

## Article

# A Study of Dispersed, Thermally Activated Limestone from Ukraine for the Safe Liming of Water Using ANN Models

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**Abstract:** Liming surface water is a fairly popular method of increasing the pH values and decreasing the concentration of phosphates and heavy metals. According to the Environmental Protection Agency (EPA) recommendations, the increase of water pH should not exceed 1.5. If surface water is the source of water supply, liming is a process that reduces water contamination. This should prevent the creation of an additional load for the water treatment plants in urban settlements. This article is an interdisciplinary research study aiming to (1) determine and compare the doses of new dispersed, thermally activated limestone and natural limestone, (2) find the relation between dose value and initial water parameters (pH, Eh and total mineralization), and (3) create an artificial neural network (ANN) model to predict changes in water pH values according to EPA recommendations. Recommended doses were obtained from experimental studies, and those of dispersed, thermally activated limestone were lower than the doses of natural limestone. Neural networks were used to predict the changes in water pH values when adding different doses of limestone with different initial water parameters using the ANN model. Four ANN models with different activation functions and loss function optimizers were tested. The best results were obtained for the network with the ReLU activation function for hidden layers of neurons and Adam's loss function optimizer ( $MAPE = 14.1\%$ ;  $R^2 = 0.847$ ). Further comparison of the results of the loss function and the results of calculating the quality metric for the training and validation dataset has shown that the created ANN can be used to solve the set research issue.

**Keywords:** limestone; water pH; surface water; artificial neural network; modelling



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## 1. Introduction

Appropriate water management is essential for implementing Sustainable Development Goals (SDGs). Management of surface water includes maintaining its desired quality. Often, surface water is the only source of water supply for cities and villages [1], where low water pH value may lead to serious problems in the water treatment system. Such water contains increased concentrations of pollutants due to sediment dissolution in the lake bottom, frequently needing additional resources (chemical reagents, additional purification equipment) for the water treatment stations [2] required to purify surface waters following the existing standards. The issue can be solved using natural limestone, a secondary product of limestone quarries [3]. The use of limestone to maintain water pH within the neutral values will allow the water treatment plant to operate without the use of additional resources [4,5]. In addition, the application of dispersed, thermally active limestone may help reduce the accumulation of secondary products of limestone quarries. Thus, land

use (use of secondary products of limestone quarries) has an impact on water use, with a cascading effect on reducing the use of additional resources for water purification.

Low pH of surface water can be a consequence of anthropogenic influence (eutrophication) and is one of the most common causes of water quality deterioration [6]. The problem of low pH value can be solved using CaO, Ca(OH)<sub>2</sub>, Na<sub>2</sub>CO<sub>3</sub>, and limestone CaCO<sub>3</sub> additives [7,8]. The use of limestone CaCO<sub>3</sub> has several advantages over CaO due to the fact that the addition of CaO increases water temperature since its interaction with water is an exothermic reaction. The use of Ca(OH)<sub>2</sub> also has a number of limitations associated with the preparation of its solution and the use of dispensers. Moreover, when added to water, its pH value can increase sharply and lead to chemical burns of flora and fauna. Na<sub>2</sub>CO<sub>3</sub> is also applied for the neutralization of natural waters [9]. The disadvantage of using this material is its high price compared to natural limestone. The use of CaCO<sub>3</sub> provides longer-lasting changes in water pH value compared with other neutralizing materials such as CaO, Ca(OH)<sub>2</sub>, Na<sub>2</sub>CO<sub>3</sub>.

When in water, limestone dissolves slowly and therefore, the increase in water pH does not cause chemical burns of flora and fauna [10]. In addition, carbonate can be an important source of CO<sub>2</sub> for photosynthesis in water, and thus crucial for the bioavailability decrease in toxic divalent cations of heavy metals (e.g., Ni, Cd, Cu, Pb, Zn) due to the formation of insoluble complexes with metals [11–14]. Furthermore, accurate liming practices often result in decreased mercury contents in fish [15].

Limestone is the best and most frequently used neutralizing agent for treating acidified surface waters worldwide, and the recommended limiting doses are indicative. The most commonly used doses of fine-grained limestone are in the range of 750–1000 kg/ha (for the typical depth of natural lakes at 1.0–1.2 m). The dose of limestone from the proposed range depends on water pH level and sediment type in the reservoir bottom. A smaller dose is applied when the bottom is sandy and a higher dose when the bottom is muddy [16].

The Swedish Liming Programme recommended another approach to determine the dose of limestone [17]. The dose depends on water pH and retention time; for lake water, the dose is 10–75 g/m<sup>3</sup>, and for watercourses—10–30 g/m<sup>3</sup>. Liming of lakes by adding doses of 10–30 g/m<sup>3</sup> should be repeated before the pH level exceeds 6. Mathematical models can be applied to calculate limestone dosages [18].

Liming of surface water is not always mandatory, but EPA recommendations should be followed if necessary. According to these recommendations, with a single application of limestone, the pH increase should not exceed 1.5 [19]. It is very important to consider that when limestone is added to water, the pH increase can also be high. Therefore, before adding limestone to surface water, all factors that increase the pH value [20] must be considered. These factors include initial pH values, total mineralization, redox potential, and limestone dose [21,22]. The pH value and total water salinity affect the solubility of minerals. When the initial pH value is lower, more minerals are dissolved, resulting in a pH increase. The same pattern is observed when the mineral is dissolved in water characterized by different mineralization. Analysis of these factors coupled with experimental studies helps determine the optimal dose of limestone needed to reduce water pH following EPA recommendations.

Trach et al. [23] studied the change in water pH in relation to the dose of natural limestone. With the increase in the initial water pH values (from 3 to 7) and mineralization (from 0.01 to 0.53 g/dm<sup>3</sup>), the required dose of limestone decreased. The results of experimental studies have shown that with lower initial water values (pH and mineralization), the required dose of limestone is also lower. To increase water pH to a neutral value (pH ≈ 7) and ensure the safety of this process, the limestone must be added to water several times at specific intervals. Such interval is necessary for adapting flora and fauna after the increase in pH and water salinity. In addition to water pH and salinity, the pH change with limestone also depends on water Eh [18]. The value of this water parameter depends mainly on the amount of organic matter.

Next, the existing experience of using ANN in ecological engineering was analyzed. The application of ANN for modelling time series of total phosphorus concentration was described for the Odra River conditions [24]. Using sensitivity analysis, a relationship was established between phosphorus concentrations and other water quality variables. Two models were created to predict phosphorus concentrations, with one input variable and fourteen input variables. Both ANN models showed good predictive ability with new datasets.

