



A Novel Low-Cost Photocatalyst: Preparation, Characterization, and Photocatalytic Properties of CeO₂-Diatomite Composites

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Article

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Abstract: Developing CeO₂-diatomite composites with highly efficient photocatalytic performance is a practical and low-cost strategy for the removal of abundant contaminants in water and wastewaters. Diatomite (D) was modified by acid treatment to obtain a more porous structure. CeO₂-diatomite composites were prepared in two different mass ratios of D to CeO₂ of 1:1 and 1:2 via a facile precipitation method. The changes in structural, morphological, optical, and thermal properties of CeO₂-diatomite composites were characterized by FTIR, XRD, ESEM-EDAX, BET surface area, TGA, PL, Raman spectroscopy, and zeta potential techniques. ESEM images presented the morphological differences of CeO₂-diatomite composites, reflecting the effect of modification as a more folded sheet form morphology and higher BET surface area. XRD analysis revealed the fluorite-type structure of CeO₂ particles in composites. Photocatalytic activities were investigated by following the degradation of methylene blue (MB) as a thiazine dye model under UVA light irradiation. CeO₂-diatomite composites exhibited irradiation time- and dose-dependent remarkable photocatalytic efficiencies, whereas composite type inconsistent variations were also noticed. The attained performance of the CeO₂-diatomite composites could strongly imply a favorable application prospect in the photocatalysis field.

Keywords: CeO₂-diatomite composites; diatomite; methylene blue; precipitation method; photocatalysis

1. Introduction

The lack of efficient wastewater treatment and the spreading of hazardous chemicals including dyes from industries into water resources is a major global issue [1]. The textile industry consumes the most dye, followed by the various dye-related industries, such as printing, and the production of paper, paint, and leather for various manufacturing purposes. Among the different textile dyes, MB is widely used as a cationic and primary thiazine dye for coloring silk, wool, cotton, and paper [2,3]. The accumulation of MB in receiving waters can have a harmful impact on the ecological system. Therefore, the removal of MB from wastewater effluent is highly crucial prior to the direct discharging into the water bodies [1,3,4]. To date, numerous approaches have been reported for the elimination of MB from wastewaters, including coagulation, flocculation, biodegradation, and adsorption treatments [3,5-7]. However, these wastewater treatments are generally not able to achieve complete dye degradation, even with a sequential combination of two or three methods. Therefore, it is necessary to focus on an alternative and efficient water treatment technique having a low cost [8-11]. Advanced oxidation processes, specifically heterogeneous photocatalysis, may degrade complex organic pollutants via reactive oxygen species (ROS) operating mainly through hydroxyl radical attack [12,13].

Recently, new catalysts using either semiconductors such as CdS [14], CuO [15], MgO [16], SnO₂[17], and Al-doped CoFe₂O₄[18], or composites such as PANI-ZnO [19]

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Copyright: © 2022 by the author. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https://creativecommons.org/license s/by/4.0/). and Ag@TiO₂/W₃[20], were developed for the photocatalytic degradation of MB. A rare earth oxide, CeO₂, received attention as a promising catalyst since it is non-toxic, non-photocorrosive, and low-cost, and has a strong oxygen ability with high chemical stability properties [21–23].

The primary events of a photocatalytic degradation process using a photocatalyst, e.g., CeO₂, can be expressed by the reactions given below (Reactions (1-6)):

Charge separation: Upon irradiation of CeO₂, a hole in the valence band (VB, h^+_{VB}) and an electron in the conduction band (CB, e^-_{CB}) are generated (Reaction (1)).

$$CeO_2 + h\nu (E > E_{bg}) \rightarrow CeO_2 (e^{-}CB + h^{+} \nu_B)$$
(1)

These charge carriers react with either the H₂O molecule (Reaction (2)) or with the hydroxyl ion OH⁻ (Reaction (3)) to form hydroxyl radical (•OH).

•OH formation via h ⁺ v_B:

$$h^{+}_{VB} + H_{2}O \rightarrow H^{+} + {}^{\bullet}OH$$
(2)

$$h + v_{\rm B} + OH^{-} \rightarrow {}^{\bullet}OH \tag{3}$$

Reduction via e-cb:

$$e^{-}CB + O_2 \rightarrow {}^{\bullet}O_2^{-} \tag{4}$$

Reaction (4) defines a reduction process indicating formation of a superoxide anion ($^{\circ}O_{2^{-}}$) followed by formation of H₂O₂ (Reactions (5) and (6)).

$$^{\bullet}\text{O}_{2^{-}} + \text{H}_{2}\text{O}_{2} \rightarrow ^{\bullet}\text{OH} + \text{O}_{2} + \text{HO}_{-}$$
(5)

$$^{\bullet}OH + ^{\bullet}OH \rightarrow H_2O_2 \tag{6}$$

Degradation of organic pollutants can be achieved through action of ROS (i.e., OH, O2⁻, H2O2, etc.) proceeding through various consecutive steps and eventually resulting in formation of innocuous products as CO2, H2O, and inorganic ions (Reaction (7)) [23,24].

$$ROS + organic pollutants \rightarrow \rightarrow CO_2 + H_2O$$
 (7)

Recently, some researchers reported the photodegradation of MB using CeO₂ as a catalyst to solve the problem of water pollution [25–30]. Considerable efforts were made to enhance CeO₂ adsorption with the addition of SiO₂ nanoparticles [23,31–35]. Modification of the morphology with SiO₂ was an advantageous step for successful separation of the photogenerated charge carriers. It is well-known that the use of SiO₂ emerged as a promising strategy to obtain a more uniform distribution and porosity of the final product [31,35].

In the past, D was used to prepare new natural silica mineral carrier materials for wastewater treatment [36–38]. Raw D is a unique ordered porous silica material with a large specific surface area, low density, high adsorption capacity, and chemical resistance. This low-cost porous material is derived from the skeleton fossils of unicellular aquatic algae known as diatoms having various shapes and sizes and consisting of amorphous silica [36,39]. Raw D consists of mineral impurities that reduce the original structural characteristics. Therefore, the removal of the existing undesired impurities is sometimes crucial to enhance the pore-size distribution and adsorption properties of D. Accordingly, acid treatment is applied as a simple and cheap approach, resulting in a modified diatomite (MD) [39–41]. Several studies have focused on the photocatalytic performance of MB using various catalysts including TiO₂-diatomite composites [42,43], Zn-diatomite composite [44], SnO₂-diatomite composite [39], CdS-diatomite composite [45], and TiO₂-ZnO-diatomite composite [46]. However, there were very few studies on the development of CeO₂-diatomite composites. Zhou and colleagues prepared a Pt single-atom catalyst on CeO₂-modified diatomite support for the hydrogenation of phe-

nylacetylene to styrene [47]. Su and co-workers proved that the synthesized modified nano-CeO₂-filled oily diatomite/polyvinylidene fluoride composites exhibited tribological behavior [48]. Therefore, this study addressed the photocatalytic performance of CeO₂-diatomite composite, which represents a significant contribution.

CeO₂-diatomite composites were synthesized via a facile precipitation method considering the photocatalytic activity of CeO₂ with the adsorption ability of diatomite as the representative material of porous silica. The present work dealt with the preparation of CeO₂-diatomite composites using raw D and MD in two different mass ratios to understand the relationship between the pretreatment method and resulting composites. Furthermore, the possible morphological, structural, optical, and thermal differences depending on the use of the diatomite type in the composite were systematically investigated. The photocatalytic properties of the prepared CeO₂-diatomite composites via two different routes were evaluated by testing MB dye for the first time.

2. Materials and Methods

2.1. Materials

Diatomite (average particle size = 5.2 µm) was used as supplied from Beg Tug Mineral, Türkiye. Chemically, D was mainly composed of SiO₂ (75.45%), inorganic constituents such as Al₂O₃ (7.15%), CaO (3.05%), Fe₂O₃ (2.35%), MgO (1.45%), and TiO₂ (0.40%), and some organic impurities. Ce(NH₄)₂(NO₃)₆, NaOH, and HCl (37%) were purchased from Sigma-Aldrich and used without further purification. All aqueous solutions were prepared using distilled water (conductivity 2×10^{-6} S/m at 25 °C).

2.2. Preparation of Modified Diatomite

Acid-modified diatomite was prepared according to the procedure reported by Jiang and colleagues [39]. A quantity of 10 g diatomite was added to 100 mL of 1 M HCl to prepare a suspension and placed in a water bath shaker for 2 h. Then, MD was filtered through a Gooch funnel equipped with a sintered glass disc, followed by extensive washings with distilled water, dried at 105 °C for 24 h, and finally calcined in a muffle furnace at 450 °C in air for 1h.

2.3. Preparation of CeO₂-Diatomite Composites

CeO₂-diatomite composites were prepared via a modified co-precipitation method [49]. The theoretical diatomite/CeO₂ mass ratios were 1:1 and 1:2 of raw D or MD to CeO₂, and these four composites were denoted as DC-11, DC-12 MDC-11, and MDC-12, respectively. The representative synthesis of DC-11 was given as follows: 1.72 g raw D was added to 100 mL of 0.1 M Ce(NH₄)₂(NO₃)₆ solution in a flat-bottomed flask and stirred vigorously for 5 min. Subsequently, the pH was adjusted to 12 by dropwise (approximately one drop/second) addition of 0.5 M NaOH with vigorous stirring by a magnetic stirrer, and stirring was continued for 2 h at room temperature. The precipitate was filtered, washed with distilled water, and then dried in an air oven at 80 °C for 24 h. Finally, the dried CeO₂-diatomite composites were calcined in a muffle furnace at 500 °C in air for 2 h. The same synthesis procedure was repeated by taking half the amount of D for the preparation of DC-12. MDC-11 and MDC-12 composites were synthesized using MD, accordingly.

