

Review

Development of Ecosystem Health Assessment (EHA) and Application Method: A Review

Shaokang Fu ^{1,2}, Lin Zhao ^{1,2,*}, Zhi Qiao ^{1,2} , Tong Sun ^{1,2}, Meng Sun ^{1,3}, Yuying Hao ^{1,2}, Siyu Hu ^{1,2} and Yanchang Zhang ³

¹ School of Environmental Science and Engineering, Tianjin University, Tianjin 300072, China; fushaokang976@tju.edu.cn (S.F.); qiaozhi@tju.edu.cn (Z.Q.); 2018214043@tju.edu.cn (T.S.); sunmeng@tju.edu.cn (M.S.); hyy_4869@tju.edu.cn (Y.H.); hsy_0116@tju.edu.cn (S.H.)

² Tianjin Engineering Center for Technology of Protection and Function Construction of Ecological Critical Zone, Tianjin 300072, China

³ School of Chemical Engineering and Technology, Tianjin University, Tianjin 300072, China; zhangyc11235@126.com

* Correspondence: zhaolin@tju.edu.cn

Abstract: Human industrialization has caused damage to ecosystems. In this context, researchers have developed several methods to assess the health of various types of ecosystems. In this paper, we evaluated the developmental history and status of ecosystem health (EH) and summarized the concept of EH. We also reviewed ecosystem health assessment (EHA) methods and analyzed the application of EHA methods. EHA methods are generally classified into biological indicator and index system method. The former method is mainly based on the number of dominant species, such as diatom, plankton, and macroinvertebrate. Results indicate that trophic diatom index (TDI), plankton index of biotic integrity (P-IBI), and Ephemeroptera, Plecoptera, and Trichoptera (EPT) are the most commonly used indices. The latter method combines multiple ecosystem metrics and reflects ecosystem processes. The pressure–state–response (PSR) model most commonly uses the index system method. For the application of EHA methods, biological indicator methods are mostly applied in rivers/streams ecosystem, while the index system is primarily involved in urban ecosystems. Therefore, the information presented in this review may be helpful for the modification of EHA methods.

Keywords: ecosystem; ecosystem health assessment (EHA); indicators; index; application



Citation: Fu, S.; Zhao, L.; Qiao, Z.; Sun, T.; Sun, M.; Hao, Y.; Hu, S.; Zhang, Y. Development of Ecosystem Health Assessment (EHA) and Application Method: A Review. *Sustainability* **2021**, *13*, 11838. <https://doi.org/10.3390/su132111838>

Academic Editors: Tina Eleršek, Anja Bubik and Nataša Mori

Received: 15 August 2021
Accepted: 23 October 2021
Published: 26 October 2021

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

1. Introduction

With the development of industrial civilization, the increase in pollution has severely affected the natural environment [1,2]. According to the assessment results of the Millennium Ecosystem Assessment project, human activities have significantly altered several ecosystems on Earth in the last 50 years of the 20th century [3]. The status of ecosystem health is closely related to sustainable human development [4], and natural ecosystems provide the materials and services needed for human survival [5]. When disturbances reduce or exceed the regulatory capacity of the system itself, the ecosystem also limits the sustainable growth of human well-being by providing lower-quality ecosystem services [6]. Maintaining ecosystem health is the primary factor in achieving sustainable socio-economic development. Consequently, the assessment of the overall health of ecosystems is of great significance.

The assessment of the ecosystem has helped us to understand the degree of the persecution of the ecosystem, identify the pollutants flowing into the ecosystem, and prevent contaminants from increasing pollution [7,8]. These pollutants would affect the development of organisms in the ecosystem and lead to toxic effects on organisms that are exposed to them [9]. For example, diatoms [10,11], planktons [12–14], and macroinvertebrates [15,16] can characterize heavy metals (Zn, Cu, Hg, and AS), microplastics, and

organic pollutants in the ocean. In addition, EHA helps to understand the impact of EH on the human living environment, which is embodied in population, economy, and social responses [17,18]. Ecosystem services reflect the interdependent relationship between human beings and ecosystems, which is critical for ecosystem health [19,20]. When ecosystems deteriorate, ecosystem services and functions are significantly reduced, reducing their social and economic benefits to a considerable extent [20–22]. Moreover, the health assessment of ecosystems provides a scientific basis for managers to improve scientific management [23], which is helpful to build sustainable livelihoods [24]. When confronted with different types of ecosystems, decision-makers make and manage decisions according to their goals. In natural ecosystems, the corresponding adjustment of the ecosystem is mainly based on the pollution level [25,26]. In the urban ecosystem, decisions are based on sustainability to maximize social and natural benefits [27,28].

This study examines the developmental history and developmental status of EH. It summarizes the current popular keywords and studies conducted by famous scholars in the field of research. We also reviewed the assessment methods of EH, which were classified into two ways. One of the methods is the biological indicator method, which mostly utilizes metrics or indexes of selected indicator organisms such as diatoms, plankton, and macroinvertebrates to characterize the ecosystem's health. The present study collected and analyzed the indexes of the above three commonly used indicator organisms. The other method is the index system method, which utilizes large-scale economic and social metrics, combines other metrics, determines weights, and thus forms a comprehensive index to assess the ecosystem's health. Comparisons between the two methods are presented in Table S1. In addition, the application of the EHA method is reviewed and analyzed.

2. The Development of Ecosystem Health

2.1. Brief History of Ecosystem Health

EH development can be divided into three periods (Figure S1). The first stage is the ideology enlightenment. As early as 1941, Leopold put forward the concept of “Land health” based on the idea of medical health [29]. The Soil and Health Association of New Zealand was established in the same year, resulting in being the first academic organization to implement EH as the research target [30]. In the 1970s, due to the development of stress ecology, researchers studied ecosystem services and ecosystem functions [31–34], laying the foundation for the proposal of EH.

The second stage is the proposal of concept. Researchers discussed the concept of EH from 1980 to 1990. The general development of this stage is presented in Table S2. In the past, researchers have mainly tried to explain the health of the ecosystem in two aspects. From a process perspective, researchers explain the ecosystem's stress and responsiveness by analyzing how the ecosystem changes from a normal state to an abnormal state. From the perspective of results, researchers pay attention to the damaged state of the ecosystem and focus on using physiology to describe the syndrome of an unhealthy ecosystem. The views of some representative researchers are summarized in Table S3. In addition, we give our definition of EH: “an ecosystem that is of great biodiversity, able to resist natural and man-made disturbances, maintain structural integrity, be self-sustaining and renewing, meet the reasonable needs of people, and serve society”.

The third stage is the development of EH. Here, the research focus shifted to methods of EHA. The general development of this stage is presented in Table S4. As early as the concept stage, researchers used ecological indicators to assess ecosystem status [35]. However, early research was limited by conditions and operability, and the results have no greater scientific significance for environmental management. For example, in an aquatic ecosystem, when there are enough species of organisms at the same trophic level, fluctuations in the number of a single organism at the same trophic level will not have much impact on the entire ecosystem [36,37]. Researchers have been working on the innovation of assessment methods for different types of ecosystems.

2.2. Research on EHA in 21st Century

In the 21st century, researchers improved and innovated assessment methods and conducted case studies in a wide range of ecosystems, including aquatic and terrestrial ecosystems. In the present study, the Web of Science core database was used to perform bibliographic retrieval on “ecosystem health assessment” in the last 20 years of the 21st century. The retrieval results were analyzed using VosViewer (version 1.6.16). Here are the search criteria: TS = (“ecosystem health”) AND (evaluate* OR assess*), database: core collection of Web of Science, language: English, paper type: Article, time: 2000–2020. The title, abstract, keywords, and authors of 1750 publications were recorded and reviewed. The number of articles published each year is shown in Figure S2. In particular, the number of articles in 2013 increased by 64% compared to the previous year. The number of articles published in 2020 decreased compared to 2019, which may be due to the impact of COVID-19.

The 1750 retrieval results were analyzed using the keyword co-occurrence in VosViewer. In the data preprocessing, the occurrence frequency was set to ten to screen 9858 keywords. A total of 268 keywords met the condition, and were cleaned according to the following criteria: keywords with different formats but the same nouns are removed (e.g., “water quality” and “water-quality”); single and plural nouns are removed (e.g., “bioindicators”, “bioindicators”, and “biological indicators”); meaningless keywords are removed (e.g., “challenge”); keywords with the same meaning but different spelling were removed (e.g., “Estuarine” and “Estuary”). Ultimately, 231 keywords were determined. The co-occurrence analysis results were shown in Figure 1a. As observed from the figure, the 231 keywords were divided into five clusters, with different colors representing different clusters. The highest occurrence frequency is “ecosystem health”, which appeared 305 times. Except ecosystem health, the term “management” appears in the most in the publications. The overall clustering effect is not apparent, indicating that there was no significant degree of classification in the research in this field. Researchers mainly conduct ecological research (such as biodiversity and management) and toxicology research (contaminant and metal). Each cluster presents a tendency of overlapping, and the keywords are closely related, indicating that the study in this direction is comprehensive.

The 1750 retrieval results were analyzed in VosViewer for the co-citation of authors. In the data preprocessing, the citation frequency was set to 25 to screen 54093 authors. A total of 165 articles met the condition. The data were cleaned as follows: the authors with the wrong format but the same author name (such as “*EPA” and “US, EPA”) were merged; the abbreviated names of the authors and the full names of the authors (such as “Costanza, R” and “Costanza, Robert”) were merged. As a result, 157 results were determined, and the co-cited analysis results are shown in Figure 1b. The 157 authors are divided into 6 clusters. The authors with the highest citation frequency are “Rapport, DJ”, whose article has been cited 356 times in total. The authors with the closest connection are “Jorgensen, Se” and “Xu, Fl”. At the same time, the fact that “Rapport, DJ” and “Costanza, Robert” often collaborate academically is also proved by this figure, as the intensity of co-citation between the two authors is 478. Five authors who have made significant contributions to this field are visible in the figure: “Karr, JR”, “US, EPA” (US Environmental Protection Agency), “Rapport, DJ”, “Costanza, Robert”, and “Jorgensen, Se”. These results reduce the blindness of reading literature and provide a basis for researchers to quickly understand the development of this field.

and other disciplines to screen out appropriate comprehensive metrics, establish an index system, and comprehensively assess the EH [5,24,39].

