly seen that all as-prepared catalysts with different amounts of CuO loadings present similar XRD patterns. The diffraction peaks ascribed to CuO were significantly shown with the increasing of CuO contents in the XRD patterns. Meanwhile, the actual CuO contents were analyzed by inductively coupled plasma atomic emission spectroscopy (ICP-AES) and summarized in Table S1. These values are in good agreement with the intended values.



**Figure 1.** XRD patterns of (**a**) pure TiO₂, (**b**) CuO/TiO₂-1, (**c**) CuO/TiO₂-2, (**d**) CuO/TiO₂-3, (**e**) CuO/TiO₂-4, (**f**) CuO/TiO₂-5, (**g**) CuO/TiO₂-6, and (**h**) pure CuO.

To thoroughly examine the element proportion and chemical state of CuO/TiO₂ catalysts, X-ray photoelectron spectroscopy (XPS) of CuO/TiO₂-5 catalyst was investigated. As expected, Cu, Ti, and O elements from CuO/TiO₂ catalyst were observed in the full spectrum from Figure 2a. Furthermore, Figure 2b displays the high resolution spectrum of Cu 2p, separated into Cu  $2p^{3/2}$  and Cu  $2p^{1/2}$  at 933.8 eV and 953.8 eV, respectively. The distance between these Cu 2p main peaks positions is 20.0 eV, which agrees well with previous reports about CuO spectrum [37]. Moreover, additional confirmation of CuO state was seen with the broad satellite peaks at a higher binding energy than the main peaks. The main peak of Cu  $2p^{3/2}$  at 933.8 eV was accompanied by two satellite peaks on the higher binding energy side at about 943.8 eV and 941.5 eV, which suggests the existence of CuO [38-41]. From this figure, we can clearly see that the main peak of Cu  $2p^{1/2}$  at 953.8 eV and its satellite peak at 962.5 eV were separated by about 9.0 eV, which also confirms the presence of CuO [42]. Both Ti  $2p^{3/2}$  and Ti 2p^{1/2} peaks at 458.7 eV and 464.4 eV, respectively, were observed (Figure 2c) with a separation of 5.7 eV, indicating that TiO₂ existed in this catalyst [43]. The obtained XPS spectrum of O 1 s was presented in Figure 2d. An obvious peak appeared at 529.7 eV, which can be indexed to  $O^{2-}$  in the CuO and TiO₂. Notably, there are other three weak O 1 s peaks. One located at 530.7 eV is identified with surface hydroxyls, which is likely the by-product from the synthesis process of CuO/TiO₂. Additionally, the remaining two peaks observed at 531.5 eV and 532.8 eV are confirmed to C=O and C-O [44,45]. A slight peak of C1 s was detected in Figure 2a, which is always observed in XPS spectra of real-world solids [46,47]. The high resolution spectrum of C 1 s was shown in Figure S1, consistent with the results of O1s.



**Figure 2.** (a) The full XPS spectrum of CuO/TiO₂-5 catalyst and high-resolution XPS spectra of (b) Cu 2p; (c) Ti 2p and (d) O 1 s.

To obtain the morphological information regarding the CuO/TiO₂ catalysts, scanning electron microscope (SEM) images were firstly shown in Figure S2. One can see that the catalysts were composed of 3-dimension NPs with a certain amount of CuO nano-floc supported on the surface of TiO₂. Further from the SEM images, for low CuO content of CuO/TiO₂ (5 wt %, 10 wt %, and 20 wt %), CuO NPs were dotted sporadically on the surface of TiO₂. Subsequently, for high CuO content of CuO/TiO₂ (40 wt %, 60 wt %, and 80 wt %), it is evident that the CuO NPs presented very covered on the TiO₂ surface forming a massive and compact layer. Pure CuO NPs presented very

porous, sponge-like structures. Moreover, transmission electron microscopy (TEM) investigation was used to gain deeper insight into the structural feature of the catalysts. TEM of CuO/TiO₂-5 was shown in Figure 3a. Numerous CuO NPs were irregularly interspersed on the TiO₂ surface, in accordance with the SEM results. High-resolution TEM images of CuO/TiO₂-5 were displayed in Figure 3b to acquire more detailed information about the structure of the catalyst. The fast Fourier transform (FFT, inset of Figure 3b) pattern of CuO and TiO₂ shows concentric rings and bright discrete diffraction spots, which are indicative of high crystallinity. From Figure 3b, distinctive lattice fringes were found in both TiO₂ and CuO NPs. The lattice fringes with interplanar spacing of 0.188 nm were assigned to the (200) plane of the TiO₂, expressed by orange lines (inserted in Figure 3b). Additionally, the lattice fringes with d-spacings of 0.256 nm corresponded to the plane (111) of CuO phase, represented by the yellow lines (inserted in Figure 3b).



Figure 3. TEM images of  $CuO/TiO_2$ -5 catalyst in (a) low magnification and (b) high magnification (the insets are FFT pattern and the detailed images of white dashed circle).

## 2.2. Catalyst Properties

The electrocatalytic activity of CuO/TiO₂-5 catalyst was carried out in the typical three-electrode system through cyclic voltammetry (CV) measurement, which was firstly tested at a scan rate of 50 mV/s in N₂-saturated and CO₂-saturated 0.5 M KHCO₃ solution, respectively, as shown in Figure 4a. In both N₂ and CO₂ atmosphere, two obvious reduction peaks were observed, corresponding to the in situ reduction of CuO to Cu₂O and sequentially Cu₂O to Cu [48]. As expected, XRD characterization could verify the formation of Cu metal from CuO/TiO₂-5 catalyst after 2 h electrolysis, depicted in Figure S3. Three significant diffraction peaks for metallic Cu (denoted by solid diamond) appeared, suggesting that CuO NPs in CuO/TiO₂-5 catalyst could be in situ electroreduced to metallic Cu-forming  $Cu/TiO_2$  in the electrolysis process and subsequently serves as an effective catalyst for  $CO_2$  reduction, verified by previous reports [16,36,47]. In the more negative potential region, an obvious increase of current density (J) after -0.65 V vs. RHE was shown, relating to the hydrogen evolution reaction (HER) in the  $N_2$ -saturated 0.5 M KHCO₃ solution (pH = 8.63, black line), while more dramatic increase of J was observed in the  $CO_2$ -saturated 0.5 M KHCO₃ solution (pH = 7.21, red line). To avoid the pH effect, a CV curve was recorded in the  $N_2$ -saturated solution with HCl solution (pH = 7.21, same as that of the CO₂-saturated 0.5 M KHCO₃ solution, blue line), which shows lower J than that in CO₂-saturated 0.5 M KHCO3 solution after -0.65 V vs. RHE. It demonstrates that CO2 reduction is more favorable than HER on  $CuO/TiO_2$ -5.



**Figure 4.** (a) CV curves of CuO/TiO₂-5 catalyst over glassy carbon electrode (GCE) in N₂-saturated without (black line) and with (blue line) HCl solution, and CO₂-saturated (red line) 0.5 M KHCO₃ solution; (b) FEs for different products over CuO/TiO₂-5 catalyst at various potentials.

The CO₂ER electrocatalytic activity on CuO/TiO₂-5 catalyst was also evaluated by controlled potential electrolysis (from -0.65 V to -1.05 V vs. RHE) in CO₂-saturated 0.5 M KHCO₃ solution. The main liquid products of the electrolysis are ethanol, acetone, and n-propanol, which were detected by ¹H NMR. The overall FEs (Figure 4b) presented a sharply incremental tendency at an applied potential range from -0.65 V to -0.85 V vs. RHE and achieved a maximum of 47.4% (37.5% for ethanol, 4.3% for acetone, and 5.6% for n-propanol) at the potential of -0.85 V vs. RHE. As the potentials shift more negatively, the overall FEs significantly decreased to 18.8% at the potential of -1.05 V vs. RHE because of the competition from HER. Furthermore, a similar variation trend for FE_{ethanol}, FE_{acetone}, and FE_{n-propanol} was observed and depicted in detail in Figure 4b. Moreover, Figure S4 shows the total current vs. time curve for the CuO/TiO₂-5 electrode at -0.85 V vs. RHE, which exhibited an initial current of 85 mA as the CuO was reduced and, subsequently, a sharp decline current in a short time. Finally, a stable current of 16 mA in the long test appeared. Notably, the FE for ethanol was maintained at approximately 35% throughout the electrolysis. This finding suggests not only efficient but also stable activity for CO₂ reduction on this electrode.

We also studied the effect of CuO contents in various CuO/TiO₂ catalysts on CO₂ER activities, as shown in Figure 5a. All CuO/TiO₂ catalysts showed two obvious reduction peaks in the CO₂-saturated 0.5 M KHCO₃ solution; however, the position and size of the reduction peaks varied greatly among the catalysts, which, assigned to in situ, reduce CuO to Cu. Noteworthily, in the more negative potential region (below -0.65 V vs. RHE), all CuO/TiO₂ catalysts show different catalytic abilities for CO₂ reduction. The J of all CuO/TiO₂ catalysts at a potential of -0.85 V vs. RHE were summarized in Table S2. CuO/TiO₂-5 showed the largest J. The dramatic difference of electrocatalytic performance on various CuO/TiO₂ catalysts might be closely related to the composition of these catalysts. Figure 5a shows that increasing the CuO content (from 5 wt % to 60 wt %) in the CuO/TiO₂ catalysts can facilitate the reduction of CO₂, leading to enhanced activity for CO₂ reduction. However, on the CuO/TiO₂-6 (with the intended CuO content of 80 wt %) catalyst for sequentially increasing the content of CuO, CO₂ reduction was subsequently suppressed. The results clearly indicate that an optimum content of CuO in the catalysts is required to achieve the maximum activity for CO₂ reduction.

Potentiostatic electrolysis (at -0.85 V vs. RHE) of CO₂ on various CuO/TiO₂ catalysts was performed in CO₂-saturated 0.5 M KHCO₃ solution to further investigate the catalytic activities of CuO/TiO₂ catalysts with different CuO contents. Figure 5b summarizes the FEs of ethanol, acetone, and n-propanol achieved over all CuO/TiO₂ catalysts. As expected, the CuO/TiO₂-5 with the intended CuO content of 60 wt % gives the highest total FE, which reached 47.4%, when compared with other CuO/TiO₂ catalysts, which is in a well agreement with CV results as shown in Figure 5a. However, the total FEs decreased along with the higher CuO content (80 wt % and even 100 wt %) of CuO/TiO₂, which is ascribed to the large number of CuO NPs that accumulated on the surface of TiO₂, as shown in Figure S2f,g, which probably generated the decrease of active areas for CO₂ER. Yet, low CuO content (5 wt %–40 wt %) of CuO/TiO₂ also exhibited slightly poor electrocatalytic performance for CO₂ER due to the low content of CuO NPs. It demonstrates further that the catalytic activity for CO₂ER strongly depends on the CuO contents of the catalysts as-prepared, which results in different electrochemistry surface areas (ECSA). To confirm this finding, ECSA of a variety of CuO/TiO₂ electrodes were examined by CV in a potential range in which the faradaic process did not occur with the double layer capacitance in N₂ atmosphere 0.1 M HClO₄ solution (Figure S5 and Table S3). As expected, CuO/TiO₂-5 catalyst showed a noticeable performance improvement for the double layer capacitance (C_{d1}), which gives the positive correlation with ECSA. The C_{d1} in Table S3 indicated the ECSA increased with increasing CuO content of CuO/TiO₂ catalysts plays a vital role in ECSA that can affect CO₂ER, which is in accordance with the results of Brunauer– Emmett–Teller (BET) specific surface area (Table S4).



**Figure 5.** (a) CV curves of various  $CuO/TiO_2$  catalysts over GCE in  $CO_2$ -saturated 0.5 M KHCO₃ solution. Scan rate, 50 mV/s; (b) FEs for different products over various  $CuO/TiO_2$  catalysts at -0.85 V vs. RHE in  $CO_2$ -saturated 0.5 M KHCO₃ aqueous solution.

