



# Article Stability and Reactivity of Guaiacylglycerol- $\beta$ -Guaiacyl Ether, a Compound Modeling $\beta$ -O-4 Linkage in Lignin

Zeinab Rabiei<sup>1</sup>, Andrew Simons<sup>1</sup>, Magdalena Folkmanova<sup>2</sup>, Tereza Vesela<sup>2</sup>, Ondrej Uhlik<sup>2</sup>, Evguenii Kozliak<sup>1,\*</sup> and Alena Kubátová<sup>1,\*</sup>

- <sup>1</sup> Chemistry Department, University of North Dakota, Grand Forks, ND 58201, USA; rabieizeynab@gmail.com (Z.R.); andrew.j.simons@und.edu (A.S.)
- <sup>2</sup> Department of Biochemistry and Microbiology, University of Chemistry and Technology, Prague, Technicka 3, 16628 Prague, Czech Republic
- \* Correspondence: evguenii.kozliak@und.edu (E.K.); alena.kubatova@und.edu (A.K.); Tel.: +1-701-7770348 (A.K.)

Abstract: Lignin, a complex and abundant biopolymer, is a major constituent of plant cell walls. Due to its chemical and structural complexity, lignin degradation is a challenging task for both natural and engineered systems. Therefore, investigation of lignin degradation using so called "model compounds" has been the focus of many research efforts in recent years. This study addresses the utility of guaiacylglycerol- $\beta$ -guaiacyl ether (G $\beta$ 2) as a model compound for evaluating the  $\beta$ -O-4 bond cleavage under diverse thermal and aqueous medium conditions. Experimental conditions included varied pH (3-10), microbial biodegradation, subcritical water environment (150-250 °C), and mild pyrolysis (150–250 °C). A high-performance liquid chromatography with high-resolution mass spectrometry was employed for accurate detection and quantification of both  $G\beta 2$  and its degradation/modification products in an aqueous environment. Pyrolysis experiments were performed using gas chromatography-mass spectrometry analysis with a pyrolyzer. The results showed that G $\beta$ 2 remained stable under exposure to moderate pH and several bacterial strains, which were successfully used previously for biodegradation of other recalcitrant pollutants. We report, for the first time, differing GB2 breakdown pathways for subcritical water treatment vs. pyrolysis under an inert atmosphere. The scientific novelty lies in the presentation of differences in the degradation pathways of GB2 during subcritical water treatment compared to pyrolysis in an inert atmosphere, with water playing a key role. The observed differences are ascribed to the suppression of homolytic reactions by water as a solvent.

**Keywords:** guaiacylglycerol- $\beta$ -guaiacyl ether; lignin model compound; subcritical water; pyrolysis; biodegradation

## 1. Introduction

Lignocellulosic biomass is a biocomposite material consisting mainly of cellulose, hemicellulose, and lignin. Lignin, as one of the main biomass constituents (15–35%) responsible for the rigidity of plant cell walls, is an important renewable feedstock for production of phenolic monomers, which are viewed as a possible replacement of fossil fuels [1].

Natural lignin is a heteropolymer consisting primarily of three phenylpropanoid monomers including *p*-hydroxyphenyl (H), guaiacyl (G), and syringyl units (S) (Scheme 1a) [2]. However, unlike regular biopolymers, e.g., proteins or carbohydrates, these common structural moieties are interconnected by a number of different strong linkages (Table 1) yielding an irregular and rather recalcitrant structure [3]. As a result, lignin typically depolymerizes into a mixture of phenolic derivatives (shown in Scheme 1b) occurring as both monomers and oligomers. Such monomers, e.g., those shown in Scheme 1b, are different from lignin



**Citation:** Rabiei, Z.; Simons, A.; Folkmanova, M.; Vesela, T.; Uhlik, O.; Kozliak, E.; Kubátová, A. Stability and Reactivity of Guaiacylglycerol-β-Guaiacyl Ether, a Compound Modeling β-O-4 Linkage in Lignin. *Separations* **2024**, *11*, 59. https:// doi.org/10.3390/separations11020059

Academic Editor: Paraskevas D. Tzanavaras

Received: 1 January 2024 Revised: 30 January 2024 Accepted: 9 February 2024 Published: 14 February 2024



**Copyright:** © 2024 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). structural monomers shown in Scheme 1a. For example, guaiacol derivatives are formed out of the guaiacyl unit [4]. These products, i.e., monoaromatic phenolic derivatives, are also called "phenolic monomers" to distinguish them from other, less useful products of lignin degradation containing several linked aromatic moieties, e.g., dimers, trimers, and higher-order oligomers [5].



**Scheme 1.** Chemical structures of phenolic monomers: (**a**). monolignols, the actual monomers of lignin synthesis and (**b**). examples of monoaromatic phenolic products of lignin degradation.

When targeting this feedstock's ultimate utilization, its structural complexity requires the use of representative model compounds reflecting specific motifs of lignin structure to effectively study lignin degradation pathways. These model compounds are usually phenolic dimers, which convert into a defined mixture of specific monomers as a result of breaking targeted linkages.

In this study, we focus on the most abundant linkage, the alkyl aryl ether  $\beta$ -*O*-4 bond (Scheme 2), which represents more than 50% of all linkages that exist in native softwood lignin [6]. A number of other, less abundant linkages occur as well, and some corresponding model compounds were also proposed and used in previous studies. However, those less abundant connections are outside of the scope of this particular study.



**Scheme 2.** Chemical structures of: (a) general model compound featuring a  $\beta$ -O-4 linkage and (b) specific model compound, guaiacylglycerol- $\beta$ -guaiacyl ether (G $\beta$ 2) used in this study.

One of the most recognized and used model compounds featuring both the guaiacyl moiety and  $\beta$ -O-4 linkage is guaiacylglycerol- $\beta$ -guaiacyl ether (Scheme 2), which is often abbreviated as G $\beta$ 2, although the alternate GGGE acronym is also used. Other model compounds with this linkage are less representative because they do not possess the guaiacyl moiety. This dissimilarity with the actual lignin structural features may affect both their reactivity and solubility in water [7,8]. A variety of approaches to lignin breakdown were applied using these model compounds including oxidative [9], reductive [10], catalytic [11], and biological degradation [12]. Those studies performed in aqueous media with or without catalysts are summarized in Table 1.