3.1. Biological Indicator Method

3.1.1. Benthic Diatoms

In temperate regions, aquatic systems are formed by various communities. However, diatom species play a vital role in these communities [40]. These species are often utilized as water quality indicator organisms, which can significantly reflect the biological integrity of the study area and the anthropogenic stressors affecting EH [41]. Diatoms respond quickly to the physical and chemical characteristics of the aquatic system [42]. Herein, DO, turbidity, nitrate, and phosphate levels were found to be significantly correlated to the diatom index ($R > 0.9$) [43]. Diatoms are considered more efficient than other communities because their production time is shorter than that of fish and macroinvertebrates [44]. Compared with other organisms, the diatom community structure provides a time-integrated indication of water quality composition [45]. Furthermore, the major advantage of diatoms is that most diatom species have a wide geographic distribution and universal applicability [46–48]. In China [49], South Africa [50], Turkey [51], Brazil [45], the United Kingdom [52], and the United States [53] there are cases of using diatoms as indicator organisms to assess ecosystem health. The integrity of these communities provides a straightforward but not an overall measure of the ecosystem, which further reflects the pollution of the ecosystem [54].

Owing to the diatom size and living conditions, the diatom index is widely used [55]. Diatom indices are based on ecological metrics, such as species abundance and diversity index, and use the formula to calculate the new index [56]. Half of the published papers about diatoms studied the diatom index [55]. Therefore, the biological diatom index is a key point in diatom research. Table S5 presents 38 diatom indices of the studies that were reviewed in this study. All the publications were relatively highly cited references which were published since 2000. It can be observed from the table that the Trophic Diatom Index (TDI) is the most commonly used, appearing 16 times in the table. It is worth noting that the index of biological integrity based on diatoms is also often implemented by researchers. The individual ecological index based on diatoms and IBI are two basic methods for evaluating the environmental conditions of rivers and streams using diatoms [57]. In terms of research areas, the diatom index has global applications. The most important application areas are concentrated in Europe, which appears ten times in the table, along with various countries in Europe which have applied the diatom index for EHA. China and North America are also popular areas of research. As for metrics, species abundance is the most used metric in the diatom index, which suggests that the biggest difference between the reference point and the damage point was species abundance. The results are shown in Figure 2a.

3.1.2. Plankton

Plankton is utilized to assess the health of aquatic ecosystems, such as lakes, rivers, and reservoirs. Plankton is a collection of species that mostly includes phytoplankton and zooplankton. Phytoplanktons are the most critical aquatic organism in aquatic ecosystems [58]. Phytoplanktons are at the first trophic level and absorb light energy for energy transfer and material transformation [59]. Their biological activity affects the circulation of several primary and trace elements, such as carbon, nitrogen, and iron [60]. However, excessive nutrients usually accelerate the growth of phytoplankton, and excessive phytoplankton populations may lead to algal blooms [61,62]. In coastal zones, phytoplankton is always a good indicator, as population growth increases the importation of nutrients into the seawater, leading to frequent HAB accidents [63]. Zooplanktons are the primary consumers of aquatic ecosystems and are indispensable in biochemical transformation process [64]. The community responds quickly to environmental variables related to climate change [65,66]. It is a vital link in the food chain and food web, and most zooplankton feed on phytoplankton [67]. Many researchers suggest that zooplanktons should be taken as biological quality elements of Water Framework Directive (WFD) [68–70].

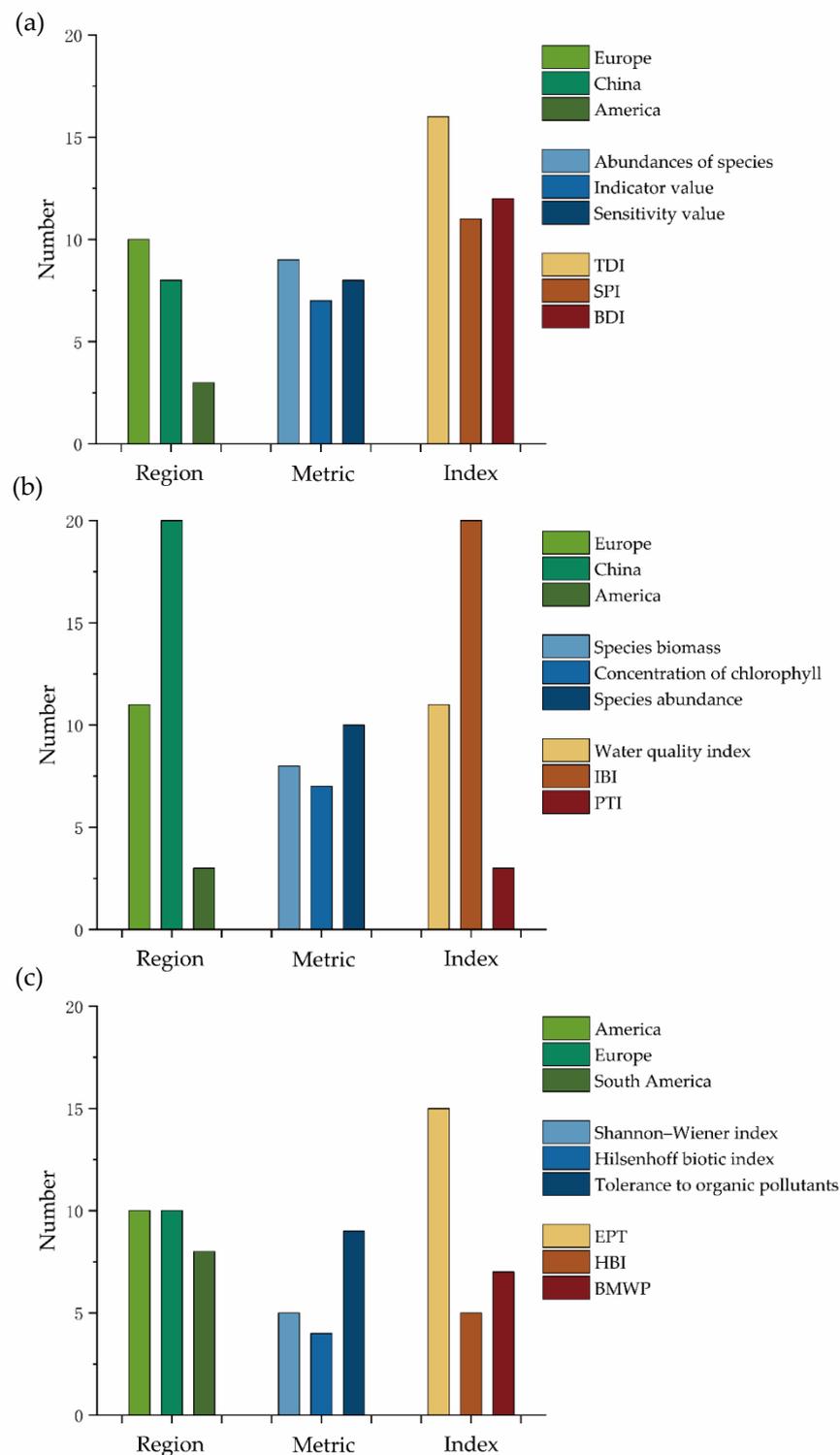


Figure 2. The most commonly used indices, metrics of indices, and where they are used most. (a) Diatom indices: BDI: Biological Diatom Index; SPI: Specific Pollution Sensitivity Index; TDI: Trophic Diatom Index. (b) Plankton indices: PTI: Phytoplankton Trophic Index; IBI: Index of Biotic Integrity. (c) Macroinvertebrate indices: BMWP: Biological Monitoring Working Party; HBI: Hilsenhoff biotic index; EPT: Ephemeroptera, Plecoptera, and Trichoptera.

The plankton-based assessment method is also mainly modified based on the Index of Biological Integrity (IBI) and has been applied globally [71–74]. Table S6 lists 38 plankton index studies reviewed in this study. All the publications were relatively highly cited

references published since 2000. Among the 38 plankton indexes in Table S6, 18 were developed or used based on the index of biological integrity (IBI). Furthermore, the next most used are the plankton trophic index (PTI) and the water quality index (WQI), which are utilized three times. It should be noted that an index such as the phytoplankton community index, which is only applied in one area (the UK), also has a high frequency of use [75,76], but it is not applicable. In terms of research areas, the most significant number of studies were carried out in China, with 20 occurrences, followed by 11 in Europe and 3 in the United States. The most frequently used metric is species abundance, including the abundance of each species and species combination. Ten studies selected this metric. In addition, species biomass and chlorophyll concentration are also frequently chosen by researchers, with frequencies of 8 and 7, respectively. The results are shown in Figure 2b.

3.1.3. Macroinvertebrate

Macroinvertebrates are mostly used to assess the health of river ecosystems. Owing to their limited mobility in river channels, they are relatively easy to collect [77]. Macroinvertebrates are consumers of medium nutrient levels in aquatic ecosystems. They act as nodes in the system and contribute to nutrient cycling, primary productivity, and material decomposition and transfer [78]. The existence of macroinvertebrate species often alters the physical surroundings or the flow of resources, thereby creating or modifying habitats and influencing other organisms in the community [79]. The activities of these organisms create dynamic sediment mosaic, effectively transport solutes into burrows, and increase sediment oxygenation [80]. In addition, macroinvertebrates are sensitive to low concentrations of pollutants in water and can be utilized to detect and assess the early state of rivers [81].