To further study the improved performance of CuO/TiO₂-5 catalyst, we compared the electrocatalytic abilities of CO₂ reduction on CuO/TiO₂-5 and CuO/C catalysts. The CuO/C was successfully synthesized and characterized by XRD (Figure S6), SEM, and TEM (Figure S7). CuO/C catalyst represents the onset potential for  $CO_2$  reduction at -0.75 V vs. RHE, corresponding to a 0.10 V more negative onset potential than CuO/TiO₂-5, as shown in Figure S8. This finding suggests that  $TiO_2$  as the supporter is more beneficial to facilitating  $CO_2ER$ . The J in  $CO_2$ -saturated solution of CuO/TiO₂-5 at the potential of -0.85 V vs. RHE was 8.307 mA/cm², which is larger than that of CuO/C (1.402 mA/cm²). Additionally, potentiostatic electrolysis in CO₂-saturated 0.5 M KHCO₃ solution for the CuO/C was performed at the potential of -0.85 V vs. RHE, which shows poor active towards CO₂ reduction. Only 16.3% for overall FE (12.9% for ethanol, 1.1% for acetone, and 2.3% for n-propanol) was analyzed by ¹H NMR. In contrast, CuO/TiO₂-5 catalyst presents excellent activity for CO₂ reduction, which achieved 47.4% for overall FE. Besides, as seen from Figure S5 and Table S3, the ECSA of CuO/TiO₂-5 reveals a larger value than the one of CuO/C. As is well known,  $CO_2$ adsorption on the active sites is the prerequisite for subsequent CO₂ reduction reaction. The relatively larger amount of CO₂ adsorption on active sites may offer more original reactants such as CO₂ [49]. The CO₂ adsorption abilities of CuO/TiO₂-5 and CuO/C were evaluated by CO₂ adsorption, which were displayed in Figure S9. CuO/TiO₂-5 catalyst possessed a more remarkably improved CO₂ adsorption capacity than CuO/C, which was 5.80 mg/g and 1.34 mg/g, respectively. It is reasonable to assume that the enhanced CO₂ adsorption capacity can make a significant contribution to the superior performance of CO₂ reduction. By comparing the electrochemical and material characterizations of CuO/TiO₂ with those of CuO/C, we gained insight that TiO₂ plays an important role in promoting the electrocatalytic performance of  $CuO/TiO_2$ . Hence, on the basis of the above analysis, a brief illustration of CO2ER over CuO/TiO2 catalysts is stated. Large amount of CO2 is first adsorbed

on TiO₂. CuO/TiO₂ catalysts could be in situ electroreduced to Cu/TiO₂ at less negative cathode potentials than the ones of CO₂ER. Then, the adsorbed CO₂ gets one electron from the achieved Cu/TiO₂ electrode and is converted to CO₂⁻, which can be dimerized to *C₂O₂⁻. Notably, the C-C bond-making step on the Cu surface is a key step for CO₂ reduction to ethanol and even the C3 product (acetone and n-propanol) [50]. Interestingly, the catalytic ability of CuO/TiO₂-5 at -0.85 V vs. RHE expresses more obvious performance than the one of 40 wt % Cu/TiO₂ (the optimal catalyst in X wt % Cu/TiO₂ system) at -1.45 V vs. RHE [34], which indicates that in situ electroreduced Cu NPs have greater activity to catalyze CO₂ reduction. The low-cost CuO/TiO₂ catalyst is an efficient alternative to expensive materials for the application of CO₂ER in industry.

## 3. Materials and Methods

#### 3.1. Materials and Instruments

Cupric acetate monohydrate  $[Cu(Ac)_2 \cdot H_2O$ , Analytical Reagent (AR) grade], ammonium carbonate  $[(NH_4)_2CO_3, AR \text{ grade}]$ , potassium bicarbonate  $[KHCO_3, AR \text{ grade}]$ , and titanium dioxide  $[TiO_2, AR \text{ grade}]$  were purchased from Sinopharm Chemical Reagent Co. (SCR, Shanghai, China) with 99% purity and used as received. Nafion[®] 117 solution (5%) and Nafion[®] 117 membrane were obtained from Dupont (Wilmington, DE, USA). Carbon paper (CP, HCP010) and conductive carbon black (C, VXC72R) were purchased from Shanghai Hesen Electrical CO. (Shanghai, China).

Crystal-phase X-ray diffraction (XRD) patterns of CuO/TiO₂ catalysts were recorded using an Ultima IV X-ray powder diffractometer (Kurary, Tokyo, Japan) equipped with Cu Kα radiation (k = 1.5406 Å). The values for actual CuO loadings of the synthesized catalysts were determined on an inductively coupled plasma atomic emission spectroscopy (ICP-AES, IRIS Intrepid II XPS, Waltham, MA, USA). The scanning electron microscope (SEM) images of the catalysts were obtained using a Hitachi S-4800 field-emission scanning electron microscope (Tokyo, Japan). Transmission electron microscopy (TEM) patterns were recorded by TECNAI G₂F30 transmission electron microscope (Tokyo, Japan). The X-ray photoelectron spectroscopy (XPS, Thermo Fisher Scientific, Waltham, MA, USA) analysis was performed on the Thermo Scientific ESCA Lab 250Xi using 200 W monochromatic Al K $\alpha$  radiation. The 500  $\mu$ m X-ray spot was used. The base pressure in the analysis chamber was about  $3 \times 10^{-10}$  mbar. Typically, the hydrocarbon C1s line at 284.8 eV from adventitious carbon was used for energy referencing. Brunauer-Emmett-Teller (BET) specific surface area was characterized by nitrogen adsorption in a BELSORP-MAX instrument (MicrotracBEL, Tokyo, Japan) after outgassing the samples for 10 h under vacuum at 573 K. CO₂ adsorption was obtained in a BELSORP-MAX instrument (MicrotracBEL, Japan) under CO2 atmosphere at 298 K. Potentiostatic electrolysis and cyclic voltammetry (CV) were performed using a CHI 660 C electrochemical station (Shanghai Chenhua Instrument Co. Ltd., Shanghai, China).

#### 3.2. Materials Synthesis

A simple and mild hydrothermal synthesis method was used to prepare six different weight ratios of CuO/TiO₂ catalysts, which are defined as CuO/TiO₂-1, CuO/TiO₂-2, CuO/TiO₂-3, CuO/TiO₂-4, CuO/TiO₂-5, and CuO/TiO₂-6, corresponding to the CuO content of 5 wt %, 10 wt %, 20 wt %, 40 wt %, 60 wt %, and 80 wt %, respectively. Typically, due to similar fabrication procedure, for simplicity, CuO/TiO₂-5 acts as the specimen to illustrate the following experiment. The detailed synthesis procedure for CuO/TiO₂-5 is depicted as follows: 0.05 M of Cu(Ac)₂ aqueous solution (60 mL) was mixed with 127 mg of TiO₂. After stirring for several minutes, an appropriate amount of 0.05 M of (NH₄)₂CO₃ aqueous solution was added drop-wise with stirring to control the synthetic rate. Then, the mixture was stirred gently for 3 h at the ambient temperature. Subsequently, the above mixture was centrifuged, washed with distilled water, and stored in a vacuum oven with 60 °C for 4 h to dry the precipitate. Finally, the collected powder was transferred to a crucible, further maintaining at 220 °C for 3 h to obtain the final CuO/TiO₂ powder. Pure CuO catalyst was acquired by the same

method, without adding TiO₂. Additionally, CuO/C catalyst with the CuO content of 60 wt % got through the similar synthetic process, but with C replacing TiO₂.

#### 3.3. Electrode Preparation and Electrochemical Test

Potentiostatic electrolysis was carried out in an H-type three-electrode cell with a piece of Nafion[®] 117 cation exchange membrane (H⁺ form) as a separator. The working electrode was manufactured using the following route: 10 mg of CuO/TiO₂ catalyst was suspended in a mixture solution with 20  $\mu$ L of 5 wt % Nafion[®] solution and 40  $\mu$ L of distilled water. After sonification for 20 min, the mixture was spread on a porous CP (2 × 2 cm) by micropipette and then dried in air. The counter and reference electrodes were Pt sheet and SCE, respectively. A 0.5 M of KHCO₃ aqueous solution serves as electrolyte, which was saturated with CO₂ by bubbling for 30 min before the electrolysis experiment. CO₂ was bubbled continuously throughout overall experiment time. CV tests were measured in a single cell system using a standard three-electrode setup at 0.5 M KHCO₃ aqueous solution under CO₂ and N₂ atmosphere with a scan rate of 50 mV/s, respectively. In this system, 2  $\mu$ L of the above prepared catalyst suspension was coated on glassy carbon electrode (GCE, diameter = 2 mm) playing as the working electrode, and a Pt mesh and an SCE acted as the counter electrode and the reference electrode, respectively. All potentials were based on RHE as reference potentials, and converted by the following equation:

$$E$$
 (vs. RHE) =  $E$  (vs. SCE) + 0.059 × pH + 0.241.

#### 3.4. Product Analysis

Liquid phase products were analyzed by ¹H-nuclear magnetic resonance (NMR) spectra recorded on an Ascend 400 (400 MHz, Bruker, Germany) spectrometer in D₂O with Me₄Si as an internal standard.

#### 4. Conclusions

In this study, a simple synthesis strategy was employed to fabricate CuO/TiO₂ material. A series of characterizations has demonstrated that CuO NPs were uniformly distributed on the surface of TiO₂. Different CuO/TiO₂ catalysts displayed dramatically different electrocatalytic performance for CO₂ER, depending strongly on the CuO contents of the catalysts as-prepared. High FEs of CO₂ reduction products (ethanol, acetone, and n-propanol) reach up to 47.4% on CuO/TiO₂-5 at the potential of -0.85 V vs. RHE among all CuO/TiO₂ and CuO/C catalysts. The CO₂ adsorption ability test suggests that CuO/TiO₂-5 exhibits more superior CO₂ adsorption performance than CuO/C, indicating that CuO/TiO₂-5 is more beneficial to CO₂ reduction. The fact that the as-synthesized CuO/TiO₂ catalysts exhibit high activity for CO₂ER is attributed to the special roles of TiO₂ that show excellent CO₂ adsorption ability. Our work may provide some concepts for designing cheap and effective catalysts for highly efficient CO₂ER.

Supplementary Materials: The following are available online at http://www.mdpi.com/2073-4344/8/4/171/, Table S1: CuO composition of the synthesized CuO/TiO2 catalysts. Figure S1: The high-resolution XPS spectrum of C1s for CuO/TiO₂-5 catalyst. Figure S2: SEM images of (a) pure TiO₂, (b) CuO/TiO₂-1, (c) CuO/TiO₂-2, (d) CuO/TiO₂-3, (e) CuO/TiO₂-4, (f) CuO/TiO₂-5, (g) CuO/TiO₂-6, and (h) pure CuO. Figure S3: XRD patterns of CuO/TiO₂-5 catalyst after 2 h electrolysis. The solid diamond represents the in situ generated Cu metal. Figure S4:  $CO_2$  reduction electrolysis data at -0.85 V vs. RHE for CuO-TiO₂-5 catalyst. Table S2: The current density (J) for all the CuO/TiO₂ catalysts in CO₂-saturated environment at the potential of -0.85 V vs. RHE. Figure S5: Plot of I vs. potential of (A) CuO/TiO2-1, (B) CuO/TiO2-2, (C) CuO/TiO2-3, (D) CuO/TiO2-4, (E) CuO/TiO2-5, (F) CuO/TiO₂-6, (G) pure CuO, and (H) CuO/C in N₂-saturated 0.1 M HClO₄ solution cycled between 0.35 and 0.55 V vs. SCE at scan rates in the range of 5–60 mV/s. Insets show plot of  $I_c$  vs. v in which the linear regressions give capacitance information. Table S3: The analysis of C_{dl} of various CuO/TiO₂ and CuO/C catalysts. Table S4: The analysis of specific surface area of CuO/TiO₂ catalysts. Figure S6: XRD patterns of CuO/C and C catalysts. Figure S7: SEM (a) and TEM (b) images of CuO/C catalyst. Figure S8: CV curve of CuO/C catalyst over glassy carbon electrode in CO₂-saturated (red line) 0.5 M KHCO₃ solution. Figure S9: CO₂ adsorption-desorption isotherms at 297 K of CuO/TiO₂-5 (red square) and CuO/C (blue triangle). Filled and empty symbols represent adsorption and desorption, respectively.