The analysis of this literature (Table 1), however, reveals two significant knowledge gaps pertaining to the use of lignin model compounds. First, while a variety of product analysis methods were applied, only a few studies addressed the analysis of both reactant (e.g., G $\beta$ 2) and products using a *single* protocol [12–14]. Namely, most of the previous studies employed gas chromatography-mass spectrometry (GC-MS) for product characterization. This method enables a reliable identification and quantification of volatile reaction products, i.e., guaiacols and other methoxyphenol derivatives [14]. However, the direct GC-MS analysis does not allow for the simultaneous determination of reactants, i.e., hydroxylated dimer model compounds, most of which are either not GC-elutable, as is G $\beta$ 2, or have low thermal stability [12–15]. The quantification was thus restricted to the breakdown products in most of the conducted studies (Table 1), thus limiting the options to conduct mass balance closure.

The above-mentioned papers that used a single GC-MS method for determination of both G $\beta$ 2 and its decomposition products employed the derivatization of hydroxyl groups [12,13]. However, this approach is not ideal for quantification, as the derivatization protocol must be optimized for any specific condition to assure its completion. Another approach is to quantify products directly by high-performance liquid chromatography with high resolution or tandem mass spectrometry (HPLC-HR MS or MS-MS) as reported by Rahimi et al. [14].

The second current knowledge gap is the shortage of baseline  $\beta$ -O-4 degradation data under moderate conditions in aqueous environments. While investigating various process conditions for model compounds' hydrolytic breakdown in water, the previous studies either used multiple additives (catalysts, co-catalysts, medium components, etc.), or conducted thermal breakdown under rather drastic conditions. In particular, various studies investigated the stability of lignin model compounds in heated aqueous media in the presence of catalysts (Table 1). Yet, little consideration was given to their stability in subcritical water (under sufficient pressure to maintain a liquid state) as a solvent.

Nagel et al. reported a  $G\beta2$  hydrothermal decomposition in water without additives as well as in aqueous media containing varying concentrations of hydroxide and carbonate bases [13]. However, considering the reaction volumes used, the reactions investigated in that study appeared to occur mainly in the gas phase (i.e., steam) rather than in liquid subcritical water, which may not be relevant as a baseline to many applications. Furthermore, the reactivities of steam and liquid water can be expected to differ, as the dielectric constant and thus the polarity are significantly reduced for steam [16]. On the other hand, the other studies referenced in Table 1 used water in its liquid phase. However, they also employed many different additives, which made it difficult to separate the influence of redox, thermal, and acid-base factors.

As for the biodegradation occurring in aqueous media under ambient conditions, the impact of pH and blank mineral medium has not been assessed. To the best of our knowledge, only two studies evaluated the biodegradation of lignin model compounds, both claiming significant G $\beta$ 2 conversion rates [12,17], although without detailed quantitative data [12]. The impact of biodegradation was evaluated in this study using three different strains. Of those, *Burkholderia xenovorans* LB400 is an example of aerobic bacteria that activate bond cleavage in polychlorinated biphenyls (PCBs) [18]. It has been shown that the extracellular media of two other aromatic hydrocarbon degrading microorganisms, *Pseudomonas putida* JAB1, and *Rhodococcus jostii* RHA1, enable the lignin breakdown occurring due to the genes encoding for the extracellular dye-decolorizing peroxidases (DyP), dypA and dypB [19,20].

**Table 1.** Literature data on model compounds featuring the  $\beta$ -O-4 linkage and summary of their degradation data in aqueous media including the products obtained and the analysis protocols.



	Table 1. Cont.			
Model Compound	Depolymerization Method & Conditions	Breakdown Products	Analysis &Quantification Approach	Ref.
Gβ2	Aqueous phase hydrodeoxygenation. MC (100 $\mu$ g), solid acid zeolite, H <sup>+</sup> -Y (300 mg), Ru/Al <sub>2</sub> O <sub>3</sub> (300 mg), H <sub>2</sub> O (30 mL), 40 bar H <sub>2</sub> , 250 °C, 4 h		GC-MS with quantification, using <i>n</i> -decane as an IS. The quantification appears to be based only on normalization. G $\beta$ 2 conversion NR. <sup>a</sup>	[22]
OH OH 2-Phenoxy-1-phenylethanol	Base promoted hydrogenolysis over metal catalysts in water. MC (10 mol%); Ni <sub>7</sub> Au <sub>3</sub> , NaOH, H <sub>2</sub> O, 10 bar H <sub>2</sub> , 130 °C, 0.5 h.	OH HO CH <sub>3</sub> HO CH <sub>3</sub> HO CH <sub>3</sub> CH S CH S CH S CH S CH S CH S CH S CH	GC-MS identification, GC-FID with quantification using standards and effective carbon number (ECN) factors. <i>n</i> -Dodecane was used as an IS.	[23]
Gβ2	Catalysis with H <sub>2</sub> donor agent (reduction); deoxygenation pathways under H <sub>2</sub> transfer conditions. MC (0.067 g/L) in 50% EtOH in H <sub>2</sub> O, Pt/C (5.12 g), 80 bar, 275 °C.	H <sub>3</sub> C CH <sub>3</sub> H <sub>3</sub> C CH <sub>3</sub>	GC-MS identification, quantification approach NR.	[24]
Gβ2	Catalytic hydrogenolysis. MC (0.5 mmol). 50% THF in $H_2O$ (20 mL), Fe-L1/C-800 catalyst (0.1 g), 10 bar $H_2$ , 240 °C, 12 h.	$\begin{array}{c} & & & \\ & & & & \\ & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & & & & \\ & &$	GC-MS with IS based quantification of products using phenol, dodecane to determine toluene. Gβ2 conversion NR	[11]
OH OH 2-Phenoxy-1-phenylethanol	<b>Catalytic redox-neutral</b> <b>C–O bond cleavage</b> . MC (0.1 mmol), Pd on metal-organic framework, EtOH/H <sub>2</sub> O at 120 °C, 6 h under N <sub>2</sub> . Pressure value NR. <sup>a</sup>	HO CH <sub>3</sub>	GC-MS quantification, the quantification approach NR, G $\beta$ 2 conversion NR. <sup>a</sup>	[25]
Ho $G\beta^2$ and its oxidized derivative	Formic acid-induced depolymerization. MCs (1 mmol), sodium formate (3 mmol), HCO <sub>2</sub> H:H <sub>2</sub> O (10:1) (5 mL), 110 °C, 12 h. Pressure value NR.	CH <sub>3</sub> CH <sub>3</sub> HO CH <sub>3</sub> HO H <sub>3</sub> C	Identification and quantification by LC-MS, LC-MS-MS and GC-MS with/without derivatization. Both products and feedstock MCs were quantified.	[14]