Generally, the health index of macroinvertebrates is divided into macroinvertebrate-based biotic indices and macroinvertebrate-based multi-metric indices [82,83]. A biotic index based on the tolerances of each observed taxa generally uses only one metric to assess river health, and the other elements need to be idealized. In comparison, multi-metric indices use several metrics to assess river health. Their development should consider the complexity of river ecosystems and they provide more detailed information for decision-makers [84]. Table S7 lists the 41 plankton index studies that were reviewed in this study. All the publications were relatively highly cited references which had been published since 2000. Among the 41 indexes counted, 15 studies utilized Ephemeroptera Plecoptera Trichoptera (EPT) as the basic index, making EPT the most commonly used basic index. Among the modifications to the EPT, most researchers added biological metrics such as the number or percentage of Trichoptera [85,86]. This enables the modified index to better reflect the internal conditions of the river and consider local characteristics. Assessment research using the macroinvertebrate index is mainly concentrated in Europe, North America, and South America. Europe and the United States are at the top level in this field. Researchers often use the EPT as a metric to calculate new indices. In addition, they commonly use tolerance metrics and Shannon–Wiener index. The results are displayed in Figure 2c.

3.1.4. Other Indicators

Several other species were selected as the biological indicators. Fish have been used as indicators of aquatic ecosystem changes mostly because of their accessible collection, handling, identification, and sensitivity to habitat loss and other environmental stressors [87–89]. According to Pérez-Domínguez [90], most studies of fish-based indices have been carried out in streams and rivers of the United States and other temperate countries. It is also one of the four key BQEs in the WFD [91]. Besides fish, Simone Ciadamidaro selected the black fly (Simuliidae) as an indicator to assess the ecosystem health of streams in urban areas because its larvae are distributed in the waters [92]. In addition, oysters [93], marine turtles [94], seagrass [95], and many other organisms were used as indicators. When the assessed target is a terrestrial natural ecosystem, other flora and fauna could be

used as alternatives. Ants could perfectly reflect the health condition in a comprehensive way [96,97]; specific species like sled dogs [98], which only live in cold zones, are a very good choice to assess arctic ecosystem health.

3.2. Index System Method

Although the biological indicator method provides a feasible and straightforward way to assess the health of aquatic ecosystems, limitations still exist. Thus, it can be inferred that: (1) Species of different ecological levels should be taken into consideration. (2) Species interactions are ignored. (3) The external pressure on the ecosystem is not well-reflected. To make corrections and better reflect the actual situation, it is necessary to establish an index system to synthesize a large amount of complex information. The vigor (V), organization (O), and resilience (R) proposed by Costanza in the early stages are the prototype of this method [99]. This method combines multiple metrics of the ecosystem and reflects the process of the ecosystem, which also aims to assess ecosystem health from the perspective of ecosystem structure, function succession process, ecological services, and product services. The index system method emphasizes the ecosystem's services to humans and the evolutionary relationship between the ecosystem and the regional environment [100]. Based on the VOR idea, researchers have developed many models and frameworks, such as PSR, DPSR, and DPSEEA, as listed in Table S8. Small-scale metrics help to understand the community's pollution level and spatial distribution, while the metric selected on the medium-scale and large-scale can be used for regional differences but ignores specific details [101]. Multi-scale metrics of toxicology, chemistry, community structure functions, and socio-economic factors can provide more comprehensive information and display a more realistic level of ecosystem health. After establishing a suitable index system, the ecosystem health index (EHI) is calculated according to the weight of each metric. Different weighting methods lead to different ranges of the EHI.

In the large scale assessment of EH, remote sensing (RS) has potential for assessing and monitoring ecosystem health at different temporal and spatial scales across extensive areas with a broad extent [102,103]. The main advantage of RS tools is that they can be used for directly detailing ecological health indicators, such as productivity, species richness [104,105], and resilience after natural and human-induced disturbances and for indirectly providing inputs for spatially explicit ecological process modeling [103–106]. In addition, RS removes barriers of scales [107]. The crucial point of using RS is the relevance of metrics that are obtained by RS to ecosystem health. An analytical framework has been provided when using remote sensing tools in EHA [108]. It is a procedure for researchers which can make metrics reflect the ecosystem health better. Land use/land cover is the most commonly used RS data [109]. Based on this, landscape indices such as PD, NP, SHID, LPI, SIDI, and MPS were calculated and widely applied in models [110–112].

3.2.1. Classification and Selection of Metrics

The establishment of the EHA index system can be conducted in two ways. The first is the internal metric of the ecosystem, including toxicology, ecology, and biochemistry; the second is the external metric of the ecosystem, such as the socio-economic metric and ecosystem service metric [113]. In the health assessment of a rapidly urbanizing urban ecosystem, ecological function and production function can be utilized as internal metrics, and economic metrics such as GDP can be utilized as external metrics [114]. Some researchers have proposed a more comprehensive index classification framework or model [115,116]. The “pressure-state-response” (PSR) model [117] proposed by the OECD in 1993 has been applied by most researchers in EHA. This model classified metrics into pressure, state, and response, which makes it easier to select metrics [118].

A more positive response generally indicates a healthier ecosystem. The selection of metrics determines the reliability of the response. Selecting appropriate metrics such as environmental protection investment [119] and sewage treatment capacity [120] will influence the assessment results. It needs to be quantified if the chosen metric is not a

measurable variable, such as public awareness of environmental protection [121]. One problem is that the number of metrics is uneven. It does not mean that the reliability of the ecosystem's health assessment results is related to the number of metrics. Zhang [120] and Sun [122] assessed the urban ecosystem of Lanzhou City based on the PSR model. The former selected 20 metrics, while the latter selected only 15 metrics. However, there is little difference between the assessment results. Detailed comparisons are presented in Table S9. The quality of the ecosystem health index is related to the quality of the selected metrics, but not to the quantity [123].

3.2.2. Ecosystem Services

Ecosystem services (ES) are a significant metric that concerns sustainability [124]. ES, such as soil retention, water yield, and carbon storage, are determined by natural resources and habitats [125]. However, ES are often ignored when conducting a traditional EHA as such an assessment focuses more on the integrity and sustainability of the actual ecosystem itself. Therefore, it is necessary to enable a link between EH and the provision of ES, and to determine how any ecosystem dysfunction relates to these services when making an EHA [126]. In recent years, with the popularity of various methods to measure ES, researchers take ES as a metric of the index system, developed the VORS framework which combines ES with vigor, organization, and resilience [127–129]. Changes in EH directly or indirectly affect the output of ES [130]. It is important to focus on the changes in ES to gain a more comprehensive understanding of EH.

4. Application in Different Ecosystems

We adopted the Web of Science Core Collection Database to illustrate the application of the two methods in different ecosystems. Here are the search criteria: database: core collection of Web of Science, time: 1 January 2010, to 31 May 2021, paper type: Article, language: English, TI = ("Ecosystem health" OR "Ecosystem integrity" OR "Ecosystem quality" OR "Ecosystem status") AND ("assess*" OR "evaluate*" OR "measure*"). In addition, an asterisk (*) was inserted at the end of the words "assess", "evaluate", and "measure" to guarantee greater search precision. Only publications with full pdf format text were analyzed. The following data were recorded for analysis: (1) year, (2) journal, (3) study country, (4) ecosystem type, (5) methods, and (6) title. In total, 143 articles were found in the WOS result. After screening, a total of 127 publications belonging to 83 journals met the criteria, which could be all retrieved from the Scopus database. As expected, the main application methods of EHA are biological indicator and index system (Figure S3). Almost half of the studies utilized the biological indicator method in the current study. The index system method is also prevalent in EHA. Other methods take up nearly 10% of the study. For the application region (Figure S4), 65 studies were carried out in China, which is the most applied region. Applications in Asia, North America, and South America are significant and regional, while applications in Europe are small and dispersed. Most of the current studies were conducted in aquatic ecosystems (Figure 3a), and approximately one-third of the studies were conducted on terrestrial ecosystems. The total number of ecosystems was greater than the number of publications because one publication was carried out in aquatic and terrestrial ecosystems. The aquatic ecosystem was reclassified into lakes, reservoirs, lagoons, rivers, streams, marine coast area/Bay/Gulf, estuaries/delta, and wetland (Figure 3b), whereas the terrestrial ecosystem was reclassified into urban, forest, grassland, land, and desert (Figure 3c). Not surprisingly, rivers/streams were the most concerning targets, followed by marine coast area/bay/gulf. By comparison, urban, rivers/streams, and marine coast area/bay/gulf were at the same level; these were more frequently assessed than others.

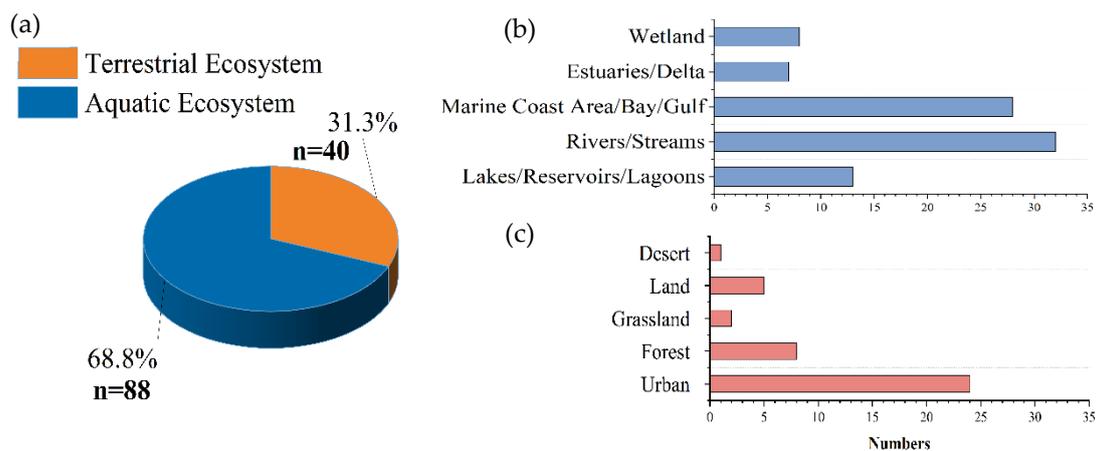


Figure 3. Summary of assessed ecosystem types in the publications found in the systematic review. All studies were classified by aquatic ecosystems and terrestrial ecosystem (a). The aquatic ecosystem (b) and terrestrial ecosystem (c) were also classified.