2.4. Characterization Techniques

Fourier transform infrared (FTIR) spectroscopy was performed with the attenuated total reflectance (ATR) technique using a Perkin Elmer Spectrum Two model FTIR and Universal ATR accessory with diamond/ZnSe crystal in a spectral region of 3600–500 cm⁻¹ with a scan resolution of 2 cm⁻¹. Dispersive Raman spectroscopy measurements were performed by a Thermo Scientific DXR Raman Microscope using Ar⁺ laser power (10 mW) at λ = 532 nm. X-ray diffraction (XRD) patterns were obtained by a Rigaku-

D/MAX-Ultima diffractometer with Cu K α radiation (λ = 1.54 Å) as an X-ray source. The accelerating and applied currents were 40 kV and 40 mA, respectively. Scanning electron microscopy in combination with energy dispersive X-ray analysis (ESEM-EDAX) was achieved on a FEI-Philips XL30 Scanning Electron Microscope-EDAX instrument. Brunauer–Emmett–Teller (BET) surface area and Barrett–Joyner–Halenda (BJH) poresize distribution measurements were obtained by N₂ adsorption at 77 K using a Quantachrome Quadrosorb SI instrument. Thermogravimetric (TG) analysis was undertaken by a Perkin Elmer model STA 600 in the temperature range of 30 to 800 °C under pure nitrogen using 10 mg of specimen. The applied heating rate and nitrogen flow rate were 10 °C/min and 20 mL/min, respectively. The photoluminescence (PL) spectra were recorded on an Edinburgh Instruments FS5 spectrofluorometer with an excitation wavelength at λ = 325 nm. Zeta potential measurements were performed in a Malvern Zetasizer Nano ZSP in the pH range of 2–12 by adding HC1 or NaOH solutions.

2.5. Assessment of Initial Adsorptive Surface Interactions

Photocatalytic time-dependent (0–480 min) adsorption experiments were carried out in a continuously shaking water bath at 25 ± 1 °C. The volume of MB (5 mg/L) solution in Erlenmeyer flasks was 25 mL and the adsorbent dose was 0.25 g/L. Following predetermined time intervals, the suspensions were filtered through 0.45 µm cellulose acetate membrane filter (Millipore HA) to obtain clear solutions for spectroscopic analysis.

2.6. Photocatalytic Activity Assessment

Photocatalytic activity experiments were carried out in a cylindrical Pyrex reaction vessel containing 50 mL MB solution illuminated from the top with a 125 W black light fluorescent lamp ($\lambda_{emis} = 300-400$ nm, $\lambda_{max} = 365$ nm). The intensity of the BLF lamp was determined as I_o=1.65 × 10¹⁶ quanta/s using potassium ferrioxalate actinometer [50]. Two different dosages of CeO₂-diatomite composites (0.25 g/L and 0.50 g/L) were used while keeping the MB concentration constant (5 mg/L). Prior to testing of photocatalytic activity, the solution was magnetically stirred thoroughly for 30 min in the dark to achieve effective surface coverage. Following each photocatalytic treatment period (0–360 nm), the solutions were immediately filtrated through 0.45 µm membrane filters. The absorption changes of MB were assessed at specific maximum absorbance wavelengths of $\lambda = 664$ nm, $\lambda = 292$ nm, and $\lambda = 246$ nm. The UV-vis measurements acquired by a Thermo Scientific Genesys 10S double-beam UV-vis spectrophotometer using 1 cm quartz cells were denoted as A₆₆₄, A₂₉₂, and A₂₄₆ parameters.

3. Results and Discussion

3.1. Characterization of the Diatomites and CeO₂-Diatomite Composites

3.1.1. FTIR Analysis

FTIR spectral features revealed the influence of acid modification on characteristic functional groups of D, MD, and CeO₂-diatomite composites (Figure 1a,b). The spectrum of D indicated a broad band near 3385 cm⁻¹ that was assigned to the free silanol group (vSiO–H) on the surface [51]. The bands at 2987 and 2903 cm⁻¹ corresponded to the asymmetric and symmetric C–H bands on CH₂ and CH₃, respectively, reflecting the presence of organic matter in diatomite [52,53]. The intense band observed at 1636 cm⁻¹ was related to the bending vibration of water (δHOH). Moreover, the characteristic silica bands at 1045 and 796 cm⁻¹ were attributed to the asymmetric stretching of Si–O–Si (vas-Si–O–Si), and the symmetric stretching vibration of Si–O (v_sSi–O) in amorphous silica, respectively [53]. The band at 523 cm⁻¹ was related to the stretching vibration of Si–O silica. This could indicate that the acid treatment had an impact on the chemical properties

of the D. The reason for this could be the change in the chemical composition of MD, resulting in an enhancement of the silica amount due to the removal of the impurity content of D.

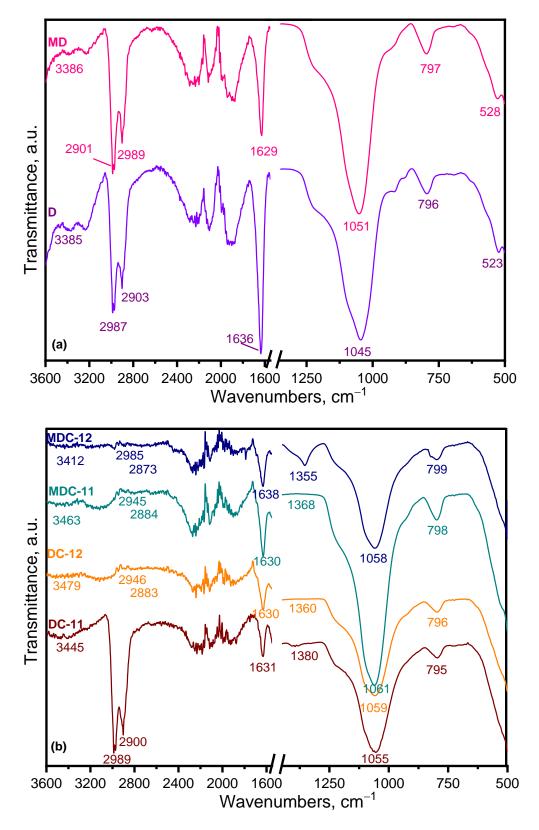
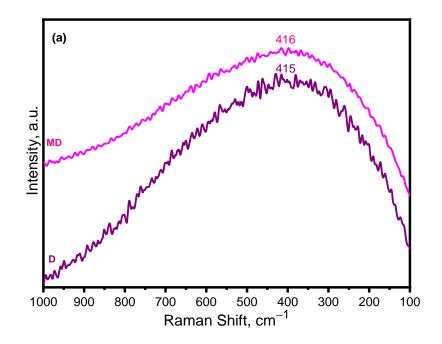


Figure 1. FTIR spectra of (a) D, MD, (b) DC-11, DC-12, MDC-11, and MDC-12.

The FTIR spectral profiles of composites are shown in Figure 1b. The weak band located in the range of 1355–1380 cm⁻¹ belonged to Ce–O–Ce vibration bond in composites, indicating the existence of CeO₂ [55,56]. In general, the positions of the prominent D bands were slightly shifted in the spectrum of CeO2-diatomite composites. The observed band shifts could be related to the incorporation of CeO2 into the framework of the SiO2 matrix [31,57]. In the presence of CeO₂, a new band related to the stretching vibration of Ce–O–Ce (vCe–O–Ce) at 1380 cm⁻¹ was observed in the spectrum of DC-11 composite. The main bands at 3445, 2989, 2900, 1631, 1380, 1055, and 795 cm⁻¹ corresponded to the free silanol group (vSiO–H), asymmetric C–H bands on CH₂ (v_{as}C–H), symmetric C–H bands on CH₃ (v_s C–H), bending vibration of water (δ HOH), asymmetric stretching of Si– O–Si (vasSi–O–Si), and symmetric stretching vibration of Si–O (vsSi–O), respectively. The characteristic Si-O-Si asymmetric and symmetric strong stretching signals of D and MD red shifted in all composites with the increasing amount of CeO₂. Although the band at 530 cm⁻¹ was also expected to shift to higher frequencies, it seems to be shielded by the shoulder of strong O-Ce-O vibrational signal centered at around 415 cm⁻¹ (not shown in the figure) as reported in the literature [51–56]. Moreover, the most interesting spectral change was observed in the stretching vibration band of Ce–O–Ce. This band was notably blue shifted in MDC-11 and MDC-12 composites compared to that of composites prepared by raw D, designated as DC-11 and DC-12, respectively.

3.1.2. Raman Spectroscopy

Raman spectra of diatomite specimens and CeO₂-diatomite composites are presented in Figure 2.



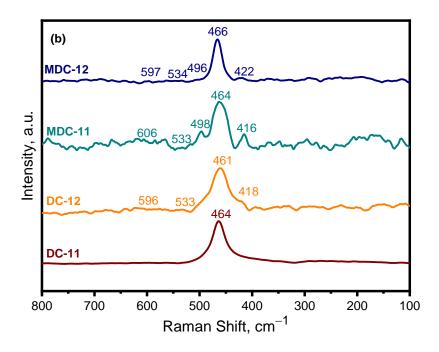
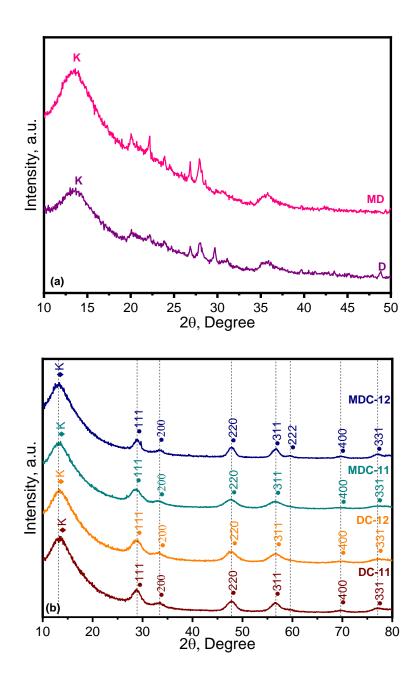


Figure 2. Raman spectra of (a) D, MD, (b) DC-11, DC-12, MDC-11, and MDC-12.

Diatomite specimens exhibited a broad band at ~415 cm⁻¹ that was assigned to the O-Si-O bending mode [58]. Raman spectra of DC-11 composite revealed a prominent band at 464 cm⁻¹ that was attributed to the triply degenerate F_{2g} vibration mode, which was assigned to a symmetric breathing mode of the oxygen atoms around Ce ions [59]. This band was in the range of 461–464 cm⁻¹ in composites, suggesting the fluorite structure of CeO₂ [60]. The observed band at ~600 cm⁻¹ corresponded to the non-degenerate longitudinal-optical mode of CeO₂ indicating the oxygen vacancies in the CeO₂ lattice [59,61]. The D₁ bands in the region between 500 and 600 cm⁻¹ could be related to the bulk oxygen vacancies in CeO₂. The band at ~420 cm⁻¹ could belong to the silica structure due to the contribution of D. The bands below 200 cm⁻¹ could be assigned to the scattering [62,63].