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## Article Hydrogen Production from Chemical Looping Reforming of Ethanol Using Ni/CeO₂ Nanorod Oxygen Carrier

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**Abstract:** Chemical looping reforming (CLR) technique is a prospective option for hydrogen production. Improving oxygen mobility and sintering resistance are still the main challenges of the development of high-performance oxygen carriers (OCs) in the CLR process. This paper explores the performance of Ni/CeO₂ nanorod (NR) as an OC in CLR of ethanol. Various characterization methods such as N₂ adsorption-desorption, X-ray diffraction (XRD), Raman spectra, transmission electron microscopy (TEM), X-ray photoelectron spectroscopy (XPS), H₂ temperature-programmed reduction (TPR), and H₂ chemisorption were utilized to study the properties of fresh OCs. The characterization results show the Ni/CeO₂-NR possesses high Ni dispersion, abundant oxygen vacancies, and strong metal-support interaction. The performance of prepared OCs was tested in a packed-bed reactor. H₂ selectivity of 80% was achieved by Ni/CeO₂-NR in 10-cycle stability test. The small particle size and abundant oxygen vacancies contributed to the water gas shift reaction, improving the catalytic activity. The covered interfacial Ni atoms closely anchored on the underlying surface oxygen vacancies on the (111) facets of CeO₂-NR, enhancing the anti-sintering capability. Moreover, the strong oxygen mobility of CeO₂-NR also effectively eliminated surface coke on the Ni particle surface.

Keywords: chemical looping reforming; hydrogen; oxygen carrier; CeO₂; nanorod

## 1. Introduction

Hydrogen is considered an efficient energy carrier that is environmentally benign [1]. Chemical looping reforming (CLR) is a prospective alternative for hydrogen production due to its energy efficiency and inherent CO₂ capture [2,3]. The fixed-bed reactor configuration CLR process (Figure 1) is performed by alternatively switching the feed gases; the oxygen carriers (OCs) are stationary and periodically exposed to redox atmosphere [4]. The key to developing a CLR process is to screen high-performance OCs. Ni-based OCs have been widely investigated because of their ability for carbon–carbon and carbon–hydrogen bonds cleavage [5–7]. Zafar et al. [8] prepared a series of OCs including Fe, Cu, Mn, and Ni supported by SiO₂ and MgAl₂O₄ and concluded that NiO supported on SiO₂ exhibited high H₂ selectivity in CLR. Löfberg et al. [9] demonstrated that Ni plays two essential roles in CLR process, i.e., the anticipated activation of reactants as well as the regulation of the oxygen supply rate from solids. Dharanipragada et al. [10] reported that the Ni-ferrites OC suffers from the loss of oxygen storage capacity due to Ni sintering in chemical looping with alcohols. Improving oxygen mobility and sintering resistance remain the major challenges of the development of Ni-based OCs in the CLR process due to the high activation energy of oxygen anion diffusion in NiO (2.23 eV in CLR) and the low Tammann temperature of Ni (691 °C) [11–14].



Figure 1. Schematic description of CLR of ethanol.

Recent research has revealed that CeO₂ exhibits excellent redox property due to abundant oxygen storage capacity and the strong capability to stabilize Ni nanoparticles because of strong metal support interaction (MSI) [15,16]. CeO₂ can readily release lattice oxygen under reducing conditions, thus creating oxygen vacancies which are associated with oxygen mobility [17]. The MSI between the Ni and CeO₂ could tune the physiochemical properties of Ni, contributing to high catalytic reactivity and stability [18]. Jiang et al. [19] proposed that the oxygen vacancies of CeO₂ could effectively eliminate surface coke deposition, activate steam, and shorten the "dead time" in the CLR process. Dou et al. [20] revealed that the surface oxygen originating from the CeO₂ lattice can oxidize coke precursors, keeping the OC surface free of coke deposition. It has been reported that MSI and the mobility of lattice oxygen show strong dependence on the morphology of CeO₂ [21]. Lykaki et al. [22] have demonstrated CeO₂ morphology dominates reducibility and oxygen mobility, which follow the sequence: nanorod (NR) > nanopolyhedra (NP) > nanocube (NC). Moreover, the apparent activation energy of these three CeO₂ shapes for the CO oxidation in a hydrogen-rich gas shows the opposite trend, implying the potential highest water gas shift (WGS) activity of NR [23].

Therefore, in this work, a Ni/CeO₂-NR OC for CLR of ethanol for hydrogen production is prepared by a hydrothermal method. The physicochemical properties are investigated by N₂ adsorption-desorption, X-ray diffraction (XRD), inductively coupled plasma optical emission spectroscopy (ICP-OES), Raman spectra, high resolution transmission electron microscopy (HRTEM), X-ray photoelectron spectroscopy (XPS), and H₂ temperature-programmed reduction (H₂-TPR). The performances of the Ni/CeO₂-NR OC are tested in a packed-bed reactor and compared with the other reference bulk OC, i.e., Ni/CeO₂.

## 2. Results and discussion

## 2.1. Characterization of OCs

Predominant physicochemical properties of fresh Ni/CeO₂-NR and CeO₂-NR are tabulated in Table 1. The introduction of Ni species has a pronounced influence on the texture of CeO₂-NR. The CeO₂-NR support showed higher Brunauer-Emmett-Teller (BET) surface area than that of Ni/CeO₂-NR. The average pore diameter and pore volume also exhibited the same trends. It has been reported that the BET surface of CeO₂-NR decreased by 13–18% after the incorporation of 7.5 wt % CuO [22]. The actual Ni content of Ni/CeO₂-NR determined by ICP-OES was 9.7 wt %.

Sample	Surface Area	Average Pore	Pore Volume	Ni Content	Ni Dispersion	Ni Crystal Size/Ni	CeO ₂ Crystal
	(m ² /g)	Diameter (nm)	(cm ³ /g)	(wt %)	(m²/g _{Ni} )	Average Particle Size (nm)	Size (nm)
Ni/CeO ₂ -NR	75.1	18.2	0.37	9.7 ¹	6.5 ²	8.9 ³ /8.1 ⁴	16.1 ³
CeO ₂ -NR	90.3	26.5	0.80	N/A	N/A	N/A	14.8

Table 1. Physicochemical properties of fresh Ni/CeO₂-NR and CeO₂-NR.

^{1.} Determined by ICP-OES; ^{2.} Determined by H₂ chemisorption; ^{3.} Determined by XRD from Ni (111) and CeO₂ (111) plane; ^{4.} Calculated from the HRTEM images.

The XRD profiles of fresh Ni/CeO₂-NR and CeO₂-NR OCs are shown in Figure 2. The reflection peaks at 28.5°, 33.1°, 47.5°, 56.3°, 59.1°, 69.4°, 76.7°, and 79.1° could be indexed to (111), (200), (220), (311), (222), (400), (331), (420), and (422) planes of a fluorite-structure CeO₂ (Space group: *Fm3m*), respectively. There were no other impurity peaks of the hexagonal structure of Ce(OH)₃ or Ce(OH)CO₃ detected, indicating the excellent crystalline purity of CeO₂. Two peaks at 44.5° and 51.8°, which could be respectively attributed to (111) and (200) facets of Ni, were observed for Ni/CeO₂-NR. The mean crystal sizes of Ni and CeO₂ were obtained from the Scherrer Equation (listed in Table 1). The crystal sizes of CeO₂ for CeO₂-NR and Ni/CeO₂-NR were 14.8 and 16.1 nm, respectively, indicating that the incorporation of Ni would not significantly change the structure characteristics of CeO₂ support. Similar crystallite sizes of CeO₂-NR prepared by the hydrothermal method have been reported in other work [22,24].



Figure 2. XRD profiles of Ni/CeO₂-NR and CeO₂-NR.

Raman spectroscopy was employed to characterize surface and bulk defects. As shown in Figure 3, both samples present four characteristic peaks. The appreciable peak at 462.8  $\rm cm^{-1}$  resulted from the first order  $F_{2g}$  mode of the fluorite cubic structure. The Raman shift at 596.1 cm⁻¹ could be assigned to the defect-induced band (D band). The relocation of O atom from the interior of tetrahedral cationic sub-lattice to the interior of ideally empty octahedral cationic sites (Frenkel interstitial sites) would lead to the deformation of the anionic lattice of  $CeO_2$  [25]. The intensity of the D band is a gauge of the distortion of ionic lattice, which results in punctual defects and oxygen vacancies [22]. Therefore, the value of  $I_D/I_{F_{2q}}$  is commensurate with the number of defect sites in ceria. The relative intensity of  $I_D/I_{F_{2q}}$  decreased after the incorporation of Ni, suggesting NiO species suppress the surface oxygen vacancy of CeO₂-NR. Li et al. [26] reported that Vanadium atom would bond to the surface of CeO2-NR by generating V–O–Ce species, thus covering oxygen defects and stabilizing adjacent Ce atoms. The inconspicuous peak at 258.1 cm⁻¹ results from the second order transverse acoustic mode (2TA) or doubly degenerate transverse optical mode (TO), which are Raman inactive in a perfect crystal [27]. The peak at  $1179.9 \text{ cm}^{-1}$  may be related to the stretching mode of the short terminal Ce=O. Moreover, the Raman shift of Ni/CeO₂-NR moved towards a low wavenumber in comparison with that of CeO2-NR, indicating the incorporation of Ni affects the symmetry of Ce-O bonds [21].



Figure 3. Visual Raman spectroscopy of Ni/CeO2-NR and CeO2-NR.

The TEM of CeO₂-NR and Ni/CeO₂-NR are illustrated in Figure 4. The Ni/CeO₂-NR maintained the original nanorod shape after Ni incorporation. The CeO₂-NR was approximately 14 nm in diameter and several hundreds nm in length. As shown in Figure 4a, the calculated interference fringe spacings (d) are 0.27 and 0.31 nm, revealing the CeO2-NR predominantly exposes the (100) and (111) facets. The preferable exposure crystal facets of CeO2-NR are consistent with those of samples derived from the CeCl₃ precursor in other studies [28,29]. Theory calculation has demonstrated that the (111) is the least reactive facet, followed by (100) and (110). In addition to the preferable crystal facets, there are other aspects degerming CeO₂ activity. Several 'dark pits' are shown in the box of Figure 4. Liu et al. [29] have reported that these dark pits are related to the surface reconstruction and defects; their study also concluded that these defects play a more important role in determining the CeO₂ activity than the exposure planes. Sayle et al. [30] have performed a molecular dynamic modeling to simulate the synthesis of CeO2-NR and discovered that the atomistic sphere model exhibits many steps on the (111) planes of CeO₂-NR. The migration of oxygen in CeO₂ occurs by a vacancy hopping mechanism; therefore, the clusters of defects are conducive to oxygen transfer. If the diffusion rate of oxygen anions becomes adequately high, a consecutive oxygen flow is generated, resulting in high reducibility. The Ni particle size distribution, which is calculated from 52 particles, was inserted in Figure 4, and the average particle size was 8.1 nm (shown in Table 1). Shen et al. [31] have elucidated that the strong interfacial anchoring effect, which exists between the surface oxygen vacancies on (111) planes of CeO₂-NR and the gold particles, only allows the gold particles to locally rotate or vibrate but not to migrate to form aggregates. Therefore, the strong MSI would significantly improve the Ni dispersion on the CeO₂-NR surface, resulting in a small particle size of Ni.