Among the studies that adopted the biological indicator method (Figure 4a), the majority of the studies were applicable to aquatic ecosystems, which accounted for 95.2% of the total publications. Rivers/streams are the ecosystems where the biological indicator methods were applied most frequently, followed by marine coast area/bay/gulf (Figure 4b). The terrestrial ecosystems where the biological indicator method was adopted accounted for 4.8% of the total publications. Wike and Bharti [97,131] both utilized the ant as a biological indicator in the study, while Jenssen [132] used lumbricoides to assess the soil properties. For studies that applied the index system method (Figure 5a), about 61.5% of the studies were on terrestrial ecosystems. Aquatic ecosystems accounted for 38.5% of the total publications. Among the studies in which the index system method was adopted with aquatic ecosystems, the majority of the studies (7) were on the marine coast area/bay/gulf (Figure 5b). Two studies utilized the PSR model in seven studies [119,133]. Three out of six studies applied the PSR model in wetland ecosystems. Among the 32 studies that adopted the index system method (Figure 5c), 24 were on urban ecosystems. Furthermore, the other eight included forest, grassland, land, and desert areas.

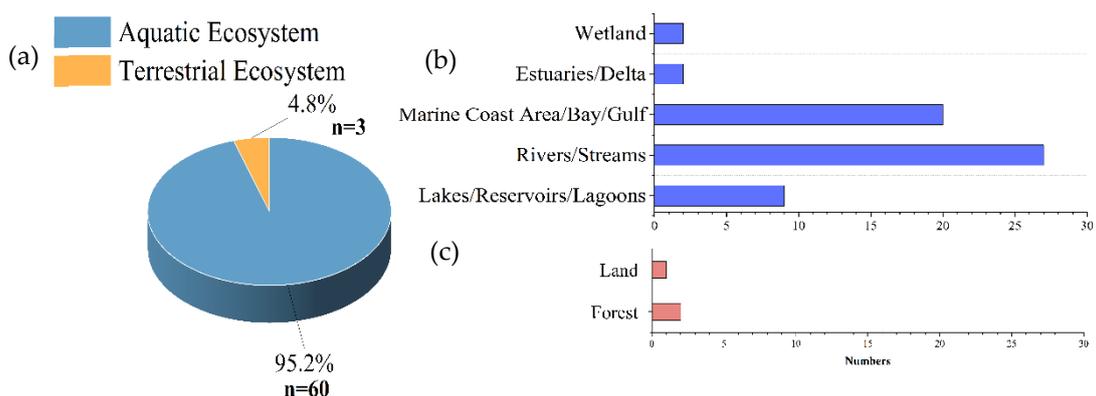


Figure 4. Summary of assessed ecosystems which adopted biological indicator method. The study was classified by aquatic ecosystem and terrestrial ecosystem (a). The aquatic ecosystem (b) and terrestrial ecosystem (c) were also classified.

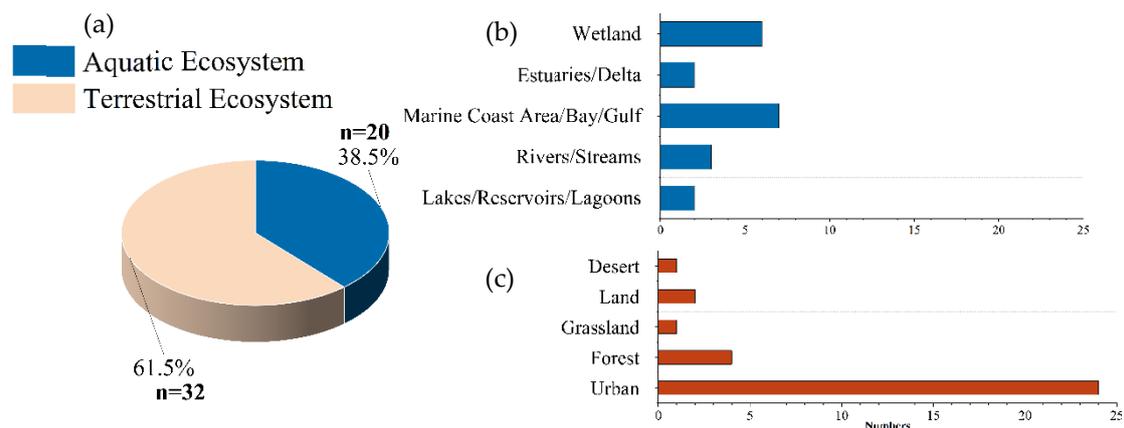


Figure 5. Summary of assessed ecosystems which adopted index system method. The study was classified by aquatic ecosystem and terrestrial ecosystem (a). The aquatic ecosystem (b) and terrestrial ecosystem (c) were also classified.

Additionally, a few researchers have attempted to innovate the EHA method [134,135]. Commonly assessed ecosystems are rivers/streams, marine coast area/bay/gulf, and urban areas. These ecosystems are heavily polluted and are highly associated with anthropogenic activities. The biological indicator method was mostly applied to aquatic ecosystems. In comparison, the index system method was applied in all kinds of ecosystems. However, urban ecosystems are more appropriate for index system methods.

5. Discussion

5.1. Biological Indicator Method

The biological indicator method focuses on environmental pollution and reflects impact of single factor to the ecosystem. The diatom, plankton, and macroinvertebrate indexes discussed in this paper are site-specific, which means that the index needs to be suitable for the conditions of the site to ensure an accurate assessment of ecosystem health. The Hilsenhoff biotic index (HBI) is a good index of organic pollutants [85] and can also be used as a metric of other indices [136]. However, HBI may not be applicable in other areas because the tolerance of species to pollutants may vary from place to place depending on the natural conditions of different habitats. The EPT index is applied to detect low levels of degradation because Ephemeroptera (mayflies), Plecoptera (stoneflies), and Trichoptera (caddisflies) are sensitive [137]. Several plankton and macroinvertebrate communities also have seasonal dynamic characteristics that would affect the community structure both upstream and downstream.

5.2. Index System Method

The index system method represents a comprehensive result. At present, we suggest that researchers focus on the index system method. It not only reflects pollution level but also indicates sustainable development. It should be a trend to adopt ecosystem services as a regular metric. A comprehensive index system should be developed as a standard [138]. Several researchers ignore the fact that the change in the ecosystem status has a certain lag, making it impossible for government departments to implement timely prevention and control. On the one hand, it makes it difficult for the study area to recover in a short time. On the other hand, it will also cause residents to suffer. Consequently, time scales should be considered based on assessment metrics. Moreover, researchers conducted the EHA, while the government departments controlled the collection and collation of data, which led to research inconvenience. Most of the spatial distribution of river basin ecosystems is across administrations units. The data available to the public are often divided by administrative unit, and the record indicators of each unit are not the same, resulting in the inability to obtain general information about the area. So far, no researchers have used two different methods to assess the ecosystem's health in the same area on the same time scale. It is

worth trying to combine the two methods to assess the same target area and compare it with the assessment result of a single method.

Supplementary Materials: The following are available online at <https://www.mdpi.com/article/10.3390/su132111838/s1>, Figure S1: The development of ecosystem health. Figure S2: Number of papers published annually from 2000 to 2020. (a) Publications annually from 2000 to 2020. (b) Publications change from 2000 to 2020. Figure S3: The number and ratio of publications which applied Biological Indicators, Index System, and Others. Figure S4: Geographic distribution of surveys found in the systematic review. The bubble sizes refer to the number of studies. Table S1: The comparisons between Biological Indicators Method and Index System Method. Table S2: The concept proposal stage. Table S3: Representative opinions of different researchers on EH. Table S4: The development stage. Table S5: List of diatom indices, origin, modifications, characteristics, application regions, and metrics. Table S6: List of plankton indices, origin, modifications, characteristics, application regions, and metrics. Table S7: List of macroinvertebrate indices, origin, modifications, characteristics, application regions, and metrics. Table S8: Models and frameworks developed by researchers based on VOR. Table S9: The metrics comparison of Zhang and Sun in EHA of Lanzhou City, China.

Author Contributions: Bibliographic retrieval, S.F., T.S., M.S., Y.H. and S.H.; writing—original draft preparation, S.F.; writing—review and editing, Z.Q. and Y.Z.; supervision, L.Z. All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Data Availability Statement: Web of Science Core Collection Database (<http://ifbicd85ae6022a1f4d78sbvbbck56c5o56cno.fiac.eds.tju.edu.cn/wos/woscc/basic-search>) (accessed on 26 July 2021).

Acknowledgments: This research was supported by the Tianjin Key Scientific and Technological Project (Grant No. 18ZXSZSF00240).

Conflicts of Interest: The authors declare no conflict of interest.