Model Compound	Depolymerization Method & Conditions	Breakdown Products	Analysis &Quantification Approach	Ref.
2-Phenoxyacetophenone	Bioconversion with a Gram-negative bacterium, Acinetobacter sp. TUS-SO1 cells with 0.1  g/L yeast extract and 1% (v/v) Tween 80 with MC (1 mM) in DMSO, 1% [v/v] at 30 °C for 72 h, shaking		HPLC-MS quantification of product and MC feedstock. GC-MS with trimethylsilylation confirmed identification.	[17]
Gβ2	<b>Bioconversion</b> ( <i>Burkholderia</i> <b>sp. ISTR5).</b> MC (2,000 mg/L), with overnight grown bacteria to obtain a final optical density of 0.1 at 30 °C for 72 h shaking.	$HO + CH_3 + CH_3 + O + O + CH_3 + O + O + O + O + O + O + O + O + O + $	GC-MS with BSTFA derivatization; identification of both products and MC, quantification approach NR.	[12]

Table 1. Cont.

<sup>a</sup> NR denotes not reported.

The present study was designed to investigate the G $\beta$ 2 stability toward both abiotic hydrolysis at moderate pH and biodegradation, as well as its thermal degradation at moderate temperatures, with and without subcritical water, using a reverse phase HPLC with high-resolution time-of-flight mass spectrometry (TOF-MS) for analysis. The results obtained in the aqueous environment were compared to breakdown pathways via mild thermolysis under inert atmosphere using pyrolysis (Pyr)-GC-MS. Therefore, we set out to obtain a "baseline" for any future studies to be conducted with this model compound when using various catalysts and more severe process conditions, ultimately leading to the development of new and improved methods for lignin utilization. As a result of using this method, this study enabled the evaluation of the impact of water as a solvent on the G $\beta$ 2 reactivity and decomposition pathways, by detecting and identifying the reaction products including any side reactions.

## 2. Materials and Methods

## 2.1. Materials

Water was obtained from Millipore Direct-Q UV, and LC-MS grade acetonitrile (ACN) was purchased from Fisher Scientific. G $\beta$ 2, used in biodegradation, abiotic hydrolysis, and subcritical water experiments, was synthesized at UND [26]. For pyrolytic experiments, this substrate was replaced with pure G $\beta$ 2 purchased from Sigma Aldrich St. Louis, MO, USA). All other standards including vanillin, 4-chloroacetophenone (used as an internal standard, IS), guaiacol, and 4-hydroxy-3-methoxycinnamaldehyde were purchased from Sigma Aldrich (see their structures in Figure S1). Three bacterial strains, *Paraburkholderia xenovorans* LB400, *Pseudomonas alcaliphila* JAB1, and *Rhodococcus jostii* RHA1 were tested for G $\beta$ 2 degradation. *Paraburkholderia xenovorans* LB400 was purchased from American Type Culture Collection, *Pseudomonas alcaliphila* JAB1 and *Rhodococcus jostii* RHA1 were obtained from the collection of strains in the Department of Biochemistry and Microbiology, UCT Prague. Bacteria were cultured in a mineral salt solution (MSS, described elsewhere [27]) with biphenyl (Sigma Aldrich) and sodium pyruvate (Sigma Aldrich) added, along with either yeast extract (Sigma Aldrich) or G $\beta$ 2.

#### 2.2. Assessment of $G\beta$ 2 Stability

For **pH effect studies**,  $G\beta2$  aqueous solutions (300 µg/mL) were prepared at three different pH values with acetic acid (0.1 M pH 2.9 rounded to 3 henceforth), ammonium acetate (0.1 M, pH 7.3 rounded to 7 henceforth), and ammonium hydroxide (0.1 M, pH 10.3 rounded to 10 henceforth), in two different container materials, i.e., glass and plastic vials. The G $\beta2$  stability was evaluated after exposure to these media for 0–7 days.

In **biodegradation experiments**, we followed the procedure used in our earlier study [27], with the essential modifications due to the use of G $\beta$ 2 as a substrate. The optimum substrate concentration and harvesting time-points were determined as follows. Bacterial cultures growing in MSS medium with sodium pyruvate and yeast extract (30 mM and 0.001%, respectively) at an exponential phase (optical density at 600 nm, OD<sub>600</sub>, of 0.5–1.0) were centrifuged at 5000× *g* for 10 min and washed twice with MSS. Cells were resuspended in MSS to a final OD<sub>600</sub> of 0.025. Autoclaved cell suspensions (121 °C for 20 min) were used as controls. Three replicates of 1 mL active or autoclaved cells or MSS were incubated with 0.1, 0.3, 0.5, 1.0, 3.0, 5.0, or 10.0 mM G $\beta$ 2 in 8 mL glass vials sealed with screw caps at 28 °C on a rotary shaker at 130 rpm. Every 24 h, the OD<sub>600</sub> of the culture was measured in microvolumes (50 µL). According to the measured growth curves, the optimal concentration of G $\beta$ 2 with which the culture reached the maximum OD<sub>600</sub> (5.0 mM) was determined. This concentration was then used during the assays described below.

The harvesting time points corresponding to the exponential phase, maximum  $OD_{600}$  and stationary phase were determined; these time points were then used during the degradation assays (DGA), which were conducted as follows. Three replicas of 1 mL active or autoclaved cells or MSS were incubated with 5.0 mM G $\beta$ 2 in 8 mL glass vials sealed with screw caps for 0, 20, 40, and 110 h at 28 °C on a rotary shaker at 130 rpm. Destructively harvested samples were stored at -20 °C.

Biodegradation of G $\beta$ 2 was assessed through a resting cell assay (RCA) using the following protocol. All strains were at least thrice re-inoculated in fresh medium with biphenyl as a sole carbon source to support the expression of degradative genes. Bacterial cultures growing in MSS with 5.0 mM biphenyl as a sole carbon source were filtered when reaching the exponential phase (OD<sub>600</sub> of 0.5–1.0) with a coffee filter to remove biphenyl crystals, centrifuged at 5000× *g* for 10 min and washed twice with MSS. Cells were resuspended in MSS to a final OD<sub>600</sub> of 0.5. Autoclaved cell suspensions (121 °C for 20 min) were used as controls. Four replicas of 1 mL active or autoclaved cells were incubated with 5 mM/1600 ppm G $\beta$ 2 in 8 mL glass vials sealed with screw caps for 0, 24, 48, and 72 h at 28 °C on a rotary shaker at 130 rpm. Destructively harvested cultures in vials were stored at –20 °C until analysis.