**Figure 4.** (a) HRTEM image of CeO₂-NR, (b) TEM image of CeO₂-NR, (c) and (d) TEM images of Ni/CeO₂-NR.

XPS was employed to investigate the valences of Ce and Ni cations. The XPS spectra of Ce 3d of Ni/CeO₂-NR (shown in Figure 5a) could be deconvoluted into two spin-orbit series, i.e., 3d_{5/2} (u) and  $3d_{3/2}(v)$ . The multiplet splitting components labeled  $u^0$ ,  $u^1$ ,  $u^2$ , v,  $v^1$  and  $v^2$  corresponds to the  $3d^{10}4f^0$  state of Ce⁴⁺, while the  $u_0$  and  $v_0$  are related to  $3d^{10}4f^1$  state of Ce³⁺. These two Ce species in Ni/CeO2-NR indicate the OC surface was partly reduced because of oxygen desorption and the formation of oxygen vacancies. It is widely accepted that oxygen vacancies are produced to maintain electrostatic balance once Ce³⁺ exists in fluorite Ce (Equation 1). The percentage of Ce³⁺ cations to the total Ce cations is determined by the area ratio of different Ce species in XPS spectra. The Ce³⁺ ratio of Ni/CeO₂-NR (16.9%) was lower than that of pure CeO₂-NR (24.3%) reported in Lykaki et al. [22], suggesting NiO species could inhibit the formation of surface-unsaturated Ce³⁺. The decrease in surface Ce³⁺ species of Ni/CeO₂-NR was consistent in the trend of  $I_D/I_{F_{2g}}$  in the Raman result. The Ni 2p XPS spectra (illustrated in Figure 5b) were characterized by two spin-orbit groups, i.e., 2p_{3/2} (855.2 and 856.4 eV) and 2p_{1/2} (873.3 eV), as well as a shake-up peak at 861.8 eV. The photoelectron peak of 2p_{3/2} over Ni/CeO₂-NR shifted towards high-binding energy in contrast to those over pure NiO (844.4 eV, reported in Lemonidou et al. [32]) and Ni/CeO₂ (854.5 eV, reported in Tang et al. [33]). This result implies there is an enhanced MSI between Ni and CeO₂-NR. The Ni/Ce atom ratio of the outer surface of  $CeO_2$ -NR (0.49) is higher than the nominal one (0.32), suggesting a Ni species enrichment from bulk to surface.

$$4Ce^{4+} + O^{2-} \to 4Ce^{4+} + \frac{2e^{-}}{\delta} + 0.5O_2 \to 2Ce^{4+} + 2Ce^{3+} + \delta + 0.5O_2 \tag{1}$$

where  $\delta$  represents an empty position derived from the removal of  $O^{2-}$  from an oxygen tetrahedral site (Ce₄O).



Figure 5. (a) XPS spectra of Ce 3d; (b) XPS spectra of Ni 2p.

H₂-TPR was performed to investigate MSI. As illustrated in Figure 6, the H₂-TPR patterns of CeO₂-NR are comprised of two reduction peaks. An inconspicuous peak at 262 °C, which can be assigned to the reduction of surface-adsorbed oxygen, was observed, while a broad peak ranging from 400 °C to 550 °C could be ascribed to the reduction of Ce⁴⁺ to Ce³⁺ [22,34]. With regard to Ni/CeO₂-NR, the TPR profiles consisted of three peaks. The small peak at 220 °C may have been attributable to the reduction of the NiO species, which slightly interacted with CeO₂-NR supports. Zhang et al. [21] have reported that this NiO species is characterized by small radius and could incorporate into CeO₂-NR surface. The second peak at 356 °C could be ascribed to the reduction of NiO species, with a strong interaction with CeO₂-NR supports. The H₂ consumption of the second peak was highest among the three peaks, indicating most of the NiO strongly interacted with the support. The third peak at 494 °C was also attributed to the reduction of Ce⁴⁺ to Ce³⁺. However, the broad reduction peak of Ni/CeO₂-NR shifted to a lower temperature than that of pure CeO₂-NR, suggesting the introduction of Ni improves the reducibility of CeO₂-NR. It has demonstrated that the promoted reduction behavior is caused by the generated Ni-Ce-O species which improves the deformation of CeO₂ [35].



Figure 6. H₂-TPR patterns of Ni/CeO₂-NR and CeO₂-NR.

#### 2.2. Activity Tests of OCs

The activity of Ni/CeO₂-NR was tested in CLR of ethanol and compared with Ni/CeO₂. Figure 7 displays the performance of both OCs in CLR: H₂ selectivity and ethanol conversion at the fuel feed step as well as the concentration of CO, CO₂ and O₂ at the air feed step. The H₂ selectivity and ethanol conversion of both OCs remained relatively stable during the process. With regard to Ni/CeO₂-NR, the average ethanol conversion remained at 88.0%, and the average H₂ selectivity was 78.9%. However, with the average ethanol conversion and H₂ selectivity of conventional Ni/CeO₂ at 72.9% and 60.5% respectively, the activity of Ni/CeO₂-NR was superior to the Ni/CeO₂. As the Ni content and other experimental conditions were the same for both OCs, the improved activity therefore resulted from the support.



Figure 7. Activity tests of (a) Ni/CeO₂-NR and (b) Ni/CeO₂.

Various factors contributed to the enhanced catalytic activity. Notably, the CeO₂-NR with high specific-to-volume ratio guarantees excellent Ni dispersion, as evidenced by TEM and TPR results. Such a one-dimensional nanostructure enables uniform and small Ni nanoparticles to be finely dispersed on supports, thus generating many accessible catalytic active sites. Additionally, the high concentration oxygen vacancies of CeO₂-NR, which was proven via Raman and XPS analysis, can activate and produce OH groups from steam. H₂ and CO₂ can be generated via the reaction between the formed OH groups as well as intermediate species, thus improving H₂ selectivity. It has been proposed that the interface of the metal/CeO₂ is the main site for steam reforming reaction [36].

The air feed process is highly oxygen consuming and is accompanied by the generation of  $CO_2$  and CO due to coke oxidation and partial oxidation. As illustrated in Figure 7, all the samples exhibited  $CO_2$  and CO evolution at the air feed stage. The integration areas of C-containing gas concentrations corresponded to the quantity of carbon deposition. Ni/CeO₂-NR exhibited small peak areas compared with the Ni/CeO₂ OC, indicating that the carbon deposition on Ni/CeO₂-NR was effectively suppressed. This is a result of the abundant oxygen vacancies and its strong oxygen storage capacity, which enhanced the oxygen mobility and thereby facilitated the removal of carbon deposition. In addition, the depleted CeO₂-NR could be partially replenished by the air feed step. The O₂ concentration increased from 0% to 23% at the end of the air feed stage, indicating the end of coke elimination and replenishment of lattice oxygen.

## 2.3. Stability Tests

The durability of Ni/CeO₂-NR and Ni/CeO₂ was tested in a 10-cycle CLR process. As shown in Figure 8, the H₂ selectivity of Ni/CeO₂ slowly declined with the increase of the running cycles, while the Ni/CeO₂ NR maintained its activity throughout the test. During the tests, the H₂ selectivity of Ni/CeO₂-NR decreased from 81.7% to 79.2%, while Ni/CeO₂ exhibited a significant decrease in H₂

selectivity from 61.8% to 51.4%. The deactivation of OCs predominantly caused Ni sintering and carbon deposition. Notably, the Ni/CeO₂-NR OC exhibited a more durable performance than the Ni/CeO₂, revealing that the CeO₂-NR supported metal was more resistant to deactivation. The superior durability of the Ni/CeO2-NR sample resulted from the strong MSI between Ni and CeO2-NR support, which improved the metal-sintering resistance. The abundant oxygen vacancies of CeO₂-NR are essential units for the anchoring of metal particles. The essential role of CeO₂ is to disperse and stabilize Ni particles over its surface oxygen vacancies, which depend on the morphology of CeO₂. As evidenced by XPS, the (111) plane of CeO₂-NR plays an important role in anchoring the Ni particles. It has been proposed that the interfacial Au atoms which are located away from the particle perimeter (covered interfacial atoms) would closely interact with the underlying surface oxygen vacancies on the (111) facets of CeO₂-NR [31]. Because this interfacial region was not involved in the reforming reaction, this strong interaction would effectively stabilize the Ni particles on CeO₂-NR. Additionally, small size Ni particles can lower the driving force for coke diffusion and thereby help to reduce the carbon deposition. Furthermore, the strong oxygen mobility of CeO₂-NR is indispensable to removing the carbon deposition at the metal surface. It has been demonstrated that the oxygen deposited in CeO₂ lattice can react with the carbon species left over from the steam reforming reaction [37]. Overall, CeO2-NR support not only promotes the anti-sintering capability of OCs but also reduces the carbon deposition, thus improving the durability of OCs. A comparison of several investigations regarding CLR over different Ni-based OCs is tabulated in Table 2. The Ni/CeO₂-NR in this work showed higher average fuel conversion in long-term tests despite the low Ni loading, indicating the strong sintering resistance and high reforming activity of this OC.



Figure 8. Stability tests of OCs.

<b>Fable 2.</b> Comparison of CLR over different OC	٦s.
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OCs	Ni Loading (wt %)	T (°C)	S/C	Tested Cycles	Reactant	Average Conversion (%)	Average H ₂ Yield (%)	References
CeNi/MCM-41	6	650	1.5	10	Glycerol	~90	6.2	[37]
LaNiO ₃ /MMT	13	650	3	10	Ethanol	~90	-	[3]
Ni/MMT	19.9	650	4	20	Ethanol	~78	-	[13]
CeNi/PSNT	24.7	650	3	10	Glycerol	~100	12.5	[19]
CeNi/SBA-15	12	650	3	14	Ethanol	~84	-	[20]
Ni/CeO2-NR	9.7	650	4	10	Ethanol	~90	-	This work

## 3. Materials and Methods

#### 3.1. Preparation of OCs

CeO₂-NR was synthesized on the basis of our previous work with small modifications [38]. Firstly, 1 g of CeCl₃ (99.9%, Aladdin, Shanghai, China) and 19.3 g of NaOH (97%, Aladdin) were dissolved in 50 mL deionized water and continuously stirred for 25 min. The resulting mixture was then put into a Teflon-lined autoclave and kept at 100 °C for 24 h. Subsequently, the sample was separated, washed, then dried at 85 °C for 10 h and finally calcined at 400 °C for another five hours. The Ni/CeO₂-NR oxygen carrier was synthesized by a wet-impregnation method. Ni(NO₃)₂·6H₂O (98%, Aladdin) of 0.2 g, equivalent to 10 wt % Ni loading, was first dispersed in 5 mL deionized water, and the prepared CeO₂-NR of 0.5 g was added into the solution. The mixture was stirred under sonication for three hours at 60 °C and then dried at 85 °C for 10 h. The as-synthesized sample was then calcined at 700 °C for two hours.

The other reference bulk  $Ni/CeO_2$  was also synthesized by a wet-impregnation method, and the procedure has been reported in Xu et al. [39]. The Ni content of this OC was also set to 10 wt %.

## 3.2. Characterization of OCs

The texuture of OCs were analyzed by a Micrometric Acusorb 2100E apparatus (Ottawa, ON, Canada) at 77 K. The OCs were degassed at 573 K for three hours before tests.

ICP-OES (Nijmegen, The Netherlands) was applied to measure the elemental composition. Before tests, the sample was solved by hydrofluoric acid solution.

XRD was performed, using a Shimadzu XRD-600 instrument (Kyoto, Japan), to identify the phase composition. The scanning  $2\theta$  degrees ranged from  $10^{\circ}$  to  $80^{\circ}$ , and the scanning rate was set to  $4^{\circ}$ /min. A graphite-filtered Cu K $\alpha$  radiation ( $\lambda$  = 1.5406 Å) was applied as a radiation source.

Raman spectra were employed by a Renishaw Spectroscopy (Gloucestershire, UK) with a visible 514 nm Ar-ion laser under ambient conditions. During the mesurement, the flowing gas was He, and the test temperature was maintained at 300  $^{\circ}$ C.

TEM was conducted, using FEI Tecnai F30 (Cleveland, OH, USA), to investigate the morphology of the Ocs. The samples were first solved in ethanol under sonication, followed by dispersal on a copper grid-supported carbon foil and dried in air.

XPS was carried out by a ThermoFisher K-Alpha system with a 150 W Al K $\alpha$  source (Waltham, MA, USA). The samples were placed on the holder, and the scanning step was set to 0.15 eV.

H₂-TPR was employed by a Micrometrics AutoChem 2920 instrument (Ottawa, ON, Canada). In a typical procedure, a sample of 90 mg was preheated at 450 °C for one hour in an Ar flow of 35 mL/min and subsequently cooled to 80 °C. A mixture of 10 vol % H₂ in Ar flow (35 mL/min) was then inserted, and the temperature was increased from 80 °C to 100 °C at a rate of 5 °C/min simultaneously. H₂ chemisorption was also performed by the same apparatus to analyze the Ni dispersion. The sample was first reduced at 700 °C in a H₂/Ar flow (30 mL/min) for one hour, followed by cooling to 100 °C under Ar flow. Subsequently, H₂ pulses were introduced until the eluted peaks of successive pulses became steady. The Ni active surface area was obtained from the H₂ adsorbption volume considering the stoichiometric ratio H_{adsorbed}/Ni_{surface} = 1 and a surface area of  $6.5 \times 10^{-20}$  m² per Ni atom [40].