Furthermore, the ANN model was applied to forecast the concentration of total nitrogen and total phosphorus in the lakes from the United States [25]. ANN models were trained, tested, and validated using three inputs (pH, conductivity and turbidity) that were statistically correlated with the output data. The study results showed that ANN modelling is a good tool for assessing nutrient concentrations in lakes.

Van et al. [26] predicted the concentration of dissolved oxygen in a river setting based on hydrological parameters (temperature, pH, turbidity, conductivity, chemical oxygen demand, biological oxygen demand, nitrate and phosphate). Multivariate regression (MLR) and backpropagation neural network (BPNN) methods were used to establish these relationships. The results of their study showed that the BPNN might accurately predict the concentration of dissolved oxygen in the water.

Stamenković et al. [27] developed an ANN model to predict the concentration of nitrates in river water. Their model was trained and tested on 26 input water quality parameters. The results obtained showed the ability of ANN models to predict the concentration of nitrates with an average absolute error of 0.53 and 0.42 mg/dm<sup>3</sup> for the test data.

Krtolica et al. [28] constructed a multilayer feed-forward ANN model using macrophytes as independent variables for each water quality variable (dissolved oxygen, nitrate-nitrogen, and orthophosphates). They selected 28 macrophytes as key water quality indicators by sensitivity analysis for one environmental variable. The developed ANN architecture presented a modelling approach that can be applied in different biological systems.

Gebler et al. [29] performed ANN modelling of macrophyte indicators depending on the physicochemical water parameters. They analyzed several indices of macrophyte diversity (species richness-N, Shannon's index-H, Simpson's index-D and Pielou's index-J), as well as the ecological status index (macrophyte index for rivers-MIR). Alkalinity, conductivity, pH, nitrate and ammonium nitrogen, reactive and total phosphorus, and biochemical oxygen demand were used as the input variables. The quality of the constructed models was assessed using the calculated errors and the Pearson correlation coefficient.

Chen et al. [30] developed three models for estimating the concentration of dissolved oxygen: backpropagation neural network (BPNN), adaptive neural fuzzy inference system (ANFIS), and multilinear regression model (MLR). The input variables of the neural network were water temperature, pH, conductivity, turbidity, suspended solids, total hardness, total alkalinity, and ammonia nitrogen. The performance of the models was assessed using the mean absolute error, root-mean-square error, and correlation coefficient. Comparative analysis of the models showed that ANN had the best performance characteristics.

In the study performed by Wen et al. [31], the authors used a three-level backpropagation ANN with a Bayesian regularization learning algorithm. The input variables of the neural network were pH, conductivity, chloride (Cl<sup>-</sup>), calcium (Ca<sup>2+</sup>), total alkalinity, total hardness, nitrate-nitrogen (NO<sub>3</sub><sup>-</sup>-N), and ammonia nitrogen (NH<sub>4</sub><sup>-</sup>-N). The values of obtained correlation coefficient and root-mean-square error showed the effectiveness of the ANN model.

Singh et al. [32] described the training, validation, and application of ANN models to calculate dissolved oxygen levels and biochemical oxygen demand. Their models used eleven input water quality variables measured in river water. The efficiency of ANN models was assessed using the coefficient of determination ( $R^2$ ) and root-mean-square error, calculated based on values of the output variables measured and calculated in the

models. The authors summarized that ANN models could be used to calculate water quality parameters.

Csábrági et al. [33] aimed at predicting the concentration of dissolved oxygen based on easily measurable parameters of water quality (pH, temperature, conductivity) and runoff. Four linear and nonlinear models were adapted, i.e., multivariate linear regression model, multilayer neural network (perceptron), radial base function neural network, and general regression neural network model. The performance of the models was evaluated using various statistical indicators. Nonlinear models gave better results than linear ones. To determine the parameter that had the most significant impact on the performance of the models, the authors conducted a sensitivity analysis, which showed that for all three neural network models, pH played the most important role in estimating the content of dissolved oxygen.

As indicated above, ANN has been successfully used for modelling various physicochemical processes in different site conditions. This is because most environmental processes are nonlinear due to their complex chemical and physical nature. Nevertheless, the task of assessing the permissible pH level when adding limestone to water has not been fully accomplished.

This article is an interdisciplinary research paper aiming at:

- (1) Determination and comparison of the doses of new dispersed, thermally activated limestone and natural limestone,
- (2) Finding the relation of dose value to initial water parameters (pH, Eh and total mineralization values).
- (3) Creation of a model using artificial neural networks (ANN) to predict changes in water pH values in line with EPA recommendations.

The doses of dispersed, thermally activated limestone were determined, taking into account EPA recommendations, the use of which has an advantage over natural limestone. The limestone application helped to reduce the accumulation of secondary products of limestone quarries.

In addition, the systematization of various scientific directions in one study is the scientific novelty of this research.

## 2. Materials and Methods

### 2.1. Materials

The material used in the study was dispersed, thermally activated limestone mined in a quarry in Ternopil, Ukraine (Figure 1). Preparation of such limestone in the Ternopil Quarry enterprise is carried out according to a specific algorithm (Figure 2). Directly at Ternopil Quarry is an accumulation of fine fractions (0–40 mm) of carbonate rocks, which do not satisfy the requirements for building mortar fillers and are unsuitable as building stone. Therefore, the enterprise produces dispersed limestone from unused limestone screenings. Such limestone is used as a Ca-rich mineral fertilizer.

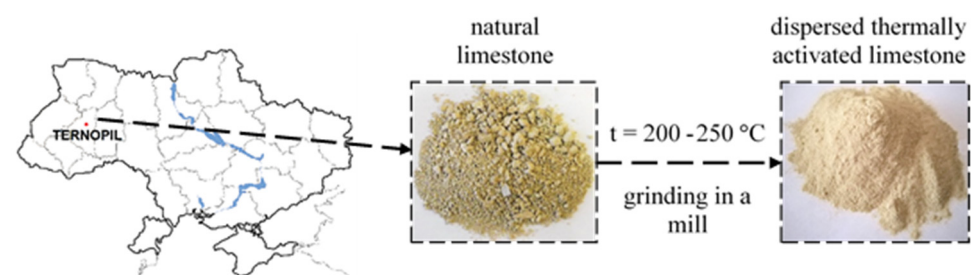
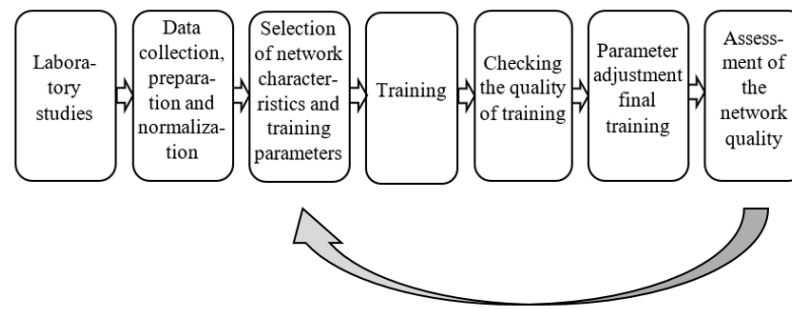


Figure 1. The object of the laboratory studies.