3.1.3. XRD Analysis

XRD analysis was performed to determine the main chemical composition of raw D and MD (Figure 3a). XRD patterns of D exhibited a broad diffraction in the region of $2\theta = 20-30^\circ$, suggesting the presence of amorphous silica. There was no remarkable alteration in the peak positions of the MD. However, a slight enhancement was observed in the peak intensities, indicating that the acid treatment affected the crystallinity of diatomite.



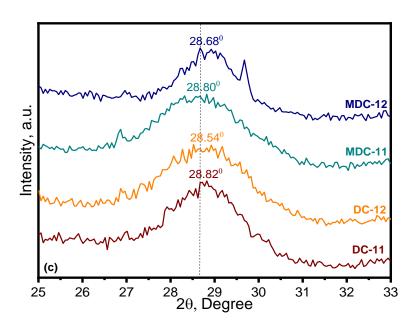


Figure 3. XRD spectra of (**a**) D, MD, (**b**) DC-11, DC-12, MDC-11, and MDC-12, (**c**) enlarged (111) plane of CeO₂.

The XRD patterns of CeO₂-diatomite composites are shown Figure 3b. The specific diffraction peaks were attributed to CeO₂ and D. For MDC-12 composite, $2\theta = 28.68^{\circ}$, 33.19°, 47.85°, 56.72°, 59.42°, 69.69°, and 77.12° corresponded to the presence of (111), (200), (220), (311), (222), (400), and (331) planes of CeO₂, while the broad diffraction peak at $2\theta = 13.23$ could be related to the kaolinite-like structure in D. The XRD analysis revealed that CeO₂ in composites was consistent with the fluorite-type structure due to the reflection plane of (111) observed at $2\theta = 28.68^{\circ}$ [64]. The pattern was well-matched with the CeO₂ standard card JCPDS NO. 43–1002, as well as with Raman data. The diffraction plane of (222) was not visible for the other three CeO₂-diatomite composites. Moreover, there were no extra diffraction peaks related to the formation of new or mixed phases between CeO₂ and D. The enlarged (111) diffraction planes of CeO₂-diatomite composites (MDC-11 and MD-11), a gradual 2 θ shift towards to higher angles was noted. Compared with DC-12, the diffraction peak of MDC-12 was also shifted to a higher angle due to acid treatment process of D.

The average crystallite sizes (D, nm) of CeO₂-diatomite composites were calculated from the peak of the (111) plane of CeO₂ using the Scherrer equation (Equation (8)) [65]:

$$D = K \lambda / (\beta \cos \theta)$$
(8)

where K = 0.9, λ is the X-ray wavelength (1.5418 Å), θ is the Bragg angle, and β is the full width at half maximum intensity (FWHM, radians).

The influence of acid treatment on the crystallite size of CeO₂-diatomite composites was also reflected as X-ray line broadening. The diffraction peaks of MDC-12 composite were obviously narrower than those of DC-12 composite, thus indicating a relatively larger crystallite size. Almost equal crystal sizes were calculated as 2.2 and 2.1 nm for DC-11 and MDC-11 composites, respectively. However, the crystallite sizes of CeO₂ in DC-12, of 1.9 nm, and MDC-12, of 3.0 nm, displayed an incremental increase, most probably due to the modification of diatomite. This result indicates that the acid treatment affected the crystallite particle sizes of CeO₂-diatomite composites as the CeO₂ amount was increased.

3.1.4. ESEM Analysis

The ESEM images of D and MD in comparison to CeO2-diatomite composites are presented in Figure 4. An orderly set pore structure in a fold sheet form can be observed in the D image (Figure 4a). The diatom debris grains were arranged dispersedly, resulting in a porous structure [66]. After acid treatment, the morphology was altered, and a more folded sheet form was formed. Moreover, the size of the diatom pores that were clearly visible was increased (Figure 4b). ESEM images of CeO2-diatomite composites revealed an irregularly shaped morphology containing both D and CeO2 particles (Figure 4c-f). The dominant morphology was the porous structure of D for DC-11 composite spreading throughout the composite. It was also clear that a distinctive almost spherical shape of CeO₂ particles could be observed [31]. Increasing the CeO₂ amount resulted in a decrease in the folded sheets of D in DC-12 composite. With respect to ESEM images of DC-11 and MDC-11 composites, a decrease in both agglomeration and particle size of CeO2 nanoparticles was evident in MDC-11. The pores of D were enlarged, and folded sheets almost disappeared. Similar morphological changes were also observed for DC-12 and MDC-12 composites and the particle size was reduced in the MDC-12 composite. The spherical particles of CeO₂ were transformed into various polyhedral particles and MD was deposited onto the particles. The observed variations could be due to the formation of chemical bonds between CeO2 and diatomite with respect to the CeO2 amount in composites.

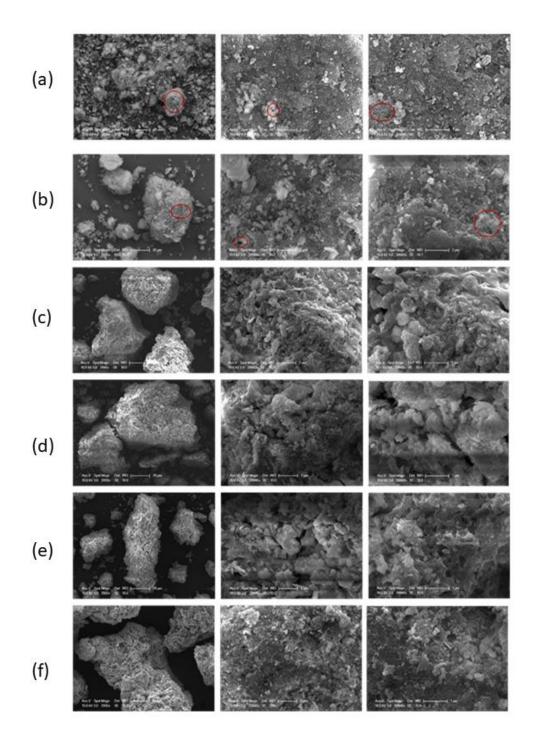


Figure 4. ESEM images of (a) D, (b) MD, (c) DC-11, (d) DC-12, (e) MDC-11, (f) MDC-12.

EDAX spectral elemental composition of D, MD, and corresponding CeO₂diatomite composites are displayed in Figure S1, in Supplementary Materials (SI Part 1, Figure S1a–f). High content of silicium was strongly confirmed by EDAX analyses of D and MD. Si content increased after modification and other elements such as C, O, Na, Mg, Al, Ca, and Fe were also observed. The presence of Cl as detected in the EDAX spectrum of MD could be due to the acid modification of D with HCl treatment. The presence of Ca, Si, and Mg elements could be in the form of CaCO₃, silicate, and Mg(OH)₂ sediments, as already reported in the materials section (Section 2.1) [67]. More importantly, EDAX spectra (Figure S1c–f) showed that the as-prepared composites were mainly composed of Si, Ce, and O elements.

3.1.5. BET Analysis

BET nitrogen adsorption/desorption isotherms and BJH pore-size distribution curves of D, MD, and representative CeO₂-diatomite composites are shown in SI Part 1, Figures S2 and S3. Average pore diameter (d), pore volume (V_P), and specific surface area (SBET) are summarized in Table 1. The nitrogen adsorption/desorption isotherms of D and MD exhibited a type II sorption behavior with an H3 hysteresis loop according to the International Union of Pure and Applied Chemistry (IUPAC) classification [68]. It was clear that the D and MD revealed the same form of isotherms and hysteresis loop. The H3 hysteresis loop was principally related with the filling and emptying of the mesopores by capillary condensation, suggesting that the pores were mostly narrow and slit-shaped, and their sizes were not well-proportioned [69,70]. In addition, the adsorption isotherm curves were significantly increased with the relatively moderate pressure, signifying a gradual pore diameter enhancement. However, acid treatment resulted in a slight enhancement in nitrogen adsorption, indicating that the overflow possessed a higher SBET. It is well-known that acid modification of D leads to an increase in surface area and adsorption capacity [71–75]. The reason for this could be the elimination of minor impurities responsible of pore blockage via acid treatment [73].

The adsorption-desorption isotherms of CeO₂-diatomite composites were classified as type IV with an H3 hysteresis loop, implying a typical unfilled mesoporous structure. However, MDC-12 composite corresponded to a typical IV adsorption isotherm with an H2 hysteresis loop due to the existence of the pore blockage or seepage [68] (SI Part 1, Figure S2).

Specimens	<i>S</i> вет *, m²/g	V _p **, cm ³ /g	D ***, nm
D	109	0.3082	4.55
MD	118	0.3478	5.05
DC-11	76	0.0954	2.29
DC-12	84	0.1578	1.73
MDC-11	127	0.2032	3.83
MDC-12	126	0.1792	3.49

Table 1. Surface characteristics properties of diatomites and CeO2-diatomite composites.

* Specific surface area data calculated by the multi-point BET method. ** Pore volume obtained from the BJH adsorption cumulative volume of pores between 0.36 and 300 nm diameter. *** Adsorption average pore diameter (4 V/A by BET).

The pore-size distribution curves based on the pore volumes of the BJH adsorption are shown in SI Part 1, Figure S3. A heterogeneous distribution of pore diameters smaller than 40 nm was evidenced for D. The DC-11, DC-12, and MDC-11 composites exhibited narrow pore-size distributions with average pore diameters of 3.82, 2.29, and 1.73 nm, respectively.

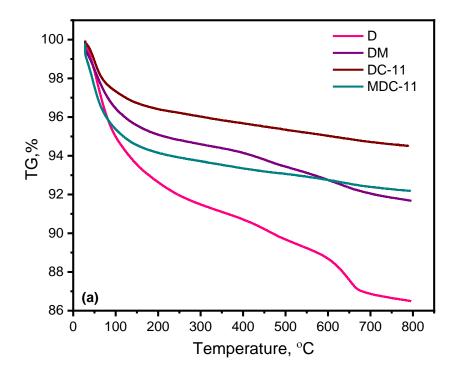
3.1.6. Thermal Analysis

The thermal stability of D and MD and respective DC-11 and MDC-11 composites were determined by TG and DTG analysis (Figure 5a,b). D and MD exhibited three weight-loss steps. The first step of TG curves in the low temperature range of 30–300 °C was attributed to the loss of adsorbed water [76]. The weight loss obtained in this step was 8.5% and 5.4% for D and MD, respectively. The observed peaks at 79 °C (D) and 58 °C (MD) in DTG curves indicated the dehydration process. The second weight loss of D in the temperature range of 350–530 °C with DTG peak was at 458 °C. This MD peak corresponding to condensation of hydroxyl groups, and the dehydroxylation process was slightly shifted to 453 °C [76,77]. The final step was related to the dehydroxylation due to the loss of OH groups surrounding the Al(VI) atom [78]. A similar DTG peak at ~600 °C was observed for M, MD, and corresponding composites (Figure 5b). The dehydration of adsorbed H₂O and dehydroxylation were also observed in TG curves of compo-

sites. The total weight loss was 13.2%, 8.02%, 5.31%, and 7.94% for D, MD, DC-11, and MDC-11, respectively.