References

1. Chapman, P.M. Future Challenges for Marine Pollution Monitoring and Assessment. *Mar. Pollut. Bull.* **2015**, *95*, 1–2. [[CrossRef](#)] [[PubMed](#)]
2. Khim, J.S.; Wang, T.; Snyder, S.A. The Yellow Sea Ecosystem: Pollution, Ecosystem Threats, and Environmental Health. *Chemosphere* **2017**, *182*, 794–796. [[CrossRef](#)]
3. Carpenter, S.R.; DeFries, R.; Dietz, T.; Mooney, H.A.; Polasky, S.; Reid, W.V.; Scholes, R.J. Millennium Ecosystem Assessment: Research Needs. *Science* **2006**, *314*, 257–258.
4. Wolf, K.L.; Blahna, D.J.; Brinkley, W.; Romolini, M. Environmental Stewardship Footprint Research: Linking Human Agency and Ecosystem Health in the Puget Sound Region. *Urban Ecosyst.* **2013**, *16*, 13–32. [[CrossRef](#)]
5. Rapport, D.J.; Maffi, L. Eco-Cultural Health, Global Health, and Sustainability. *Ecol. Res.* **2011**, *26*, 1039–1049. [[CrossRef](#)] [[PubMed](#)]
6. Villns, A.; Norkko, J.; Hietanen, S.; Josefson, A.B.; Lukkari, K.; Norkko, A. The Role of Recurrent Disturbances for Ecosystem Multifunctionality. *Ecology* **2013**, *94*, 2275–2287. [[CrossRef](#)]
7. Hader, D.P.; Banaszak, A.T.; Villafane, V.E.; Narvarte, M.A.; Gonzalez, R.A.; Helbling, E.W. Anthropogenic Pollution of Aquatic Ecosystems: Emerging Problems with Global Implications. *Sci. Total Environ.* **2020**, *713*, 136586. [[CrossRef](#)]
8. Rocha, L.; Hegoburu, C.; Torremorell, A.; Feijoo, C.; Navarro, E.; Fernandez, H.R. Use of Ecosystem Health Indicators for Assessing Anthropogenic Impacts on Freshwaters in Argentina: A Review. *Environ. Monit. Assess.* **2020**, *192*, 611. [[CrossRef](#)]
9. Backhaus, T.; Snape, J.; Lazorchak, J. The impact of chemical pollution on biodiversity and ecosystem services: The need for an improved understanding. *Integr. Environ. Assess. Manag.* **2012**, *8*, 575–576. [[CrossRef](#)] [[PubMed](#)]
10. Aguirre-Rubí, J.R.; Ortiz-Zarragoitia, M.; Izagirre, U.; Etxebarria, N.; Espinoza, F.; Marigómez, I. Prospective Biomonitoring and Sentinel Bivalve Species for Pollution Monitoring and Ecosystem Health Disturbance Assessment in Mangrove-Lined Nicaraguan Coasts. *Sci. Total Environ.* **2019**, *649*, 186–200. [[CrossRef](#)]
11. Łuczyńska, J.; Paszczyk, B.; Łuczyński, M.J. Fish as a Bioindicator of Heavy Metals Pollution in Aquatic Ecosystem of Pluszne Lake, Poland, and risk Assessment for Consumer's Health. *Ecotoxicol. Environ. Saf.* **2018**, *153*, 60–67. [[CrossRef](#)]

12. Jung, J.-W.; Park, J.-W.; Eo, S.; Choi, J.; Song, Y.K.; Cho, Y.; Hong, S.H.; Shim, W.J. Ecological Risk Assessment of Microplastics in Coastal, Shelf, and Deep Sea Waters with a Consideration of Environmentally Relevant Size and Shape. *Environ. Pollut.* **2021**, *270*, 116217. [[CrossRef](#)] [[PubMed](#)]
13. Nava, V.; Leoni, B. A Critical Review of Interactions between Microplastics, Microalgae and Aquatic Ecosystem Function. *Water Res.* **2021**, *188*, 116476. [[CrossRef](#)]
14. Vanapalli, K.R.; Dubey, B.K.; Sarmah, A.K.; Bhattacharya, J. Assessment of Microplastic Pollution in the Aquatic Ecosystems—an Indian Perspective. *Case Stud. Chem. Environ. Eng.* **2021**, *3*, 100071. [[CrossRef](#)]
15. Chessman, B.; Williams, S.; Besley, C. Bioassessment of Streams with Macroinvertebrates: Effect of Sampled Habitat and Taxonomic Resolution. *J. N. Am. Benthol. Soc.* **2007**, *26*, 546–565. [[CrossRef](#)]
16. Forio, M.A.E.; Lock, K.; Radam, E.D.; Bande, M.; Asio, V.; Goethals, P.L. Assessment and Analysis of Ecological Quality, Macroinvertebrate Communities and Diversity in Rivers of a Multifunctional Tropical Island. *Ecol. Indic.* **2017**, *77*, 228–238. [[CrossRef](#)]
17. Costanza, R.; De Groot, R.; Sutton, P.; Van Der Ploeg, S.; Anderson, S.; Kubiszewski, I.; Farber, S.; Turner, R.K. Changes in the Global Value of Ecosystem Services. *Glob. Environ. Chang.* **2014**, *26*, 152–158. [[CrossRef](#)]
18. Rapport, D.J.; Costanza, R.; McMichael, A.J. Assessing Ecosystem Health. *Trends Ecol. Evol.* **1998**, *13*, 397–402. [[CrossRef](#)]
19. Costanza, R.; d’Arge, R.; De Groot, R.; Farber, S.; Grasso, M.; Hannon, B.; Limburg, K.; Naeem, S.; O’neill, R.V.; Paruelo, J. The Value of the World’s Ecosystem Services and Natural Capital. *Nature* **1997**, *387*, 253–260. [[CrossRef](#)]
20. Costanza, R. Ecosystem Health and Ecological Engineering. *Ecol. Eng.* **2012**, *45*, 24–29. [[CrossRef](#)]
21. Li, W.; Xie, S.; Wang, Y.; Huang, J.; Cheng, X. Effects of Urban Expansion on Ecosystem Health in Southwest China from a Multi-Perspective Analysis. *J. Clean. Prod.* **2021**, *294*, 126341. [[CrossRef](#)]
22. Curran, S.; Kumar, A.; Lutz, W.; Williams, M. Interactions between Coastal and Marine Ecosystems and Human Population Systems: Perspectives on How Consumption Mediates this Interaction. *Ambio* **2002**, *31*, 264–268. [[CrossRef](#)]
23. Moiseenko, T.; Voinov, A.; Megorsky, V.; Gashkina, N.; Kudriavtseva, L.; Vandish, O.; Sharov, A.; Sharova, Y.; Koroleva, I. Ecosystem and Human Health Assessment to Define Environmental Management Strategies: The Case of Long-Term Human Impacts on an Arctic Lake. *Sci. Total Environ.* **2006**, *369*, 1–20. [[CrossRef](#)]
24. Connell, D.J. Sustainable Livelihoods and Ecosystem Health: Exploring Methodological Relations as a Source of Synergy. *EcoHealth* **2010**, *7*, 351–360. [[CrossRef](#)]
25. An, K.-G.; Lee, J.-Y.; Kumar, H.K.; Lee, S.-J.; Hwang, S.-J.; Kim, B.-H.; Park, Y.-S.; Shin, K.-H.; Park, S.; Um, H.-Y. Control of Algal Scum Using Top-Down Biomanipulation Approaches and Ecosystem Health Assessments for Efficient Reservoir Management. *Water Air Soil Pollut.* **2009**, *205*, 3–24. [[CrossRef](#)]
26. Horwitz, P.; Finlayson, C. Wetlands as Settings for Human Health: Incorporating Ecosystem Services and Health Impact Assessment into Water Resource Management. *BioScience* **2011**, *61*, 678–688. [[CrossRef](#)]
27. Su, M.; Chen, B.; Yang, Z. Implication of Ecosystem Health Assessment for Urban Management. *Procedia Environ. Sci.* **2010**, *2*, 674–680. [[CrossRef](#)]
28. Su, M.; Yang, Z.; Chen, B.; Liu, G.; Zhang, Y.; Zhang, L.; Xu, L.; Zhao, Y. Urban Ecosystem Health Assessment and Its Application in Management: A Multi-Scale Perspective. *Entropy* **2012**, *15*, 1–9. [[CrossRef](#)]
29. Leopold, A. Wilderness as a Land Laboratory. *Living Wilderness* **1941**, *6*, 3.
30. Hodgson, J. *Grazing Management. Science into Practice*; Longman Group UK Ltd.: London, UK, 1990.
31. Bradshaw, A.D.; Barrett, G.W.; Rosenberg, R. Stress Effects on Natural Ecosystems. *J. Appl. Ecol.* **1982**, *19*, 988. [[CrossRef](#)]
32. Barrett, G.W.; Van Dyne, G.M.; Odum, E.P. Stress Ecology. *BioScience* **1976**, *26*, 192–194. [[CrossRef](#)]
33. Jernelöv, A.; Rosenberg, R. Stress Tolerance of Ecosystems. *Environ. Conserv.* **1976**, *3*, 43–46. [[CrossRef](#)]
34. Woodwell, G.M. Effects of Pollution on the Structure and Physiology of Ecosystems. *Science* **1970**, *168*, 429–433. [[CrossRef](#)]
35. Karr, J.; Fausch, K.; Angermeier, P. *Assessing Biological Integrity in Running Water Waters: A Method and Its Rational*; Illinois Natural History Survey Special Publication: Champaign, IL, USA, 1986; Volume 5.
36. Norton, B.G. A New Paradigm for Environmental Management. In *Ecosystem Health: New Goals for Environmental Management*; Island Press: Washington, DC, USA, 1992; pp. 23–41.
37. Reynoldson, T.B.; Metcalfe-Smith, J.L. An Overview of the Assessment of Aquatic Ecosystem Health Using Benthic Invertebrates. *J. Aquat. Ecosyst. Health* **1992**, *1*, 295–308. [[CrossRef](#)]
38. Marshall, F.E.; Banks, K.; Cook, G.S. Ecosystem Indicators for Southeast Florida Beaches. *Ecol. Indic.* **2014**, *44*, 81–91. [[CrossRef](#)]
39. Wu, Z.; Chen, R.; Meadows, M.E.; Liu, X. Application of the Ocean Health Index to Assess Ecosystem Health for the Coastal Areas of Shanghai, China. *Ecol. Indic.* **2021**, *126*, 107650. [[CrossRef](#)]
40. Oeding, S.; Taffs, K.H. Are Diatoms a Reliable and Valuable Bio-Indicator to Assess Sub-Tropical River Ecosystem Health? *Hydrobiologia* **2015**, *758*, 151–169. [[CrossRef](#)]
41. Tan, X.; Sheldon, F.; Bunn, S.E.; Zhang, Q. Using Diatom Indices for Water Quality Assessment in a Subtropical River, China. *Environ. Sci. Pollut. Res.* **2013**, *20*, 4164–4175. [[CrossRef](#)]
42. Xiang, Z.; Chen, H.; Li, C.; Yin, X.; Xu, Z.; Zhang, Y. Application of Diatom Index in Assessment of Aquatic Ecosystem Health in Taizi River, China. *J. Dalian Ocean. Univ.* **2016**, *31*, 416–425.
43. Bere, T. Challenges of Diatom-Based Biological Monitoring and Assessment of Streams in Developing Countries. *Environ. Sci. Pollut. Res.* **2016**, *23*, 5477–5486. [[CrossRef](#)] [[PubMed](#)]