**Figure 2.** Research algorithm.

The raw materials for the production of dispersed limestone are primary and secondary crushing screenings with sizes up to 40 mm. Wet limestone is loaded into a shaft dryer. The gas temperature at the dryer entrance is 200–250 °C, and at the dryer exit-100–120 °C. A pipe is used for drying and separating the material that is 1 mm in size. Particles larger than 1 mm settle in the pipe and are sent for grinding in a hammer mill, after which they return to the drying pipe. This technological process produces dispersed, thermally activated limestone with a grain size below 0.001 mm.

## 2.2. Methods

As input values of the model, the following parameters were used: pH, redox potential (Eh), total dissolved solids (TDS), and various doses of dispersed, thermally activated limestone. For experimental studies, the applied doses of limestone were at 0.025, 0.05, and 0.1 g/dm<sup>3</sup>. Initial water pH values were 4.0, 5.0, and 6.0, Eh was 120, 280, and 420 mV, TDS was 0.02, 0.240, and 0.56 g/dm<sup>3</sup>. Experimental studies were performed at a temperature of 15 ± 0.5 °C.

Experimental studies were carried out under static conditions. A magnetic stirrer (model JNE FK-25W) was used to mix water and the analyzed limestone at a mixing speed of 0–2400 rpm and a maximum mixing volume of 3000 mL.

The changes in pH and TDS were determined using a multimeter (Milwaukee MW802, Rocky Mount, NC, USA). Eh was measured using an ORP meter (Milwaukee MW500 PRO, Rocky Mount, NC, USA). The MW802 has TDS measuring ranges up to 4000 ppm (4 g/dm<sup>3</sup>). The pH measuring ranged from 0.00 to 14.00. Graduation was set as 0.10 pH, 10 ppm (0.01 g/dm<sup>3</sup>). Accuracy (25 °C): ±0.20 pH, TDS: ±0.02% of full scale. MW500 PRO measures Eh in a ±1000 mV range. The accuracy is at (25 °C): ±5 mV. Total dissolved solids (TDS) represent a measure of the dissolved combined content of all inorganic and organic substances present in a liquid in molecular, ionized, or microgranular (colloidal sol) suspended form (g/dm<sup>3</sup>). TDS concentrations are often reported in parts per million (ppm) or mg/L. Water TDS concentrations were determined using a digital meter. Before starting a series of measurements, the multimeter was calibrated using three buffer solutions (pH values at 4.00, 6.86 and 9.01), according to the recommendations for their use to ensure measurement accuracy. To adjust the required initial pH values, Eh and TDS, 0.1 M HCl, 0.1 M NaOH, and a potassium humate solution were applied. For experimental studies of the dependence of limestone dissolution on water Eh, a 10% solution of potassium humate was prepared. Chemical characteristics of potassium humate were as follows: water solubility (dry basis) —100%, humic acid (dry basis)—60%, fulvic acid (dry basis)—10.0–15.0 %, and potassium (K<sub>2</sub>O dry basis)—10.0%–12.0%, moisture—15.0%.

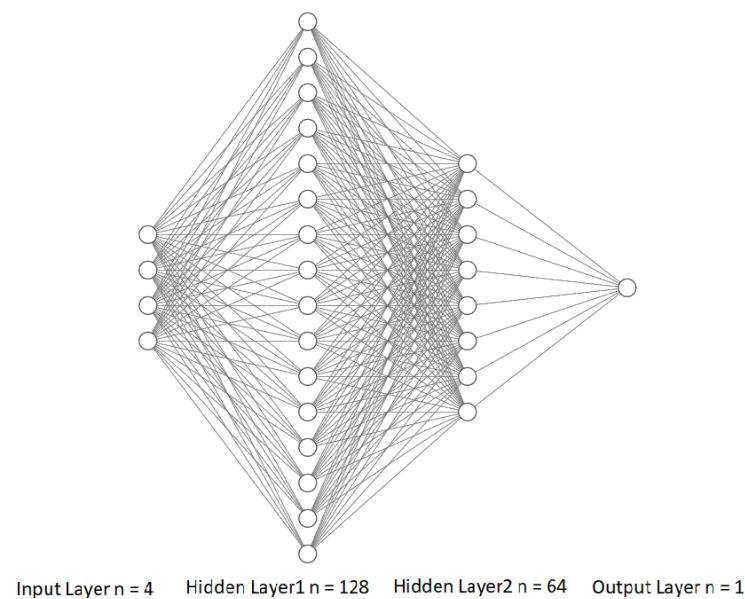
## 2.3. ANN Modelling

ANN can be thought of as a system of interacting artificial neurons. Connected in a large network with controlled interactions, neurons can solve fairly complex problems. The advantage of neural networks over mathematical methods is that they can look for patterns in fuzzy data, learn and systematize solutions. A neural network can generalize

and highlight hidden dependencies between input and output data. Once trained, the network can predict future values based on previous values and various existing factors.

Within the framework of this study, data were collected, analyzed, and prepared, four ANN models were created, and a comparative analysis of their effectiveness was carried out. Normalization of input variables and ANN modelling, training, and testing were performed using the Keras library in the Python programming language. The results were visualized using the Matplotlib library in the Python programming language.

The ANN architecture proposed in this study has one input layer, two hidden layers and one output layer of neurons. The neural network architecture is a specific neuron arrangement and connection. The scheme of ANN architecture used in this study is shown in Figure 3. The input layer consists of four variables, the first hidden layer has 128 neurons, the second hidden layer has 64 neurons, and the output layer consists of one neuron.



**Figure 3.** The architecture of ANN used in this research.

The number of hidden layers and neurons was determined empirically. Initially, one hidden layer of neurons was used ( $n_1 = 128$ ); in this case, the model showed a large error (MAPE = 48.3%). It was decided to increase the number of hidden layers and neurons ( $n_1 = 128$ ,  $n_2 = 64$ ). This led to an improvement in the ANN model performance with a slight increase in the calculation time.