3.1.7. PL Analysis

Diatomites were composed of surfaces rich in reactive silanol (Si-OH) groups [79-81]. Figure 6 reveals PL spectra of D, MD, and CeO₂-composites. The PL spectrum of D exhibited three main peaks at 400 nm (3.01 eV), 485 nm (2.56 eV), and 564 nm (2.20 eV), while three peaks belonging to MD were located at 398 nm (3.12 eV), 482 nm (2.57 eV), and 560 nm (2.21 eV) [82]. It was demonstrated that the origin of visible emission in the blue and green regions was related to the different defect states, such as oxygen defect centers, indicating non-bridging oxygen hole centers or neutral oxygen vacancy and selftrapped excitons [80]. The broad peaks appeared at three main regions of 365-435 nm, 450–505 nm, and 520–600 nm in the PL spectra of composites with different CeO2 amounts. The observed peaks in PL spectra could belong to the characteristic peaks of CeO₂ or D that were also observed in CeO₂-diatomite composites with a blue shift. DC-11 composite showed a strong violet/blue light emission peak at 387 nm (3.20 eV), which corresponded to the band-edge free exciton luminescence. The blue-green region of visible range at 467 nm (2.66 eV) was the bound exciton luminescence. The third emission peak was observed at 564 nm (2.20 eV) in the green visible region [83,84]. The emission peaks in the range of 400–500 nm corresponded to the different defect levels between the localized Ce 4f⁰ state and O 2p valence band [83,85]. PL emission is useful to understand the recombination possibility of excited electrons and holes in photocatalyst. In general, a lower PL intensity indicates a lower recombination rate of electron and hole pairs. Furthermore, an intense PL intensity reveals a higher number of defects and/or oxygen vacancies. The observed PL intensity trend with the CeO₂ amount in CeO₂-diatomite composites was nonlinear. The reason for this could be particle size, recombination, and presence of defects [23,70,84].



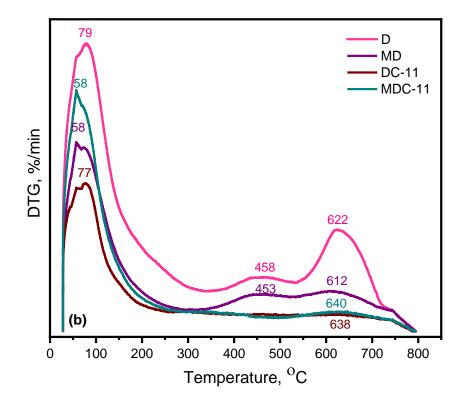


Figure 5. (a) TG, (b) DTG curves for D, DM, DC-11, and MDC-11.

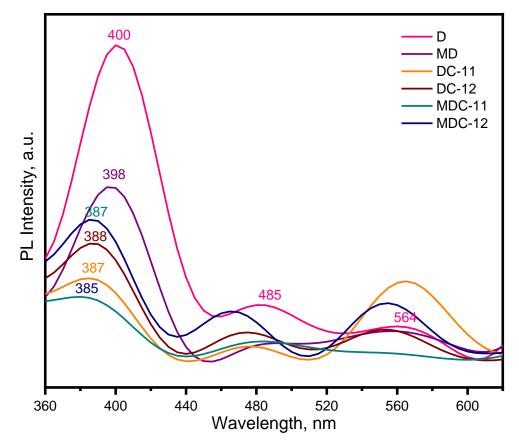


Figure 6. PL spectra of D, MD, DC-11, DC-12, MDC-11, and MDC-12.

3.1.8. Zeta Potential Analysis

pH-dependent variations in zeta potential (ζ , mV), as an efficient tool denoting the surface charge, are shown in SI Part 1, Figure S4. It was obvious that the zeta potentials of D and MD had negative magnitudes in the pH range of pH = 2-12. The surface of MD was remarkably more negatively charged throughout the pH range compared to D, indicating the presence of a higher number of functional groups. The zeta potentials were dominantly negative in a wide pH range, excluding the presence of any isoelectric potential (IEP) representing pH of zero-point charge (pHpzc) [86]. This result implies the independent pH behavior of D and MD particles. Moreover, the monitored permanent negative charge could be related to the existence of substituted isomorphous aluminosilicate structure, and this observation was consistent with the chemical/mineralogical composition of D [87,88]. Similar results were also reported indicating the absence of IEP [76,89]. In contrast, in several studies, IEP of diatomite was measured as pH_{pzc}=2 [90,91]. The reason for this could be related to the structure of the diatomite types that contained different minerals, with various ratios correlating with the zeta potential analyzed. The pH_{pzc} values of DC-11, DC-12, MDC-11, and MDC-12 composites were 6.07, 4.57, 5.38, and 5.28, respectively. The results indicate that no remarkable effect of different CeO₂ contents was noticed on the surface charges of the MDC-11 and MDC-12 composites. The reason for this could be the favorable attraction between the positive charge of CeO₂ (pH_{pzc}=~7) and negative charge of MD [23].

3.2. Assessment of Photocatalytic Activity

3.2.1. Preliminary Experiments

The UV-vis absorption spectra of MB revealed four absorption peaks. An intense absorption peak at λ_{max} = 664 nm (A₆₆₄) assigned to MB monomer with a shoulder peak at λ = 612 nm attributed to MB dimer were composed of a sulfur–nitrogen conjugated system expressing the chromophore groups of MB. The two other observed peaks at λ = 292 nm (A₂₉₂) and λ = 246 nm (A₂₄₆) were related to the substituted benzene rings, namely, the phenothiazine unit in MB structure [3]. Photocatalytic degradation of MB was expressed by descriptive parameters such as A₆₆₄, A₂₉₂, and A₂₄₆.

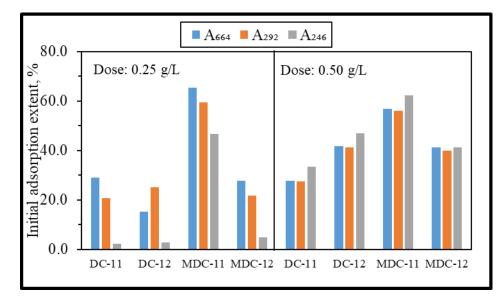
The preliminary experiments displayed direct photolysis of MB and surface interactions, revealing MB adsorption onto CeO₂-diatomite composites. As a representative parameter of MB, A₆₆₄ was used unless otherwise stated [92].

Variations in UV-vis spectral features of MB under irradiation due to direct photodegradation were formerly reported by Turkten and colleagues [19]. Accordingly, direct photodegradation of MB was extremely slow due to the non-overlapping trend of UVvis absorption spectra with the BLF lamp emission spectra.

The effect of pre-contact time prior to initiation of light exposure was determined under dark conditions. D and DC-11 composite (0.25 g/L) were chosen as representative basic specimens and adsorption of MB (5 mg/L) was investigated for a range of contact time (t = 0–8 h). The initial adsorption amount was determined as ~20%, as related to the prevailing interactions between the negatively charged surface of the specimens and cationic dye MB [93]. Slight variations (\pm 10%) were attained during the whole period of dark interactions. Fundamentally, the extent of initial adsorption should sufficiently disclose the free surface area for light absorption required for the initiation of the photocatalytic reaction. Therefore, as an initial dark contact time, 30 min was chosen for photocatalytic treatment of MB using all specimens.

3.2.2. Photocatalytic Degradation of MB using CeO₂-Diatomite Composites

The photocatalytic degradation of MB was studied using CeO₂-diatomite composites using two different doses (0.25 and 0.50 g/L). As an effective constraint of the surface-oriented nature of photocatalysis prior to initiation of light exposure, the dose-



dependent initial dark adsorption extents of all descriptive parameters are presented in Figure 7.

Figure 7. Dose dependent initial dark adsorption extents of MB onto DC-11, DC-12, MDC-11, and MDC-12, as expressed by all descriptive parameters.

From a general perspective, coverage of MB onto composites increased with respect to increasing dose for the MDC-11 specimen. Upon use of the lower dose of 0.25 g/L, all descriptive parameters displayed a consistently decreasing trend, irrespective of the composite type. However, for the higher dose of 0.50 g/L, all descriptive parameters displayed similar roles towards the surface of the photocatalyst specimens. The MDC-11 specimen expressed a remarkably different trend as evidenced by the highest surface interaction, most probably due to the counter-balancing effect related to the equal compositional ratio of MD to CeO₂.

Upon initiation of irradiation, during the whole duration of photocatalysis (up to 300 min), UV-vis spectral features of MB were followed, as shown in SI Part II, Figure S5 and Figure S6. The gradual decrement in the A₆₆₄ peak during photocatalysis implied that the sulfur–nitrogen conjugated system was degraded. A blue shift was noticed due to the prolonged exposure of UVA illumination. This result indicated a probable route of N-demethylated degradation of MB derivatives [19,94,95]. A similar progressively decreasing trend in A₂₉₂ and A₂₄₆ was also recorded. This outcome indicated the occurrence of oxidative destruction of MB, resulting in a ring-opening reaction of the phenothiazine species. The absence of new peaks suggested that there was no formation of new intermediates converted from the destroyed phenothiazine structure of MB [94,96]. At 300 min of irradiation, all descriptive parameters were diminished and almost a complete disappearance of the A₆₁₂ was found upon use of the MDC-11 composite. This could be ascribed to the cleavage of benzene rings and the heteropoly aromatic linkage due to the degradation of MB.

The photocatalytic degradation of MB in the presence of the prepared CeO₂diatomite composites followed the pseudo-first-order rate model expressed by Equation (9):

Rate (R) =
$$-dA/dt = kA$$
 (9)

where R is the pseudo-first-order rate (cm⁻¹min⁻¹); A_0 is the initial absorbance of MB expressed as $A_{664,0}$, $A_{292,0}$ and $A_{246,0}$; A is the absorbance of MB expressed as A_{664} , A_{292} , and A_{246} at time t; t is the irradiation time (min); and k is the pseudo-first-order reaction rate constant (min⁻¹). All UV-vis parameters, i.e., A_{664} , A_{292} , and A_{246} , exhibited logarithmic decay profiles, revealing kinetic model parameters as presented in Table 2.