For analysis, a sample aliquot (0.1 mL) was transferred to an Eppendorf tube and extracted with 60% MeOH in water (0.5 mL). The extraction was completed using vortexing and refrigeration for 15 and 10 min, respectively. Then, an aliquot of the extract (0.1 mL) was filtered using a syringe filter and 0.9 mL of water and IS were added.

The **subcritical water experiments** were conducted in a lab scale batch reactor previously described [28]. Briefly, the system consisted of a GC oven (Hewlett-Packard GC5890) and rotor made of a Leeson permanent magnet DC gear motor (Grafton, WI, USA) equipped with a Dayton DC speed control. The rotary part was implemented in the thermally insulated door of the oven. The heated part of the rotor contained five holders to stir Parker vessels rated to 517 bars (Cleveland, OH, USA) at an approximate rate of 3 rpm to provide sufficient mixing.

An aqueous solution of G $\beta$ 2 without additives was prepared using 0.7 µL of G $\beta$ 2 (final concentration 280 µg/mL) and water with a total volume of ~2.4 mL. The experiment was set up in a way that ensured sufficient headspace while maintaining the pressure required to ensure that water was present in subcritical state and **safe operation (preventing overpressurization**), according to water/steam equilibrium tables (Table 2), similar to our earlier work [28]. The specific experimental details are included in the supplementary information. The total reaction time was one hour. Three different temperatures, 150, 200,

and 250 °C, were applied. Five replicates of each experiment were performed with the variance shown as standard deviation.

Table 2.	Saturation	conditions	for water	[29]	ŀ
----------	------------	------------	-----------	------	---

Reaction temperature (°C)	150	200	250
Water saturation pressure (bar)	4.780	15.55	39.76
Liquid phase density (g/cm <sup>3</sup> )	0.9167	0.8647	0.7989
Vapor density (g/cm <sup>3</sup> )	0.002545	0.007861	0.01997

For **mild pyrolysis experiments**, G $\beta$ 2 was used as a solid. The procedure was similar to that used in earlier studies [28]. Stainless steel Eco-cups were cleaned using a butane blowtorch until glowing hot to remove any impurities that could be carried over from previous experiments. The cups were allowed to cool to room temperature. Each analyte was individually placed in a prepared Eco-cup and weighed (30–60 µg) using a microbalance. Following the introduction of the analyte, a precleaned quartz filter (heated at 600 °C for 24 h) was placed on top of each sample in each Eco-cup.

All experiments were conducted in triplicate unless otherwise stated. The data (concentrations and/or degradation %) were reported as mean values  $\pm$  one standard deviation. The *t*-test with 95% confidence was used to assess the statistical significance of the obtained values compared to the blanks/controls.

#### 2.3. HPLC-TOF-MS and Pyr-GC-MS Analyses

The analyses, except for the pyrolysis experiment, were performed using a reverse phase HPLC Agilent 1100 Series system equipped with a diode array detector (DAD) and a TOF-MS detector, Agilent G1969A (Agilent Technologies, Santa Clara, CA, USA), with an electrospray ionization (ESI) source. The target analytes were separated using a gradient elution program in 30% ACN in water as described below. The separations were performed using an Agilent Zorbax Eclipse Plus C18,  $2.1 \times 159$  mm, 3.5-µm column. The injection volume was 20  $\mu$ L. The mobile phase consisted of 2.5 mM formic acid in water (solvent A) and 2.5 mM formic acid in acetonitrile (solvent B). The binary gradient program started with 0 min, 30% B (0–0.1 min), followed by a linear gradient to 90% B (0.1–10 min), then 90% B (10–13 min). The last step was conducted from 13 to 16 min with a gradient to 30% B followed by an equilibration 7 min hold preparing the system for the next analysis. The total run time was 23 min. The ESI was performed (after the optimization described in Table S1) in the positive ion mode, with a capillary voltage, 5000 V; fragmentor voltage, 125 V; nebulizer pressure (nitrogen), 25 psig; drying gas (N<sub>2</sub>) flow rate, 12 L min<sup>-1</sup>; gas temperature, 350 °C. The HPLC-TOF-MS data were recorded across the mass range of 50-1000 m/z.

The results were processed with Agilent MassHunter software version 10.0. Based on the obtained HPLC-TOF-MS data, i.e., peak retention times and mass spectra, and characteristic ions, G $\beta$ 2 and all products were identified and quantified (Supporting Figures S1 and S2). The quantification was done using standards when available. For the rest of the analytes, the response factors of vanillin and G $\beta$ 2 were used for quantification of monomers and dimers, respectively. For the quantitative data processing, the representative ions of target analytes were extracted within the  $\pm$ 0.03 amu range (Table 3).

The pyrolysis experiments were performed using online **pyrolysis with GC-MS.** The instrument consisted of a Frontier 3030D pyrolyzer with an autosampler connected to an Agilent 7890 GC instrument with an Agilent 5975C MS detector having an electron ionization source. The pyrolysis was performed at selected temperatures (150, 200, 250 °C) for 30 s. The GC inlet temperature was set at 300 °C and the analysis was operated in 1:20 split mode with a helium flow rate of 1 mL/min. The GC oven temperature was held constant at 300 °C, the column was 25 m long with 0.25 mm I.D. and a 0.25  $\mu$ m stationary phase composed of 5% phenyl polydimethylsiloxane (Ultra ALLOY<sup>+</sup>-5, Frontier). The MS

analysis was carried out without a solvent delay in the 10-650 m/z range. Based on the obtained Pyr-GC-MS data, i.e., peak retention times and mass spectra, all products were identified either by the NIST library or pure chromatographic standards (HMC, G, and V) (Table 4).

**Table 3.** HPLC-TOF-MS analysis of  $G\beta 2$  and products of its hydrotreatment breakdown: retention times and the ions used for identification of the corresponding compounds.