In the first stage of ANN use, input layer neurons get an input signal and send an output signal to the hidden layer neurons. The neurons in the hidden layer calculate activation and send a signal to the neurons in the output layer. For an ANN, the total input signal is calculated by the formula [34]:

$$net = \sum_{i=1}^n x_i w_i \quad (1)$$

where:  $x_i$  are input variables,  $w_i$  are weights.

The neurons of the output layer calculate activation and send the input signal. The activation function is used to calculate the neuron activation. In this study, ANN models were tested using two activation functions: the sigmoid (logistic) function and the ReLU (Rectified Linear Unit).

The sigmoid function is a smooth, monotonically increasing nonlinear function, which is defined as [34]:

$$f(x) = \frac{1}{1 + e^{(-x)}} \quad (2)$$

This function can be used in neural networks with many layers, as well as train these networks using the backpropagation method. The sigmoid function has the advantage of normalizing the output to the range (0, 1). This is useful when the resulting layer value must represent the probability of a random variable.

Recently, the ReLU activation function has gained wide popularity. If the argument takes a negative value when calculating this function, then the function is equal to 0; in the case of a positive value of the argument, the function returns the number itself:

$$f(x) = \begin{cases} 0 & \text{for } x < 0 \\ x & \text{for } x \geq 0. \end{cases} \quad (3)$$

The advantage of ReLU is activation sparsity. In networks with a large number of neurons, the use of the sigmoid function entails the activation of almost all neurons, which can affect the model performance. When using ReLU, the number of included neurons becomes smaller due to the function characteristics, and the network itself becomes more productive.

The next stage of ANN operation is the calculation of the loss (error) function. The loss function measures “how good” the neural network is for a given training set and expected responses. The loss function is responsible for assessing how well the model predicts the real value, and building the model comes down to solving the problem of minimizing the value of this function at each stage. The calculation of the loss (error) includes comparing the actual and target values for each neuron. The ANN learns until it achieves the global minimum error between the actual and target data.

In this study, the Mean Square Error (MSE) was used to calculate the loss function:

$$MSE = \frac{1}{n} \sum_{i=1}^n (Y'_i - Y_i)^2 \quad (4)$$

where:

$Y'_i$  is the output calculated by the model,  $Y_i$  is the target output.

The squared deviation is calculated for each dataset, after which the resulting values are summed up and divided by the total number of datasets. The closer the obtained value is to zero, the more accurate the model is. This calculation method is highly sensitive to outliers in the sample or samples where the range of values is large.

For the appearance of overfitting, the early stop method was used [35]. Training the network stops when the monitored metric no longer shows improvement. The monitored metric was validation loss. A training loop checks whether the loss no longer decreases at the end of every epoch. In such a case, the training stops.

Various methods can be used to minimize the loss function. In this study, two gradient descent methods were tested to optimize ANN training: Stochastic Gradient Descent (SGD) and Adaptive Moment Estimation (Adam).

The SGD algorithm updates the neural network weights using a single training sample at each step [36]. SGD does not perform unnecessary calculations since the loss function is calculated not for the entire training set but only for one example. This contributes to the algorithm learning much faster. However, because at each step of the algorithm, the gradient is calculated based on different sets of initial data, updates of the weight coefficients are accompanied by frequent fluctuations of the objective function. Thus, on the one hand, SGD allows to move to potentially better local minima quickly, but on the other hand, large fluctuations significantly slow down the convergence.

According to Kingma and Ba [37], Adam’s method is computationally efficient, requires little memory, is invariant to diagonal scaling of gradients, and is well suited for large data and parameter problems. The rule for updating the weights in Adam’s method is determined by using estimates of two different moments; the first uses the previously calculated values of partial derivatives, and the second uses their squares.

In this study, to assess the performance of the ANN models presented, the quality metric Mean Absolute Error (*MAE*) was used. *MAE* is the average sum of the absolute values of the difference between the real and predicted values [38]:

$$MAE = \frac{1}{n} \sum_{i=1}^n |Y'_i - Y_i| \quad (5)$$

The metric functions are similar to loss functions, except that the results of the use of the metric function are not applied in the model training.

*MAE* is similar to *MSE* in many ways but is less sensitive to outliers. The main reason is that in *MSE* when squaring errors, the outliers (which usually have higher errors than other samples) dominate the final error and affect model parameters.

One of the primary tasks of this study was the choice of the optimal ANN model. For comparative analysis of the accuracy of four neural networks, the Mean Absolute Percentage Error (*MAPE*) and the Coefficient of determination ( $R^2$ ) were used.

*MAPE* is a ratio defined by Formula (6) [39]:

$$MAPE = \frac{100\%}{n} \sum_{i=1}^n \left| \frac{Y_i - Y'_i}{Y_i} \right| \quad (6)$$

$R^2$  is a statistical measure used in various models, for example, for predicting future outcomes or testing hypotheses based on other related information.  $R^2$  provides a measure of how well the observed results are reproduced by the model, based on the proportion of the total variation in the results explained by the model. The values of the coefficient of determination belong to the interval [0, 1].  $R^2$  can be calculated by the formula [40]:

$$R^2 = 1 - \frac{\sum (Y_i - Y'_i)^2}{\sum (Y_i - \bar{Y}_i)^2} \quad (7)$$

where:

$\bar{Y}_i$ —is the mean of the target output data.

After ANN training, the process proceeds to the validation stage.

The aim of this stage is to guarantee that the ANN can settle resumptive data on the training stage [41]. The next phase is the testing stage that involves checking the network's performance on data that was never seen during the previous steps [42].

### 3. Results and Discussion

The results of the experimental studies carried out using the described methodology were assembled in tables. A fragment of the research results is presented in Table 1.