	First-C	First-Order Kinetic Parameters Dose: 0.25 g/L		First-Order Kinetic Parameters Dose: 0.50 g/L		
A664	kx10 ⁻³ , min ⁻¹	t1/2, min	Rx10 ⁻³ , cm ⁻¹ min ⁻¹	kx10 ⁻² , min ⁻¹	t1/2, min	Rx10 ⁻³ , cm ⁻¹ min ⁻¹
DC-11	2.24	309	2.15	5.91	117	5.68
DC-12	3.54	196	3.40	8.98	77	8.63
MDC-11	10.2	68	9.84	9.00	77	8.65
MDC-12	3.05	227	2.93	5.36	129	5.15
A292						
DC-11	1.93	359	1.05	5.00	139	2.72
DC-12	2.76	251	1.50	6.56	106	3.57
MDC-11	2.76	251	1.50	6.11	113	3.32
MDC-12	2.49	278	1.35	4.33	160	2.35
A246						
DC-11	1.42	488	0.344	6.19	112	1.50
DC-12	1.83	379	0.443	7.41	94	1.80
MDC-11	3.77	184	0.913	8.58	81	2.07
MDC-12	1.64	423	0.397	4.22	164	1.02

The MDC-11 composite expressed the highest decolorization rate constant upon use of two different doses, and thus the highest photocatalytic activity. The reason for this could be the larger specific surface area of the MDC-11 composite ($S_{BET} = 127 \text{ m}^2/\text{g}$), providing more available active sites, and hence favorable for the fast adsorption of MB, resulting in a high photocatalytic activity [97].

The photocatalytic degradation rate constants of CeO₂-diatomite composites using a 0.25 g/L photocatalyst dose can be presented in decreasing order as:

A₆₆₄: MDC-11 > DC-12 > MDC-12 > DC-11

A292: MDC-11 = DC-12 > MDC-12 > DC-11

A246: MDC-11 > DC-12 > MDC-12 > DC-11

Upon use of a 0.50 g/L photocatalyst dose, the trend in photocatalytic degradation rate constants can be given in the following order as:

A664: MDC-11 > DC-12 > DC-11 > MDC-12

A292: DC-12 > MDC-11 > DC-11 > MDC-12

A246: MDC-11 > DC-12 > DC-11 > MDC-12

The degree of MB decolorization/degradation using CeO₂-diatomite composites, as presented in Figure 8, was also calculated by the represented Equations (10–12):

Decolorization₆₆₄, % = ((A_{664,0} - A₆₆₄)/ A_{664,0})) × 100 (10)

Degradation₂₉₂, % = $((A_{292,0} - A_{292})/A_{292,0})) \times 100$ (11)

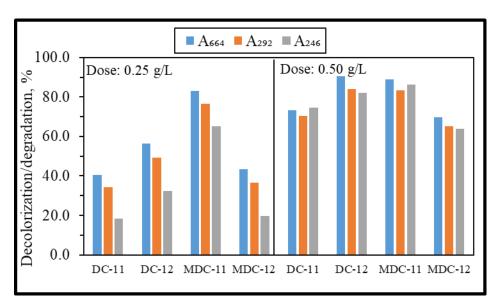


Figure 8. Dose-dependent decolorization/degradation efficiencies of MB onto DC-11, DC-12, MDC-11, and MDC-12 as expressed by all descriptive parameters.

The degree of MB decolorization/degradation using 0.25 g/L of CeO₂-diatomite composites for an irradiation period of 120 min can be presented in a decreasing order as:

A₆₆₄: MDC-11 > DC-12 > MDC-12 > DC-11 A₂₉₂: MDC-11 > DC-12 > MDC-12 > DC-11 A₂₄₆: MDC-11 > DC-12 > MDC-12 > DC-11

The degree of MB decolorization/degradation using 0.50 g/L of CeO₂-diatomite composites for an irradiation period of 120 min can be presented in the following order as:

A₆₆₄: DC-12 > MDC-11 > DC-11 > MDC-12 A₂₉₂: DC-12 > MDC-11 > DC-11 > MDC-12 A₂₄₆: MDC-11 > DC-12 > DC-11 > MDC-12

Upon use of 0.25 g/L, the effect of diatomite modification was found to be more pronounced for DC-11 and MDC-11 specimens. The lowest decolorization/degradation efficiency attained for DC-11 can be expressed by the folded diatomite structure (Figure 4a) with a high agglomeration resulting in a negative effect on the photocatalytic efficiency [98].

An increased CeO₂ ratio resulted in a retardation effect, as observed for DC-12 and MDC-12; the reason of this could be the lower surface coverage related to the surface morphological properties (Table 1, Figure 7). However, the presence of a higher photocatalyst dose (0.50 g/L) resulted in a pronounced effect due to both diatomite modification and CeO₂ content, as well as the available surface area (Figure 8).

For simplicity purposes, as an indicative parameter of MB, A₆₆₄ was selected for comparative presentation of kinetics and removal of MB upon use of all photocatalyst specimens (Figure 9).

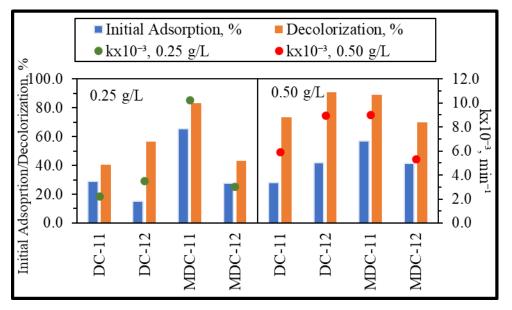


Figure 9. Comparison of dose-dependent initial adsorption extents and decolorization efficiencies MB, A₆₆₄ with respect to first-order kinetic rate constants.

As can be visualized, upon use of 0.25 g/L, first-order rate constants revealed an inconsistent relationship with surface coverage extents, whereas decolorization efficiencies represented a close resemblance irrespective of the photocatalyst type. The highest rate constant of MDC-11 could not be directly correlated with the surface coverage extent, although decolorization efficiency was found to be in good order. On the other hand, upon use of 0.50 g/L, first-order rate constants indicated differing tendencies with respect to initial adsorption extents; however, they displayed similarity to the observed decolorization efficiencies. Increasing the photocatalyst dose substantially increased degradation rates, especially in the MDC-11 specimen, which could be related to the degradation mechanism resulting in co-operating adsorption based on Langmuir– Hinshelwood kinetics and, in solution, operating in the Eley–Rideal kinetic model.

It should be emphasized that photocatalytic degradation of methylene blue should be followed by all descriptive spectroscopic parameters rather than sole decolorization (A₆₆₄) due to hydroxylation leading to the aromatic ring opening [92].

4. Conclusions

CeO₂-composites were prepared by a simple precipitation method using diatomite and modified diatomite as representative silica sources. The effect of acid treatment resulted in CeO₂-composites having different morphological, structural, optical, and thermal features. The application pretreatment of diatomite not only led to the formation of a less-agglomerated folded sheet structure, but also enhanced the surface area and porosity of MDC-11 composite compared to DC-11 composite. The presence of the FTIR bands belonging to the Ce–O–Ce vibration bond and silica groups confirmed the coupled binary system of CeO₂ and SiO₂ in composites. The XRD and Raman analysis indicated that CeO₂ resembled a fluorite-type structure in the composites. The average crystallite sizes of CeO₂ in the composites determined through Scherrer's equation were in the range of 1.9–3.0 nm. The thermal behavior of the composites was related to the dehydration of adsorbed water and dihydroxylation.

The photocatalytic activity of the prepared CeO₂-diatomite composites was examined through the degradation of MB, represented by three descriptive absorbances, i.e., A₆₆₄, A₂₉₂, and A₂₄₆. The highest photocatalytic efficiency of MB was obtained using the MDC-11 composite due to the high surface area of 127 m²/g, resulting in enhanced surface contact between MB and the composite. This study originally described a systematic comparison effect of acid treatment of diatomite on the preparation of CeO₂-diatomite composites, as well as photocatalytic activities. Based on the attained results using MB as the substrate, CeO₂-diatomite composites could be a promising new affordable and low-cost candidate for photocatalytic applications.

Supplementary Materials: The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/w14213373/s1, Figure S1: EDAX spectra of (a) D, (b) MD, (c) DC-11, (d) DC-12, (e) MDC-11; figure S2: BET adsorption-desorption isotherms of (a) D, (b) MD, (c) DC-11, (d) MDC-11; Figure S3: Pore size distributions of (a) D, (b) MD, (c) DC-11, (d) MDC-11; Figure S3: Pore size distributions of (a) D, (b) MD, (c) DC-11, (d) MDC-11; Figure S3: Pore size distributions of (a) D, (b) MD, (c) DC-11, (d) MDC-11; Figure S4: Zeta potential plots of (a) diatomites, (b) CeO2-diatomite composites as a function of pH; Figure S5: UV-vis absorption spectra of MB using 0.25 mg/mL dose of (a) DC-11, (b) DC-12, (c) MDC-12; Figure S6: UV-vis absorption spectra of MB using 0.50 mg/mL dose of (a) DC-11, (b) DC-12, (c) MDC-11, and (d) MDC-12.

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Data Availability Statement: The data presented in the study is available on request from the corresponding author.