Compound Name, Abbreviation	Retention Time, min	Monoisotopic Mass, m/z	Quantification Ion [M + H] <sup>+</sup> <i>m</i> /z	Confirmation Ion $[M + Na]^+$ m/z	ID & Quant. Confirmation *
Guaiacylglycerol-β-guaiacyl ether, Gβ2	7.1	320.1260	321.1333	343.1152	Standard
Vanillin, V	5.9	152.0473	153.0546	175.0366	Standard
Guaiacol, G	6.9	124.0524	125.0597	147.0416	Standard
4-Hydroxy-3-methoxy cinnamaldehyde, HMC	7.8	178.0630	179.0703	201.0522	Standard
1-(2'-Methoxyphenoxy)-2-(4'- hydroxy-3'- methoxyphenyl)ethane, D1	8.4	274.1205	275.1278	297.1097	Tentative
3-Hydroxy-2-phenoxy-1-(4'- hydroxy-3'-methoxyphenyl)-1- propanone, D2	8.9	288.0998	289.1071	311.0890	Tentative
2-(2'-Methoxyphenoxy)-3-(4'- hydroxy-3'- methoxyphenyl)propanal, D3	9.5	302.1154	303.1227	325.1046	Tentative

\* The identification was either confirmed by standards or if unavailable, the tentative identification was based on high-resolution TOF-MS spectra with mass accuracy < 10 ppm. The tentatively identified dimers were quantified using the response factor for G $\beta$ 2.

Table 4. Retention times and	d molecular ions of	f compounds identified	ed in the Pyr-GC	<ul> <li>MS experiments.</li> </ul>
		1	2	1

Compound Name, Abbreviation	Retention Time (min)	Molecular Ion $m/z$	Identification Type *
Vanillin, V	6.7	152	Standard
Guaiacol, G	4.9	124.1	Standard
2-(2'-Methoxyphenoxy)-3-(4'-hydroxy- 3'-methoxyphenyl)propanol, D3	11.3	302.1	Tentative
<i>Cis/trans</i> -1,2-Di(4'-hydroxy-3'- methoxyphenyl) ethene, D4	11.6, 11.8, 12.8 for isomers	272.1	Tentative
4-Hydroxy-3-methoxy cinnamaldehyde	7.6	178.1	Standard
Coniferyl alcohol, CA	9.3	180.1	Tentative
Homovanillin, H	6.3	166.1	Tentative

\* The identification was either confirmed by standards or, if unavailable, the tentative identification was based on the MS NIST 2020 library.

## 3. Results

## *3.1. Stability at Different pH*

The first step was the assessment of  $G\beta2$  stability under the conditions relevant to biodegradation, i.e., at room temperature and non-extreme pH values. The pH dependence of  $G\beta2$  concentration before and after room-temperature hydrolysis is shown in Figure 1. The results obtained in both glass and plastic vials yielded similar data, thus indicating insignificant loss of target species due to sorption. The findings demonstrated that pH variation has only a minor impact on  $G\beta2$  in terms of hydrolysis. The results obtained after seven days, in both glass and plastic containers at pH 3 and 7, showed statistically insignificant changes, if any. However, in glass containers at pH 10, we observed a statistically significant decrease in the G $\beta$ 2 amount, which suggests an influence of adsorption in basic media on G $\beta$ 2 hydrolysis. One may conclude that G $\beta$ 2 is not prone to room-temperature hydrolysis under both acidic and neutral conditions.



**Figure 1.** Effect of pH on G $\beta$ 2 concentrations in aqueous solutions with two container materials: (a) initial concentration and (b) concentrations after 7 days. The initial concentrations of G $\beta$ 2 used for experiments were ~300 ppm, the data shown reflect the dilution for analysis. The results are presented based on triplicates  $\pm$  one standard deviation.

### 3.2. Evaluation of G $\beta$ 2 Biodegradation

 $G\beta2$  biodegradation was investigated for three different bacterial strains previously shown to break down aromatic pollutants [18]. No statistically significant degradation was observed for JAB1 after 72 h of incubation. Under similar conditions with LB400 and RHA1, the G $\beta2$  amount decreased by ~20% compared to the control with dead cells (Figure 2), showing a minor, yet statistically significant effect.



**Figure 2.**  $G\beta 2$  amounts in media with live biomass, taken as the ratios to those with dead biomass obtained under similar conditions at the same time, following average of 24–72 h incubation with control (medium with dead cells), *Pseudomonas putida* JAB1 strain, *Burkholderia xenovorans* LB400 strain, and *Rhodococcus jostii* RHA1 strain.

The loss of  $G\beta2$  occurred within the first 24 h, and  $G\beta2$  concentrations did not decrease further. Thus, despite extensive optimization of growth curves for all strains and

ing biodegradation. This is likely due to the recalcitrance of ether bonds in G $\beta$ 2. Other studies using *Burkholderia* sp. ISTR5 (R5) strain claimed successful lignin breakdown [12], yet a closer look at these published results shows that they were only qualitative, reporting the qualitatively assessed feedstock loss and products obtained by GC-MS following derivatization—but without quantification. Thus, the extent of biodegradation is difficult to determine, considering that response factors are different for the targeted analytes.

However, from the standpoint of creating a baseline for biodegradation experiments, the specific goal of this study has been fulfilled. Namely, no abiotic degradation was observed due to wall or medium effects, or adsorption to inactive biomass. This conclusion was confirmed by the lack of pH influence discussed in the previous section. G $\beta$ 2 proved to be more recalcitrant toward biodegradation than anticipated. Therefore, it appears to be suitable as a model substrate to assess lignin biodegradation.

## 3.3. Stability in Subcritical Water

In contrast to the observed stability of G $\beta$ 2 at moderate pH or under biodegradation conditions, its significant breakdown was observed in subcritical water at moderate temperatures of 150, 200, and 250 °C (Figure 3).



**Figure 3.** HPLC-TOF-MS total ion current chromatograms (TIC) ( $\mathbf{a}$ - $\mathbf{f}$ ) of ( $\mathbf{a}$ ) the G $\beta$ 2 sample before any treatment, ( $\mathbf{b}$ ) G $\beta$ 2 kept in the reaction vessel in water for 1 h at ambient conditions, then treated at ( $\mathbf{c}$ ) 150 °C, ( $\mathbf{d}$ ) 200 °C, and ( $\mathbf{e}$ ) 250 °C in subcritical water. ( $\mathbf{f}$ ) Extracted ion chromatograms (EIC) of the G $\beta$ 2 sample treated at 250 °C. For abbreviations of peak labels and extracted ions, see Table 3. G $\beta$ 2-m is an impurity in the synthesized G $\beta$ 2.

The HPLC-TOF-MS results demonstrated that  $G\beta 2$  undergoes a sizable conversion, even at 150 °C, while at 200 and 250 °C its conversion was complete. The same degradation products were observed at all three temperatures tested, but with different abundance. An extracted ion chromatogram confirming the identification of those products is provided as panel (f) of Figure 3.