**Table 1.** Results of experimental studies of water pH increase using dispersed, thermally activated limestone.

pH	$\Delta$ pH	Eh, mV	$\Delta$ Eh	TDS, g/dm <sup>3</sup>	$\Delta$ TDS	CaCO <sub>3</sub> , g/dm <sup>3</sup>	t, Equilibrium Time, Min
4.10	3.95	120	−105	0.02	70	0.1	80
4.10	3.78	120	−90	0.02	40	0.05	80
4.10	2.36	120	−72	0.02	20	0.025	50
5.01	3.78	120	−92	0.02	50	0.1	70
5.01	3.52	120	−79	0.02	40	0.05	70
5.01	2.3	120	−60	0.02	20	0.025	40
6.02	2.88	120	−71	0.02	40	0.1	50
6.02	1.97	120	−62	0.02	20	0.05	40
6.02	1.78	120	−50	0.02	10	0.025	30

The obtained results show the possibility of using dispersed, thermally activated limestone to increase the pH of surface waters. The increase in water pH was influenced by initial water pH, TDS, and the dose of limestone. The prepared model solutions



with different water Eh values did not significantly affect the increase in water pH when limestone was added at various doses to water with different pH and TDS values. Thus, analysis of the obtained results made it possible to establish the recommended doses of the studied limestone, presented in Table 2.

**Table 2.** Recommended doses of dispersed, thermally activated limestone in surface water.

pH	Water with TDS 0.02 g/dm <sup>3</sup>	Water with TDS 0.24 g/dm <sup>3</sup>	Water with TDS 0.56 g/dm <sup>3</sup>
	Doses of CaCO <sub>3</sub> , g/m <sup>3</sup>	Doses of CaCO <sub>3</sub> , g/m <sup>3</sup>	Doses of CaCO <sub>3</sub> , g/m <sup>3</sup>
4	25 (three times a year)	25 (twice a year)	25 (twice a year)
5	25 (twice a year)	50 (once a year)	25 (once a year)
6	50 (once a year)	50 (once a year)	25 (once a year)

It is important to remember that each water body is unique and individual in terms of its physical and chemical composition. In addition, the terrain, climate, and anthropogenic factors affect the physicochemical composition of the reservoir. Therefore, when deciding on the implementation of reservoir liming, in addition to the proposed doses of dispersed, thermally activated limestone (Table 2), the results of monitoring a specific water body should also be considered. By combining monitoring data and results of experimental studies, it is possible to make a determination regarding the required dose of the limestone.

After laboratory studies, a matrix was formed from the obtained data, which contained 336 rows and five columns (four columns (pH, Eh, TDS, doses CaCO<sub>3</sub>) input variables, one column ( $\Delta$ pH) target output data). The ANNs model effectiveness is depended on the data and their preparation. After the stage of data collection, they were normalized. Normalization is a procedure of data processing when the values are brought to a certain specified range [43]. To normalize the data, the arithmetic mean and standard deviation were calculated for the input data. Then the arithmetic mean was subtracted from the input data, and the result was divided by the standard deviation. The data matrix was split into three datasets: 60% was the training set, 20% was the validation set, and 20% was the testing set. The data were trained, validated, and tested using four ANN models.

Table 3 shows the characteristics of the ANN models and a comparison of their performance. Model performance testing was carried out for 20, 50, and 100 epochs. The number of epochs indicates how many times the model was exposed to training. Thus, an epoch is one pass forward or backwards for all learning examples. An increase in the number of epochs over 100 did not lead to any improvement in the model performance.

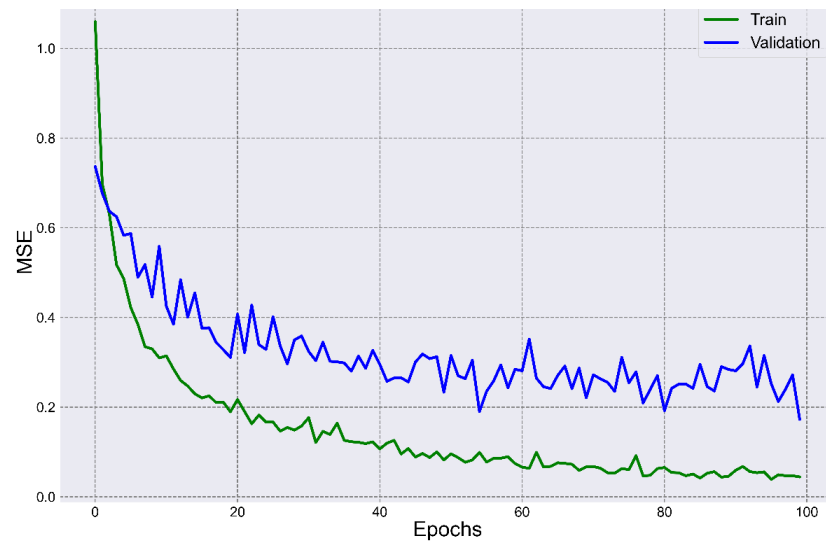
**Table 3.** Comparison of the ANN models performance.

Parameters	Models											
	ANN 1			ANN 2			ANN 3			ANN 4		
activator	Sigmoid			Sigmoid			ReLU			ReLU		
optimizer	Adam			SGD			Adam			SGD		
epochs	20	50	100	20	50	100	20	50	100	20	50	100
R <sup>2</sup>	0.137	0.224	0.351	0.184	0.184	0.132	0.555	0.771	0.847	0.667	0.794	0.747
MAPE, %	45.6	41.9	46.1	46.1	46.1	43.9	28.6	18.9	14.1	26.6	17.9	17.8

Analysis of MAPE and R<sup>2</sup> indicators showed that the ANN 3 network achieved the best performance results (lowest Mean Absolute Percentage Error and highest Coefficient of determination). This network used the ReLU activation function for neurons hidden layers and the Adam loss function optimizer. The best performance was achieved for 100 epochs: MAPE = 14.1%; R<sup>2</sup> = 0.847.

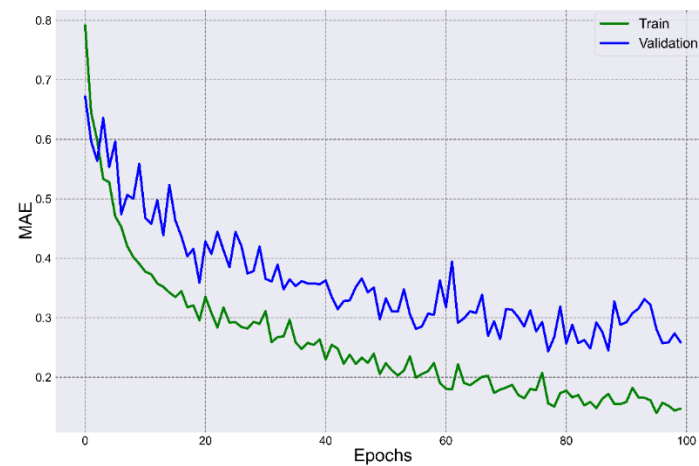
The research results are presented for the ANN 3 network and allow assessment of its adequacy and possible use in forecasting new datasets. Figure 4 shows the comparison of the loss function results for the training and validation datasets. The maximum MSE in

the first epoch was: training set—1.0597; validation set—0.7914. The minimum *MSE* in the 100th epoch were training set—0.0443 and validation set—0.1725.