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References

- Waghchaure, R.H.; Adole, V.A.; Jagdale, B.S. Photocatalytic degradation of methylene blue, rhodamine B, methyl orange and Eriochrome black T dyes by modified ZnO nanocatalysts: A concise review. *Inorg. Chem. Commun.* 2022, 143, 109764. https://doi.org/10.1016/j.inoche.2022.109764.
- Hamad, H.N.; Idrus, S. Recent developments in the application of bio-waste-derived adsorbents for the removal of methylene blue from wastewater: A review. *Polymers* 2022, 14, 783.
- 3. Khan, I.; Saeed, K.; Zekker, I.; Zhang, B.; Hendi, A.H.; Ahmad, A.; Ahmad, S.; Zada, N.; Ahmad, H.; Shah, L.A.; et al. Review on methylene blue: Its properties, uses, toxicity and photodegradation. *Water* **2022**, *14*, 242.
- 4. Rafatullah, M.; Sulaiman, O.; Hashim, R.; Ahmad, A. Adsorption of methylene blue on low-cost adsorbents: A review. J. *Hazard. Mater.* 2010, 177, 70–80. https://doi.org/10.1016/j.jhazmat.2009.12.047.
- 5. Kuang, Y.; Zhang, X.; Zhou, S. Adsorption of methylene blue in water onto activated carbon by surfactant modification. *Water* **2020**, *12*, 587.
- Labena, A.; Abdelhamid, A.E.; Amin, A.S.; Husien, S.; Hamid, L.; Safwat, G.; Diab, A.; Gobouri, A.A.; Azab, E. Removal of methylene blue and congo red using adsorptive membrane impregnated with dried *Ulva fasciata* and *Sargassum dentifolium*. *Plants* 2021, 10, 384.
- Verma, A.K.; Dash, R.R.; Bhunia, P. A review on chemical coagulation/flocculation technologies for removal of colour from textile wastewaters. J. Environ. Manage. 2012, 93, 154–168. https://doi.org/10.1016/j.jenvman.2011.09.012.
- 8. Gomes, J.; Lincho, J.; Domingues, E.; Quinta-Ferreira, R.M.; Martins, R.C. N–TiO₂ Photocatalysts: A Review of Their Characteristics and Capacity for Emerging Contaminants Removal. *Water* **2019**, *11*, 373.
- Kumar, M.S.; Sonawane, S.H.; Pandit, A.B. Degradation of methylene blue dye in aqueous solution using hydrodynamic cavitation based hybrid advanced oxidation processes. *Chem. Eng. Process. Process Intensif.* 2017, 122, 288–295. https://doi.org/10.1016/j.cep.2017.09.009.
- 10. Kurian, M. Advanced oxidation processes and nanomaterials—A review. *Cleaner Eng. Technol.* 2021, 2, 100090. https://doi.org/10.1016/j.clet.2021.100090.
- 11. Modi, S.; Yadav, V.K.; Gacem, A.; Ali, I.H.; Dave, D.; Khan, S.H.; Yadav, K.K.; Rather, S.-u.; Ahn, Y.; Son, C.T.; et al. Recent and Emerging Trends in Remediation of Methylene Blue Dye from Wastewater by Using Zinc Oxide Nanoparticles. *Water* **2022**, *14*, 1749.
- 12. Hoffmann, M.R.; Martin, S.T.; Choi, W.; Bahnemann, D.W. Environmental Applications of Semiconductor Photocatalysis. *Chem. Rev.* **1995**, *95*, 69–96. https://doi.org/10.1021/cr00033a004.
- 13. Turchi, C.S.; Ollis, D.F. Photocatalytic degradation of organic water contaminants: Mechanisms involving hydroxyl radical attack. *J. Catal.* **1990**, *122*, 178–192. https://doi.org/10.1016/0021-9517(90)90269-P.
- 14. Gharbani, P.; Mehrizad, A.; Mosavi, S.A. Optimization, kinetics and thermodynamics studies for photocatalytic degradation of Methylene Blue using cadmium selenide nanoparticles. *npJ. Clean Water* **2022**, *5*, 34. https://doi.org/10.1038/s41545-022-00178-x.
- 15. George, A.; Magimai Antoni Raj, D.; Venci, X.; Dhayal Raj, A.; Albert Irudayaraj, A.; Josephine, R.L.; John Sundaram, S.; Al-Mohaimeed, A.M.; Al Farraj, D.A.; Chen, T.-W.; et al. Photocatalytic effect of CuO nanoparticles flower-like 3D nanostructures

under visible light irradiation with the degradation of methylene blue (MB) dye for environmental application. *Environ. Res.* **2022**, 203, 111880. https://doi.org/10.1016/j.envres.2021.111880.

- Ahmad, A.; Khan, M.; Khan, S.; Luque, R.; Almutairi, T.M.; Karami, A.M. Bio-construction of MgO nanoparticles using Texas sage plant extract for catalytical degradation of methylene blue via photocatalysis. *Int. J. Environ. Sci. Technol.* 2022, 1–12. https://doi.org/10.1007/s13762-022-04090-2.
- Perumal, V.; Inmozhi, C.; Uthrakumar, R.; Robert, R.; Chandrasekar, M.; Mohamed, S.B.; Honey, S.; Raja, A.; Al-Mekhlafi, F.A.; Kaviyarasu, K. Enhancing the photocatalytic performance of surface—Treated SnO₂ hierarchical nanorods against methylene blue dye under solar irradiation and biological degradation. *Environ. Res.* 2022, 209, 112821. https://doi.org/10.1016/j.envres.2022.112821.
- Abbas, N.; Rubab, N.; Sadiq, N.; Manzoor, S.; Khan, M.I.; Fernandez Garcia, J.; Barbosa Aragao, I.; Tariq, M.; Akhtar, Z.; Yasmin, G. Aluminum-Doped Cobalt Ferrite as an Efficient Photocatalyst for the Abatement of Methylene Blue. *Water* 2020, 12, 2285.
- 19. Turkten, N.; Karatas, Y.; Bekbolet, M. Preparation of PANI Modified ZnO Composites via Different Methods: Structural, Morphological and Photocatalytic Properties. *Water* **2021**, *13*, 1025. https://doi.org/10.3390/w13081025.
- Basumatary, B.; Basumatary, R.; Ramchiary, A.; Konwar, D. Evaluation of Ag@TiO₂/WO₃ heterojunction photocatalyst for enhanced photocatalytic activity towards methylene blue degradation. *Chemosphere* 2022, 286, 131848. https://doi.org/10.1016/j.chemosphere.2021.131848.
- Birben, N.C.; Paganini, M.C.; Calza, P.; Bekbolet, M. Photocatalytic degradation of humic acid using a novel photocatalyst: Cedoped ZnO. *Photochem. Photobiol. Sci.* 2017, 16, 24–30. https://doi.org/10.1039/C6PP00216A.
- 22. Chen, F.; Cao, Y.; Jia, D. Preparation and photocatalytic property of CeO₂ lamellar. *Appl. Surf. Sci.* 2011, 257, 9226–9231. https://doi.org/10.1016/j.apsusc.2011.06.009.
- 23. Phanichphant, S.; Nakaruk, A.; Channei, D. Photocatalytic activity of the binary composite CeO₂/SiO₂ for degradation of dye. *Appl. Surf. Sci.* 2016, 387, 214–220. https://doi.org/10.1016/j.apsusc.2016.06.072.
- 24. Litter, M.I. Heterogeneous photocatalysis: Transition metal ions in photocatalytic systems. *Appl. Catal. B* 1999, 23, 89–114. https://doi.org/10.1016/S0926-3373(99)00069-7.
- Cheng, Z.; Luo, S.; Liu, Z.; Zhang, Y.; Liao, Y.; Guo, M.; Nguyen, T.T. Visible-light-driven hierarchical porous CeO₂ derived from wood for effective photocatalytic degradation of methylene blue. *Opt. Mater.* 2022, 129, 112429. https://doi.org/10.1016/j.optmat.2022.112429.
- Jasim, S.A.; Machek, P.; Abdelbasset, W.K.; Jarosova, M.; Majdi, H.S.; Khalaji, A.D. Solution combustion synthesis of CeO₂ nanoparticles for excellent photocatalytic degradation of methylene blue. *Appl. Phys. A* 2022, 128, 475. https://doi.org/10.1007/s00339-022-05532-x.
- 27. Majumder, D.; Chakraborty, I.; Mandal, K.; Roy, S. Facet-Dependent Photodegradation of Methylene Blue Using Pristine CeO₂ Nanostructures. *ACS Omega* **2019**, *4*, 4243–4251. https://doi.org/10.1021/acsomega.8b03298.
- Pouretedal, H.R.; Kadkhodaie, A. Synthetic CeO₂ Nanoparticle Catalysis of Methylene Blue Photodegradation: Kinetics and Mechanism. *Chinese J. Catal.* 2010, *31*, 1328–1334. https://doi.org/10.1016/S1872-2067(10)60121-0.
- Reddy Yadav, L.S.; Lingaraju, K.; Daruka Prasad, B.; Kavitha, C.; Banuprakash, G.; Nagaraju, G. Synthesis of CeO₂ nanoparticles: Photocatalytic and antibacterial activities. *Eur. Phys. J. Plus* 2017, *132*, 239. https://doi.org/10.1140/epjp/i2017-11462-4.
- Saadoon, S.J.; Jarosova, M.; Machek, P.; Kadhim, M.M.; Ali, M.H.; Khalaji, A.D. Methylene blue photodegradation using assynthesized CeO₂ nanoparticles. J. Chin. Chem. Soc. 2022, 69, 280–288. https://doi.org/10.1002/jccs.202100476.
- Christy, E.J.S.; Alagar, R.; Dhanu, M.; Pius, A. Porous nonhierarchical CeO₂/SiO₂ monolith for effective degradation of organic pollutants. *Environ. Nanotechnol. Monit. Manag.* 2020, 14, 100365. https://doi.org/10.1016/j.enmm.2020.100365.
- Dhmees, A.S.; Rashad, A.M.; Eliwa, A.A.; Zawrah, M.F. Preparation and characterization of nano SiO₂@CeO₂ extracted from blast furnace slag and uranium extraction waste for wastewater treatment. *Ceram. Int.* 2019, 45, 7309–7317. https://doi.org/10.1016/j.ceramint.2019.01.014.
- Farrukh, M.A.; Butt, K.M.; Chong, K.-K.; Chang, W.S. Photoluminescence emission behavior on the reduced band gap of Fe doping in CeO₂-SiO₂ nanocomposite and photophysical properties. *J. Saudi Chem. Soc.* 2019, 23, 561–575. https://doi.org/10.1016/j.jscs.2018.10.002.
- Mohamed, R.M.; Aazam, E.S. Synthesis and Characterization of CeO₂-SiO₂ Nanoparticles by Microwave-Assisted Irradiation Method for Photocatalytic Oxidation of Methylene Blue Dye. Int. J. Photoenergy 2012, 2012, 928760. https://doi.org/10.1155/2012/928760.
- 35. Rani, N.; Ahlawat, R.; Goswami, B. Annealing effect on bandgap energy and photocatalytic properties of CeO₂-SiO₂ nanocomposite prepared by sol-gel technique. *Mater. Chem. Phys.* **2020**, 241, 122401. https://doi.org/10.1016/j.matchemphys.2019.122401.
- 36. Li, X.; Simon, U.; Bekheet, M.F.; Gurlo, A. Mineral-Supported Photocatalysts: A Review of Materials, Mechanisms and Environmental Applications. *Energies* **2022**, *15*, 5607.
- Mishra, A.; Mehta, A.; Basu, S. Clay supported TiO₂ nanoparticles for photocatalytic degradation of environmental pollutants: A review. J. Environ. Chem. Eng. 2018, 6, 6088–6107. https://doi.org/10.1016/j.jece.2018.09.029.