As seen in Figure 3, HMC produced from G $\beta$ 2 decomposition was detected at all three temperatures, signifying that the C $_{\beta}$ –O bond was cleaved. Furthermore, G (which is the other product of the target bond cleavage) was also present among products, confirming that the C $_{\beta}$ –O bond is cleaved during the decomposition process. However, the other observed products, D1, D2, and D3, retained the C $_{\beta}$ –O bond. In line with previous studies, V was also identified as a significant product [14].

Based on the observed molecular ions as protonated or sodium adducts (Table 3) detected with high-resolution TOF, we proposed the breakdown pathways shown in Scheme 3. Three different pathways appear to take place. First, there is the cleavage of a C–C bond, which resulted in producing V. Only traces of corresponding fragment were observed. Second, there is the cleavage of the targeted C–O ether bond, that resulted in G and HMC production. Several modified G $\beta$ 2 products formed with bond cleavages only in side chains, i.e., dimers, D1–D3, were also observed.



Scheme 3. Proposed breakdown products and pathway following the G<sub>β</sub>2 treatment in subcritical water.

Reactant and product quantification was performed as described in the Section 2 to close mass balance. Figure 4 shows the determined amounts of both G $\beta$ 2 and its thermal degradation products. G $\beta$ 2 showed some stability at 150 °C whereas it was virtually gone at 200 and 250 °C. Conversely, the amounts of guaiacol (G) and vanillin (V) produced increased with temperature, as expected, assuming that higher temperature promotes decomposition reactions. However, unexpectedly, these two products of ether bond cleavage represented only two minor reaction pathways.



**Figure 4.** Reactant (G $\beta$ 2) and product amounts following a 1 h treatment in subcritical water (150, 200, and 300 °C. The results are presented based on triplicates  $\pm$  one standard deviation.

The formation of dimer products (D2–D3), which was found to be the major reaction pathway, indicates the occurrence of significant side reactions that modify the G $\beta$ 2 structure in subcritical water without affecting the  $\beta$ -O-4 bond. D2 was more abundant at higher temperatures, whereas D3 formation showed a maximum at 200 °C. Thus, this chemical appears to be less stable than D2. Bond dissociation enthalpy assessment indicated that the C–O bond strength of the  $\beta$ -O-4 linkage is approximately 70 kcal·mol<sup>-1</sup> in both the native lignin and its common model compounds, and the presence of various substituents has merely a minimal impact on this value, within 3 kcal·mol<sup>-1</sup> [30]. However, this increment may be sufficient to shift the minimum degradation temperature, as in the case of D2 vs. D3.

One other factor appears to be important in effecting the varied  $\beta$ -O-4 bond stability in different compounds. The low G $\beta$ 2 conversion with ether bond scission in subcritical water reported in this study contrasts with much higher values observed with water at much lower pressure, i.e., steam [13]. Nagel and Zhang observed similar products with the intact  $\beta$ -O-4 bond—but only in minor amounts [13]. This difference indicates that G $\beta$ 2 is significantly more stable toward high temperatures in a polar solvent, i.e., subcritical water.

Presumably, a polar solvent tends to suppress homolytic reactions, while the heterolytic ether bond cleavage requires a greater energy. Consistent with this conclusion, Dou et al. reported that an altered reaction pathway for zeolite-catalyzed cleavage was observed when water was added to the ethanol solution [30]. It is of note that G $\beta$ 2 modification products with intact  $\beta$ -O-4 bonds appear to be characteristic for aqueous media as a liquid-phase solvent. One needs to be aware of this feature when using G $\beta$ 2 as a lignin model compound at high temperatures.

#### 3.4. Stability toward Mild Solvent-Free Pyrolysis

To confirm the assumption that subcritical water protects  $G\beta 2$  from homolytic cleavage, a pyrolysis experiment was conducted without the presence of water in an inert helium atmosphere, using Pyr-GC-MS.

Figure 5 shows the Pyr-GC-MS TIC chromatograms obtained at three temperatures, 150, 200, and 250 °C. This information complements the other study, in which the pyrolysis of G $\beta$ 2 was investigated at much higher temperatures, 450, 550, 650, and 750 °C [31]. In contrast to our study, no dimers were detected in those high severity conditions, being replaced with catechols resulting from the scission of the other ether bond, within the methoxy group of G $\beta$ 2 [31]. The other study that investigated this process did not find any products but guaiacol at 200 and even 300 °C. Perhaps, the use of transfer lines in the Pyr-GC-MS setup precluded the detection of dimers and other less volatile products [32].



**Figure 5.** Pyr-GC-MS TIC chromatograms of the pure G $\beta$ 2 samples following the moderate pyrolysis at (a) 150 °C, (b) 200 °C and (c) 250 °C under inert helium atmosphere.

At moderate thermal conditions, G was the main product (Figure 5), indicating a smooth ether bond cleavage at 200 and 250 °C. This cleavage was presumably homolytic, as no medium of high polarity, i.e., water, was added. However, besides the main product, guaiacol, the obtained chromatograms showed a different product composition compared to the subcritical water treatment. Namely, the abundant peaks of dimers maintaining the  $\beta$ -O-4 bond, D1–D3, were not observed. Only D3 was recovered in small amounts. In contrast, a different dimer, D4, was formed, occurring as three isomers. This dimer is proposed to be an alkene derivative, presumably formed as a result of a bimolecular reaction between guaiacol and coniferyl alcohol (or its derivatives) as shown in Scheme 4. Most likely, a guaiacyl radical is formed as an intermediate of guaiacol formation, which may attach itself to the second product at several positions. Thus, given the different functional groups observed in the dimer products, our conclusion on the significant solvent effect of water was confirmed.



**Scheme 4.** Proposed Gβ2 moderate temperature pyrolysis breakdown pathways.

Furthermore, the guaiacol abundance proved to be greater than expected. In addition to G, the second product of the  $\beta$ -O-4 bond cleavage, CA, was formed. However, this second product was formed in smaller amounts than G, even when the other monomer derivatives of CA, e.g., HMC, were taken into account. The observed excess yield of G was apparently due to its additional formation caused by the CC bond cleavage (presumably, homolytic) occurring on the other side of the G $\beta$ 2 molecule, as shown in Scheme 4. Low selectivity in bond breaking is characteristic for homolytic reactions, so this phenomenon should be expected. However, the indistinguishability of guaiacol molecules formed by the cleavage of two different bonds is an undesired feature for studies targeting specifically a  $\beta$ -O-4 bond scission. This issue could be potentially addressed by using isotopically labeled model compounds, yet this was beyond the scope of this study. One has to beware of this issue when using model compounds with  $\beta$ -O-4 bonds in non-polar solvents or solvent-free pyrolytic reactions.