**Figure 4.** Comparison of loss function results (*MSE*).

Figure 5 shows a comparison of the results of calculating the quality metric for the training and validation datasets. The maximum *MAE* value in the first epoch was training set—0.7914 and validation set—0.6721. The minimum *MAE* for the training set (0.1399) was reached in epoch 96 and the validation set (0.2439) in epoch 78.



**Figure 5.** Comparison of the results of calculating the quality metric (*MAE*).

The histogram of errors (Figure 6) shows that most of the errors fall within the range from 0 to 0.25, although there are some anomalous errors.

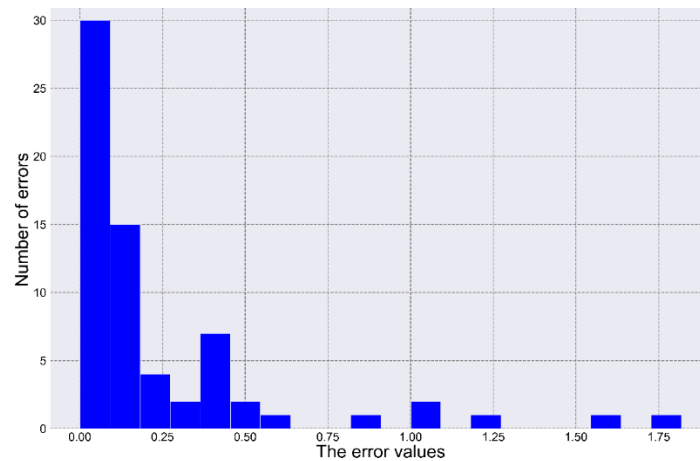


Figure 6. Histogram of errors.

Figure 7 shows the scatter of the predicted values relative to the target values.

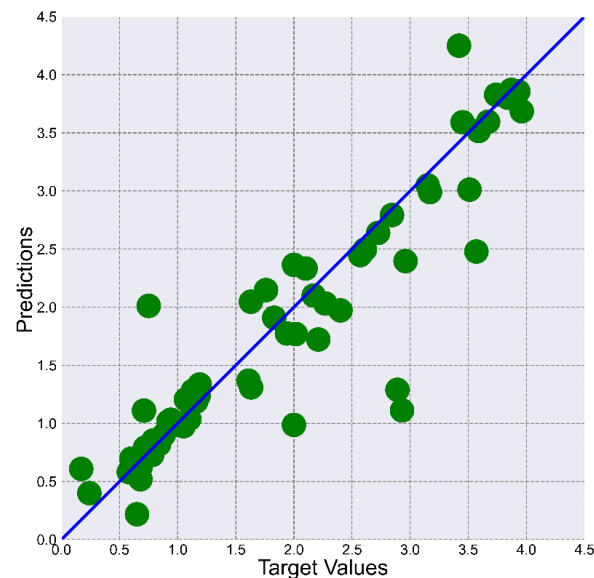


Figure 7. Scatter of predicted and target data values.

In general, the prediction values fit reasonably well against the target values, except for a few outliers. Most likely, there are significant anomalies in the output for these values that require data cleanup. The results of the  $R^2$  calculation also serve as indirect confirmation of the anomalies in the data. For example, there are input datasets whose  $R^2$  values are  $-4.73$  and  $-1.60$ .

One of the main methods to analyze the ANN model is the comparison of various indicators (mean absolute percentage error,  $MAPE$ ; root mean square error,  $RMSE$ ; mean square error and  $MSE$ ), and the coefficients of determination ( $R^2$ ) and regression ( $R$ ). Tijanić, Car-Pušić and Špera [44] argued that  $R^2$  and  $MAPE$  are the most often applied estimators of the model accuracy. The authors set the goal of their study to predict the cost of road construction. The values of  $R^2$  and  $MAPE$  indicators were  $0.959$  and  $13\%$ , respectively. Mounter et al. [45] investigated the capabilities of ANN to improve the accuracy of long-term predictions of building energy consumption. The  $MAPE$  indicator was  $12.03\%$ . Kulisz et al. [46] aimed to study ANN ability to model the water quality index in groundwater. They used the coefficient of determination ( $R^2 = 0.998$ ) as a measure of ANN accuracy. Szul, Nęcka & Mathia [47] used five different ANN to predict energy consumption. To assess the quality of the developed models, the  $MAPE$  and  $R^2$  indices were used, the values of

which were 13.7% and 0.8, respectively. Ding et al. [48] used the ANN model to predict the power output of a PV system. Their results showed 10.06% and 18.9% *MAPE* forecast errors during sunny and rainy days, respectively.

Thus, comparing the *MAPE* and  $R^2$  indicators of the created ANN model ( $MAPE = 14.1\%$ ;  $R^2 = 0.847$ ) with other cases, it can be stated that, in general, the degree of agreement between the measured and modelled data is satisfactory.

The low level of the ANN model results (*MAPE* and  $R^2$ ) were associated with two causes. Firstly, the peculiar ability of limestone to increase pH values faster at lower doses. When limestone is added to water, such water becomes a complex dynamic system.

The water pH value is the ratio of  $H^+$  to  $OH^-$ . The peculiarity of such a system is as follows:



The pH value increases due to  $CaCO_3$  dissolution. The peculiarity of the  $CaCO_3$  property is that its dissolution occurs up to a certain point. Then, a reverse process begins, which is the interaction of  $Ca^{2+}$  with  $CO_3^{2-}$  and the formation of  $CaCO_3$ . Secondly, the measurement of the values of water quality parameters (pH, Eh, TDS) was influenced by the permissible metrological errors of the measuring instruments.

Based on the data obtained, the authors believe that the created ANN can be used to achieve the study task and could later be applied to predict changes in the water pH value when adding various doses of limestone.