44. Hering, D.; Johnson, R.K.; Kramm, S.; Schmutz, S.; Szoszkiewicz, K.; Verdonshot, P.F.M. Assessment of European Streams with Diatoms, Macrophytes, Macroinvertebrates and Fish: A Comparative Metric-Based Analysis of Organism Response to Stress. *Freshw. Biol.* **2006**, *51*, 1757–1785. [[CrossRef](#)]
45. Bere, T.; Tundisi, J.G. Applicability of Borrowed Diatom-Based Water Quality Assessment Indices in Streams around São Carlos-SP, Brazil. *Hydrobiologia* **2011**, *673*, 179–192. [[CrossRef](#)]
46. Rodríguez-Alcalá, O.; Blanco, S.; García-Girón, J.; Jeppesen, E.; Irvine, K.; Nöges, P.; Nöges, T.; Gross, E.M.; Bécares, E. Large-Scale Geographical and Environmental Drivers of Shallow Lake Diatom Metacommunities across Europe. *Sci. Total Environ.* **2020**, *707*, 135887. [[CrossRef](#)]
47. Taylor, J.C.; Prygiel, J.; Vosloo, A.; De La Rey, P.A.; Van Rensburg, L. Can diatom-Based Pollution Indices Be Used for Biomonitoring in South Africa? A Case Study of the Crocodile West and Marico Water Management Area. *Hydrobiologia* **2007**, *592*, 455–464. [[CrossRef](#)]
48. Vishnyakov, V.S.; Kulikovskiy, M.S.; Genkal, S.I.; Dorofeyuk, N.I.; Lange-Bertalot, H.; Kuznetsova, I.V. Taxonomy and Geographical Distribution of the Diatom Genus *Epithemia* Kützing in Water Bodies of Central Asia. *Inland Water Biol.* **2014**, *7*, 318–330. [[CrossRef](#)]
49. Xue, H.; Zheng, B.; Meng, F.; Wang, Y.; Zhang, L.; Cheng, P. Assessment of Aquatic Ecosystem Health of the Wutong River Based on Benthic Diatoms. *Water* **2019**, *11*, 727. [[CrossRef](#)]
50. Dalu, T.; Bere, T.; Froneman, P.W. Assessment of Water Quality Based on Diatom Indices in a Small Temperate River System, Kowie River, South Africa. *Water* **2016**, *42*, 183.
51. Rıdvan, E.S.; Sophia, B.; Cuuml neyt, N.S.; Kadir, C.O. Ecological Assessment of Great Lota Lake (Turkey) on the Base of Diatom Communities. *Afr. J. Biotechnol.* **2013**, *12*, 453–464. [[CrossRef](#)]
52. Bennion, H.; Kelly, M.G.; Juggins, S.; Yallop, M.L.; Burgess, A.; Jamieson, J.; Krokowski, J. Assessment of Ecological Status in UK Lakes Using Benthic Diatoms. *Freshw. Sci.* **2014**, *33*, 639–654. [[CrossRef](#)]
53. Kireta, A.R.; Reavie, E.D.; Sgro, G.V.; Angradi, T.R.; Bolgrien, D.W.; Jicha, T.M.; Hill, B.H. Assessing the Condition of the Missouri, Ohio, and Upper Mississippi Rivers (USA) using Diatom-Based Indicators. *Hydrobiologia* **2012**, *691*, 171–188. [[CrossRef](#)]
54. Desrosiers, C.; Leflaive, J.; Eulin, A.; Ten-Hage, L. Bioindicators in Marine Waters: Benthic Diatoms as a Tool to Assess Water Quality from Eutrophic to Oligotrophic Coastal Ecosystems. *Ecol. Indic.* **2013**, *32*, 25–34. [[CrossRef](#)]
55. Rimet, F. Recent Views on River Pollution and Diatoms. *Hydrobiologia* **2012**, *683*, 1–24. [[CrossRef](#)]
56. Lobo, M.T.M.P.S.; Scalize, P.S.; Kraus, C.N.; Da Silva, W.J.; Garnier, J.; Marques, D.D.M.; Bonnet, M.-P.; Nogueira, I.D.S. Biological Index Based on Epiphytic Diatom Assemblages is More Restrictive than the Physicochemical Index in Water Assessment on an Amazon Floodplain, Brazil. *Environ. Sci. Pollut. Res.* **2020**, *27*, 10642–10657. [[CrossRef](#)]
57. Stevenson, R.J.; Pan, Y.; Van Dam, H. Assessing Environmental Conditions in Rivers and Streams with Diatoms. *Diatoms Appl. Environ. Earth Sci.* **1999**, *1*, 57–85.
58. Feng, B.; Zhang, M.; Chen, J.; Xu, J.; Xiao, B.; Zhou, M.; Zhang, M. Reduction in the Phytoplankton Index of Biotic Integrity in Riverine Ecosystems Driven by Industrial Activities, Dam Construction and Mining: A Case Study in the Ganjiang River, China. *Ecol. Indic.* **2021**, *120*, 106907. [[CrossRef](#)]
59. Cai, W.; Zhou, Z.; Xia, J.; Wang, W.; Dou, C.; Zeng, Z. An Advanced Index of Ecological Integrity (IEI) for Assessing Ecological Efficiency of Restoration Revetments in River Plain. *Ecol. Indic.* **2020**, *108*, 105762. [[CrossRef](#)]
60. Falkowski, P.G.; Barber, R.T.; Smetacek, V. Biogeochemical Controls and Feedbacks on Ocean Primary Production. *Science* **1998**, *281*, 200–206. [[CrossRef](#)] [[PubMed](#)]
61. Scholz-Starke, B.; Bo, L.; Holbach, A.; Norra, S.; Floehr, T.; Hollert, H.; Roß-Nickoll, M.; Schäffer, A.; Ottermanns, R. Simulation-based Assessment of the Impact of Fertiliser and Herbicide Application on Freshwater Ecosystems at the Three Gorges Reservoir in China. *Sci. Total Environ.* **2018**, *639*, 286–303. [[CrossRef](#)]
62. O’Boyle, S.; McDermott, G.; Silke, J.; Cusack, C. Potential Impact of an Exceptional Bloom of *Karenia* Mikimotoi on Dissolved Oxygen Levels in Waters off Western Ireland. *Harmful Algae* **2016**, *53*, 77–85. [[CrossRef](#)] [[PubMed](#)]
63. Herrera-Silveira, J.A.; Morales-Ojeda, S.M. Evaluation of the Health Status of a Coastal Ecosystem in Southeast Mexico: Assessment of Water Quality, Phytoplankton and Submerged Aquatic Vegetation. *Mar. Pollut. Bull.* **2009**, *59*, 72–86. [[CrossRef](#)]
64. Liu, Y.; Yu, N.; Feng, D.; Xiong, Z.; Jiang, X.; Li, E.; Chen, L. Annual Variations of Zooplankton Community Structure in Shanghai Downtown Rivers. *Chin. J. Ecol.* **2010**, *29*, 370–376.
65. Beaugrand, G.; Edwards, M.; Legendre, L. Marine Biodiversity, Ecosystem Functioning, and Carbon Cycles. *Proc. Natl. Acad. Sci. USA* **2010**, *107*, 10120–10124. [[CrossRef](#)] [[PubMed](#)]
66. Bucklin, A.; Lindeque, P.; Rodriguez-Ezpeleta, N.; Albaina, A.; Lehtiniemi, M. Metabarcoding of Marine Zooplankton: Prospects, Progress and Pitfalls. *J. Plankton Res.* **2016**, *38*, 393–400. [[CrossRef](#)]
67. Shao, N.F.; Yang, S.T.; Sun, Y.; Gai, Y.; Zhao, C.S.; Wang, F.; Yin, X.; Dong, B. Assessing Aquatic Ecosystem Health through the Analysis of Plankton Biodiversity. *Mar. Freshw. Res.* **2019**, *70*, 647. [[CrossRef](#)]
68. Davidson, T.A.; Bennion, H.; Jeppesen, E.; Clarke, G.H.; Sayer, C.D.; Morley, D.; Odgaard, B.V.; Rasmussen, P.; Rawcliffe, R.; Salgado, J.; et al. The Role of Cladocerans in Tracking Long-Term Change in Shallow Lake Trophic Status. *Hydrobiologia* **2011**, *676*, 299–315. [[CrossRef](#)]