#### 4. Conclusions

This study investigated the efficiency of  $\beta$ -O-4 bond scission in moderate conditions using guaiacylglycerol- $\beta$ -guaiacyl ether (G $\beta$ 2) as the model compound. It appears to be suitable for biodegradation studies, because no artificial abiotic G $\beta$ 2 cleavage was observed. In the case of solvent-free pyrolysis, one has to be aware that guaiacol is formed from G $\beta$ 2, not only due to the  $\beta$ -O-4 bond scission, but also as a result of C–C bond cleavage on the other site of its molecule. However, this study showed only a limited G $\beta$ 2 applicability for thermal hydrolytic  $\beta$ -O-4 bond scission at 150–250 °C (i.e., in subcritical water), due to the occurrence of significant side reactions yielding other phenolic dimers, G $\beta$ 2 derivatives featuring much higher  $\beta$ -O-4 bond stability. The occurrence of such side reactions was shown to be due to the solvent effect of water protecting the reactant from homolytic thermal breakdown. The pathways of G $\beta$ 2 thermal degradation with and without subcritical water as a solvent were rather different, presumably reflecting the difference between heterolytic and homolytic bond cleavage, respectively. Extrapolating these results to lignin, the pathways of lignin depolymerization in subcritical water and as a result of solvent-free pyrolysis appear to have a different nature, leading to different products. The observed formation of the deoxygenated dimer as a result of G $\beta$ 2 pyrolysis (as opposed to subcritical water treatment) indicates that repolymerization of intermediates of lignin breakdown in that system may occur more readily.

**Supplementary Materials:** The following supporting information can be downloaded at: https: //www.mdpi.com/article/10.3390/separations11020059/s1, Figure S1: Structures of the compounds considered in this study; Figure S2: Mass spectra of the compounds used or observed in this study. Text: Experimental details for subcritical water treatment experiments. Table S1: Optimization of SI conditions with acquired responses for target ions dissolved in various solvent and electrolyte systems in positive/negative mode and varying fragmentor and capillary voltages performed using direct infusion.

**Author Contributions:** Conceptualization, A.K. and E.K.; methodology, Z.R.; software, Z.R.; validation, Z.R.; investigation, Z.R., M.F., T.V. and A.S.; data curation, O.U. and A.K.; resources M.F. and T.V.; writing—original draft preparation, Z.R.; writing—review and editing, A.K. and E.K.; project administration, A.K.; funding acquisition, A.K. and O.U. All authors have read and agreed to the published version of the manuscript.

**Funding:** The collaboration was enabled through National Science Foundation (NSF) IRES award # 1658615. The project was supported through U.S. Department of Energy's (DOE) Office of Energy Efficiency and Renewable Energy (EERE) Award DE-EE0009257. T.V. and M.F. were supported through ERASMUS, A.S. was supported through NSF S-STEM award #1742269. M.F., T.V., and O.U. received funding from the Czech Science Foundation (CSF) under grant no. 20-00291S. Any opinions, findings, conclusions, or recommendations expressed in this material are those of the author(s) and do not necessarily reflect the views of the NSF, DOE, or CSF.

**Data Availability Statement:** The raw data supporting the conclusions of this article will be made available by the authors on request.

**Acknowledgments:** We would like to express our gratitude to Irina Smoliakova for consultation on degrada tion pathways. Furthermore, we would like to acknowledge Nafisa Bala for operation and support with Pyr-GC-MS, and Chirby Ambo for HPLC work on biodegradation products.

Conflicts of Interest: The authors declare no conflicts of interest.

### References

- Guo, H.; Zhao, Y.; Chang, J.-S.; Lee, D.-J. Lignin to value-added products: Research updates and prospects. *Bioresour. Technol.* 2023, 384, 129294. [CrossRef]
- Jia, S.; Liu, X.; Li, X.; Sun, C.; Cao, X.; Liu, W.; Guo, G.; Bi, C. Modification of lignin composition by ectopic expressing wheat TAF5H1 led to decreased salt tolerance in transgenic arabidopsis plants. J. Plant Physiol. 2023, 287, 153997. [CrossRef] [PubMed]
- 3. Lupoi, J.S.; Singh, S.; Parthasarathi, R.; Simmons, B.A.; Henry, R.J. Recent innovations in analytical methods for the qualitative and quantitative assessment of lignin. *Renew. Sustain. Energy Rev.* **2015**, *49*, 871–906. [CrossRef]
- Zhang, R.; Qi, Y.; Ma, C.; Ge, J.; Hu, Q.; Yue, F.-J.; Li, S.-L.; Volmer, D.A. Characterization of lignin compounds at the molecular level: Mass spectrometry analysis and raw data processing. *Molecules* 2021, 26, 178. [CrossRef] [PubMed]
- Sun, Z.; Fridrich, B.; de Santi, A.; Elangovan, S.; Barta, K. Bright side of lignin depolymerization: Toward new platform chemicals. *Chem. Rev.* 2018, 118, 614–678. [CrossRef] [PubMed]
- Chakar, F.S.; Ragauskas, A.J. Review of current and future softwood kraft lignin process chemistry. *Ind. Crops Prod.* 2004, 20, 131–141. [CrossRef]
- Lahive, C.W.; Kamer, P.C.; Lancefield, C.S.; Deuss, P.J. An introduction to model compounds of lignin linking motifs; synthesis and selection considerations for reactivity studies. *ChemSusChem* 2020, 13, 4238–4265. [CrossRef]