#### 4. Conclusions

The presented approach allowed us to determine and compare the doses of new dispersed, thermally activated limestone and natural limestone. It was revealed that the dose of dispersed, thermally activated limestone was lower than the dose of natural limestone. It was also found that the increase in water pH was influenced by the initial water pH, TDS, and the dose of limestone. Therefore, recommended doses of dispersed, thermally activated limestone in surface water were established. Moreover, four ANN models were tested that varied in the use of different activation functions and the loss function optimizers. For a comparative analysis of ANN accuracy, the Mean Absolute Percentage Error (*MAPE*) and Coefficient of determination ( $R^2$ ) were used. The best performance results had the network with the ReLU activation function for neurons hidden layers and the Adam loss function optimizer ( $MAPE = 14.1\%$ ;  $R^2 = 0.847$ ). The study had limitations that could influence the results of the ANN model. The first was the peculiarity of limestone, which can increase pH values faster at lower doses. The second showed that the measurement of the values of water quality parameters (pH, Eh, TDS) was influenced by the permissible metrological errors of the measuring instruments.

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#### References

1. Wichowski, P.; Rutkowska, G.; Kamiński, N.; Trach, Y. Analysis of Water Consumption in the Campus of Warsaw University of Life Sciences—SGGW in Years 2012–2016. *J. Ecol. Eng.* **2019**, *20*, 193–202. [[CrossRef](#)]
2. Michel, M.; Tytkowska, M.; Reczek, L.; Trach, Y.; Siwiec, T. Technological Conditions for the Coagulation of Wastewater from Cosmetic Industry. *J. Ecol. Eng.* **2019**, *20*, 78–85. [[CrossRef](#)]

3. Lirski, A.; Wolnicki, J. Rola wapnowania w stawowej i jeziornej produkcji rybackiej oraz ochronie wód. *Nawozy Nawożenie* **2006**, *8*, 130–139.
4. Appelberg, M. Restructuring of Fish Assemblages in Swedish Lakes Following Amelioration of Acid Stress through Liming. *Restor. Ecol.* **1998**, *6*, 343–352. [[CrossRef](#)]
5. Hutorowicz, A.; Kapusta, A.; Krzywosz, T.; Hutorowicz, J. The Ichthyofauna of the Dystrophic Lake Smolak (Northern Poland) in Light of Selected Physical and Chemical Water Conditions Thirty Years after the Conclusion of Liming and Fertilization. *Fish. Aquat. Life* **2005**, *13*, 207–225.
6. Bengtsson, B.; Norling, I.; Pierrou, U. National Survey Data on Sportfishing, Economics, and Environmental Changes in Sweden. *Am. Fish. Soc. Symp.* **1991**, *12*, 74–83.
7. Driscoll, C.T.; Fordham, G.F.; Ayling, W.A.; Oliver, L.M. The chemical response of acidic lakes to calcium carbonate treatment. *Lake Reserv. Manag.* **1987**, *3*, 404–411. [[CrossRef](#)]
8. Leoni, B.; Morabito, G.; Rogora, M.; Pollastro, D.; Mosello, R.; Arisci, S.; Forasacco, E.; Garibaldi, L. Response of Planktonic Communities to Calcium Hydroxide Addition in a Hardwater Eutrophic Lake: Results from a Mesocosm Experiment. *Limnology* **2007**, *8*, 121–130. [[CrossRef](#)]
9. Lessmark, O. *Kalkningar i Kronobergs Län 1977–1989. En Utvärdering av Sjöalkningars Kemiska och Biologiska Effekt; Miljöfakta i Kronobergs Län: Kronoberg, Sweden, 1990; Volume 1, p. 70.*
10. Angeler, D.G.; Drakare, S.; Johnson, R.K.; Köhler, S.; Vrede, T. Managing Ecosystems without Prior Knowledge: Pathological Outcomes of Lake Liming. *Ecol. Soc.* **2017**, *22*, art44. [[CrossRef](#)]
11. Andersson, M.P.; Sakuma, H.; Stipp, S.L.S. Strontium, Nickel, Cadmium, and Lead Substitution into Calcite, Studied by Density Functional Theory. *Langmuir* **2014**, *30*, 6129–6133. [[CrossRef](#)]
12. Di Lorenzo, F.; Cametti, G.; Vanhecke, D.; Churakov, S.V. The Role of Interfaces in Controlling Pb<sup>2+</sup> Removal by Calcium Carbonate Minerals. *Cryst. Growth Des.* **2020**, *20*, 6157–6169. [[CrossRef](#)]
13. Hoffmann, U.; Stipp, S.L.S. The Behavior of Ni<sup>2+</sup> on Calcite Surfaces. *Geochim. Cosmochim. Acta* **2001**, *65*, 4131–4139. [[CrossRef](#)]
14. Dickson, W. Liming Toxicity of Aluminium to Fish. *Vatten* **1983**, *39*(4), 400–404.
15. Dickson, W.; Brodin, Y. Strategies and methods for freshwater liming. In *Liming of Acidified Surface Waters*; Springer: Berlin/Heidelberg, Germany, 1995; pp. 81–124.
16. Adámek, Z.; Pardo, M.A.; Vilizzi, L.; Roberts, J. Successful Reproduction of Common Carp *Cyprinus Carpio* in Irrigation Waterways. *Fish. Manag. Ecol.* **2015**, *22*, 279–285. [[CrossRef](#)]
17. Svenson, T.; Dickson, W.; Hellberg, J.; Moberg, G.; Munthe, N. The Swedish Liming Programme. *Water Air Soil Pollut.* **1995**, *85*, 1003–1008. [[CrossRef](#)]
18. Henrikson, L.; Brodin, Y.W. (Eds.) *Liming of Acidified Surface Waters*; Springer: Berlin/Heidelberg, Germany, 1995; ISBN 978-3-642-79311-0.
19. Bernes, C. *Acidification and Liming of Swedish Freshwaters*; Swedish Environmental Protection Agency: Solna, Sweden, 1991; ISBN 978-91-620-1109-3.
20. Müller, B.; Meyer, J.S.; Gächter, R. Alkalinity Regulation in Calcium Carbonate-Buffered Lakes: Alkalinity Regulation in Calcium Carbonate-Buffered Lakes. *Limnol. Oceanogr.* **2016**, *61*, 341–352. [[CrossRef](#)]
21. Stumm, W.; Morgan, J.J. *Aquatic Chemistry: Chemical Equilibria and Rates in Natural Waters*, 3rd ed.; Environmental science and technology; Wiley: New York, NY, USA, 1996; ISBN 978-0-471-51184-7.
22. Håkanson, L. A General Management Model to Optimize Lake Liming Operations: Management Model for Lake Liming. *Lakes Reserv. Res. Manag.* **2003**, *8*, 105–140. [[CrossRef](#)]
23. Trach, Y.; Melnychuk, V.; Melnychuk, G.; Mazur, Ł.; Podlasek, A.; Vaverková, M.D.; Koda, E. Using Local Mineral Materials for the Rehabilitation of the Ustya River—A Case Study. *Desalination Water Treat.* **2021**, *232*, 346–356. [[CrossRef](#)]
24. Możejko, J.; Gniot, R. Application of Neural Networks for the Prediction of Total Phosphorus Concentrations in Surface Waters. *Pol. J. Environ. Stud.* **2008**, *17*, 363–368.
25. Sinshaw, T.A.; Surbeck, C.Q.; Yasarer, H.; Najjar, Y. Artificial Neural Network for Prediction of Total Nitrogen and Phosphorus in US Lakes. *J. Environ. Eng.* **2019**, *145*, 04019032. [[CrossRef](#)]
26. Van, T.P.; Dao, M.P.; Han, P.N.N. Modelling of Dissolved Oxygen in Thi Vai River Water Incorporating Artificial Neural Network and Multivariable Regression. *Org. Med. Chem. Int. J.* **2017**, *4*, 74–81.
27. Stamenković, L.J.; Mrazovac Kurilić, S.; Presburger Ulniković, V. Prediction of Nitrate Concentration in Danube River Water by Using Artificial Neural Networks. *Water Supply* **2020**, *20*, 2119–2132. [[CrossRef](#)]
28. Krtolica, I.; Cvijanović, D.; Obradović, Đ.; Novković, M.; Milošević, D.; Savić, D.; Vojinović-Miloradov, M.; Radulović, S. Water Quality and Macrophytes in the Danube River: Artificial Neural Network Modelling. *Ecol. Indic.* **2021**, *121*, 107076. [[CrossRef](#)]
29. Gebler, D.; Kayzer, D.; Szoszkiewicz, K.; Budka, A. Artificial Neural Network Modelling of Macrophyte Indices Based on Physico-Chemical Characteristics of Water. *Hydrobiologia* **2014**, *737*, 215–224. [[CrossRef](#)]
30. Chen, W.-B.; Liu, W.-C. Artificial Neural Network Modeling of Dissolved Oxygen in Reservoir. *Environ. Monit. Assess.* **2014**, *186*, 1203–1217. [[CrossRef](#)] [[PubMed](#)]
31. Wen, X.; Fang, J.; Diao, M.; Zhang, C. Artificial Neural Network Modeling of Dissolved Oxygen in the Heihe River, Northwestern China. *Environ. Monit. Assess.* **2013**, *185*, 4361–4371. [[CrossRef](#)] [[PubMed](#)]