- Tang, X.; Tang, R.; Xiong, S.; Zheng, J.; Li, L.; Zhou, Z.; Gong, D.; Deng, Y.; Su, L.; Liao, C. Application of natural minerals in photocatalytic degradation of organic pollutants: A review. *Sci. Total Environ.* 2022, *812*, 152434. https://doi.org/10.1016/j.scitotenv.2021.152434.
- 39. Jiang, H.; Wang, R.; Wang, D.; Hong, X.; Yang, S. SnO₂/Diatomite Composite Prepared by Solvothermal Reaction for Low-Cost Photocatalysts. *Catalysts* **2019**, *9*, 1060.
- 40. Xiong, C.; Ren, Q.; Chen, S.; Liu, X.; Jin, Z.; Ding, Y. A multifunctional Ag₃PO₄/Fe₃O₄/Diatomite composites: Photocatalysis, adsorption and sterilization. *Mater. Today Commun.* **2021**, *28*, 102695. https://doi.org/10.1016/j.mtcomm.2021.102695.
- Zhang, G.; Cai, D.; Wang, M.; Zhang, C.; Zhang, J.; Wu, Z. Microstructural modification of diatomite by acid treatment, high-speed shear, and ultrasound. *Microporous Mesoporous Mater.* 2013, 165, 106–112. https://doi.org/10.1016/j.micromeso.2012.08.005.
- Cherrak, R.; Hadjel, M.; Benderdouche, N.; Bellayer, S.; Traisnel, M. Treatment of recalcitrant organic pollutants in water by heterogeneous catalysis using a mixed material (TiO₂-diatomite of algeria). *Desalination Water Treat.* 2016, 57, 17139–17148. https://doi.org/10.1080/19443994.2016.1162201.
- Zhang, G.; Wang, B.; Sun, Z.; Zheng, S.; Liu, S. A comparative study of different diatomite-supported TiO₂ composites and their photocatalytic performance for dye degradation. *Desalination Water Treat.* 2016, 57, 17512–17522. https://doi.org/10.1080/19443994.2015.1085449.
- Tanniratt, P.; Wasanapiarnpong, T.; Mongkolkachit, C.; Sujaridworakun, P. Utilization of industrial wastes for preparation of high performance ZnO/diatomite hybrid photocatalyst. *Ceram. Int.* 2016, 42, 17605–17609. https://doi.org/10.1016/j.ceramint.2016.08.074.
- Tajmiri, S.; Hosseini, M.R.; Azimi, E. Combined photocatalytic-adsorptive removal of water contaminants using a biologically prepared CdS-diatomite nanocomposite. *Mater. Chem. Phys.* 2021, 258, 123913. https://doi.org/10.1016/j.matchemphys.2020.123913.
- 46. Yang, B.; Ma, Z.; Wang, Q.; Yang, J. Synthesis and Photoelectrocatalytic Applications of TiO₂/ZnO/Diatomite Composites. *Catalysts* **2022**, *12*, 268.
- 47. Zhou, Y.; Xi, W.; Xie, Z.; You, Z.; Jiang, X.; Han, B.; Lang, R.; Wu, C. High-Loading Pt Single-Atom Catalyst on CeO₂-Modified Diatomite Support. *Chem. Asian J.* **2021**, *16*, 2622–2625. https://doi.org/10.1002/asia.202100730.
- Su, C.; Wang, H.; Li, M.; Qu, Y.; Zhu, Y. Tribological behavior and characterization analysis of modified nano-CeO₂ filled oily diatomite/PVDF composites. *Tribol. Int.* 2019, 130, 299–307. https://doi.org/10.1016/j.triboint.2018.10.007.
- Seeharaj, P.; Kongmun, P.; Paiplod, P.; Prakobmit, S.; Sriwong, C.; Kim-Lohsoontorn, P.; Vittayakorn, N. Ultrasonicallyassisted surface modified TiO₂/rGO/CeO₂ heterojunction photocatalysts for conversion of CO₂ to methanol and ethanol. *Ultrason. Sonochem.* 2019, 58, 104657. https://doi.org/10.1016/j.ultsonch.2019.104657.
- 50. Hatchard, C.G.; Parker, C.A.; Bowen, E.J. A new sensitive chemical actinometer II. Potassium ferrioxalate as a standard chemical actinometer. *Proc. R. Soc. Lond. A* **1956**, 235, 518–536. https://doi.org/10.1098/rspa.1956.0102.
- Dai, X.; Zeng, H.; Jin, C.; Rao, J.; Liu, X.; Li, K.; Zhang, Y.; Yu, Y.; Zhang, Y. 2D–3D graphene-coated diatomite as a support toward growing ZnO for advanced photocatalytic degradation of methylene blue. *RSC Adv.* 2021, *11*, 38505–38514. https://doi.org/10.1039/D1RA07708B.
- Ma, L.; Xu, H.; Xie, Q.; Chen, N.; Yu, Q.; Li, C. Mechanism of As(V) adsorption from aqueous solution by chitosan-modified diatomite adsorbent. J. Dispersion Sci. Technol. 2021, 43, 1512–1524. https://doi.org/10.1080/01932691.2021.1876592.
- 53. Pu, X.; Dang, Q.; Liu, C.; Xu, Q.; Li, B.; Ji, X.; Liu, H.; Ma, Y.; Zhang, B.; Cha, D. Selective capture of mercury(II) in aqueous media using nanoporous diatomite modified by allyl thiourea. *J. Mater. Sci.* **2022**, *57*, 9246–9264. https://doi.org/10.1007/s10853-022-07245-1.
- Nguyen, Q.-B.; Vahabi, H.; Rios de Anda, A.; Versace, D.-L.; Langlois, V.; Perrot, C.; Nguyen, V.-H.; Naili, S.; Renard, E. Dual UV-Thermal Curing of Biobased Resorcinol Epoxy Resin-Diatomite Composites with Improved Acoustic Performance and Attractive Flame Retardancy Behavior. *Sustain. Chem.* 2021, *2*, 24–48.
- Othman, A.; Vargo, P.; Andreescu, S. Recyclable Adsorbents Based on Ceria Nanostructures on Mesoporous Silica Beads for the Removal and Recovery of Phosphate from Eutrophic Waters. ACS Appl. Nano Mater. 2019, 2, 7008–7018. https://doi.org/10.1021/acsanm.9b01512.
- Xie, L.; Ren, Z.; Zhu, P.; Xu, J.; Luo, D.; Lin, J. A novel CeO₂–TiO₂/PANI/NiFe₂O₄ magnetic photocatalyst: Preparation, characterization and photodegradation of tetracycline hydrochloride under visible light. *J. Solid State Chem.* 2021, 300, 122208. https://doi.org/10.1016/j.jssc.2021.122208.
- Laha, S.C.; Mukherjee, P.; Sainkar, S.R.; Kumar, R. Cerium Containing MCM-41-Type Mesoporous Materials and their Acidic and Redox Catalytic Properties. J. Catal. 2002, 207, 213–223. https://doi.org/10.1006/jcat.2002.3516.
- 58. Yao, G.; Lei, J.; Zhang, X.; Sun, Z.; Zheng, S.; Komarneni, S. Mechanism of zeolite X crystallization from diatomite. *Mater. Res. Bull.* **2018**, 107, 132–138. https://doi.org/10.1016/j.materresbull.2018.07.021.
- Reddy, B.M.; Khan, A.; Yamada, Y.; Kobayashi, T.; Loridant, S.; Volta, J.-C. Surface Characterization of CeO₂/SiO₂ and V₂O₅/CeO₂/SiO₂ Catalysts by Raman, XPS, and Other Techniques. J. Phys. Chem. B 2002, 106, 10964–10972. https://doi.org/10.1021/jp021195v.
- 60. Mamontov, G.V.; Grabchenko, M.V.; Sobolev, V.I.; Zaikovskii, V.I.; Vodyankina, O.V. Ethanol dehydrogenation over Ag-CeO₂/SiO₂ catalyst: Role of Ag-CeO₂ interface. *Appl. Catal. A* **2016**, *528*, 161–167. https://doi.org/10.1016/j.apcata.2016.10.005.