- 8. Vangeel, T.; Schutyser, W.; Renders, T.; Sels, B.F. Perspective on lignin oxidation: Advances, challenges, and future directions. *Lignin Chem.* **2020**, *376*, 53–68. [CrossRef]
- 9. Zhang, C.; Li, H.; Lu, J.; Zhang, X.; MacArthur, K.E.; Heggen, M.; Wang, F. Promoting lignin depolymerization and restraining the condensation via an oxidation—hydrogenation strategy. *ACS Catal.* **2017**, *7*, 3419–3429. [CrossRef]
- Zhang, B.; Qi, Z.; Li, X.; Ji, J.; Luo, W.; Li, C.; Wang, A.; Zhang, T. ReO<sub>x</sub>/AC-catalyzed cleavage of C–O bonds in lignin model compounds and alkaline lignins. ACS Sustain. Chem. Eng. 2019, 7, 208–215. [CrossRef]
- 11. Li, J.; Sun, H.; Liu, J.-X.; Zhang, J.-J.; Li, Z.-X.; Fu, Y. Selective reductive cleavage of C-O bond in lignin model compounds over nitrogen-doped carbon-supported iron catalysts. *Mol. Catal.* 2018, 45, 236–245. [CrossRef]
- 12. Morya, R.; Kumar, M.; Thakur, I.S. Bioconversion of guaiacylglycerol-β-guaiacyl ether into vanillin by *Burkholderia* sp. ISTR5: A genomic and proteomic approach. *Int. Biodet. Biodegr.* **2021**, *161*, 105231. [CrossRef]
- 13. Nagel, E.; Zhang, C. Hydrothermal decomposition of a lignin dimer under neutral and basic conditions: A mechanism study. *Ind. Eng. Chem. Res.* **2019**, *58*, 18866–18880. [CrossRef]
- 14. Rahimi, A.; Ulbrich, A.; Coon, J.J.; Stahl, S.S. Formic-acid-induced depolymerization of oxidized lignin to aromatics. *Nature* **2014**, 515, 249–252. [CrossRef] [PubMed]
- Kozliak, E.I.; Kubátová, A.; Artemyeva, A.A.; Nagel, E.; Zhang, C.; Rajappagowda, R.B.; Smirnova, A.L. Thermal liquefaction of lignin to aromatics: Efficiency, selectivity, and product analysis. ACS Sustain. Chem. Eng. 2016, 4, 5106–5122. [CrossRef]
- 16. Pawliszyn, J. Sampling and Sample Preparation in Field and Laboratory: Fundamentals and New Directions in Sample Preparation; Elsevier: Amsterdam, The Netherlands, 2002.
- 17. Oya, S.; Tonegawa, S.; Nakagawa, H.; Habe, H.; Furuya, T. Isolation and characterization of microorganisms capable of cleaving the ether bond of 2-phenoxyacetophenone. *Sci. Rep.* **2022**, *12*, 2874. [CrossRef] [PubMed]
- Parnell, J.J.; Park, J.; Denef, V.; Tsoi, T.; Hashsham, S.; Quensen, J., III; Tiedje, J.M. Coping with polychlorinated biphenyl (PCB) toxicity: Physiological and genome-wide responses of *Burkholderia xenovorans* LB400 to PCB-mediated stress. *Appl. Environ. Microbiol.* 2006, 72, 6607–6614. [CrossRef]
- 19. Bugg, T.D.H.; Rahmanpour, R. Enzymatic conversion of lignin into renewable chemicals. *Curr. Op. Chem. Biol.* **2015**, *29*, 10–17. [CrossRef]
- Ahmad, M.; Roberts, J.N.; Hardiman, E.M.; Singh, R.; Eltis, L.D.; Bugg, T.D.H. Identification of DypB from *Rhodococcus jostii* RHA1 as a lignin peroxidase. *Biochemistry* 2011, 50, 5096–5107. [CrossRef]
- He, J.; Zhao, C.; Lercher, J.A. Ni-catalyzed cleavage of aryl ethers in the aqueous phase. J. Am. Chem. Soc. 2012, 134, 20768–20775. [CrossRef]
- 22. Wang, H.; Ruan, H.; Pei, H.; Wang, H.; Chen, X.; Tucker, M.P.; Cort, J.R.; Yang, B. Biomass-derived lignin to jet fuel range hydrocarbons via aqueous phase hydrodeoxygenation. *Green Chem.* **2015**, *17*, 5131–5135. [CrossRef]
- 23. Konnerth, H.; Zhang, J.; Ma, D.; Prechtl, M.H.G.; Yan, N. Base promoted hydrogenolysis of lignin model compounds and organosolv lignin over metal catalysts in water. *Chem. Eng. Sci.* 2015, 123, 155–163. [CrossRef]
- 24. Besse, X.; Schuurman, Y.; Guilhaume, N. Reactivity of lignin model compounds through hydrogen transfer catalysis in ethanol/water mixtures. *Appl. Catal. B Environ.* **2017**, 209, 265–272. [CrossRef]
- Ahsan Usman, M.; Naeem, M.; Saeed, M.; Zaheer, M. Catalytic C-O bond cleavage in a β-O-4 lignin model through intermolecular hydrogen transfer. *Inorg. Chim. Acta* 2021, 521, 120305. [CrossRef]
- Schumaker, J. Quantification of Lignin Hydroxyl Containing Functional Groups via <sup>31</sup>P[<sup>1</sup>H] NMR Spectroscopy and Synthesis of Lignin Degradation Standards. M.S. Thesis, University of North Dakota, Grand Forks, ND, USA, 2017.
- Uhlik, O.; Jecna, K.; Mackova, M.; Vlcek, C.; Hroudova, M.; Demnerova, K.; Paces, V.; Macek, T. Biphenyl-metabolizing bacteria in the rhizosphere of horseradish and bulk soil contaminated by polychlorinated biphenyls as revealed by stable isotope probing. *Appl. Environ. Microbiol.* 2009, 75, 6471–6477. [CrossRef] [PubMed]
- 28. Voeller, K.; Bílek, H.; Kreft, J.; Dostálková, A.; Kozliak, E.; Kubátová, A. Thermal carbon analysis enabling comprehensive characterization of lignin and its degradation products. *ACS Sustain. Chem. Eng.* **2017**, *5*, 10334–10341. [CrossRef]
- 29. Harvey, A.H.; Peskin, A.P.; Klein, S.A. *NIST/ASME Steam Formulation Database 10*; National Institute of Standards and Technology: Gaithersburg, MD, USA, 1996.
- Jasiukaitytė-Grojzdek, E.; Huš, M.; Grilc, M.; Likozar, B. Acid-catalysed α-O-4 aryl-ether bond cleavage in methanol/(aqueous) ethanol: Understanding depolymerisation of a lignin model compound during organosolv pretreatment. *Sci. Rep.* 2020, 10, 11037. [CrossRef] [PubMed]
- 31. Wang, L.; Yin, J.; Jiang, J.; Zhang, Y.; Song, M.; Zhang, R.; Dong, Z.; Yang, H.; Yu, H. Revealing G-lignin model compounds pyrolysis behavior: β-O-4 and 5-5′ dimer and trimer. *Fuel* **2022**, *317*, 123531. [CrossRef]
- Liu, Y.; Lyu, G.; Ji, X.; Yang, G.; Chen, J.; Lucia, L.A. Analytical pyrolysis pathways of guaiacyl glycerol-β-guaiacyl ether by py-gc/ms. *BioRes.* 2016, 11, 5816–5828. [CrossRef]

**Disclaimer/Publisher's Note:** The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.