32. Singh, K.P.; Basant, A.; Malik, A.; Jain, G. Artificial Neural Network Modeling of the River Water Quality—A Case Study. *Ecol. Model.* **2009**, *220*, 888–895. [[CrossRef](#)]
33. Csábrági, A.; Molnár, S.; Tanos, P.; Kovács, J. Application of Artificial Neural Networks to the Forecasting of Dissolved Oxygen Content in the Hungarian Section of the River Danube. *Ecol. Eng.* **2017**, *100*, 63–72. [[CrossRef](#)]
34. Juszczyk, M.; Zima, K.; Lelek, W. Forecasting of sports fields construction costs aided by ensembles of neural networks. *J. Civ. Eng. Manag.* **2019**, *25*, 715–729. [[CrossRef](#)]
35. Trach, R.; Trach, Y.; Lendo-Siwicka, M. Using ANN to Predict the Impact of Communication Factors on the Rework Cost in Construction Projects. *Energies* **2021**, *14*, 4376. [[CrossRef](#)]
36. Ruder, S. An Overview of Gradient Descent Optimization Algorithms. *arXiv* **2017**, arXiv:1609.04747.
37. Kingma, D.P.; Ba, J. Adam. A Method for Stochastic Optimization. *arXiv* **2017**, arXiv:1412.6980.
38. Willmott, C.J.; Matsuura, K. Advantages of the Mean Absolute Error (MAE) over the Root Mean Square Error (RMSE) in Assessing Average Model Performance. *Clim. Res.* **2005**, *30*, 79–82. [[CrossRef](#)]
39. De Myttenaere, A.; Golden, B.; Grand, B.L.; Rossi, F. Mean Absolute Percentage Error for Regression Models. *Neurocomputing* **2016**, *192*, 38–48. [[CrossRef](#)]
40. Hughes, A.J.; Grawoig, D. *Statistics: A Foundation for Analysis*; Addison-Wesley: Reading, UK, 1971; ISBN 978-0-201-03021-1.
41. Zhang, G.; Hu, M.Y.; Eddy Patuwo, B.; Indro, D.C. Artificial Neural Networks in Bankruptcy Prediction: General Framework and Cross-Validation Analysis. *Eur. J. Oper. Res.* **1999**, *116*, 16–32. [[CrossRef](#)]
42. Nicał, A.; Anysz, H. The Quality Management in Precast Concrete Production and Delivery Processes Supported by Association Analysis. *Int. J. Environ. Sci. Technol.* **2020**, *17*, 577–590. [[CrossRef](#)]
43. Koda, E.; Osiński, P.; Kołanka, T. Flow numerical modeling for efficiency assessment of vertical barriers in landfills. In *Coupled Phenomena in Environmental Geotechnics*; Manassero, M., Dominijanni, A., Foti, S., Musso, G., Eds.; CRC Press/Balkema: London, UK, 2013; pp. 693–698.
44. Tijanić, K.; Car-Pušić, D.; Šperac, M. Cost Estimation in Road Construction Using Artificial Neural Network. *Neural Comput. Applic.* **2020**, *32*, 9343–9355. [[CrossRef](#)]
45. Mounter, W.; Ogwumike, C.; Dawood, H.; Dawood, N. Machine Learning and Data Segmentation for Building Energy Use Prediction—A Comparative Study. *Energies* **2021**, *14*, 5947. [[CrossRef](#)]
46. Kulisz, M.; Kujawska, J.; Przysucha, B.; Cel, W. Forecasting Water Quality Index in Groundwater Using Artificial Neural Network. *Energies* **2021**, *14*, 5875. [[CrossRef](#)]
47. Szul, T.; Nęcka, K.; Mathia, T.G. Neural Methods Comparison for Prediction of Heating Energy Based on Few Hundreds Enhanced Buildings in Four Season’s Climate. *Energies* **2020**, *13*, 5453. [[CrossRef](#)]
48. Ding, M.; Wang, L.; Bi, R. An ANN-Based Approach for Forecasting the Power Output of Photovoltaic System. *Procedia Environ. Sci.* **2011**, *11*, 1308–1315. [[CrossRef](#)]