- 61. Weber, W.H.; Hass, K.C.; McBride, J.R. Raman study of CeO₂: Second-order scattering, lattice dynamics, and particle-size effects. *Phys. Rev. B* **1993**, *48*, 178–185. https://doi.org/10.1103/PhysRevB.48.178.
- Grabchenko, M.V.; Mamontov, G.V.; Zaikovskii, V.I.; La Parola, V.; Liotta, L.F.; Vodyankina, O.V. Design of Ag-CeO₂/SiO₂ catalyst for oxidative dehydrogenation of ethanol: Control of Ag-CeO₂ interfacial interaction. *Catal. Today* 2019, 333, 2–9. https://doi.org/10.1016/j.cattod.2018.05.014.
- 63. Humbert, B.; Burneau, A.; Gallas, J.P.; Lavalley, J.C. Origin of the Raman bands, D1 and D2, in high surface area and vitreous silicas. *J. Non-Cryst. Solids* **1992**, *143*, 75–83. https://doi.org/10.1016/S0022-3093(05)80555-1.
- Cerrato, E.; Calza, P.; Cristina Paganini, M. Photocatalytic reductive and oxidative ability study of pristine ZnO and CeO2-ZnO heterojunction impregnated with Cu2O. J. Photochem. Photobiol. A 2022, 427, 113775. https://doi.org/10.1016/j.jphotochem.2022.113775.
- 65. Scherrer, P. Estimation of the size and internal structure of colloidal particles by means of röntgen. *Nachr. Ges. Wiss. Göttingen* **1918**, *2*, 96–100.
- 66. Gao, L.; Luo, Y.; Kang, Y.; Gao, M.; Abdulhafidh, O. Experimental Study on Physical Mechanical Properties and Microstructure of Diatomite Soil in Zhejiang Province, China. *Appl. Sci.* **2022**, *12*, 387.
- Zuo, X.; Wang, L.; He, J.; Li, Z.; Yu, S. SEM-EDX studies of SiO2/PVDF membranes fouling in electrodialysis of polymerflooding produced wastewater: Diatomite, APAM and crude oil. *Desalination* 2014, 347, 43–51. https://doi.org/10.1016/j.desal.2014.05.020.
- Sing, K.S.W. Reporting physisorption data for gas/solid systems with special reference to the determination of surface area and porosity (Recommendations 1984). *Pure Appl. Chem.* 1985, 57, 603–619. https://doi.org/10.1351/pac198557040603.
- 69. Rouquerol, F.; Rouquerol, J.; Sing, K.S.W.; Llewellyn, P.; Maurin, G. *Adsorption by Powders and Porous Solids*, 2nd ed.; Academic Press: Cambridge, MA, USA, 2014; https://doi.org/10.1016/C2010-0-66232-8.
- Yurdakal, S.; Garlisi, C.; Özcan, L.; Bellardita, M.; Palmisano, G. Chapter 4–(Photo)catalyst Characterization Techniques: Adsorption Isotherms and BET, SEM, FTIR, UV–Vis, Photoluminescence, and Electrochemical Characterizations. In *Heterogeneous Photocatalysis*; Marci, G., Palmisano, L., Eds; Elsevier: Amsterdam, The Netherlands, 2019; pp. 87–152. https://doi.org/10.1016/B978-0-444-64015-4.00004-3.
- Azimi Pirsaraei, S.R.; Asilian Mahabadi, H.; Jonidi Jafari, A. Airborne toluene degradation by using manganese oxide supported on a modified natural diatomite. J. Porous Mater. 2016, 23, 1015–1024. https://doi.org/10.1007/s10934-016-0159-2.
- Benkacem, T.; Hamdi, B.; Chamayou, A.; Balard, H.; Calvet, R. Physicochemical characterization of a diatomaceous upon an acid treatment: A focus on surface properties by inverse gas chromatography. *Powder Technol.* 2016, 294, 498–507. https://doi.org/10.1016/j.powtec.2016.03.006.
- 73. Inchaurrondo, N.; Ramos, C.P.; Žerjav, G.; Font, J.; Pintar, A.; Haure, P. Modified diatomites for Fenton-like oxidation of phenol. *Microporous Mesoporous Mater.* 2017, 239, 396–408. https://doi.org/10.1016/j.micromeso.2016.10.026.
- 74. Mazidi, M.; Behbahani, R.M.; Fazeli, A. Screening of treated diatomaceous earth to apply as V₂O₅ catalyst support. *Mater. Res. Innovations* **2017**, *21*, 269–278. https://doi.org/10.1080/14328917.2016.1211478.
- 75. Xia, Y.; Li, F.; Jiang, Y.; Xia, M.; Xue, B.; Li, Y. Interface actions between TiO₂ and porous diatomite on the structure and photocatalytic activity of TiO₂-diatomite. *Appl. Surf. Sci.* **2014**, *303*, 290–296. https://doi.org/10.1016/j.apsusc.2014.02.169.
- 76. Sprynskyy, M.; Kowalkowski, T.; Tutu, H.; Cukrowska, E.M.; Buszewski, B. Ionic liquid modified diatomite as a new effective adsorbent for uranium ions removal from aqueous solution. *Colloids Surf. A Physicochem. Eng. Asp.* 2015, 465, 159–167. https://doi.org/10.1016/j.colsurfa.2014.10.042.
- 77. Caliskan, N.; Kul, A.R.; Alkan, S.; Sogut, E.G.; Alacabey, İ. Adsorption of Zinc (II) on diatomite and manganese-oxidediatomite: equilibrium study. Hazard. 2011, modified А kinetic and J. Mater. 193. 27 - 36. https://doi.org/10.1016/j.jhazmat.2011.06.058.
- 78. Ilia, I.; Stamatakis, M.; Perraki, T. Mineralogy and technical properties of clayey diatomites from north and central Greece. *Open Geosci.* **2009**, *1*, 393–403. https://doi.org/10.2478/v10085-009-0034-3.
- Kong, X.; Chong, X.; Squire, K.; Wang, A.X. Microfluidic diatomite analytical devices for illicit drug sensing with ppb-Level sensitivity. Sens. Actuators B Chem. 2018, 259, 587–595. https://doi.org/10.1016/j.snb.2017.12.038.
- 80. Rea, I.; De Stefano, L. Recent Advances on Diatom-Based Biosensors. Sensors 2019, 19, 5208.
- Viji, S.; Anbazhagi, M.; Ponpandian, N.; Mangalaraj, D.; Jeyanthi, S.; Santhanam, P.; Devi, A.S.; Viswanathan, C. Diatom-Based Label-Free Optical Biosensor for Biomolecules. *Appl. Biochem. Biotechnol.* 2014, 174, 1166–1173. https://doi.org/10.1007/s12010-014-1040-x.
- Rea, I.; Martucci, N.M.; De Stefano, L.; Ruggiero, I.; Terracciano, M.; Dardano, P.; Migliaccio, N.; Arcari, P.; Taté, R.; Rendina, I.; et al. Diatomite biosilica nanocarriers for siRNA transport inside cancer cells. *Biochim. Biophys. Acta Gen. Subj.* 2014, 1840, 3393–3403. https://doi.org/10.1016/j.bbagen.2014.09.009.
- 83. Choudhary, S.; Sahu, K.; Bisht, A.; Singhal, R.; Mohapatra, S. Template-free and surfactant-free synthesis of CeO₂ nanodiscs with enhanced photocatalytic activity. *Appl. Surf. Sci.* **2020**, *503*, 144102. https://doi.org/10.1016/j.apsusc.2019.144102.
- Kaviyarasu, K.; Fuku, X.; Mola, G.T.; Manikandan, E.; Kennedy, J.; Maaza, M. Photoluminescence of well-aligned ZnO doped CeO₂ nanoplatelets by a solvothermal route. *Mater. Lett.* 2016, *183*, 351–354. https://doi.org/10.1016/j.matlet.2016.07.143.
- Wang, G.; Mu, Q.; Chen, T.; Wang, Y. Synthesis, characterization and photoluminescence of CeO₂ nanoparticles by a facile method at room temperature. J. Alloys Compd. 2010, 493, 202–207. https://doi.org/10.1016/j.jallcom.2009.12.053.

- Parks, G.A. Aqueous Surface Chemistry of Oxides and Complex Oxide Minerals. In *Equilibrium Concepts in Natural Water* Systems; Advances in Chemistry; American Chemical Society: Washington, DC, USA, 1967; Volume 67, pp. 121–160.
- Nosrati, A.; Larsson, M.; Lindén, J.B.; Zihao, Z.; Addai-Mensah, J.; Nydén, M. Polyethyleneimine functionalized mesoporous diatomite particles for selective copper recovery from aqueous media. *Int. J. Miner. Process.* 2017, 166, 29–36. https://doi.org/10.1016/j.minpro.2017.07.001.
- 88. Sondi, I.; Bišćan, J.; Pravdić, V. Electrokinetics of Pure Clay Minerals Revisited. J. Colloid Interface Sci. 1996, 178, 514–522. https://doi.org/10.1006/jcis.1996.0146.
- Ye, X.; Kang, S.; Wang, H.; Li, H.; Zhang, Y.; Wang, G.; Zhao, H. Modified natural diatomite and its enhanced immobilization of lead, copper and cadmium in simulated contaminated soils. *J. Hazard. Mater.* 2015, 289, 210–218. https://doi.org/10.1016/j.jhazmat.2015.02.052.
- 90. Gao, B.; Jiang, P.; An, F.; Zhao, S.; Ge, Z. Studies on the surface modification of diatomite with polyethyleneimine and trapping effect of the modified diatomite for phenol. Appl. Surf. Sci. 2005, 250. 273-279. https://doi.org/10.1016/j.apsusc.2005.02.119.
- 91. Sun, Z.; Yang, X.; Zhang, G.; Zheng, S.; Frost, R.L. A novel method for purification of low grade diatomite powders in centrifugal fields. *Int. J. Miner. Process.* 2013, 125, 18–26. https://doi.org/10.1016/j.minpro.2013.09.005.
- 92. Houas, A.; Lachheb, H.; Ksibi, M.; Elaloui, E.; Guillard, C.; Herrmann, J.M. Photocatalytic degradation pathway of methylene blue in water. *Appl. Catal. B* **2001**, *31*, 145–157.
- 93. Al-Qodah, Z.; Lafi, W.K.; Al-Anber, Z.; Al-Shannag, M.; Harahsheh, A. Adsorption of methylene blue by acid and heat treated diatomaceous silica. *Desalination* 2007, 217, 212–224. https://doi.org/10.1016/j.desal.2007.03.003.
- Zhang, T.; Oyama, T.; Aoshima, A.; Hidaka, H.; Zhao, J.; Serpone, N. Photooxidative N-demethylation of methylene blue in aqueous TiO₂ dispersions under UV irradiation. *J. Photochem. Photobiol. A* 2001, 140, 163–172. https://doi.org/10.1016/s1010-6030(01)00398-7.
- Zhang, T.; Oyama, T.k.; Horikoshi, S.; Hidaka, H.; Zhao, J.; Serpone, N. Photocatalyzed N-demethylation and degradation of methylene blue in titania dispersions exposed to concentrated sunlight. *Sol. Energy Mater. Sol. Cells* 2002, 73, 287–303. https://doi.org/10.1016/s0927-0248(01)00215-x.
- Lin, J.; Luo, Z.; Liu, J.; Li, P. Photocatalytic degradation of methylene blue in aqueous solution by using ZnO-SnO₂ nanocomposites. *Mater. Sci. Semicond. Process.* 2018, 87, 24–31. https://doi.org/10.1016/j.mssp.2018.07.003.
- 97. Hao, C.; Li, J.; Zhang, Z.; Ji, Y.; Zhan, H.; Xiao, F.; Wang, D.; Liu, B.; Su, F. Enhancement of photocatalytic properties of TiO₂ nanoparticles doped with CeO₂ and supported on SiO₂ for phenol degradation. *Appl. Surf. Sci.* 2015, 331, 17–26. https://doi.org/10.1016/j.apsusc.2015.01.069.
- 98. Melcher, J.; Barth, N.; Schilde, C.; Kwade, A.; Bahnemann, D. Influence of TiO₂ agglomerate and aggregate sizes on photocatalytic activity. *J. Mater. Sci.* **2016**, *52*, 1047–1056. https://doi.org/10.1007/s10853-016-0400-z.