69. Jeppesen, E.; Nøges, P.; Davidson, T.; Haberman, J.; Nøges, T.; Blank, K.; Lauridsen, T.; Søndergaard, M.; Sayer, C.; Laugaste, R.; et al. Zooplankton as Indicators in Lakes: A Scientific-Based Plea for Including Zooplankton in the Ecological Quality Assessment of Lakes According to the European Water Framework Directive (WFD). *Hydrobiologia* **2011**, *676*, 279–297. [[CrossRef](#)]
70. Irvine, K.; Caroni, R. The Potential of Zooplankton Communities for Ecological Assessment of Lakes: Redundant Concept or Political Oversight? In *Biology and Environment: Proceedings of the Royal Irish Academy*; Royal Irish Academy: Dublin, Ireland, 2010; Volume 110, pp. 35–53.
71. Baek, S.H.; Son, M.; Kim, D.; Choi, H.-W.; Kim, Y.-O. Assessing the Ecosystem Health Status of Korea Gwangyang and Jinhae bays based on a Planktonic Index of Biotic Integrity (P-IBI). *Ocean Sci. J.* **2014**, *49*, 291–311. [[CrossRef](#)]
72. Houssou, A.M.; Adjahouinou, D.C.; Bonou, C.A.; Montchowui, E. Plankton Index of Biotic Integrity (P-IBI) for Assessing Ecosystem Health within the Ouémé River Basin, Republic of Benin. *Afr. J. Aquat. Sci.* **2020**, *45*, 452–465. [[CrossRef](#)]
73. Kane, D.D.; Gordon, S.I.; Munawar, M.; Charlton, M.N.; Culver, D.A. The Planktonic Index of Biotic Integrity (P-IBI): An Approach for Assessing Lake Ecosystem Health. *Ecol. Indic.* **2009**, *9*, 1234–1247. [[CrossRef](#)]
74. Rivera, I.N.; Souza, K.M.; Souza, C.P.; Lopes, R.M. Free-Living and Plankton-Associated Vibrios: Assessment in Ballast Water, Harbor Areas, and Coastal Ecosystems in Brazil. *Front. Microbiol.* **2012**, *3*, 443. [[CrossRef](#)] [[PubMed](#)]
75. Tett, P.; Carreira, C.; Mills, D.K.; Van Leeuwen, S.; Foden, J.; Bresnan, E.; Gowen, R.J. Use of a Phytoplankton Community Index to Assess the Health of Coastal Waters. *ICES J. Mar. Sci.* **2008**, *65*, 1475–1482. [[CrossRef](#)]
76. Whyte, C.; Davidson, K.; Gilpin, L.; Mitchell, E.; Moschonas, G.; McNeill, S.; Tett, P. Tracking Changes to a Microplankton Community in a North Atlantic Sea Loch Using the Microplankton Index PI(mp). *ICES J. Mar. Sci.* **2017**, *74*, 311–325. [[CrossRef](#)]
77. Hou, Y.; Kong, F.; Li, Y.; Xi, M.; Yu, Z. Key Factors of the Studies on Benthic Macroinvertebrate in Coastal Wetlands: Methods and Biodiversity. *Ecohydrol. Hydrobiol.* **2020**, *20*, 424–436. [[CrossRef](#)]
78. Fierro, P.; Bertrán, C.; Tapia, J.; Hauenstein, E.; Peña-Cortés, F.; Vergara, C.; Cerna, C.; Vargas-Chacoff, L. Effects of Local Land-Use on Riparian Vegetation, Water Quality, and the Functional Organization of Macroinvertebrate Assemblages. *Sci. Total Environ.* **2017**, *609*, 724–734. [[CrossRef](#)]
79. Gogina, M.; Zettler, M.L. Diversity and Distribution of Benthic Macrofauna in the Baltic Sea: Data Inventory and its Use for Species Distribution Modelling and Prediction. *J. Sea Res.* **2010**, *64*, 313–321. [[CrossRef](#)]
80. Noman, A.; Mamunur, R.; Islam, M.S.; Hossain, M.B. Spatial and Seasonal Distribution of Intertidal Macrobenthos with their Biomass and Functional Feeding Guilds in the Naf River Estuary, Bangladesh. *J. Oceanol. Limnol.* **2019**, *37*, 1010–1023. [[CrossRef](#)]
81. Guimarães, A.T.B.; Rodrigues, A.S.D.L.; Pereira, P.S.; Silva, F.G.; Malafaia, G. Toxicity of Polystyrene Nanoplastics in Dragonfly Larvae: An Insight on how these Pollutants Can Affect Benthic Macroinvertebrates. *Sci. Total Environ.* **2021**, *752*, 141936. [[CrossRef](#)]
82. Herman, M.R.; Nejadhashemi, A.P. A Review of Macroinvertebrate—And Fish-Based Stream Health Indices. *Ecohydrol. Hydrobiol.* **2015**, *15*, 53–67. [[CrossRef](#)]
83. Ollis, D.J.; Dallas, H.; Esler, K.; Boucher, C. Bioassessment of the Ecological Integrity of River Ecosystems Using Aquatic Macroinvertebrates: An Overview with a Focus on South Africa. *Afr. J. Aquat. Sci.* **2010**, *31*, 205–227. [[CrossRef](#)]
84. Rakocinski, C.F. Evaluating Macrobenthic Process Indicators in Relation to Organic Enrichment and Hypoxia. *Ecol. Indic.* **2012**, *13*, 1–12. [[CrossRef](#)]
85. Butcher, J.T.; Stewart, P.M.; Simon, T.P. A Benthic Community Index for Streams in the Northern Lakes and Forests Ecoregion. *Ecol. Indic.* **2003**, *3*, 181–193. [[CrossRef](#)]
86. Houston, L.; Barbour, M.; Lenat, D.; Penrose, D. A Multi-Agency Comparison of Aquatic Macroinvertebrate-Based Stream Bioassessment Methodologies. *Ecol. Indic.* **2002**, *1*, 279–292. [[CrossRef](#)]
87. Breine, J.; Quataert, P.; Stevens, M.; Ollevier, F.; Volckaert, F.A.; Bergh, E.V.D.; Maes, J. A Zone-Specific Fish-Based Biotic Index as a Management Tool for the Zeeschelde Estuary (Belgium). *Mar. Pollut. Bull.* **2010**, *60*, 1099–1112. [[CrossRef](#)] [[PubMed](#)]
88. Martinho, F.; Nyitrai, D.; Crespo, D.; Pardal, M.A. Efficacy of Single and Multi-Metric Fish-Based Indices in Tracking Anthropogenic Pressures in Estuaries: An 8-Year Case Study. *Mar. Pollut. Bull.* **2015**, *101*, 153–162. [[CrossRef](#)]
89. Viana, A.P.; Frédou, F.L.; Frédou, T. Measuring the Ecological Integrity of an Industrial District in the Amazon Estuary, Brazil. *Mar. Pollut. Bull.* **2012**, *64*, 489–499. [[CrossRef](#)] [[PubMed](#)]
90. Pérez-Domínguez, R.; Maci, S.; Courrat, A.; Lepage, M.; Borja, A.; Uriarte, A.; Neto, J.M.; Cabral, H.; Raykov, V.; Franco, A.; et al. Current Developments on Fish-Based Indices to Assess Ecological-Quality Status of Estuaries and Lagoons. *Ecol. Indic.* **2012**, *23*, 34–45. [[CrossRef](#)]
91. Van Hoey, G.; Borja, A.; Birchenough, S.; Buhl-Mortensen, L.; Degraer, S.; Fleischer, D.; Kerckhof, F.; Magni, P.; Muxika, I.; Reiss, H.; et al. The Use of Benthic Indicators in Europe: From the Water Framework Directive to the Marine Strategy Framework Directive. *Mar. Pollut. Bull.* **2010**, *60*, 2187–2196. [[CrossRef](#)] [[PubMed](#)]
92. Ciadamidaro, S.; Mancini, L.; Rivoecchi, L. Black flies (Diptera, Simuliidae) as Ecological Indicators of Stream Ecosystem Health in an Urbanizing Area (Rome, Italy). *Ann. Dell'ist. Super. Sanita* **2016**, *52*, 269–276. [[CrossRef](#)]
93. Aguirre-Rubí, J.; Luna-Acosta, A.; Ortiz-Zarragoitia, M.; Zaldibar, B.; Izagirre, U.; Ahrens, M.J.; Villamil, L.; Marigómez, I. Assessment of Ecosystem Health Disturbance in Mangrove-Lined Caribbean Coastal Systems Using the Oyster *Crassostrea Rhizophorae* as Sentinel Species. *Sci. Total Environ.* **2018**, *618*, 718–735. [[CrossRef](#)]