#### 3.3. Activity and Stability Tests

The performance of CLR of ethanol by Ni/CeO₂-NR and reference conventional bulk OCs, i.e., Ni/CeO₂, were conducted in a packed-bed reactor, whose schematic was described in our previous work [41,42]. The tested OCs of 0.5 g were loaded at the center of the quartz reactor. During a fuel feed step, an ethanol solution (4 mL/h) with a steam to carbon ratio (S/C) of four was preheated to 150 °C and then inserted into the reactor in a N₂ flow of 180 mL/min. The reacter temperature of the fuel feed step was 650 °C. An Agilent 7890A (Santa Clara, CA, USA) chromatography with two detectors,

thermal conductivity detector (TCD) and flame ionization detector (FID), was applied to verify the effluents. TCD with a TDX-01 column was applied to detect N₂, H₂, CO, CO₂, and CH₄; the FID with a Porapak-Q column was applied to measure the concentration of C₃H₈O₃, H₂O, and CH₃CHO. In an air feed step, the air flow was 100 mL/min, and the reaction temperature was also 650 °C. The exhausted gases were detected by another GC (Agilent 7890A) with a TCD detector. A 5A molecular sieve column was used to detect the O₂, and the TDX-01 column was applied to detect CO₂, CO, and N₂. The durations of the fuel feed step and the air feed step were 60 and 10 min, respectively. A N₂ purge process of five minutes was performed between the fuel feed step and the air feed step to eliminate the residue gas in the reactor.

The conversion and H₂ selectivity were calculated as follows:

$$X = \frac{F_{in} - F_{out}}{F_{in}} \times 100\%$$
⁽²⁾

$$S_{H_2} = \frac{1}{6} \times \frac{moles \ H_2 \ produced}{moles \ ethanol \ feed \times X} \times 100\%$$
(3)

## 4. Conclusions

A Ni/CeO₂-NR OC was synthesized by hydrothermal method and tested in CLR of ethanol process in this work. H₂ selectivity of 80% was achieved by Ni/CeO₂-NR in a 10-cycle stability test. The characterization results show that the Ni/CeO₂-NR possesses high Ni dispersion, abundant oxygen vacancies, and strong MSI. The small particle size and abundant oxygen vacancies contributed to the WGS reaction, thus improving the catalytic activity. The buried interfacial Ni atoms strongly anchored on the underlying surface oxygen vacancies on the (111) facets of CeO₂-NR, therefore enhancing the anti-sintering capability. Moreover, the strong oxygen mobility of CeO₂-NR also effectively eliminated surface coke on the Ni particle surface.

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## Article Experimental Clarification of the RWGS Reaction Effect in H₂O/CO₂ SOEC Co-Electrolysis Conditions

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Abstract: In the present investigation, modified X-Ni/GDC electrodes (where X = Au, Mo, and Fe) are studied, in the form of half-electrolyte supported cells, for their performance in the RWGS through catalytic-kinetic measurements. The samples were tested at open circuit potential conditions in order to elucidate their catalytic activity towards the production of CO ( $r_{co}$ ), which is one of the products of the H₂O/CO₂ co-electrolysis reaction. Physicochemical characterization is also presented, in which the samples were examined in the form of powders and as half cells with BET, H₂-TPR, Air-TPO and TGA re-oxidation measurements in the presence of H₂O. In brief, it was found that the rate of the produced CO ( $r_{co}$ ) increases by increasing the operating temperature and the partial pressure of H₂ in the reaction mixture. In addition, the first results revealed that Fe and Mo modification enhances the catalytic production of CO, since the 2wt% Fe-Ni/GDC and 3wt% Mo-Ni/GDC electrodes were proven to perform better compared to the other samples, in the whole studied temperature range (800–900 °C), reaching thermodynamic equilibrium. Furthermore, carbon formation was not detected.

Keywords: SOECs; RWGS reaction kinetics; Au–Mo–Fe-Ni/GDC electrodes; high temperature  $H_2O/CO_2$  co-electrolysis

## 1. Introduction

Solid oxide electrolysis is a contemporary process for CO₂ capture/recycling, which is proven as an attractive method to provide CO₂ neutral synthetic hydrocarbon fuels. In particular, co-electrolysis of H₂O and CO₂ in a solid oxide electrolysis cell (SOEC) yields synthesis gas (CO + H₂), which in turn can be used towards the formation of various types of synthetic fuels [1–3] by applying the Fischer-Tropsch process. According to thermodynamics, SOECs offer benefits for endothermic reactions, such as H₂O and/or CO₂ electrolysis at temperatures >700 °C, because a larger part of the required electrical energy can be substituted by thermal energy [4,5]. In addition, high temperature can lead to a significant decrease in the internal resistance of the cell and acceleration of the electrode reaction processes due to fast reaction kinetics [4–7].

 $H_2O/CO_2$  Co-electrolysis is a much more complex process compared to pure steam or  $CO_2$  electrolysis. This is because three reactions take place simultaneously, namely  $H_2O$  electrolysis,  $CO_2$  electrolysis, and the catalytic Reverse Water Gas Shift reaction (RWGS). More specifically, at the cathode of SOECs, two electrochemical reactions take place in parallel at the triple phase boundaries, i.e.,  $H_2O$  and  $CO_2$  electrochemical reductions. The oxygen ions ( $O^{2-}$ ), produced by these reactions, are moved to the anode, through an oxygen ion-conducting electrolyte (Yttria-Stabilized Zirconia – YSZ), where oxygen ( $O_2$ ) gas is formed [8,9]:

Cathode: 
$$H_2O + 2e^- \rightarrow H_2 + O^{2-} \Delta H_{800\,^\circ C} = 151.8 \text{ kJ/mol}$$
 (1)

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$$CO_2 + 2e^- \rightarrow CO + O^{2-} \Delta H_{800\ ^\circ C} = 185.9 \text{ kJ/mol}$$
 (2)

Anode: 
$$O^{2-} \to \frac{1}{2} O_2 + 2e^-$$
 (3)

Besides the electrolysis reactions mentioned above, the reversible water gas shift (RWGS) catalytic reaction also occurs at the cathode with the co-existence of  $CO_2$  and  $H_2O$ :

Cathode: 
$$H_2 + CO_2 \rightarrow H_2O + CO \quad \Delta H_{800 \ ^\circ C} = 36.8 \text{ kJ/mol}$$
 (4)

It thus appears that CO can be produced either electrocatalytically by the reduction of  $CO_2$  or catalytically via the RWGS reaction. Questionable conclusions have been proposed in the literature about the role of the RWGS reaction for CO production. Some studies mention that CO is mainly produced electrocatalytically and RWGS has a small participation [10–15], while other studies mention that CO is absolutely produced catalytically via the RWGS reaction [16,17]. However, it has not been shown conclusively to which degree the RWGS is responsible for CO production in an SOEC.

Specifically, Mogensen and co-researchers found that the performance of co-electrolysis on an Ni/YSZ cathode varied between the highest pure H₂O electrolysis and the lowest CO₂ electrolysis, being much closer to the prior process [13,14]. This implies the significant contribution of the RWGS reaction in the H₂O/CO₂ SOEC co-electrolysis process and also confirms the co-existence of CO₂ electrolysis. The above conclusion was supported by recently published results on computational modelling of the direct and indirect (with electro-generated H₂) reduction of CO₂ [18,19]. The direction of the Water Gas Shift (WGS) reaction could be forward or backward, depending on the operating conditions, indicating that at high temperatures (>838 °C), CO was produced electrochemically and the WGS was shifted in reverse towards H₂ and CO₂.

In contrast, Yue and Irvine [20] suggested that the RWGS reaction did not contribute critically in catalysing  $H_2O/CO_2$  co-electrolysis on s (La,Sr)(Cr,Mn)O₃ (LSCM) perovskite based cathode. The LSCM cathode displayed higher catalytic activity for pure CO₂ electrolysis than pure  $H_2O$  electrolysis, and the performance for  $H_2O/CO_2$  co-electrolysis was much closer to pure CO₂ electrolysis.

On the other hand, Bae et al., Hartvigsen et al., Kee et al., and Zhao et al. considered that CO production on Ni-based cathodes was mainly from the RWGS reaction, whereas the electrochemical processes are dominated by  $H_2O$  electrolysis with identical performances for  $H_2O$  electrolysis and  $H_2O/CO_2$  co-electrolysis [16,17,21].

According to the above mentioned reports, until now, no agreement has been reached regarding the role of the RWGS reaction in the production of CO. As a result, it is crucial to quantify the degree of CO formation for each reaction. The discrepancies regarding the CO production route can be affected both by the structural characteristics of the electrode/catalyst (specific surface area, reducibility and re-oxidation behaviour, porosity, particle size, ionic and electronic conductivity), but also by the operating conditions (gas composition, temperature, etc.) [9,22]. As reported by Li's study [10], the heterogeneous thermochemical reactions occur at the external surface of the cathode and they are 20–100 times faster than the electrochemical reactions, which occur close to the electrolyte at the three phase boundaries. Furthermore, structural modifications of the cathode could enhance mass transport and promote CO production through the catalytic RWGS reaction, resulting in  $H_2O/CO_2$  co-electrolysis performance close to that of  $H_2O$  electrolysis.

The majority of studies on high temperature  $H_2O/CO_2$  co-electrolysis utilize Ni-containing ceramic cathodes with YSZ and Gadolinia-Doped Ceria (GDC), similarly to the case of  $H_2O$  electrolysis [10–14]. Ni-based materials indeed are cheap and exhibit porous structure, high electronic conductivity, appropriate catalytic activity, and a similar Thermal Expansion Coefficient (TEC) with the electrolyte. Consequently, they could act as excellent SOECs cathodes for  $H_2O/CO_2$  co-electrolysis [7,22–26].

However, SOECs comprising Ni-based cathodes face some critical degradation issues, which are more pronounced with an increasing current density [3,12,14,27–33]. The main reasons for degradation have been reported to be microstructural changes that take place after prolonged co-electrolysis,

resulting in passivation and blocking of the Three-Phase Boundaries (TPB) area [12,13,28,30,31,33,34]. Post-mortem microscopy investigations in the Ni/YSZ electrode have shown irreversible damages of the electrode's microstructure, such as loss of Ni-YSZ contact, decomposition of YSZ, Ni grain growth, loss of Ni percolation (loss of Ni–Ni contact), and even migration of Ni from the fuel electrode [28,30,31,33]. Taking into account the above, it is critical to develop alternative cathode materials for H₂O/CO₂ co-electrolysis with improved structural properties.

Previous studies reported that the electrocatalytic efficiency of SOECs can be improved by alloying transition metals with the state of the art (SoA) Ni catalyst [35–39]. Modification by means of alloying may be a promising strategy to promote the catalytic activity due to ligand and strain effects that change the electronic structure [40] of the active element. Recent studies focused on the Ni–Co alloy with Sm-doped ceria (Ni–Co/SDC) as an alternative material, with enhanced performance for H₂O electrolysis in SOECs [36]. Specifically, the addition of Co increased the intrinsic catalytic activity of pure Ni and simultaneously expanded the active reaction region [36,41]. Other reports have shown that an Ni–Fe bimetallic cathode, mixed with  $Ba_{0.6}La_{0.4}CoO_3$  on  $La_{0.9}Sr_{0.1}Ga_{0.8}Mg_{0.2}O_3$  (LSGM) electrolysis [39], due to prevention of Ni aggregation. Furthermore, the enhanced performance of the Ni–Fe formulations could be attributed to the increased stability and improved catalytic activity, due to the addition of Fe, as this was concluded by combining experimental results with theoretical Density Functional Theory (DFT) calculations [41].

Our research group have studied the effect of Au and/or Mo doping on the physicochemical and catalytic properties of Ni/GDC for the reactions of catalytic CH₄ dissociation and methane steam reforming [23,42–45] in the presence of H₂ and H₂S impurities. These modifications resulted in SOFC fuel electrodes with high sulfur and carbon tolerance as well as with improved electrocatalytic activity. Regarding the doping level of Au and Mo, the amount of 3wt% was indicated as an appropriate loading for achieving the most promising results under the examined SOFC conditions. Recently, the 3wt% Au-Ni/GDC electrode was tested under SOEC conditions for the H₂O electrolysis process and exhibited improved electrocatalytic performance compared to Ni/GDC [46]. The enhanced performance of the Au-doped cathode was attributed to the creation of a surface Ni–Au solid solution, which causes weaker interplay of absorbed H₂O_{ads} and O_{ads} species with the modified cathode. The result is a durable and resistant electrode to surface oxidation with an improved "three phase boundaries" area, especially at temperatures higher than 800 °C [46].