94. Aguirre, A.A.; Lutz, P. Marine Turtles as Sentinels of Ecosystem Health: Is Fibropapillomatosis an Indicator? *EcoHealth* **2004**, *1*, 275–283. [[CrossRef](#)]
95. Montefalcone, M. Ecosystem Health Assessment Using the Mediterranean Seagrass *Posidonia Oceanica*: A Review. *Ecol. Indic.* **2009**, *9*, 595–604. [[CrossRef](#)]
96. Andersen, A.N.; Majer, J.D. Ants Show the Way Down Under: Invertebrates as Bioindicators in Land Management. *Front. Ecol. Environ.* **2004**, *2*, 291–298. [[CrossRef](#)]
97. Wike, L.D.; Martin, F.D.; Paller, M.H.; Nelson, E.A. Impact of Forest Seral Stage on use of Ant Communities for Rapid Assessment of Terrestrial Ecosystem Health. *J. Insect Sci.* **2010**, *10*, 1–16. [[CrossRef](#)] [[PubMed](#)]
98. Sonne, C.; Letcher, R.J.; Jenssen, B.M.; Desforges, J.-P.; Eulaers, I.; Andersen-Ranberg, E.; Gustavson, K.; Bossi, R.; Styrihave, B.; Sinding, M.-H.S.; et al. Sled Dogs as Sentinel Species for Monitoring Arctic Ecosystem Health. In *Pets as Sentinels, Forecasters and Promoters of Human Health*; Springer: Cham, Switzerland, 2020; pp. 21–45. [[CrossRef](#)]
99. Costanza, R. Toward an Operational Definition of Ecosystem Health. *Ecosyst. Health: New Goals Environ. Manag.* **1992**, *239*, 269.
100. Xu, F.-L.; Zhao, Z.-Y.; Zhan, W.; Zhao, S.-S.; Dawson, R.; Tao, S. An Ecosystem Health Index Methodology (EHIM) for Lake Ecosystem Health Assessment. *Ecol. Model.* **2005**, *188*, 327–339. [[CrossRef](#)]
101. Su, M.; Fath, B.D.; Yang, Z. Urban Ecosystem Health Assessment: A Review. *Sci. Total Environ.* **2010**, *408*, 2425–2434. [[CrossRef](#)]
102. Ludwig, J.A.; Bastin, G.N.; Chewings, V.H.; Eager, R.W.; Liedloff, A.C. Leakiness: A New Index for Monitoring the Health of Arid and Semiarid Landscapes Using Remotely Sensed Vegetation Cover and Elevation Data. *Ecol. Indic.* **2007**, *7*, 442–454. [[CrossRef](#)]
103. Qiao, Z.; Wu, C.; Zhao, D.; Xu, X.; Yang, J.; Feng, L.; Sun, Z.; Liu, L. Determining the Boundary and Probability of Surface Urban Heat Island Footprint Based on a Logistic Model. *Remote Sens.* **2019**, *11*, 1368. [[CrossRef](#)]
104. Kerr, J.T.; Ostrovsky, M. From Space to Species: Ecological Applications for Remote Sensing. *Trends Ecol. Evol.* **2003**, *18*, 299–305.
105. Hilker, T.; Coops, N.C.; Wulder, M.A.; Black, T.A.; Guy, R.D. The Use of Remote Sensing in Light Use Efficiency Based Models of Gross Primary Production: A Review of Current Status and Future Requirements. *Sci. Total Environ.* **2008**, *404*, 411–423.
106. Qiao, Z.; Liu, L.; Qin, Y.; Xu, X.; Wang, B.; Liu, Z. The Impact of Urban Renewal on Land Surface Temperature Changes: A Case Study in the Main City of Guangzhou, China. *Remote Sens.* **2020**, *12*, 794. [[CrossRef](#)]
107. Cao, C.; Xu, M.; Chen, W.; Tian, R. A Framework for Diagnosis of Environmental Health based on Remote Sensing. In Proceedings of the Conference on Land Surface Remote Sensing, Kyoto, Japan, 21 November 2012.
108. Li, Z.; Xu, D.; Guo, X. Remote Sensing of Ecosystem Health: Opportunities, Challenges, and Future Perspectives. *Sensors* **2014**, *14*, 21117–21139. [[CrossRef](#)]
109. Grecchi, R.C.; Gwyn, Q.H.J.; Bénié, G.B.; Formaggio, A.R.; Fahl, F.C. Land Use and Land Cover Changes in the Brazilian Cerrado: A Multidisciplinary Approach to Assess the Impacts of Agricultural Expansion. *Appl. Geogr.* **2014**, *55*, 300–312. [[CrossRef](#)]
110. Sun, T.; Lin, W.; Chen, G.; Guo, P.; Zeng, Y. Wetland Ecosystem Health Assessment through Integrating Remote Sensing and Inventory Data with an Assessment Model for the Hangzhou Bay, China. *Sci. Total Environ.* **2016**, *566–567*, 627–640. [[CrossRef](#)] [[PubMed](#)]
111. Wang, H.; Hou, P.; Jiang, J.; Xiao, R.; Zhai, J.; Fu, Z.; Hou, J. Ecosystem Health Assessment of Shennongjia National Park, China. *Sustainability* **2020**, *12*, 7672. [[CrossRef](#)]
112. Yue, H.; Liu, Y.; Li, Y.; Lu, Y. Eco-Environmental Quality Assessment in China's 35 Major Cities Based on Remote Sensing Ecological Index. *IEEE Access* **2019**, *7*, 51295–51311. [[CrossRef](#)]
113. Pan, W.; Huang, H.; Yao, P.; Zheng, P. Assessment Methods of Small Watershed Ecosystem Health. *Pol. J. Environ. Stud.* **2021**, *30*, 1749–1769. [[CrossRef](#)]
114. Shen, W.; Zheng, Z.; Pan, L.; Qin, Y.; Li, Y. A Integrated Method For Assessing The Urban Ecosystem Health Of Rapid Urbanized Area in China based on SFPHD Framework. *Ecol. Indic.* **2020**, *121*, 107071. [[CrossRef](#)]
115. Bell, S. DPSIR=A Problem Structuring Method? An Exploration from the “Imagine” Approach. *Eur. J. Oper. Res.* **2012**, *222*, 350–360. [[CrossRef](#)]
116. Gregory, A.J.; Atkins, J.P.; Burdon, D.; Elliott, M. A Problem Structuring Method for Ecosystem-Based Management: The DPSIR Modelling Process. *Eur. J. Oper. Res.* **2013**, *227*, 558–569. [[CrossRef](#)]
117. Ramos, T.B.; Alves, I.; Subtil, R.; de Melo, J.J. Environmental Performance Policy Indicators for the Public Sector: The Case of the Defence Sector. *J. Environ. Manag.* **2007**, *82*, 410–432. [[CrossRef](#)] [[PubMed](#)]
118. Song, Q.; Wang, H.; Wen, F.; Ledwich, G.; Xue, Y. Pressure State Response-Based Method for Evaluating Social Benefits from Smart Grid Development. *J. Energy Eng.* **2015**, *141*, 04014020. [[CrossRef](#)]
119. Sun, B.; Tang, J.; Yu, D.; Song, Z.; Wang, P. Ecosystem Health Assessment: A PSR Analysis Combining AHP and FCE Methods for Jiaozhou Bay, China. *Ocean Coast. Manag.* **2019**, *168*, 41–50. [[CrossRef](#)]
120. Zhang, X.; Shi, P. The Assessment of Urban Ecosystem Health Based on PSR Model-A Case Study of Lanzhou City. *J. Arid. Land Resour. Environ.* **2010**, *24*, 77–82.
121. Wang, Y.-T.; Wang, Y.-S.; Wu, M.-L.; Sun, C.-C.; Gu, J.-D. Assessing Ecological Health of Mangrove Ecosystems along South China Coast by the Pressure–State–Response (PSR) Model. *Ecotoxicology* **2021**, *30*, 622–631. [[CrossRef](#)]
122. Sun, J.; Wei, F. The Health Assessment of the Urban Ecosystem and Analysis of Coordination for the City of Lanzhou. *J. Geo. Inf. Sci.* **2017**, *19*, 511–517.
123. Niu, M.; Wang, J. Discussion over Health Assessment Indicator System of Ecosystem in Yellow River Estuary Area. *Water Resour. Prot.* **2016**, *32*, 57–63.

124. Zhao, B.; Kreuter, U.; Li, B.; Ma, Z.; Chen, J.; Nakagoshi, N. An ecosystem service value assessment of land-use change on Chongming Island, China. *Land Use Policy* **2004**, *21*, 139–148. [[CrossRef](#)]
125. Butler, C.; Oluoch-Kosura, W. Linking Future Ecosystem Services and Future Human Well-being. *Ecol. Soc.* **2006**, *11*, 30. [[CrossRef](#)]
126. Pan, Z.; He, J.; Liu, D.; Wang, J.; Guo, X. Ecosystem Health Assessment Based on Ecological Integrity and Ecosystem Services Demand in the Middle Reaches of the Yangtze River Economic Belt, China. *Sci. Total Environ.* **2021**, *774*, 144837. [[CrossRef](#)]
127. Peng, J.; Liu, Y.; Wu, J.; Lv, H.; Hu, X. Linking Ecosystem Services and Landscape Patterns to Assess Urban Ecosystem Health: A Case Study in Shenzhen City, China. *Landsc. Urban Plan.* **2015**, *143*, 56–68. [[CrossRef](#)]
128. Liu, R.; Dong, X.; Zhang, P.; Zhang, Y.; Wang, X.; Gao, Y. Study on the Sustainable Development of an Arid Basin Based on the Coupling Process of Ecosystem Health and Human Wellbeing Under Land Use Change—A Case Study in the Manas River Basin, Xinjiang, China. *Sustainability* **2020**, *12*, 1201. [[CrossRef](#)]
129. Mallick, J.; AlQadhi, S.; Talukdar, S.; Pradhan, B.; Bindajam, A.; Islam, A.; Dajam, A. A Novel Technique for Modeling Ecosystem Health Condition: A Case Study in Saudi Arabia. *Remote Sens.* **2021**, *13*, 2632. [[CrossRef](#)]
130. Toro, I.D.; Ribbons, R.R.; Pelini, S.L. The Little Things that Run the World Revisited: A Review of Ant-Mediated Ecosystem Services and Disservices (Hymenoptera: Formicidae). *Myrmecol. News* **2014**, *17*, 133–146.
131. Bharti, H.; Bharti, M.; Pfeiffer, M. Ants as Bioindicators of Ecosystem Health in Shivalik Mountains of Himalayas: Assessment of Species Diversity and Invasive Species. *Asian Myrmecol.* **2016**, *8*, 65–79.
132. Jenssen, M.; Nickel, S.; Schütze, G.; Schröder, W. Reference States of Forest Ecosystem Types and Feasibility of Biocenotic Indication of Ecological Soil Condition as Part of Ecosystem Integrity and Services Assessment. *Environ. Sci. Eur.* **2021**, *33*, 1–18. [[CrossRef](#)]
133. Harwell, M.A.; Gentile, J.H.; McKinney, L.D.; Tunnell, J.W.; Dennison, W.C.; Kelsey, R.; Stanzel, K.M.; Stunz, G.W.; Withers, K.; Tunnell, J. Conceptual Framework for Assessing Ecosystem Health. *Integr. Environ. Assess. Manag.* **2019**, *15*, 544–564. [[CrossRef](#)]
134. Mariano, D.A.; dos Santos, C.A.; Wardlow, B.; Anderson, M.C.; Schiltmeyer, A.V.; Tadesse, T.; Svoboda, M.D. Use of Remote Sensing Indicators to Assess Effects of Drought and Human-Induced Land Degradation on Ecosystem Health in Northeastern Brazil. *Remote Sens. Environ.* **2018**, *213*, 129–143. [[CrossRef](#)]
135. Stoll, S.; Frenzel, M.; Burkhard, B.; Adamescu, M.; Augustaitis, A.; Baeßler, C.; Bonet, F.J.; Carranza, M.L.; Cazacu, C.; Cosor, G.L.; et al. Assessment of Ecosystem Integrity and Service Gradients Across Europe Using the LTER Europe Network. *Ecol. Model.* **2015**, *295*, 75–87. [[CrossRef](#)]
136. Haase, R.; Nolte, U. The Invertebrate Species Index (ISI) for Streams in Southeast Queensland, Australia. *Ecol. Indic.* **2008**, *8*, 599–613. [[CrossRef](#)]
137. Couceiro, S.; Hamada, N.; Forsberg, B.; Pimentel, T.; Luz, S. A Macroinvertebrate Multimetric Index to Evaluate the Biological Condition of Streams in the Central Amazon Region of Brazil. *Ecol. Indic.* **2012**, *18*, 118–125. [[CrossRef](#)]
138. Costanza, R.; Patten, B.C. Defining and Predicting Sustainability. *Ecol. Econ.* **1995**, *15*, 193–196. [[CrossRef](#)]