The presented investigation deals with modified X-Ni/GDC electrodes (where X = Au, Mo, Fe), in the form of half-Electrolyte Supported Cells (ESCs), for their performance in the RWGS through catalytic-kinetic measurements. Ni/GDC, 3wt% Au-Ni/GDC, 3wt% Mo-Ni/GDC, 3wt% Au–3wt% Mo-Ni/GDC, and 2wt% Fe-Ni/GDC modified cathodes were tested at Open Circuit Potential (OCP) conditions to elucidate their catalytic activity towards the production of CO ( $r_{co}$ ), which is one of the products from the H₂O/CO₂ co-electrolysis reaction. The latter approach is considered as an attempt to create a reference profile for the catalytic performance of the candidate electrodes, by applying the same H₂O/CO₂/H₂ feed conditions as those under the co-electrolysis operational mode. The samples, both in the form of powders and as half cells, were physicochemically characterized, including specific redox stability measurements in the presence of H₂O.

## 2. Results and Discussion

#### 2.1. Physicochemical Characterization

#### 2.1.1. Specific Surface Area Values of Au–Mo–Fe-Modified Powders

The specific surface area values (SSA) for the modified-NiO/GDC samples, calcined in air at 600 and 1100  $^{\circ}$ C, as well as after reduction with H₂ at 850  $^{\circ}$ C, are presented in Table 1.

The first remark is that all samples exhibit quite similar and low SSA values, which decrease further by increasing the calcination temperature from 600  $^{\circ}$ C to 1100  $^{\circ}$ C and after H₂ reduction at

850 °C. Furthermore, Mo and Au–Mo modification slightly decrease the SSA of Ni/GDC, while the addition of Fe causes a significant increase in the SSA values up to approximately 70%. Taking into account the generally low values and the  $\pm 0.2 \text{ m}^2 \text{ g}^{-1}$  accuracy limits of the measurements, it can be concluded that the reduced Mo and Au–Mo-Ni/GDC powders have similar SSAs with Ni/GDC. On the other hand, the 70% increase in the SSA of the Fe modified sample can be primarily ascribed to the formation of FeO_x species, in the oxidized form of the sample, which inhibit the decrease of SSA during reduction.

**Table 1.** BET Specific Surface Area (SSA_{BET}) of commercial NiO/GDC, and X-modified NiO/GDC samples (where X = Au, Mo, and Fe). The measurements were performed on powders, calcined at 600 °C, 1100 °C, and after H₂-reduction at 850 °C. Error/accuracy of SSA_{BET} =  $\pm 0.2$  m² g⁻¹.

	$SSA_{BET}$ (m ² g ⁻¹ )					
Sample Powder	T = 600 °C, (oxidized)	T = 1100 $^{\circ}$ C, (oxidized)	After H ₂ -reduction at T = 850 °C			
NiO/GDC	5.3	2.9	2.0			
3Au-NiO/GDC	5.4	2.7	2.1			
3Mo-NiO/GDC	5.0	2.2	1.7			
3Au-3Mo-NiO/GDC	5.1	2.4	1.8			
2Fe-NiO/GDC	8.8	4.0	3.4			

2.1.2. H₂ Reducibility and air Re-Oxidation Behaviour of Au–Mo–Fe-Modified Powders

Temperature Programmed Reduction followed by Oxidation (H₂-TPR, Air-TPO) measurements were performed, in order to investigate the reducibility and re-oxidation behavior of the prepared composite powders, as well as the existence of possible bulk effects due to the modification with Au, Mo, and Fe. The corresponding H₂-TPR and Air-TPO TGA profiles of modified-NiO/GDC samples, calcined at 1100 °C, are presented in Figure 1.



Figure 1. Cont.



**Figure 1.** TGA (**A**) H₂-TPR (12 vol.% H₂/Ar), and (**B**) O₂-TPO (33 vol.% air) profiles of X-modified NiO/GDC samples (where X= Au, Mo, and Fe). The measurements were performed on powders, calcined at 1100 °C.  $F_{total} = 60 \text{ cm}^3/\text{min}$  (STP conditions: 0 °C, 1 atm).

 $H_2$ -TPR and air-TPO measurements showed different reduction and oxidation behavior in each of the examined modified samples. In particular, two main reduction peaks (Figure 1A) and one oxidation peak (Figure 1B) are mainly observed, which can be attributed to the reduction of NiO species to Ni and the reverse [42,47]. Moreover, the presence of Au does not seem to affect significantly the reduction of NiO. On the other hand, MoO₃ and, to a certain extent, Fe inhibit the NiO reduction, implying a stronger Ni–O bond. The inhibition effect of MoO₃ is also observed in the case of the ternary 3Au–3Mo-NiO/GDC sample and indicates a bulk interaction between Ni–Au–Mo, which has been thoroughly investigated in previous studies [23,42,43]. In particular, the second broad peak ("shoulder") at 580 °C, which is observed for the 3Mo-NiO/GDC and 3Au–3Mo-NiO/GDC samples, may be associated with the reduction of MoO₃ species, which are reduced at higher temperatures in the range of 600–770 °C [42,47]. The presence of Fe seems to affect the main reduction peak of NiO, in a similar way as MoO₃, suggesting the possible formation of a solid solution between Ni and Fe. The latter observation is currently under further clarification.

Upon re-oxidation of the samples, the unmodified Ni/GDC as well as the Au modified sample are oxidized practically at the same temperature with a similar TPO profile. The reduced state of the Fe, Mo-, and Au–Mo- modified samples proved to be more resistant to re-oxidation in 33 vol.% air/Ar (6.7 vol.%  $O_2$ /Ar), since they had to reach higher temperature for complete re-oxidation.

## 2.1.3. H₂O Re-Oxidation Profiles of Au-Mo-Fe-Modified Powders

Isothermal-TGA measurements, under 15.5 vol.%  $H_2O/Ar$  conditions, were carried out at 650 °C, 700 °C, 750 °C, and 800 °C and the results are depicted in Figures 2 and 3. These measurements investigate the activity of the powders for the  $H_2O$  dissociation reaction and their concomitant
re-oxidation rate.  $H_2O$  acts as an oxidative agent and interacts with the Ni atoms on the surface of each sample towards  $H_2$  and NiO [reaction (5)] [46].

$$H_2O + Ni \rightarrow H_2 + NiO$$
 (5)

Thereafter, there is a progressive diffusion of the absorbed oxygen species ( $O_{ads}$ ) from the surface in the bulk phase of the sample, which is oxidized further [46,48–50]. Figure 2 depicts the H₂O re-oxidation profiles, as an increase of the weight ( $\Delta$ wt%), of the pre-reduced Ni/GDC, 3wt% Au-Ni/GDC, 3wt% Mo-Ni/GDC, and 2wt% Fe-Ni/GDC samples in the temperature range between 650–800 °C.



**Figure 2.** Isothermal TG analysis of the pre-reduced, (**A**) Ni/GDC, (**B**) 3wt% Au-Ni/GDC, (**C**) 3wt% Mo-Ni/GDC, (**D**) 3wt% Au–3wt% Mo-Ni/GDC, and (**E**) 2wt% Fe-Ni/GDC samples in the temperature range between 650–800 °C. (**F**) Comparative TG profiles of the samples at 800 °C. Feed Conditions: 15.5 vol.% H₂O/Ar,  $F_{total} = 100 \text{ cm}^3/\text{min}$  (STP conditions: 0 °C, 1 atm).



**Figure 3.** Arrhenius plots of the H₂O re-oxidation rate (resulting from the slope of the isothermal TG profiles in Figure 2) for: Ni/GDC, 3wt% Au-Ni/GDC, 3wt% Mo-Ni/GDC, 3wt% Au-3wt% Mo-Ni/GDC, and 2wt% Fe-Ni/GDC Feed Conditions: 15.5 vol.% H₂O/Ar,  $F_{total} = 110 \text{ cm}^3/\text{min}$  (STP conditions: 0 °C, 1 atm).

The re-oxidation profiles of the samples in Figure 2 can be separated in three sections [46]. The first section corresponds to the initial sharp increase in weight and is ascribed [46] to the complete, bulk, re-oxidation of the partially reduced CeO₂ in GDC and specifically of Ce³⁺ to Ce⁴⁺. The second section occurs during the following 12–25 min of the reaction, where the samples, except for 2Fe-Ni/GDC, keep their reduced state without any changes in their weight. This step is associated [46] to the dissociation of bulk NiH species, which are formed during the prior reduction period, and maintains the samples reduction. NiH dissociation is considered as an activated process, because by decreasing the operating temperature from 800 °C to 650 °C, the specific "delay" period increases from 12 min to 25 min. The first two sections are quite similar for the pre-reduced powders, with identical low SSA values (Table 1). On the other hand, the initial re-oxidation behavior, in the first section, of the pre-reduced 2Fe-Ni/GDC is more intense, indicating the strong effect of H₂O.

In the third section of the TG profiles, further re-oxidation takes place [reaction (5)] and the main discrepancies are detected among the samples. In the range of 650–800 °C, the ternary 3Au–3Mo-Ni/GDC sample is the most tolerant in bulk re-oxidation by H₂O, while the binary 2Fe-Ni/GDC sample is the least tolerant. The decrease in temperature ( $800 \rightarrow 650$  °C) leads to inhibition of the bulk re-oxidation, as observed by the reduced slopes in the TG profiles, whereas the trend among the samples does not change. The high tolerance of the 3Au–3Mo-Ni/GDC sample against bulk re-oxidation by H₂O is attributed to the synergistic interaction of nickel with gold and molybdenum [23,42].

The slope on the linear part of the third section in the TG curves represents the intrinsic dissociation rate of  $H_2O$  and correspondingly the re-oxidation rate of Ni atoms on each sample [reaction (5)] [46]. Figure 3 depicts the Arrhenius plots of the oxidation rates of the samples, which were calculated from the slopes on the linear parts of the TG profiles (Figure 2). The Arrhenius plots (Figure 3) show that the binary 2Fe-Ni/GDC sample has the highest  $H_2O$  re-oxidation rates, while the ternary 3Au–3Mo-Ni/GDC has the lowest. This is further confirmed by the calculated apparent activation energies ( $E_{a,ap}$ ), which are defined from the Arrhenius plots and are reported in Table 2.

Sample Powder	$E_{a, app}$ (kJ mol ⁻¹ )	
Ni/GDC [46]	41	
3Au-Ni/GDC [46]	56	
3Mo-Ni/GDC	45	

63

20

3Au-3Mo-Ni/GDC

2Fe-Ni/GDC

**Table 2.** Apparent activation energies (kJ mol⁻¹) for the dissociation of H₂O on commercial Ni/GDC [46], and modified Ni/GDC samples. The values correspond to the slopes of the Arrhenius plots in Figure 3.

The calculated values verify that the binary 2Fe-Ni/GDC sample exhibits the lowest  $E_{a,ap}$  (20 kJ/mol) for the catalytic dissociation of H₂O, whereas the ternary 3Au–3Mo-Ni/GDC sample exhibits the highest  $E_{a,ap}$  (63 kJ/mol). The above behavior can be interpreted by the fact that Ni-Fe interaction reinforces the bond of H₂O_{ads}, resulting in a higher re-oxidation rate (lower  $E_{a,ap}$ ). On the other hand, following the same interpretation, Ni–Au–Mo interaction weakens the bond of H₂O_{ads}, resulting in a lower re-oxidation rate (higher  $E_{a,ap}$ ) [46].

The calculated  $E_{a,ap}$  values are in good agreement with our previous study [46] and other studies [51–53] in the literature, which focus on theoretical investigations. The binary 2Fe-Ni/GDC sample is the most active for the thermochemical dissociation of H₂O and is less resistant against bulk re-oxidation. On the contrary, the ternary 3Au–3Mo-Ni/GDC sample exhibits the lowest activity for the thermochemical dissociation of H₂O and the highest tolerance against bulk re-oxidation.

Taking into account the above,  $H_2O$ , apart from being the main reactant of the co-electrolysis process, is also considered as a potential poisoning agent of the electrode. This means that the bonding strength of the adsorbed oxygen species, which result from  $H_2O$  decomposition, may induce the re-oxidation of the electrode and finally the deactivation of the sample [46,53]. The  $H_2O$  poisoning effect in electrolysis and co-electrolysis processes can also be correlated to the CH₄ poisoning effect in SOFC applications, where degradation is enhanced by the strong adsorption bond of CH₄ on the Ni surface [23,42–45,54,55]. Thus, the stronger binding of the adsorbed oxygen species, which result from the catalytic dissociation of  $H_2O_{ads}$ , can similarly cause a poisoning effect on the Ni surface, leading to faster re-oxidation and finally deactivation of the sample [46,53]. According to Besenbacher et al. [56], the surface solid solution between Au and Ni induces significant modifications in the electronic properties of Ni (Fermi level, work function, and d-band center) and affects the bonding strength of the adsorbed species on the surface of the sample.

In this respect, the Au- and Au–Mo- doping of Ni should shift the d-band center to lower energies, with respect to the Fermi level of nickel, thus inhibiting the interaction and the dissociation of H₂O. Therefore, the 3Au-Ni/GDC and 3Au–3Mo-Ni/GDC samples are suggested to be the least active samples for the dissociation of H₂O, having the highest  $E_{a,app}$  and consequently the lowest binding energies for both H₂O and O_{ads}. The binary 2Fe-Ni/GDC is suggested to be the most active sample for the dissociation of H₂O, with the lowest  $E_{a,app}$ , resulting in O_{ads} species with the highest binding energy and the potential to cause faster re-oxidation/poisoning of the electrode [46,53] during SOEC operation.

#### 2.2. Catalytic-Kinetic Measurements of the RWGS Reaction

All the prepared samples, in the form of half-electrolyte supported cells, were tested at Open Circuit Potential (OCP) conditions to elucidate their catalytic activity towards the production of CO, which is one of the products from the  $H_2O/CO_2$  co-electrolysis reaction. Furthermore, the effect of the modification (type of dopant: Au, Mo, Fe) on the catalytic activity for the production of CO, through the RWGS reaction, was also investigated. The latter approach is considered as an attempt to create a reference profile for the catalytic performance of the candidate electrodes, under the same experimental conditions as in co-electrolysis (including the presence of current collector). This is an important step, because it will provide detailed experimental feedback on the possible contribution of the RWGS reaction to the production of CO and the extent of this contribution to the electrochemical reduction of CO₂, during the co-electrolysis mode.

In regards to the "homogenous" catalytic production of CO, the rate was found to be negligible, in the range of  $r_{co,homogenous} = 0.05 \ \mu mol/s$ , for a fuel feed comprising 30 vol.% He – 24.5 vol.% H₂O – 24.5 vol.% CO₂ – 21 vol.% H₂. Concerning the current collector, Ni-mesh shows some catalytic activity for the production of CO ( $r_{CO}$ ), which increases by increasing the operating temperature and the partial pressure of H₂. However, comparative  $r_{CO}$  measurements (see Figure S1 in Supplementary Material) of an Ni/GDC electrode with and without the presence of Ni mesh as the current collector, as well as of the bare Ni mesh, suggest that there is no direct catalytic correlation/contribution of the Ni mesh to the activity of the electrocatalysts. This is mainly explained by the fact that Ni/GDC is a porous electrocatalyst with SSA_{BET} and thus with more active sites, compared to the metallic nickel mesh, which does not possess similar properties. Consequently, from the point where they co-exist as the electrode and current collector, the catalytic activity is mainly attributed to the electrocatalyst.

Figure 4 presents the comparison of the produced  $r_{CO}$  for the Ni/GDC and for modified-Ni/GDC electrodes, under three different H₂O-CO₂-H₂ mixtures. The corresponding %CO₂ conversions are depicted in Figure 5. The values are low enough to be considered in the differential region, apart from the case of 2Fe-Ni/GDC in the reaction mixture (A), which is relatively high. The % conversion of CO₂ was calculated by the following formula:

$$CO_2 \text{ conversion } (\%) = \frac{F_{CO}^{\text{out}}}{F_{CO_2}^{\text{in}}} \cdot 100$$
(6)

where:  $F_{CO_2}^{out}$  and  $F_{CO_2}^{in}$  correspond to the rate (µmol/s) of the produced CO and of the introduced CO₂ in the reactants feed, respectively.



**Figure 4.** CO production rates ( $\mu$ mol s⁻¹ g⁻¹) on ESCs comprising: Ni/GDC, 3Au-Ni/GDC, 3Mo-Ni/GDC, 3Au-3Mo-Ni/GDC, and 2Fe-Ni/GDC, as fuel electrodes, in the range of 800–900 °C under three different mixtures: (**A**) 24.5% H₂O – 24.5% CO₂ – 21% H₂ (P_{H2}/P_{CO2} = 0.86), (**B**) 28% H₂O – 28% CO₂ – 14% H₂ (P_{H2}/P_{CO2} = 0.50), and (**C**) 31.5% H₂O – 31.5% CO₂ – 7% H₂ (P_{H2}/P_{CO2} = 0.22). Dilution of He: 30 vol.% and F_{total} = 140 cm³/min (STP conditions: 0 °C, 1 atm) in all cases. The dash line (––) corresponds to the r_{CO} values in thermodynamic equilibrium. All studied electrodes have similar loading in the range of 10–12 mg/cm².



**Figure 5.** The corresponding %CO₂ conversions of ESCs comprising: Ni/GDC, 3Au-Ni/GDC, 3Mo-Ni/GDC, 3Au-3Mo-Ni/GDC, and 2Fe-Ni/GDC, as fuel electrodes, in the range of 800–900 °C under three different mixtures: (**A**) 24.5% H₂O – 24.5% CO₂ – 21% H₂ (P_{H2}/P_{CO2} = 0.86), (**B**) 28% H₂O – 28% CO₂ – 14% H₂ (P_{H2}/P_{CO2} = 0.50), and (**C**) 31.5% H₂O – 31.5% CO₂ – 7% H₂ (P_{H2}/P_{CO2} = 0.22). Dilution of He: 30 vol.% and F_{total} = 140 cm³/min (at STP conditions: 0 °C, 1 atm) in all cases. All studied electrodes have similar loading in the range of 10–12 mg/cm².

It is shown (Figures 4 and 5) that 2Fe-Ni/GDC and 3Mo-Ni/GDC have the highest catalytic activity of the examined electrodes for the RWGS reaction. In fact, 2Fe-Ni/GDC is the most active in terms of the produced CO. The above performance is observed for all applied fuel feeds ( $P_{H2}/P_{CO_2} = 0.86, 0.50$  and 0.22), whereas it is enhanced by increasing (i) the operating temperature and (ii) the partial pressure of H₂ in the fuel feed. The enhanced catalytic performance of the Fe-modified sample, can be primarily ascribed to the possible stronger adsorption and consequent catalytic dissociation of CO₂ on the active sites of the catalyst. This first conclusion is going to be further examined with specific CO₂ Temperature Programmed Desorption (TPD) measurements.

Another noteworthy remark is that the produced CO rates were compared to the thermodynamic equilibrium rates (dash line in Figure 4) for the three different  $H_2O-CO_2-H_2$  reaction mixtures. The equilibrium rates were calculated by using the equilibrium constant (K_{eq}) formula that is reported in [57,58]. It was found (Figure 4) that Ni/GDC, the binary Au-, and the ternary Au–Mo- modified samples exhibit CO production rates, which are lower than the thermodynamic equilibrium for all the examined reaction conditions. The performances of 2Fe-Ni/GDC and 3Mo-Ni/GDC are closer to the equilibrium, but cannot be considered as thermodynamically limited.

The results from the measurements in Figure 4 can also be presented as Arrhenius plots, depicted in Figure 6, where the derived apparent activation energies ( $E_{a, app}$ ) for the production of CO and consequently for the RWGS reaction are listed in Table 3.



**Figure 6.** Arrhenius plots of the CO production rates ( $\mu$ mol s⁻¹ g⁻¹) on ESCs comprising: Ni/GDC, 3Au-Ni/GDC, 3Au-3Mo-Ni/GDC, and 2Fe-Ni/GDC, in the range of 800–900 °C, under three different mixtures: (**A**) 24.5% H₂O – 24.5% CO₂ – 21% H₂ (P_{H2}/P_{CO2} = 0.86), (**B**) 28% H₂O – 28% CO₂ – 14% H₂ (P_{H2}/P_{CO2} = 0.50), and (**C**) 31.5% H₂O – 31.5% CO₂ – 7% H₂ (P_{H2}/P_{CO2} = 0.22). Dilution of He: 30 vol.% and F_{total} = 140 cm³/min (at STP conditions: 0 °C, 1 atm) in all cases.

 $\begin{array}{l} \mbox{Table 3. Apparent activation energies } (E_{a,app}, kJ/mol) \mbox{ for the RWGS reaction on ESCs for three different mixtures: (A) 24.5% H_2O - 24.5% CO_2 - 21% H_2 (P_{H2}/P_{CO_2} = 0.86), (B) 28% H_2O - 28% CO_2 - 14% H_2 (P_{H2}/P_{CO_2} = 0.50), \mbox{ and (C) 31.5% H_2O - 31.5% CO_2 - 7% H_2 (P_{H2}/P_{CO_2} = 0.22). } \end{array}$ 

Sample	${ m E}_{ m a,  apparent}$ (kJ mol $^{-1}$ ) and $A^{*}$ ( $\mu mol  s^{-1}  g^{-1}$ ), per reaction mixture					
	$P_{\rm H2}/P_{\rm CO_2} = 0.86$		$P_{\rm H2}/P_{\rm CO_2} = 0.50$		$P_{\rm H2}/P_{\rm CO_2} = 0.22$	
Ni/GDC	20	$2.3 \times 10^3$	22	$3.6  imes 10^3$	20	$1.0  imes 10^3$
3Au-Ni/GDC	85	$1.9  imes 10^6$	74	$5.0  imes 10^5$	46	$1.5  imes 10^4$
3Mo-Ni/GDC	27	$6.5  imes 10^3$	14	$1.1  imes 10^3$	21	$1.4  imes 10^3$
3Au-3Mo-Ni/GDC	43	$1.7  imes 10^4$	45	$1.5  imes 10^4$	50	$1.6  imes 10^4$
2Fe-Ni/GDC	21	$3.3  imes 10^3$	19	$1.6  imes 10^3$	21	$1.3  imes 10^3$

^{*}*A*: The pre-exponential factor in the Arrhenius equation,  $\mathbf{r} = \mathbf{A} \cdot \exp(-\frac{E_{a,app}}{R \cdot T})$ .

The Arrhenius plots and the calculated  $E_{a, app}$  show that 2Fe-Ni/GDC, 3Mo-Ni/GDC, and Ni/GDC have practically the same and the lowest  $E_{a, app}$ . However, the Fe- and Mo- modified samples exhibit higher pre-exponential factors, compared to Ni/GDC, which explains the higher production rates of CO. 3Au-Ni/GDC shows overall the highest  $E_{a, app}$ , which is an additional indication for its worst catalytic activity. Finally, the ternary 3Au–3Mo-Ni/GDC sample shows an apparent activation energy, which lies between that for 2Fe-Ni/GDC and 3Au-Ni/GDC. However, the catalytic activity of the ternary sample is the lowest, due to the lower pre-exponential factor compared to the binary Au-modified sample. According to the knowledge of the authors, there are no literature data available for experimentally measured  $E_{a, app}$ , for similar samples and reaction conditions. The so far available data come from theoretical investigations and there is a recent study from Cho et al. [41], who performed DFT calculations to evaluate the ability of various transition metals to increase the activity of Ni for the H₂O/CO₂ co-electrolysis. In particular, they computed the activation energies of specific elementary reaction steps and in the case of the RWGS on Ni(111), the  $E_a$  was found to be approximately 46 kJ/mole, which is very close to the values that were experimentally calculated in the present study.

The effect of H₂ partial pressure on the catalytic rate of CO production is further verified in Figure 7 for all samples at 900 °C and 800 °C. The 2Fe-Ni/GDC sample is the most active and 3Au–3Mo-Ni/GDC the least one. The 3Au-Ni/GDC catalyst at 900 °C shows similar performance with that of Ni/GDC. In addition, by decreasing the temperature at 800 °C, the catalytic activity of 3Au-Ni/GDC exhibits the highest reduction. This is apparent from the significant decrease in the slope of the corresponding curve (Figure 7) from 800 °C to 900 °C and can be further explained from the calculated  $E_{a, app}$  (Table 3), which is the highest from all samples. Finally, the kinetic behavior of all samples suggests that the production rate of CO exhibits a positive order dependence on the partial pressure of H₂.



**Figure 7.** Steady state effect of the H₂ partial pressure (pH₂, kPa) on the CO production rates (µmol s⁻¹ g⁻¹) on ESCs comprising: Ni/GDC, 3Au-Ni/GDC, 3Mo-Ni/GDC, 3Au-3Mo-Ni/GDC, and 2Fe-Ni/GDC, for 900 °C and 800 °C. The pH₂O/pCO₂ ratios are fixed and their values are presented in Table 3.

The fact that in the majority of the samples there is no trend in the  $E_{a, app}$ , by varying the pH₂/pCO₂ ratios, primarily shows that all samples, apart from the binary Au-modified, have similar intrinsic catalytic activity for the RWGS reaction. This is also concluded from the remark that the effect of the H₂ partial pressure on the CO production rate is almost linear. On the other hand, in the case of 3Au-Ni/GDC, there is an increase in  $E_{a, app}$ , by increasing the pH₂/pCO₂ ratio. This is corroborated from the results in Figure 7, where r_{CO} seems to reach a plateau at high pH₂. The latter remarks indicate that the majority of the samples have actives sites that follow a similar reaction mechanism, in regards to the dissociative adsorption of H₂ and CO₂ towards the RWGS. On the other hand, the reaction mechanism seems to be different in the case of 3Au-Ni/GDC. In particular, it is implied that the high coverage of adsorbed H₂ may inhibit the CO₂ dissociative adsorption and thus decreases the intrinsic

catalytic activity of the specific sample. This suggestion is going to be further clarified through the currently performed  $H_2O/CO_2$  co-electrolysis measurements.

Carbon formation, through the Boudouard reaction (6), was also investigated and the results are presented in Figure 8. The slight scattering of the experimentally measured rates in combination with the theoretical values, where  $r[CO_2]_{inlet}$  is equal to  $r[CO + CO_2]_{outlet}$ , suggest that there is no carbon deposition under the examined reaction conditions. This result is also in accordance with the fact that the Boudouard reaction is not thermodynamically favored above 750 °C [59].



**Figure 8.** Investigation of the Boudouard reaction at 900 °C and 800 °C under OCP conditions on ESCs comprising: Ni/GDC, 3Au-Ni/GDC, 3Mo-Ni/GDC, 3Au-3Mo-Ni/GDC, and 2Fe-Ni/GDC. Theoretical case where  $[CO_2]_{inlet}$  is equal to  $[CO+CO_2]_{outlet}$  is also depicted with a solid line.

Previous studies [1,59,60] reported that at realistic CO₂/CO concentrations, during CO₂ electrolysis in the range of 650–750 °C, the equilibrium of (6) is shifted towards CO production (reaction (7)) and therefore carbon will not be formed, catalytically, during CO₂ electrolysis. In the case where the cell is operated at OCP conditions, as in our case, any coke deposited within the porous, modified or not, Ni/GDC electrode would be oxidized to CO according to the reverse Boudouard reaction (7) and thereby removed. Furthermore, the addition of steam in the feed is reported to remove carbon depositions according to reaction (8) [1,8]:

Boudouard reaction: 
$$2CO \rightarrow CO_2 + C$$
 (7)

Reverse Boudouard reaction:  $CO_2 + C \rightarrow 2CO$  (8)

Reaction of coke with 
$$H_2O$$
:  $C + H_2O \rightarrow CO + H_2$  (9)

## 3. Materials and Methods

# 3.1. Preparation of Electrocatalysts

The modified cathode powders were prepared via the Deposition-Precipitation (D.P.) and Deposition-Co Precipitation (D.CP.) methods by using the commercial NiO/GDC cermet (65wt% NiO-35wt% GDC, Marion Technologies, Verniolle, France) as the support. The precursors for the 3wt% Au-NiO/GDC, 3wt% Mo-NiO/GDC, 3wt% Au-3wt% Mo-NiO/GDC, and 2wt% Fe-NiO/GDC samples were the HAuCl₄ (Sigma-Aldrich) or/and (NH₄)₆Mo₇O₂₄ (Sigma-Aldrich, St. Louis, MO, USA) and Fe(NO₃)₃x9H₂O (Sigma-Aldrich, St. Louis, MO, USA) solutions, respectively. Full details about the synthesis can be found elsewhere [42,43]. After filtering, the precipitate was dried at 110 °C for 24 h. All dried powders were calcined at 600 °C/90 min and a part of them at 1100 °C/75 min. The first batch was used for the paste preparation, which is described in the following paragraph. The batch at 1100 °C was used for the physicochemical characterization. In this way, the prepared catalysts were studied at similar thermal stress as the calcined electrode-electrolyte assemblies.

## 3.2. Preparation of Electrolyte-Supported Half Cells

The electrolyte-supported half cells consisted of circular shaped planar 8YSZ electrolyte (by Kerafol) with a 25 mm diameter and 300  $\mu$ m thickness. The fuel electrode was deposited by using the screen printing technique as reported in previous studies [23,46]. In particular, a paste was prepared by using an appropriate amount of the electrocatalyst (modified NiO/GDC powder), terpineol (Sigma-Aldrich, St. Louis, MO, USA) as the dispersant, and PVB (polyvinylbutyral, Sigma-Aldrich, St. Louis, MO, USA) as binder. After the deposition of the paste, the cell was sintered at 1150 °C with a heating/cooling ramp rate of 2 °C/min. The last temperature is the lowest possible in order to obtain proper adherence of the electrode on the electrolyte, whereas it is equivalent with the calcination temperature of the characterized powders (1100 °C). The examined electrodes were approximately 20  $\mu$ m thick and their loading varied in the region of 10–12 mg/cm² with a 1.8 cm² geometric surface area (Figure 9). The prepared half cells were adjusted on a ceramic YSZ tube and sealed airtight with a glass sealing material manufactured by Kerafol.





**Figure 9.** Images of a (half) electrolyte supported cell (ESC) prepared by screen printing: (A) NiO/GDC electrode and (B) SEM cross section perpendicular to the NiO/GDC//YSZ (half) electrolyte-supported cell.

#### 3.3. Physicochemical Characterization

The samples, in the form of powders, were characterized with BET,  $H_2$ -TPR, Air-TPO, and TGA re-oxidation measurements in the presence of  $H_2O$ .

The BET specific surface area values of the samples were determined from the adsorption isotherms of nitrogen at -196 °C, recorded with a Micromeritics TriStar 3000 apparatus (Micromeritics, Norcross, GA, USA).

The re-oxidation measurements in the presence of H₂O, as well as H₂-TPR and Air-TPO measurements were carried out with a TA Q50 instrument. The H₂O re-oxidation properties of the powders were studied with TGA measurements, at a constant temperature in the range of 650–800 °C. Before the measurement, the samples were pre-reduced in-situ with 80 vol.% H₂/Ar at 800 °C for 100 min and then the feed was changed to 15.5 vol.% H₂O/Ar. The total flow rate (at STP conditions: 0 °C, 1 atm) was 100 cm³/min and the loading of the measured samples was approximately 50 mg. Steam was introduced in the reactor by passing Ar through a saturator, which was maintained at a fixed temperature (65 °C).

### 3.4. Catalytic-Kinetic Measurements

The prepared half cells were catalytically investigated at Open Circuit Potential (OCP) conditions for the RWGS reaction, in the presence of Ni mesh. The catalytic experiments were accomplished at temperatures between 800–900 °C under various H₂O/CO₂/H₂ mixtures by keeping in all cases the P_{H₂O}/P_{CO₂} ratio constant. In regards to the experimental part of these measurements, H₂O was added and handled in the feed in the form of steam, as in the SOEC measurements. Before its evaporation, liquid H₂O was pressurized in a container and circulated in the system by means of a liquid water mass flow controller. Then, liquid H₂O was evaporated through lines and valves, heated at 160 °C, to prevent water condensation. The flow rate was fixed at 140 cm³/min (at STP conditions: 0 °C, 1 atm), avoiding any mass transfer limitation effects in the reactor. Reactants and products were determined by using an on-line gas chromatograph (Varian CP-3800) with a thermal conductivity detector. Further details regarding the experimental parameters are indicated in the corresponding Figures.

### 4. Conclusions

The presented study deals with the kinetic investigation of Ni-based (modified or not) electrodes towards their performance for the RWGS reaction. The samples were examined in the form of electrolyte-supported (half) cells and the measured kinetic parameter was the production rate of CO. The main objective was to clarify the effect of the modification on the catalytic activity for the RWGS, which is considered as a key reaction for the CO production under  $H_2O/CO_2$  co-electrolysis operation. The reaction conditions were similar to those that are applied under co-electrolysis mode.

Redox stability measurements in the presence of H₂O showed that the ternary 3Au–3Mo-Ni/GDC electrode is the least active sample for the dissociation of H₂O, having the highest  $E_{a,app}$  and consequently the lowest binding energy for the H₂O_{ads}. On the other hand, the binary 2Fe-Ni/GDC is the most active sample for the dissociation of H₂O, thus having the potential to experience faster re-oxidation. Complementary characterization suggests that the interaction of Ni and Fe (FeO_x species in the oxidized form of the sample) during the H₂-reduction process increases the SSA_{BET} and affects the bulk properties of the binary Ni-Fe/GDC. The interaction can be realized through the possible formation of an Ni-Fe solid solution, which is currently under further clarification and may be responsible for enhancing the catalytic dissociation of H₂O_{ads}.

The kinetic study of the candidate electrocatalysts showed that Au modification inhibits the catalytic production of CO, through the RWGS reaction, while modification with Fe or Mo induces an enhancement of  $r_{CO}$ . In fact, the 2wt% Fe-Ni/GDC sample is the most active both in terms of %CO₂ conversion and of the produced CO. In addition, a negative synergy was observed for the ternary Au–Mo-Ni modified sample. Specifically, the 2wt% Fe-Ni/GDC and 3wt% Mo-Ni/GDC samples

showed similar apparent activation energy for the RWGS reaction as that of Ni/GDC, while the 3wt% Au-Ni/GDC and 3wt% Au-3wt% Mo-Ni/GDC samples showed lower  $E_{a,app}$ . The above performance is observed for all applied fuel feeds ( $P_{H2}/P_{CO_2} = 0.86$ , 0.50, and 0.22), whereas it is enhanced by increasing (i) the operating temperature and (ii) the partial pressure of  $H_2$ .

Consequently, the kinetic behavior of all samples suggests that the production rate of CO exhibits a positive order dependence on the partial pressure of  $H_2$ , whereas carbon formation was not detected. In the case of the most active 2Fe-Ni/GDC sample, it is proposed that Fe-modification may enhance the catalytic dissociative adsorption of CO₂ towards the production of CO and this is further investigated. Finally, it is worth mentioning that the RWGS catalytic performance of both Fe- and Mo- modified samples is close to the equilibrium, but cannot be considered as thermodynamically limited.

Overall, the presented results correspond to a reference catalytic profile of the examined modified Ni/GDC samples for the RWGS reaction. These candidate electrocatalysts are currently being examined, as full electrolyte supported cells, in SOEC measurements for the  $H_2O/CO_2$  co-electrolysis reaction to elucidate any additional effects by the applied current. Moreover, further investigation of the Fe-modification on NiO/GDC for the  $H_2O/CO_2$  electrolysis processes will occur in the future.

**Supplementary Materials:** The following are available online at http://www.mdpi.com/2073-4344/9/2/151/s1, Figure S1. Kinetic study of the R.W.G.S. reaction on the nickel mesh ( $\blacksquare$ ), Ni/GDC with Ni mesh ( $\blacklozenge$ ) and Ni/GDC without Ni mesh ( $\triangle$ ), in the range of 800–900 °C, under two different mixtures: (A) 24.5% H₂O – 24.5% CO₂ – 21% H₂ (P_{H2}/P_{CO2} = 0.86) and (B) 28% H₂O – 28% CO₂ – 14% H₂ (P_{H2}/P_{CO2} = 0.50). Dilution of He: 30 vol.% and F_{total} = 140 cm³/min (at STP conditions: 0 °C, 1 atm) in all cases. The loading of Ni/GDC is 11 mg/cm².

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